Revised Draft

Phase III Remedial Investigation Report for Soil, Sediment, and Surface Water at RVAAP-29 Upper and Lower Cobbs Ponds

> Former Ravenna Army Ammunition Plant Portage and Trumbull Counties, Ohio

Contract No. W912QR-15-C-0046

Prepared for:



US Army Corps of Engineers®

U.S. Army Corps of Engineers Louisville District

Prepared by:



March 17, 2017

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CONTRACTOR STATEMENT OF INDEPENDENT TECHNICAL REVIEW

Leidos has completed the Revised Draft Phase III Remedial Investigation Report for Soil, Sediment, and Surface Water at RVAAP-29 Upper and Lower Cobbs Ponds at the Former Ravenna Army Ammunition Plant, Portage and Trumbull Counties, Ohio. Notice is hereby given that an independent technical review has been conducted that is appropriate to the level of risk and complexity inherent in the project. During the independent technical review, compliance with established policy principles and procedures, utilizing justified and valid assumptions, was verified. This included review of data quality objectives; technical assumptions; methods, procedures, and materials to be used; the appropriateness of data used and level of data obtained; and reasonableness of the results, including whether the product meets the customer's needs consistent with law and existing U.S. Army Corps of Engineers (USACE) policy.

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Michael Bolen, P.G. Independent Technical Review Team Leader

Significant concerns and the explanation of the resolution are as follows:

Internal Leidos Independent Technical Review comments are recorded on a Document Review Record per Leidos standard operating procedure ESE A3.1 Document Review. This Document Review Record is maintained in the project file. Changes to the report addressing the comments have been verified by the Study/Design Team Leader. As noted above, all concerns resulting from independent technical review of the project have been considered.

Lisa Jones-Bateman Senior Program Manager

3/17/17 Date

3/17/17 Date

3/17/17 Date

2/17/17

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Documentation of Ohio EPA Concurrence of Final Document

(Documentation to be provided once concurrence is issued.)

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Contract No. W912QR-15-C-0046

Prepared for: U.S. Army Corps of Engineers Louisville District

Prepared by: Leidos 8866 Commons Boulevard, Suite 201 Twinsburg, Ohio 44087

March 17, 2017

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ACRONYMS AND ABBREVIATIONS

| amsl | Above Mean Sea Level |
|--------------|---|
| AOC | Area of Concern |
| Army | U.S. Department of the Army |
| AT123D | Analytical Transient 1-, 2-, 3- Dimensional |
| bgs | Below Ground Surface |
| BHC | Hexachlorocyclohexane |
| BHHRA | Baseline Human Health Risk Assessment |
| Camp Ravenna | Camp Ravenna Joint Military Training Center |
| CERCLA | Comprehensive Environmental Response, Compensation, and Liability Act |
| CMCOC | Contaminant Migration Chemical of Concern |
| СМСОРС | Contaminant Migration Chemical of Potential Concern |
| COC | Chemical of Concern |
| COPC | Chemical of Potential Concern |
| COPEC | Chemical of Potential Ecological Concern |
| CSEM | Conceptual Site Exposure Model |
| CSI | Community Similarity Index |
| CSM | Conceptual Site Model |
| CUG | Cleanup Goal |
| DAF | Dilution Attenuation Factor |
| DERR | Division of Environmental Response and Revitalization |
| DFFO | Directors Final Findings and Orders |
| DNT | Dinitrotoluene |
| DQO | Data Quality Objective |
| EcoSSL | Ecological Soil Screening Level |
| EPC | Exposure Point Concentration |
| ERA | Ecological Risk Assessment |
| ESL | Ecological Screening Level |
| ESV | Ecological Screening Value |
| EU | Exposure Unit |
| f_{oc} | Mass Fraction of the Organic Carbon Soil Content |
| FS | Feasibility Study |
| FWCUG | Facility-wide Cleanup Goal |
| FWERWP | Facility-wide Ecological Risk Work Plan |
| FWGWMP | Facility-wide Groundwater Monitoring Plan |
| FWHHRAM | Facility-wide Human Health Risk Assessors Manual |
| FWSAP | Facility-wide Sampling and Analysis Plan |
| gpm | Gallons Per Minute |
| GSSL | Generic Soil Screening Level |
| HELP | Hydrologic Evaluation of Landfill Performance |
| HHRA | Human Health Risk Assessment |
| HLC | Henry's Law Constant |
| | |

| HQHazard QuotientILCRIncremental Lifetime Cancer RiskINRMPIntegrated Natural Resource Management PlanIRISIntegrated Risk Information SystemISMIncremental Sampling MethodologyK_a'SoilWater Partitioning CoefficientK_ssWater/Organic Carbon Partition CoefficientK_wWater/Organic Carbon Partition CoefficientLoAELLowest Observed Adverse Effect LevelL-QHEILake Eric Shoreline Qualitative Habitat Evaluation IndexMCLMaximum Contaminant LevelMDCMaximum Detected ConcentrationMDLMethod Detection LimiNCPNational Oil and Hazardous Substances Pollution Contingency PlanNOAELNo beserved Adverse Effect LevelOHARNGOhio Army National GuardOhia EPAOhio King Zone AverageOMZAOutside Mixing Zone AverageOMZAOutside Mixing Zone AverageOMZAOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBTPersistent, Bioaccumulative, and ToxiePCProbable Effect ConcentrationPBTPersistent, Bioaccumulative, and ToxiePCQuality AssuranceQCQuality SuranceQCQuality SuranceQCQuality SuranceQCQuality Habitat Evaluation IndexRRecommended Daily AllowanceRDIRecommended Daily Allowance <th>HMX</th> <th>Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine</th> | HMX | Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine | |
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| ILCRIncremental Lifetime Cancer RiskINRMPIntegrated Natural Resource Management PlanIRISIntegrated Risk Information SystemISMIncremental Sampling MethodologyK ₄ Soil/Water Partitioning CoefficientK ₄ sDistribution CoefficientK ₈ sOctanol-Water Partition CoefficientLOAELLowest Observed Adverse Effect LevelL-QHEILake Eric Shoreline Qualitative Habitat Evaluation IndexMCLMaximum Detected ConcentrationMDLMethod Detection LimitNOAELNo Observed Adverse Effect LevelNOAELNo Observed Adverse Effect LevelOttanol Oil and Hazardous Substances Pollution Contingency PlanNOAELNo Observed Adverse Effect LevelOHARNGOhio Army National GuardOhio Erivonmental Protection AgencyOMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone AverageOMZMOutside Mixing Zone AverageOMZMOutside Mixing Zone AverageOMAENPerformance-based Acquisition 2008 Supplemental InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxiePCBProlowed PlanPRGPerlohiniary Remediation GoalQAQuality AssuranceQCQuality ControlQIIEIQualitative Habitat Evaluation IndexRRetardation Factor <td></td> <td colspan="2">•</td> | | • | |
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| MDCMaximum Detected ConcentrationMDLMethod Detection LimitNCPNational Oil and Hazardous Substances Pollution Contingency PlanNOAELNo Observed Adverse Effect LevelOHARNGOhio Army National GuardOhio EPAOhio Environmental Protection AgencyOMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 SIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBProposed PlanPRGProposed PlanPRGQuality AssuranceQCQuality AssuranceQCQuality AssuranceQDARetardation FactorRDARecommended Daily AllowanceRDJRecommended Daily IntakeRDJRecommended Daily IntakeRDJRecommended Daily IntakeRDJReference DoseRIRemedial InvestigationRDDRecord of Decision | L-QHEI | Lake Erie Shoreline Qualitative Habitat Evaluation Index | |
| MDLMethod Detection LimitNCPNational Oil and Hazardous Substances Pollution Contingency PlanNOAELNo Observed Adverse Effect LevelOHARNGOhio Army National GuardOhio EPAOhio Environmental Protection AgencyOMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial Investigation Sampling and Analysis Plan Addendum No. 1PBTPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPolycyclic ConcentrationPECPolybable Effect ConcentrationPPAPoposed PlanPRGProlosable Effect ConcentrationQQQuality AssuranceQQQuality ControlQLHEIQualitative Habitat Evaluation IndexRRecommended Daily AllowanceRDARecommended Daily AllowanceRDJRecommended Daily IntakeRDJReference DoseRIRemedial InvestigationRDDReference DoseRIRendial InvestigationRDDRecord of Decision | MCL | Maximum Contaminant Level | |
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| NOAELNo Observed Adverse Effect LevelOHARNGOhio Army National GuardOhio EPAOhio Environmental Protection AgencyOMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQuality ControlQHEIRecommended Daily AllowanceRDARecommended Daily IntakeRDARecommended Daily IntakeRDARetraction FactorRDARetrence DoseRIRemedial InvestigationRDDRecord of Decision | MDL | Method Detection Limit | |
| OHARNGOhio Army National GuardOhio EPAOhio Environmental Protection AgencyOMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPCCProbabe Effect ConcentrationPFCProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRecommended Daily AllowanceRDARecommended Daily AllowanceRDARecommended Daily IntakeRDMReference DoseRIReindial InvestigationRDAReference DoseRIRecord Doleision | NCP | National Oil and Hazardous Substances Pollution Contingency Plan | |
| Ohio EPAOhio Evironmental Protection AgencyOMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQuality ControlQHEIRecommended Daily AllowanceRDARecommended Daily IntakeRDXHexahydro-1,3,5-trinizro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRecord of Decision | NOAEL | No Observed Adverse Effect Level | |
| OMZAOutside Mixing Zone AverageOMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRecommended Daily AllowanceRDARecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRenedial InvestigationRODRecord of Decision | OHARNG | Ohio Army National Guard | |
| OMZMOutside Mixing Zone MaximumORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRDARecommended Daily AllowanceRDJRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRenedial InvestigationRODRecord of Decision | Ohio EPA | Ohio Environmental Protection Agency | |
| ORAMOhio Rapid Assessment MethodPAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRecommended Daily AllowanceRDARecommended Daily IntakeRDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | OMZA | Outside Mixing Zone Average | |
| PAHPolycyclic Aromatic HydrocarbonPBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRIDReference DoseRIRecord of Decision | OMZM | Outside Mixing Zone Maximum | |
| PBA08 RIPerformance-based Acquisition 2008 Remedial InvestigationPBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | ORAM | Ohio Rapid Assessment Method | |
| PBA08 SAPPerformance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRecommended Daily AllowanceRDARecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRecord of Decision | РАН | Polycyclic Aromatic Hydrocarbon | |
| Analysis Plan Addendum No. 1PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRecord of Decision | PBA08 RI | Performance-based Acquisition 2008 Remedial Investigation | |
| PBTPersistent, Bioaccumulative, and ToxicPCBPolychlorinated BiphenylPECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQuality ControlRDARetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineRfDReference DoseRIRemedial InvestigationRODRecord of Decision | PBA08 SAP | Performance-based Acquisition 2008 Supplemental Investigation Sampling and | |
| PCBPolychlorinated BiphenylPECProbable Effect ConcentrationPFProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRIRemedial InvestigationRODRecord of Decision | | Analysis Plan Addendum No. 1 | |
| PECProbable Effect ConcentrationPPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRecord of Decision | PBT | Persistent, Bioaccumulative, and Toxic | |
| PPProposed PlanPRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDXRecommended Daily IntakeRDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | PCB | Polychlorinated Biphenyl | |
| PRGPreliminary Remediation GoalQAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | PEC | Probable Effect Concentration | |
| QAQuality AssuranceQCQuality ControlQHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | РР | Proposed Plan | |
| QCQuality ControlQHEIQualitative Habitat Evaluation IndexRQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | PRG | Preliminary Remediation Goal | |
| QHEIQualitative Habitat Evaluation IndexRRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | QA | Quality Assurance | |
| RRetardation FactorRDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | QC | Quality Control | |
| RDARecommended Daily AllowanceRDIRecommended Daily IntakeRDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | QHEI | Qualitative Habitat Evaluation Index | |
| RDIRecommended Daily IntakeRDXHexahydro-1,3,5-trinitro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | R | Retardation Factor | |
| RDXHexahydro-1,3,5-triazineREIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | RDA | Recommended Daily Allowance | |
| REIMSRavenna Environmental Information Management SystemRfDReference DoseRIRemedial InvestigationRODRecord of Decision | RDI | Recommended Daily Intake | |
| RfDReference DoseRIRemedial InvestigationRODRecord of Decision | RDX | Hexahydro-1,3,5-trinitro-1,3,5-triazine | |
| RIRemedial InvestigationRODRecord of Decision | REIMS | Ravenna Environmental Information Management System | |
| ROD Record of Decision | | | |
| | | - | |
| RSI Regional Screening Level | | | |
| | RSL | Regional Screening Level | |

| DVAAD | Devenue Amore Amore Diant |
|--------|--|
| RVAAP | Ravenna Army Ammunition Plant |
| SAIC | Science Applications International Corporation |
| SERA | Screening Ecological Risk Assessment |
| SESOIL | Seasonal Soil Compartment Model |
| SMCL | Secondary Maximum Contaminant Level |
| SOR | Sum of Ratio |
| SRC | Site-related Contaminant |
| SRV | Sediment Reference Value |
| SSL | Soil Screening Level |
| SSSL | Site-Specific Soil Screening Level |
| SVOC | Semi-volatile Organic Compound |
| TAL | Target Analyte List |
| TEC | Threshold Effect Concentration |
| TNT | 2,4,6-Trinitrotoluene |
| TR | Target Risk |
| UCL | Upper Confidence Limit |
| URF | Unit Risk Factor |
| USACE | U.S. Army Corps of Engineers |
| USEPA | U.S. Environmental Protection Agency |
| USP&FO | U.S. Property and Fiscal Officer |
| VOC | Volatile Organic Compound |
| WOE | Weight-of-Evidence |
| WQC | Water Quality Criteria |
| WQS | Water Quality Standards |
| | |

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1 2

EXECUTIVE SUMMARY

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ES.1 INTRODUCTION AND SCOPE

5 This document has been revised by Leidos under the U.S. Army Corps of Engineers Louisville 6 District Contract Number W912QR-15-C-0046. This Phase III Remedial Investigation (RI) Report 7 addresses soil, sediment, and surface water at the Upper and Lower Cobbs Ponds area of concern 8 (AOC) within the former Ravenna Army Ammunition Plant (RVAAP) [now known as Camp 9 Ravenna Joint Military Training Center (Camp Ravenna)] in Portage and Trumbull counties, Ohio.

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11 This report has been prepared in accordance with the requirements of the Ohio Environmental 12 Protection Agency (Ohio EPA) Director's Final Findings and Orders (DFFO) for RVAAP, dated June 13 10, 2004 (Ohio EPA 2004). The DFFO requires conformance with the Comprehensive Environmental 14 Response, Compensation, and Liability Act (CERCLA) and National Contingency Plan to implement 15 an RI to characterize the AOC; develop a feasibility study (FS) (if remediation is necessary); and 16 evaluate remedial alternatives to address contamination presenting unacceptable risk to human health 17 and the environment, present a preferred alternative in a proposed plan (PP), and document 18 stakeholder selection and acceptance of the preferred final remedy in a record of decision (ROD). The 19 following sections present the site history, scope of this report, and an explanation of the evaluation 20 of future use.

21

22 ES.1.1 Site History

23

24 The Upper and Lower Cobbs Ponds AOC is located in the east-central portion of Camp Ravenna. The 25 AOC is located east of Paris-Windham Road and south of Remalia Road, north of Load Line 12, and 26 northwest of Load Line 3. The AOC is approximately 39 acres and is comprised of the Backwater 27 Area (5.2 acres), Upper Cobbs Pond (9.4 acres), Lower Cobbs Pond (6.4 acres), and surrounding 28 pond banks (18 acres). Upper Cobbs Pond ranges from 3–8 ft in depth, and Lower Cobbs Pond ranges 29 from 2–7 ft in depth (MKM 2005). The Backwater Area is located to the south (upstream) of Upper 30 Cobbs Pond and was created through beaver dam construction activity. The Backwater Area is 31 located south of Track 33, while Upper and Lower Cobbs Ponds are located north of Track 33. Cobbs 32 Pond Road separates Upper Cobbs Pond from Lower Cobbs Pond.

33

34 Upper Cobbs Pond and Lower Cobbs Pond were constructed in 1940–1941, expanding a natural 35 drainage conveyance to receive effluent discharge and to serve as the unlined sedimentation basins 36 for Load Lines 3 and 12. From 1941–1971, the ponds received effluent from the Load Lines 3 and 12 37 sawdust filtration units, wash water, storm water runoff, and surface water runoff. Rinsate from 38 demilitarization operations at Load Lines 3 and 12 was initially allowed to flow out of buildings and 39 directly onto the ground or to drainage ditches, which ultimately discharged to Upper Cobbs Pond 40 and Lower Cobbs Pond. Load Line 12 was formerly utilized for producing ammonium nitrate from 1941-1943 and 1946-1950. From 1951-1961, explosive melt-out and demilitarization activities 41 42 occurred at Load Line 12. Since there were no wash water collection tanks or settling ponds in Load 43 Line 12 during these operations, all residues, dusts, and spills were washed into the drainage system 44 that eventually discharged into Upper Cobbs Pond.

1 From 1965–1967, Hercules Alcor, Inc. leased Building FF-19 in Load Line 12 for producing 2 aluminum chloride. On November 15, 1966, a fish kill occurred at Lower Cobbs Pond as a result of 3 improper handling of aluminum chloride during manufacturing operations. The bulk of the aluminum 4 chloride was collected and disposed at Ramsdell Quarry Landfill (RVAAP-01). The pond, receiving 5 the contaminating waste from drainage ditches, was settled, drained, and the contaminants were removed to Ramsdell Quarry. Contaminated metals were flashed at a burning ground to a 5X 6 7 condition (USATHAMA 1978 and 1982, RVAAP 2000). The U.S. Department of the Army (Army) 8 terminated the lease early on December 3, 1967 due to environmental concerns related to air 9 emissions and wastewater discharges to Upper and Lower Cobbs Ponds.

10

The Upper and Lower Cobbs Ponds AOC is currently utilized for recreational purposes, as it contains abundant fish and wildlife. The Ohio Army National Guard (OHARNG) and facility workers use the ponds for fishing. Currently, due to the fact that the investigative activities are still ongoing at the AOC under the restoration program, fishing in the ponds is catch and release. The pond bank areas are lightly vegetated and covered with grass. A picnic area, pavilion, and playground are located approximately 100 ft from the edge of the eastern bank of Upper Cobbs Pond.

- 17
- 18 ES.1.2 Scope
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The scope of this Phase III RI Report is to perform a CERCLA evaluation of soil, sediment, and surface water at the Upper and Lower Cobbs Ponds AOC. The media of concern associated with the AOC are surface soil [0–1 ft below ground surface (bgs)], subsurface soil (1–13 ft bgs), sediment, and surface water. This report does not include a full evaluation of groundwater; it will be evaluated as an individual AOC for the entire facility. However, the potential for soil and sediment contaminants to leach to and migrate in groundwater is evaluated in this RI Report to determine whether soil remedial actions to protect groundwater may be necessary.

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ES.1.3 Evaluation of Future Use

In February 2014, the Army and Ohio EPA amended the risk assessment process to address changes
 in the RVAAP restoration program. The *Final Technical Memorandum: Land Uses and Revised Risk Assessment Process for the RVAAP Installation Restoration Program* (ARNG 2014) (herein referred
 to as the Technical Memorandum) identified the three Categorical Land Uses and Representative
 Receptors below to be considered during the RI phase of the CERCLA process.

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- 1. Unrestricted (Residential) Land Use Resident Receptor (Adult and Child) (formerly called Resident Farmer).
- 2. Military Training Land Use National Guard Trainee.
- 39 3. Commercial/Industrial Land Use Industrial Receptor [U.S. Environmental Protection
 40 Agency's (USEPA) Composite Worker].
- 41

42 Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp

43 Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then

the AOC is also considered to have met the requirements of the other Land Uses (i.e., Industrial and
 Military Training), and those other Land Uses do not require evaluation.

3 4

5

ES.2 FINDINGS AND RECOMMENDATIONS OF THE REMEDIAL INVESTIGATION

This section presents the data used in the RI, contaminant nature and extent, fate and transport, human
health risk assessment (HHRA), and ecological risk assessment (ERA), followed by the conclusions
of the RI.

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- 10 11

ES.2.1 Data Use and Sample Selection Process

Quality-assured sample data from the RIs [1996 Phase I RI, 2001 Phase II RI, and 2010 Performance-12 13 Based Acquisition 2008 Remedial Investigation (PBA08 RI)] and the 2003 Facility-wide Biological and Water Quality Study (FWBWOS) were used to evaluate nature and extent of contamination at the 14 15 Upper and Lower Cobbs Ponds AOC. All available sample data were evaluated to determine suitability for use in various key RI data screens and evaluations (i.e., nature and extent, fate and 16 17 transport, risk assessment). Evaluating the data's suitability for use in the PBA08 RI involved two primary considerations: (1) whether the data represented current AOC conditions, and (2) sample 18 19 collection methods [e.g., discrete vs. incremental sampling methodology (ISM)].

20

Samples from the Phase I and II RIs were evaluated to determine if conditions had changed substantively between earlier characterization efforts and the 2010 PBA08 RI. Data collected in 2010 as part of the PBA08 RI focused on delineating the extent of contaminants identified in surface soil (0–1 ft bgs), subsurface soil (1–13 ft bgs), sediment, and surface water. The PBA08 RI sampled locations with the greatest likelihood of contamination (e.g., adjacent to historical operations areas or within sediment accumulation areas) and analyzed for chemicals identified in historical investigations.

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29 ES.2.2 Summary of Contaminant Nature and Extent

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31 Nature and extent of contamination in surface soil (0-1 ft bgs), subsurface soil (greater than 1 ft bgs), 32 sediment, and surface water was evaluated in the RI. Data from the RIs (1996 Phase I RI, 2001 Phase II RI, and 2010 PBA08 RI) and the 2003 FWBWQS effectively characterized the nature and extent of 33 34 the contamination at the AOC. Figure ES-1 shows the sample locations used to conduct this RI. To 35 support the evaluation of nature and extent of contamination, site-related contaminant (SRC) 36 concentrations were compared to screening levels (SLs) corresponding to the lowest facility-wide 37 cleanup goal (FWCUG) for the Resident Receptor (Adult and Child) and National Guard Trainee at a 38 target hazard quotient (HQ) of 0.1 or target risk (TR) of 1E-06, as presented in the Facility-wide 39 Human Health Cleanup Goals for the Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 40 2010a) (herein referred to as the FWCUG Report). It can be concluded that the vertical and horizontal 41 extent of contamination is defined, and no further sampling is needed to evaluate the AOC.

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1 ES.2.2.1 Surface Soil

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3 The predominant SRCs for surface soil at the Pond Bank were inorganic chemicals and semi-volatile 4 organic compounds (SVOCs); the majority of which were polycyclic aromatic hydrocarbons (PAHs). 5 A total of 18 inorganic chemicals (16 metals, cyanide, and nitrate/nitrite) were identified as SRCs in surface soil. Elevated inorganic chemical concentrations above background concentrations were 6 7 observed in surface soil throughout the Pond Bank, and only 1 of 24 surface soil samples collected 8 exhibited no inorganic chemicals above background concentrations (ULCPss-009). However, the 9 majority of elevated inorganic chemical concentrations occurred at levels marginally in excess of 10 background concentrations, as all but two concentrations occurred at magnitudes less than twice their respective background concentrations. As such, no apparent spatial trend is evident in the distribution

- 11 12
- 13

Sixteen SVOCs, predominantly PAHs, were identified as SRCs in surface soil at the Pond Bank. The highest frequency and magnitudes of detection were observed at CPCsb-035, located on the western bank of Upper Cobbs Pond, immediately south of Cobbs Pond Road. The pesticide betahexachlorocyclohexane (BHC) and propellant nitrocellulose were also identified as SRCs in surface soil at the Pond Bank. PAH concentrations are an order of magnitude higher in sediment than those observed in the adjacent Pond Bank surface soil.

of inorganic chemicals throughout the surface soil in the Pond Bank.

20

One propellant (nitrocellulose) and one pesticide (BHC) were identified as SRCs in surface soil at the Pond Bank. Neither exceeded their respective SLs. No explosives, volatile organic compounds (VOCs), or polychlorinated biphenyls (PCBs) were detected or identified as SRCs in Pond Bank surface soil surrounding Upper and Lower Cobbs Pond.

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26 ES.2.2.2 Subsurface Soil

Four inorganic chemicals (cadmium, lead, selenium, and silver) were identified as SRCs in subsurface soil samples at the Pond Bank. Lead and selenium were detected above the background concentrations at only one sample location. Selenium was detected at 1.6 mg/kg, slightly above the background concentration of 1.5 mg/kg, and lead was detected at 23.3 mg/kg above the background concentration of 19.1 mg/kg. As such, no apparent spatial trend in the distribution of inorganic chemicals throughout the subsurface soil in the Pond Bank is evident.

34

35 A total of 13 SVOCs, 12 of which were PAHs, were detected and identified as SRCs in subsurface 36 soil at the Pond Bank. Of the 13 SVOCs, 12 were detected from the 1–4 ft bgs interval at CPCsb-035; 37 the majority of which were observed at low, estimated concentrations below laboratory reporting 38 limits. Sample location CPCsb-035 is on the western bank of Upper Cobbs Pond and near Cobbs 39 Pond Road. PAH concentrations in the subsurface soil samples were generally comparable to those 40 observed in the corresponding surface soil sample (i.e., within the same order of magnitude). One 41 propellant (nitrocellulose) was identified as an SRC in subsurface soil at the Pond Bank. No 42 explosives, VOCs, PCBs, or pesticides were detected or identified as SRCs in Pond Bank subsurface 43 soil surrounding Upper and Lower Cobbs Pond.

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1 ES.2.2.3 <u>Sediment</u>

3 Backwater Area

4

2

5 The Backwater Area was evaluated using 14 surface sediment and 2 subsurface sediment samples. Three explosives [2.4.6-trinitrotoluene (TNT); 2.6-dinitrotoluene (DNT); and tetry] and one 6 7 propellant (nitrocellulose) were detected in surface sediment. All of the detected concentrations were 8 below the Resident Receptor (Adult and Child) FWCUG and regional screening level (RSL) at a TR 9 of 1E-06, HQ of 0.1. No explosives or propellants were detected in subsurface sediment in the Backwater Area. In surface sediment, aluminum, cobalt, and cyanide concentrations exceeded a TR 10 of 1E-06, HO of 0.1, but not a TR of 1E-05, HO of 1. Only the cobalt concentration exceeded a TR of 11 1E-06, HQ of 0.1 in subsurface sediment in one subsurface soil location, but not at a TR of 1E-05, 12 13 HO of 1. Five PAHs [benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, 14 dibenzo(a,h)anthracene, and indeno(1,2,3-cd)pyrene] exceeded their respective SLs in surface 15 sediment. Four of the five PAHs were detected at their maximum concentrations in the Phase II RI 16 sample ULCPsd-015. The detected concentrations in the adjacent PBA08 RI sample CPCsd-047 17 indicate that concentrations have declined since the 2001 Phase II RI sampling event. Benzo(a)pyrene concentrations at ULCPsd-006, ULCPsd-015, and CPCsd-047 exceeded the Resident Receptor (Adult 18 19 and Child) FWCUG at a TR of 1E-05, HO of 1. Only benzo(a)pyrene was detected above its 20 respective SL (0.022 mg/kg) at PBA08 RI sample location CPCsd-047 in the 0.5-2 ft bgs interval. 21 The concentration was below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, 22 HQ of 1. Three VOCs (acetone, methylene chloride, and 2-butanone) were detected below their 23 respective SLs in surface and subsurface sediment, respectively. PCB-1254 also was detected in the 24 Phase II RI surface sediment samples ULCPsd-006 and ULCPsd-011; however, it was not detected in 25 the PBA08 RI samples. No pesticides were detected in surface sediment. Pesticides and PCBs were not detected in the Backwater Area subsurface sediment. 26

27

28 Upper Cobbs Pond

29

30 Five surface discrete sediment samples, one surface ISM sample, and one subsurface sediment sample 31 were used to evaluate Upper Cobbs Pond. Five explosives [1,3-dinitrobenzene; TNT; 4-amino-2.6-32 DNT; octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine (HMX); and tetryl] and one propellant 33 (nitrocellulose) were detected in surface sediment at PBA08 RI sample location CPCsd-046. All 34 detected concentrations were below the Resident Receptor (Adult and Child) FWCUG and RSL at a 35 TR of 1E-06, HO of 0.1. No explosives or propellants were detected in subsurface sediment. In 36 surface sediment, aluminum, hexavalent chromium, and cobalt concentrations exceeded a TR of 1E-37 06, HQ of 0.1, but not a TR of 1E-05, HQ of 1. Only the cobalt concentration exceeded a TR of 1E-06, HO of 0.1 in the subsurface sediment, but not at a TR of 1E-05, HQ of 1. Ten SVOCs were 38 39 detected in surface and subsurface sediment at the PBA08 RI sample location CPCsd-046. Two 40 additional SVOCs (acenaphthylene and anthracene) were detected in subsurface sediment. The 10 41 SVOCs occurred at higher concentrations in surface sediment than in the corresponding subsurface 42 sediment sample. Benzo(a)pyrene was detected above its respective SL in surface and subsurface 43 sediment at CPCsd-046; however, the concentrations were below the Resident Receptor (Adult and 44 Child) FWCUG at a TR of 1E-05, HQ of 1. Two VOCs (acetone and 2-butanone) were detected in

1 surface sediment below their respective SLs, both were detected at their maximum concentrations at

2 PBA08 RI location CPCsd-046. These VOCs also were detected at the 2001 Phase II RI sample

3 location ULCPsd-020 at lower concentrations than CPCsd-046. PCB-1254 also was detected in

4 surface sediment at Upper Cobbs Pond in ULCPsd-020 below its respective SL. PCB-1254 was not

5 detected in the PBA08 RI sample. Pesticides were not detected in surface sediment at Upper Cobbs

6 Pond. The VOC 2-butanone was detected in subsurface sediment below the TR of 1E-05, HQ of 1.

- 7 The pesticide delta-BHC does not have an SL. PCBs were not detected in subsurface sediment.
- 8

9 Lower Cobbs Pond

10

Lower Cobbs Pond was evaluated using eight discrete surface sediment samples, one surface ISM 11 12 sample, and one subsurface sediment sample. Two explosives (HMX and tetryl) and one propellant 13 (nitrocellulose) were detected in surface sediment, and the explosive HMX and propellant 14 nitrocellulose were detected in subsurface sediment at Lower Cobbs Pond. The concentrations were 15 below the Resident Receptor (Adult and Child) FWCUG and RSL at a TR of 1E-06, HO of 0.1. In 16 surface sediment, aluminum, arsenic, hexavalent chromium, cobalt, and cyanide concentrations 17 exceeded a TR of 1E-06, HQ of 0.1, but only arsenic exceeded a TR of 1E-05, HQ of 1. Only the 18 cobalt concentration exceeded a TR of 1E-06, HO of 0.1 in subsurface sediment, but not at a TR of 19 1E-05, HO of 1. A total of 11 SVOCs were detected in surface sediment in Lower Cobbs Pond and 18 20 SVOCs were detected in subsurface sediment. The 11 SVOC SRCs that were identified for surface 21 sediment also were observed in subsurface sediment and occurred at their highest concentrations in 22 the PBA08 RI subsurface sediment sample CPCsd-045. Benzo(a)pyrene concentrations exceeded its 23 respective SL in surface and subsurface sediment at CPCsd-045; however, only the subsurface 24 sediment concentration exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, 25 HQ of 1. Three VOCs (2-butanone, acetone, and carbon disulfide) were detected in surface sediment 26 at Lower Cobbs Pond. All three VOCs were detected at PBA08 RI location CPCsd-045, with 2-27 butanone and carbon disulfide observed at their maximum concentrations in this sample location. 2-28 Butanone also was detected in subsurface sediment. All of the VOCs detected in surface and 29 subsurface soil were below their respective SLs. No PCBs or pesticides were detected in surface or 30 subsurface sediment in Lower Cobbs Pond.

31

32 ES.2.2.4 Surface Water

33

34 Backwater Area

35

36 Four surface water samples were used to evaluate the Backwater Area. One explosive (4-amino-2,6-37 DNT) was detected in surface water at a low, estimated concentration below laboratory reporting 38 limits. No propellants were detected in the Backwater Area surface water. Four inorganic chemical 39 concentrations (arsenic, cobalt, manganese, and selenium) exceeded the RSL at a TR of 1E-06, HQ of 40 0.1. Manganese concentrations at Phase II RI sample locations ULCPsw-001 and ULCPsw-002 and 41 cobalt at ULCPsw-001 exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, 42 HQ of 1. Three VOCs (acetone, chloromethane, and toluene) were detected in surface water for the 43 Backwater Area, all of which were detected in ULCPsw-002. No propellants, SVOCs, PCBs, or 44 pesticides were detected in surface water at the Backwater Area. Elevated inorganic chemical 1 concentrations and 4-amino-2,6-DNT are likely attributable to continuing upstream sources as these

2 constituents are observed in the PBA08 RI locations immediately downstream of the confluence of

3 the drainage channels from Load Lines 3 and 12.

4

5 Upper Cobbs Pond

6

7 Surface water in Upper Cobbs Pond was evaluated using five surface water samples. One explosive 8 (4-amino-2,6-DNT) was detected at a low, estimated concentration in two surface water samples. The 9 concentration was below the Resident Receptor (Adult and Child) FWCUG and RSL at a TR of 1E-10 06, HO of 0.1. No surface water concentrations for inorganic chemicals in Upper Cobbs Pond 11 exceeded the SL at a TR of 1E-06, HO of 0.1. The SVOCs bis(2-ethylhexyl)phthalate and di-n-butyl 12 phthalate were detected at low, estimated concentrations below the laboratory reporting limit at 13 PBA08 RI sample location CPCsw-046 and the August 2003 FWBWQS location, respectively. No 14 VOCs, PCBs, or pesticides were detected in surface water at Upper Cobbs Pond.

16 Lower Cobbs Pond

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15

18 Lower Cobbs Pond surface water was evaluated using seven surface water samples. The explosive 4-19 amino-2,6-DNT was identified as an SRC in surface water at Lower Cobbs Pond. 4-Amino-2,6-DNT 20 was detected at a concentration of 0.00029J mg/L in the August 2003 FWBWQS sample at Lower 21 Cobbs Pond. 4-Amino-2,6-DNT was not detected in the subsequent PBA08 RI samples taken at 22 CPCsw-044 and CPCsw-045 in 2010. No propellants were detected. Only manganese was detected at 23 a concentration which exceeded the Resident Receptor (Adult and Child) FWCUG and RSL at a TR 24 of 1E-06, HQ of 0.1 at one sample location. Two SVOCs (benzenemethanol and butyl benzyl 25 phthalate) were detected at low, estimated concentrations at PBA08 RI location CPCsw-045, located 26 at the center of the aggregate. No VOCs, PCBs, or pesticides were detected in surface water in Lower 27 Cobbs Pond.

28

29 ES.2.3 Summary of Contaminant Fate and Transport

30

All SRCs identified in surface soil, subsurface soil, and sediment at the Upper and Lower Cobbs Ponds AOC were evaluated through the stepwise contaminant fate and transport evaluation. The evaluation included analyzing leaching and migration from soil and sediment to groundwater and determining whether contamination present in soil and sediment may potentially impact groundwater quality at the site.

36

37 Maximum SRC concentrations identified in surface and subsurface soil were evaluated using a series 38 of generic screening steps to identify initial contaminant migration chemicals of potential concern 39 (CMCOPCs). Initial CMCOPCs for soil were further evaluated using the Seasonal Soil Compartment 40 model to predict leaching concentrations and identify final CMCOPCs based on RVAAP facility-41 wide background concentrations and the lowest risk-based screening criteria among USEPA 42 maximum contaminant levels (MCLs), USEPA tap water RSLs, or RVAAP groundwater FWCUGs 43 for the Resident Receptor Adult. These final CMCOPCs were evaluated using the Analytical 44 Transient 1-, 2-, 3-Dimensional model to predict groundwater concentrations beneath source areas

- 1 and at the nearest downgradient groundwater receptor to the AOC (e.g., Cobbs Ponds) and identify
- 2 contaminant migration chemicals of concern (CMCOCs).
- 3

A sediment screening analysis was performed for sediment samples at the AOC. Chemical-specific dilution attenuation factors (DAFs) were calculated using co-located surface water and sediment concentrations for identified sediment SRCs. These DAFs were used in the sediment screening analysis to identify the initial CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria. The initial CMCOPCs for sediment were further evaluated by comparing their maximum detected surface water concentrations against their respective MCL/RSLs to identify the final sediment CMCOPCs.

11

12 The modeling results identified the following CMCOCs for soil and final CMCOPCs for sediment:

- 13
- 14 15

16

• Arsenic, nickel, selenium, and thallium in soil were predicted to exceed the screening criteria in groundwater beneath the source area, and only arsenic was predicted to exceed the screening criteria in groundwater at the downgradient receptor location.

Hexavalent chromium in Upper and Lower Cobbs Ponds sediment was predicted to exceed
 the screening criteria in groundwater beneath the source areas.

19

A qualitative assessment of the sample results and considerations of the limitations and assumptions of the models were performed to identify if any soil CMCOCs are present in soil that may impact the groundwater beneath the source or at the downstream receptor location (Cobbs Ponds) or if any CMCOPCs in sediment may impact groundwater beneath Cobbs Ponds. This qualitative assessment concluded that the remaining soil CMCOCs and sediment CMCOPCs are not expected to adversely impact groundwater at the AOC. No further action is required of soil and sediment at the Upper and Lower Cobbs Ponds AOC for the protection of groundwater.

27

28 29

ES.2.4 Summary and Conclusions of the Human Health Risk Assessment

The HHRA identified chemicals of concern (COCs) and conducted risk management analysis to determine if COCs pose unacceptable risk to the Resident Receptor. If there is no unacceptable risk to the Resident Receptor, it can be concluded that there is no unacceptable risk to the National Guard Trainee and Industrial Receptor.

34

Media of concern at the Upper and Lower Cobbs Ponds AOC are surface soil, subsurface soil, surface water, and sediment. Soil data were aggregated into surface and subsurface soil. Surface water and sediment data were aggregated into three exposure units (EUs) (Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond).

39

40 No COCs were identified in soil. Several PAHs were identified as COCs in sediment at the backwater

41 area. The maximum detected concentration (MDC) of benzo(a)pyrene (0.89 mg/kg) at the Backwater

42 Area exceeds the FWCUG (0.221 mg/kg). The MDC is at ULCPsd-015, which was collected in 2001.

- 43 Sample ULCPsd-047 was collected at approximately the same location in 2010 and showed a
- 44 benzo(a)pyrene concentration of 0.39 mg/kg. Detected benzo(a)pyrene concentrations in the other

1 Backwater Area samples ranged from 0.063–0.23 mg/kg. Three additional PAHs, present below 2 FWCUGs, were identified as COCs based on the sum-of-ratios analysis. Benzo(a)pyrene, and by 3 association the other PAHs, were not identified as COCs to be carried forward for potential 4 remediation because the elevated PAH concentrations are located near a culvert between Load Line 3 5 Road and Track 33. The source of the PAHs in sediment is most likely runoff from the road and railroad bed. The FWCUGs for exposure to sediment are the same as the FWCUGs for soil (i.e., 6 7 exposure is assumed to be the same for soil in a residential yard and sediment in a pond). However, 8 the MDC at the backwater area (0.89 mg/kg) is below a more realistic (but still conservative) 9 calculated RSL of 9 mg/kg.

10

11 The exposure point concentration (EPC) of arsenic in sediment at Lower Cobbs Pond (26.9 mg/kg) exceeds the FWCUG for the Resident Receptor Adult of 4.25 mg/kg, the facility-wide background 12 13 concentration of 19.5 mg/kg, and the Ohio EPA sediment reference value (SRV) of 25 mg/kg. The 14 MDC of arsenic (34.3 mg/kg) at ULCPsd-026, collected in 2001, is the only sample concentration 15 greater than the facility-wide background concentration and Ohio EPA SRV. Reported arsenic 16 concentrations in the other Lower Cobbs Pond samples ranged from 5.1-20.2 mg/kg. Arsenic was not 17 identified as a COC for potential remediation because the reported concentrations appear to represent 18 naturally occurring levels in sediment.

19

The EPC of cobalt (0.01 mg/L) is less than two times the tap water RSL of 0.006 mg/L. The Backwater Area is a shallow pond created through beaver dam construction activity upstream of Upper Cobbs Pond. This limited shallow surface water is not a potential source of residential drinking water; therefore, this low exceedance of the tap water RSL does not warrant potential remediation in this EU.

25

The EPC of manganese in surface water at the Backwater Area EU exceeds the FWCUG for the Resident Receptor (Adult and Child) of 6.326 mg/L. The EPC is strongly influenced by the elevated concentration reported in a single sample collected in 2001. The manganese concentrations in all samples collected in 2010 are below the facility-wide background concentration of 0.391 mg/L. Because recent sample data indicate manganese is present at naturally occurring background concentrations, manganese does not warrant potential remediation in this EU.

32

Based on the risk management analysis, no COCs were identified to be carried forth in an FS for the
Resident Receptor (Adult and Child) in any of the media of concern at the Upper and Lower Cobbs
Pond AOC; therefore, no other receptors were evaluated and no further action is recommended from a
human health risk perspective.

- 37
- 38

ES.2.5 Summary and Conclusions of the Ecological Risk Assessment

39

40 The Level I ERA presents important ecological resources on or near the AOC and evaluates the 41 potential for current contamination to impact ecological resources. There is chemical contamination 42 present in surface soil, sediment, and surface water at Upper and Lower Cobbs Ponds AOC. This 43 contamination was identified using historical and PBA08 RI data and data from the FWBWQS. 44 Although the ponds constitute much of the AOC, red maple successional forest (dominant vegetation type); dry, mid-successional, cold-deciduous shrubland; three types of herbaceous communities; and three additional types of forests were observed on the 39 acres in the AOC. There are important and significant ecological resources in the AOC. Specifically, a state-listed threatened species (Least Bittern) and its habitat, wetlands, and surface water (i.e., ponds) are present and near contamination.

- 5 These findings invoked a requirement of a Level II ERA.
- 6

7 The Level II ERA evaluated integrated chemicals of potential ecological concern (COPECs) in soil, sediment, and surface water. Integrated COPECs are identified by screening PBA08 RI, FWBWQS, 8 9 and historical data sets against ecological screening values. Twelve integrated COPECs were 10 identified for soil. Forty integrated COPECs were identified for sediment. Eight integrated COPECs 11 were identified for surface water. The integrated soil, sediment, and surface water COPECs were further evaluated with technical and refinement factors in Step 3A. The factors in Step 3A showed 12 13 there are no integrated COPECs that are of ecological concern requiring remediation or further evaluation. Consequently, the Level II Screening ERA for Upper and Lower Cobbs Ponds AOC 14 15 concludes with a recommendation that no further action is required to be protective of important 16 ecological resources.

- 17
- 18

ES.2.6 Recommendation of the Remedial Investigation

19

20 Based on the investigation results, the Upper and Lower Cobbs Ponds AOC has been adequately 21 characterized and the recommended path forward is no further action for soil, sediment, and surface 22 water to attain Unrestricted (Residential) Land Use. Further investigation is not warranted at this 23 AOC for the following reasons: (1) the current nature and extent of impacted media has been 24 sufficiently characterized; (2) the fate and transport modeling did not identify soil CMCOCs requiring 25 further evaluation or remediation to protect groundwater; (3) there are no CERCLA-release-related 26 human health COCs identified in soil, sediment, or surface water requiring further evaluation in an FS 27 or additional remediation; and (4) remedial actions to protect ecological resources are not warranted.

28

The next step in the CERCLA process is to prepare a PP to solicit public input with respect to no further action for soil, sediment, and surface water. The PP will briefly summarize the history, characteristics, risks, and the basis for no further action. Comments on the PP received from state and

32 federal agencies and the public will be considered in preparing a ROD to document the final remedy.

33 The ROD will also include a responsiveness summary addressing comments received on the PP.

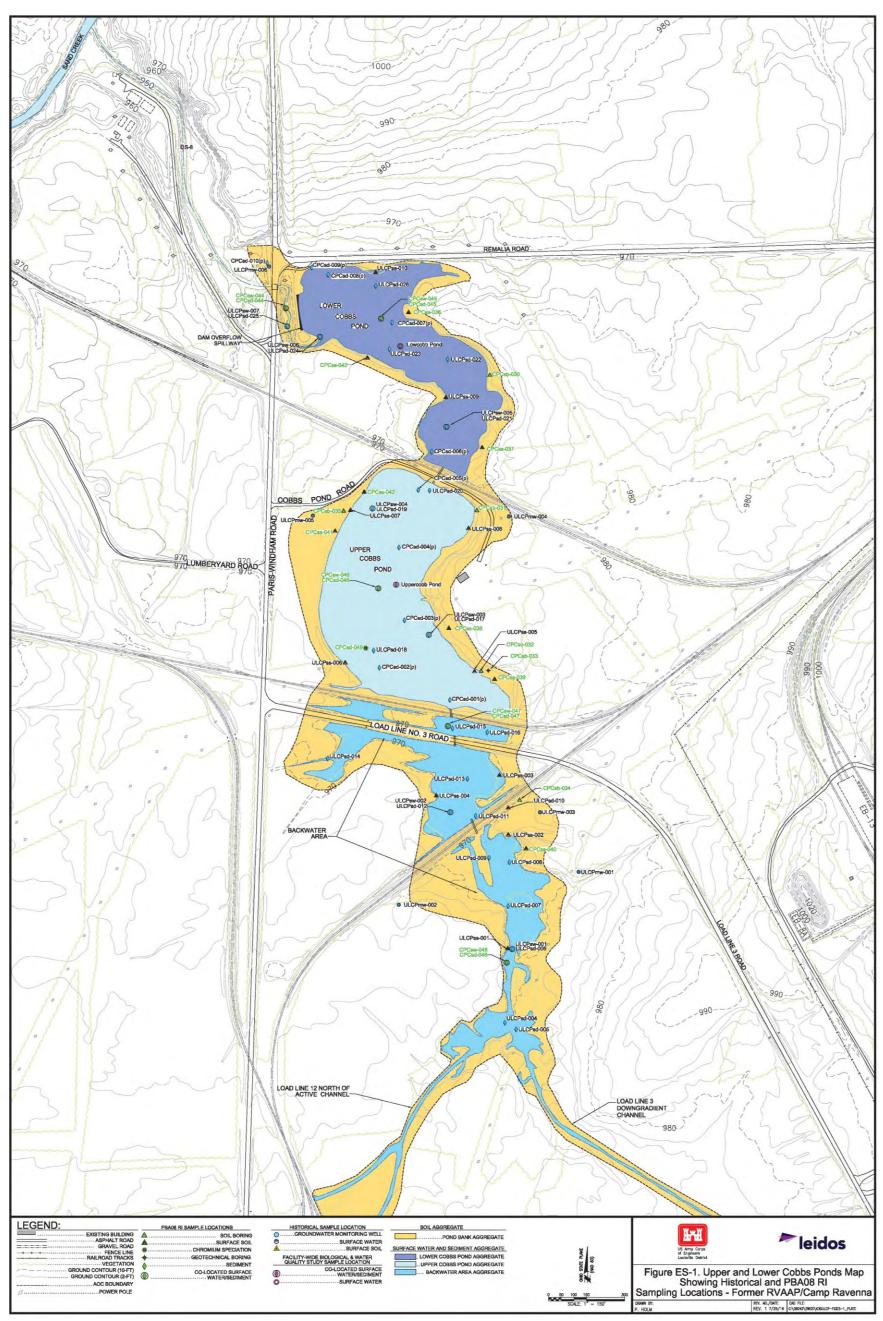


Figure ES-1. Upper and Lower Cobbs Ponds Map Showing Historical and PBA08 RI Sampling Locations - Former RVAAP/Camp Ravenna

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1 1.0 INTRODUCTION

2

This document was revised by Leidos under the U.S. Army Corps of Engineers (USACE) Louisville
District Contract Number W912QR-15-C-0046. This Phase III Remedial Investigation (RI) Report
addresses soil, sediment, and surface water at the Upper and Lower Cobbs Ponds area of concern
(AOC) within the former Ravenna Army Ammunition Plant (RVAAP) [now known as Camp
Ravenna Joint Military Training Center (Camp Ravenna)] in Portage and Trumbull counties, Ohio
(Figures 1-1 and 1-2). The Upper and Lower Cobbs Ponds AOC is designated as AOC RVAAP-29.
This report has been prepared in accordance with the requirements of the Ohio Environmental

11 Protection Agency (Ohio EPA) Director's Final Findings and Orders (DFFO) for RVAAP, dated 12 June 10, 2004 (Ohio EPA 2004). The DFFO requires conformance with the Comprehensive 13 Environmental Response, Compensation, and Liability Act (CERCLA) and National Contingency 14 Plan (NCP) to implement an RI to characterize the AOC, develop a feasibility study (FS) (if 15 remediation is necessary) to evaluate remedial alternatives to address contamination presenting 16 unacceptable risk to human health and the environment, present a preferred remedial alternative in a 17 proposed plan (PP), and document stakeholder selection and acceptance of the preferred final remedy 18 in a record of decision (ROD).

19 20

This document includes the following:

21 22

• A description of the operational history and environmental setting for the AOC.

- A summary of all historical assessments and investigations at the Upper and Lower Cobbs
 Ponds AOC.
- A description of the nature and extent of contamination, including the identification of site related contaminants (SRCs) by screening applicable data against background, essential
 human nutrients, and frequency of detection/weight-of-evidence (WOE) screening.
- An evaluation of contaminant fate and transport by identifying contaminant migration
 chemicals of potential concern (CMCOPCs) and contaminant migration chemicals of concern
 (CMCOCs) that may pose a future threat to groundwater.
- A human health risk assessment (HHRA) to identify chemicals of potential concern (COPCs)
 and chemicals of concern (COCs).
- An ecological risk assessment (ERA) to identify chemicals of potential ecological concern
 (COPECs) and chemicals of ecological concern.
- Conclusions of the Phase III RI Report.
- 36

37 1.1 PURPOSE

38

39 The purpose of this report is to use available RI data to evaluate the nature and extent of 40 contamination; fate and transport of contaminants in the environment; and risk assessments for 41 surface and subsurface soil, sediment, and surface water at the Upper and Lower Cobbs Ponds AOC. 42 This report also provides a summary of the Performance-Based Acquisition 2008 Remedial 43 Investigation (PBA08 RI) at the Upper and Lower Cobbs Ponds AOC that was performed to 1 supplement data from previous sampling events. Depending on the results of the evaluations 2 contained in this report, a conclusion of no further action is provided or a recommendation to 3 complete an FS to evaluate potential remedies and future actions will be made.

4

1.2 SCOPE

5 6

7 The scope of this report to present: (1) the nature and extent of contamination, fate and transport of 8 contaminants in the environment, and risk assessments for surface soil, subsurface soil, sediment, and 9 surface water at the AOC; (2) the results of the evaluation of remedial alternatives for meeting 10 remedial action objectives for any CERCLA-related COCs identified in the media at the AOC; and 11 (3) a conclusion of no further action or a preferred alternative to present to the public in a PP.

12

13 For the purposes of this report, the term "surface soil" includes dry sediment. Dry sediment refers to 14 unconsolidated inorganic and organic material within conveyances, ditches, or low-lying areas that 15 occasionally may be covered with water, usually following a precipitation event or due to snowmelt. 16 Dry sediment is not covered with water for extended periods and typically is dry within seven days of 17 a precipitation event. Dry sediment does not function as a permanent habitat for aquatic organisms, 18 although it may serve as a natural medium for the growth of terrestrial organisms. Dry sediment is 19 addressed the same as surface soil [0-1 ft below ground surface (bgs)] in terms of contaminant nature 20 and extent, fate and transport, and risk exposure models. The term "sediment," as used in this report, 21 refers to wet sediment within conveyances, ditches, wetlands, or water bodies that is inundated for 22 extended periods. This terminology is consistent with the Facility-wide Human Health Cleanup 23 Goals for the Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 2010a), herein referred to as 24 the FWCUG Report.

25

Potential impacts to groundwater from soil (e.g., contaminant leaching) are evaluated in this report, as protectiveness to groundwater is included in the fate and transport analysis and the evaluation of remedial alternatives for these media. Groundwater will be evaluated as an individual AOC for the entire facility (designated as RVAAP-66) and addressed in a separate RI/FS Report.

30

32

31 1.3 REPORT ORGANIZATION

This report is organized in accordance with Ohio EPA and U.S. Environmental Protection Agency (USEPA) CERCLA RI/FS guidance and applicable USACE guidance. The following is a summary of the components of the report and a list of appendices:

- 36
- Section 2.0 provides a description and history of the former RVAAP and the AOC, presents
 potential sources of contamination, presents potential receptors, and summarizes co-located
 or proximate sites.
- Section 3.0 describes the environmental setting at Camp Ravenna and the Upper and Lower
 Cobbs Ponds AOC, including the geology, hydrogeology, climate, and receptor population.
- Section 4.0 summarizes previous assessments and investigations at the Upper and Lower
 Cobbs Ponds AOC, as well as the data used to support the Phase III RI.
- Section 5.0 discusses the occurrence and distribution of contamination at the AOC.

- Section 6.0 presents an evaluation of contaminant fate and transport.
- Section 7.0 includes the methods and results of the HHRA and ERA.
 - Section 8.0 provides the conclusions and recommendations of the Phase III RI.
- Section 9.0 summarizes the framework for conducting the necessary agency and public
 involvement activities.
- Section 10.0 provides a list of references used to develop this report.
- 7 Appendices:
- 8 Appendix A: Field Sampling Logs,
- 9 Appendix B: Project Quality Assurance Summary,
- 10 Appendix C: Data Quality Control Summary Report,
- 11 Appendix D: Laboratory Analytical Results and Chains-of-Custody,
- 12 Appendix E: Fate and Transport Modeling Results,
- 13 Appendix F: Investigation-derived Waste Management Reports,
- 14 Appendix G: Human Health Risk Assessment Tables,
- 15 Appendix H: Ecological Risk Assessment Information and Data, and
- 16 Appendix I: PBA08 Remedial Investigation Summary.
- 17

3

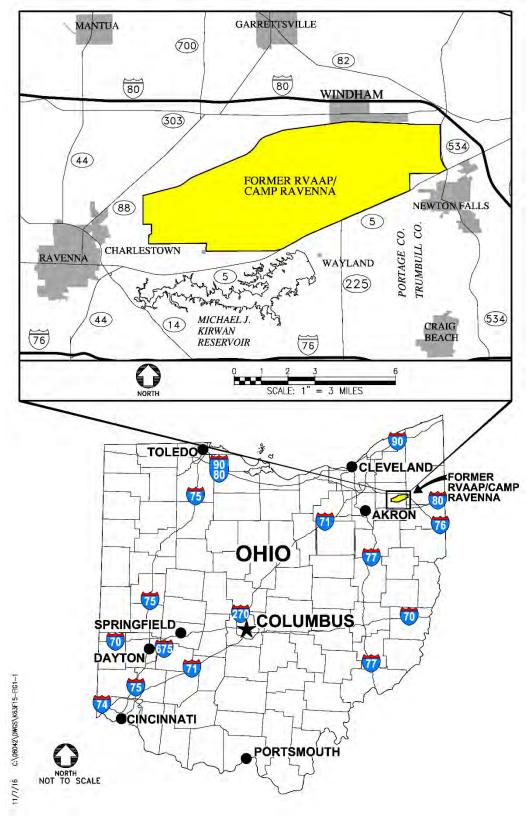


Figure 1-1. General Location and Orientation of Camp Ravenna

1 2

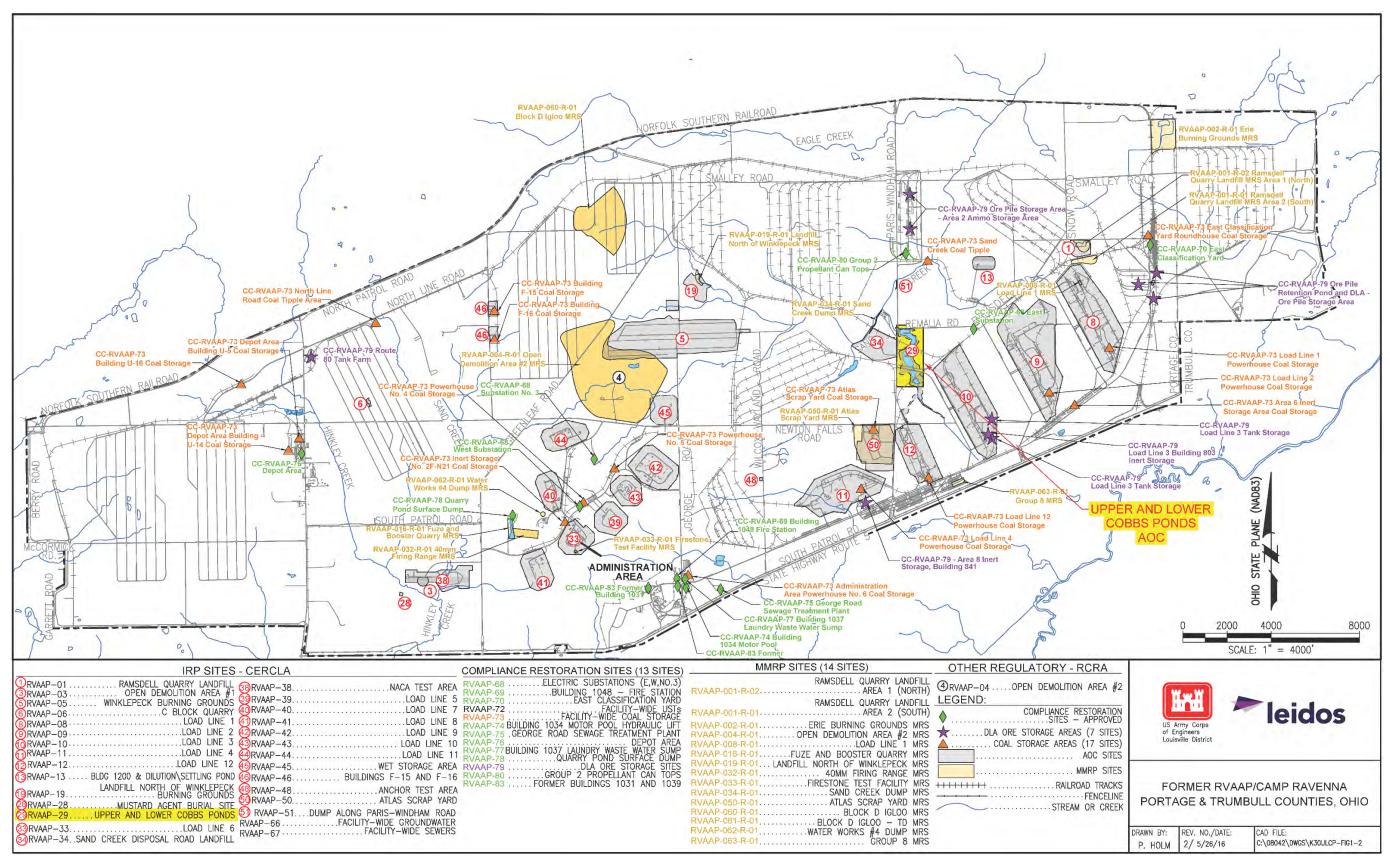


Figure 1–2. Location of AOCs and Munitions Response Sites at Camp Ravenna

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1 2.0 BACKGROUND

2

3 This section provides a description of the facility. In addition, it summarizes the Upper and Lower 4 Cobbs Ponds AOC's operational history, potential sources, potential human health receptors and 5 ecological resources, and co-located or proximate sites.

6 7

2.1

8

9

10

2.1.1 General Facility Description

FACILITY-WIDE BACKGROUND INFORMATION

11 The facility, consisting of 21,683 acres, is located in northeastern Ohio within Portage and Trumbull 12 counties, approximately 4.8 kilometers (3 miles) east/northeast of the city of Ravenna and 13 approximately 1.6 kilometers (1 mile) northwest of the city of Newton Falls (Figure 1-1). The facility, 14 previously known as RVAAP, was formerly used as a load, assemble, and pack facility for munitions 15 production. As of September 2013, administrative accountability for the entire acreage of the facility 16 has been transferred to the U.S. Property and Fiscal Officer (USP&FO) for Ohio and subsequently 17 licensed to the Ohio Army National Guard (OHARNG) for use as a military training site (Camp 18 Ravenna). References in this document to RVAAP relate to previous activities at the facility as 19 related to former munitions production activities or to activities being conducted under the 20 restoration/cleanup program.

21

22 In 1978, the U.S. Army Toxic and Hazardous Materials Agency conducted an Installation Assessment 23 of RVAAP to review the potential for contaminant releases at multiple former operations areas, as 24 documented in the Installation Assessment of Ravenna Army Ammunition Plant (USATHAMA 25 1978). The assessment identified that pink water and washout of residue, dusts, and spills at Load 26 Lines 3 and 12 were discharged into Upper Cobbs Pond through surface drainage channels. Load 27 Line 3 was used for 2,4,6-trinitrotoluene (TNT) and composition B a combination of TNT and 28 hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX)] melt-pour activities, and Load Line 12 was used for 29 ammonium nitrate production. Both load lines were also used for demilitarization activities.

30

31 The former RVAAP received bulk TNT product during operational activities and did not 32 manufacture/produce dinitrotoluene (DNT) or TNT. A facility where DNT is manufactured will have 33 the following isomers of DNT in the finished product: 2,4-DNT; 2,6-DNT; 2,5-DNT; 3,4-DNT; 2,3-34 DNT; and 3,5-DNT. This is not applicable to the former RVAAP. Degradation of TNT to 2,4-DNT 35 occurs in soil; however, 2,4-DNT and 2,6-DNT do not degrade to the lesser isomers. It is the U.S. 36 Department of the Army's (Army) position that testing DNT isomers other than 2,4-DNT and 2,6-37 DNT is unnecessary and has no additional value of being protective to human health and the 38 environment at the former RVAAP (RVAAP 2013).

39

40 2.1.2 Demography and Land Use

41

42 Camp Ravenna occupies east-central Portage County and southwestern Trumbull County. Census
 43 projections for 2010 indicated the populations of Portage and Trumbull counties are 161,419 and

1 210,312, respectively. Population centers closest to Camp Ravenna are Ravenna, with a population of

- 2 11,724, and Newton Falls, with a population of 4,795.
- 3

The facility is located in a rural area and is not close to any major industrial or developed areas. Approximately 55% of Portage County, in which the majority of Camp Ravenna is located, consists of either woodland or farmland acreage. The closest major recreational area, the Michael J. Kirwan Reservoir (also known as West Branch Reservoir), is located adjacent to the western half of Camp Ravenna, south of State Route 5.

9

Camp Ravenna is federally owned and is licensed to OHARNG for use as a military training site. Restoration activities at Camp Ravenna are managed by the Army National Guard and OHARNG. Training and related activities at Camp Ravenna include field operations and bivouac training, range firing activities, convoy training, maintaining equipment, C-130 aircraft drop zone operations, helicopter operations, and storing heavy equipment.

UPPER AND LOWER COBBS PONDS AOC BACKGROUND INFORMATION

- 15
- 16

2.2

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18 2.2.1 Operational History

The Upper and Lower Cobbs Ponds AOC is located in the east-central portion of Camp Ravenna. The AOC is located east of Paris-Windham Road and south of Remalia Road, north of Load Line 12, and northwest of Load Line 3 (Figures 1-2, 2-1, and 2-2). The AOC is approximately 39 acres and is comprised of approximately 18 acres of pond banks and the surface water features presented below from upstream to downstream:

25 26

27

28

• Backwater Area, 5.2 acres.

- Upper Cobbs Pond, 9.4 acres, ranges from 3–8 ft in depth, and is also called Big Cobbs Pond or South Cobbs Pond.
- Lower Cobbs Pond, 6.4 acres, 2–7 ft in depth, and is also called Little Cobbs Pond or North
 Cobbs Pond.
- 31

Three railroad trackbeds dissect the AOC (the railroad tracks have been removed) and are described
 below:

- 34 35
- The trackbed for Track 25 separates Lower Cobbs Pond from Upper Cobbs Pond.
- The trackbed for Track 33 is north of Load Line 3 Road and separates Upper Cobbs Pond
 from the Backwater Area.
- 38 39

40

•

Upper Cobbs Pond and Lower Cobbs Pond were constructed in 1940-1941, expanding a natural

The trackbed for Track FA runs through the Backwater Area.

- 41 drainage conveyance to receive effluent discharge and to serve as the unlined sedimentation basins 42 for Load Lines 3 and 12. The surface water conveyance is depicted in Figure 2-1. The ponds were
- 43 utilized as unlined sedimentation basins after improvements, such as flow control structures, were

1 constructed throughout the complex. The Backwater Area did not exist when the facility was

- 2 operational, rather it is a product of beaver dam construction activity (Jacobs 1989).
- 3

Flow between the Backwater Area and Upper Cobbs Pond is controlled through a culvert located
under the Track 33 and Load Line 3 Road crossings. Surface water in Lower Cobbs Pond discharges
to a dam overflow spillway adjacent to the intersection of Remalia and Paris Windham Roads.

7

8 From 1941–1971, the ponds received effluent from the Load Lines 3 and 12 sawdust filtration units, 9 wash water, storm water runoff, and surface water runoff. Rinsate from demilitarization operations at 10 Load Lines 3 and 12 was initially allowed to flow out of buildings and directly onto the ground or to drainage ditches which ultimately discharged to Upper Cobbs Pond and Lower Cobbs Pond. Load 11 12 Line 12 was formerly utilized for producing ammonium nitrate from 1941–1943 and 1946–1950. 13 From 1951–1961, explosive melt-out and demilitarization activities occurred at Load Line 12. Since 14 there were no wash water collection tanks or settling ponds in Load Line 12 during these operations, 15 all residues, dusts, and spills were washed into the drainage system that eventually discharged into 16 Upper Cobbs Pond.

17

During Load Line 3's operational history, 6.5 million munitions were produced that generated approximately 79,248 gal of pink water each month when the facility was operating at full capacity (USACE 1996a). The contaminated pink water washout at Load Lines 3 and 12 was collected in concrete settling sumps and treated through sawdust filters before being discharged into Upper Cobbs Pond through surface drainage channels.

23

24 From 1965-1967, Hercules Alcor, Inc. leased Building FF-19 in Load Line 12 for producing 25 aluminum chloride. On November 15, 1966, a fish kill occurred at Lower Cobbs Pond as a result of improper handling of aluminum chloride during manufacturing operations. The bulk of the aluminum 26 27 chloride was collected and disposed at Ramsdell Quarry Landfill (RVAAP-01). The pond, receiving 28 the contaminating waste from drainage ditches, was settled, drained, and the contaminants were 29 removed to Ramsdell Quarry. Contaminated metals were flashed at a burning ground to a 5X 30 condition (USATHAMA 1978 and 1982, RVAAP 2000). The Army terminated the lease early on 31 December 3, 1967 due to environmental concerns related to air emissions and wastewater discharges 32 to Upper and Lower Cobbs Ponds.

33

34 2.2.2 Potential Sources

35

36 No primary sources are present within the AOC. The existing features include two sedimentation 37 basins (ponds), primary drainage conveyances, and flow control structures (e.g., dam overflow 38 spillway).

39

The 1978 Installation Assessment identified the pink wastewater and washout of residue, dusts, and spills at Load Lines 3 and 12, which were discharged through surface drainage channels towards Upper Cobbs Pond (USATHAMA 1978), as contaminants. Potential site-specific contaminants include TNT; RDX; octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine (HMX); nitrate; nitrocellulose; and heavy metals (i.e., lead, chromium, mercury, and arsenic) from the pink water. Other 1 contaminants identified include washout from explosives melt-out and demilitarization activities at

- 2 Load Lines 3 and 12 and ammonium nitrate and aluminum chloride from Load Line 12.
- 3 4

5

Many other chemicals were analyzed in the site investigations and are discussed in this report.

6 2.2.3 AOC Boundary

The Upper and Lower Cobbs Ponds AOC is located in the east-central portion of Camp Ravenna. The AOC is located east of Paris-Windham Road and south of Remalia Road, north of Load Line 12, and northwest of Load Line 3 (Figures 1-2 and 2-1). No fences or perimeter boundaries exist at the AOC; however, the AOC boundary encompasses the limits of investigation of the Phase I and II RIs and PBA08 RI, as presented in Figure 2-2. The AOC is approximately 39 acres and is comprised of the Backwater Area (5.2 acres), Upper Cobbs Pond (9.4 acres), Lower Cobbs Pond (6.4 acres), and surrounding pond banks (18 acres).

15

16 2.2.4 Spatial Aggregates

17

18 Upper and Lower Cobbs Ponds AOC data were aggregated for evaluating contaminant nature and 19 extent, human health, and the environment. Spatial aggregates established for this evaluation are 20 discussed below and are presented on Figure 2-2.

21

22 Surface and subsurface soil within the geographic area of the AOC was evaluated as an AOC-wide 23 spatial aggregate, designated as the Pond Bank aggregate. The Pond Bank aggregate is defined as the 24 perimeter of the natural shorelines surrounding Upper Cobbs Pond and Lower Cobbs Pond. The Pond 25 Bank also extends from the Backwater Area to the south at the confluence of the channels exiting 26 Load Lines 3 and 12 and north to the intersection of Remalia Road and Paris-Windham Road. The 27 areas encompassed within the Pond Bank are dry the majority of the time, although they may be 28 intermittently wet in association with heavy rainfall events when they may receive runoff from the 29 surrounding areas or when water levels within the AOC are elevated.

30

31 Sediment and surface water spatial aggregates identified were the Backwater Area, Upper Cobbs 32 Pond, and Lower Cobbs Pond. Sediment samples included in the risk assessments are limited to 33 sediment samples collected with a starting depth of 0 ft bgs (e.g., 0-0.5 or 0-1.75 ft bgs). All three of 34 these aggregates are perennially inundated water bodies. The Backwater Area is in the southern 35 (upstream) portion of the complex and extends from the confluence of the drainage channels exiting 36 Load Lines 3 and 12 and reaching north to Track 33. Upper Cobbs Pond is the portion of the water 37 body north of Track 33, extending to Cobbs Pond Road. Lower Cobbs Pond (furthest downstream) is 38 north of the Cobbs Pond Road, bounded by Remalia Road on the north and Paris-Windham Road to 39 the west.

40

41 **2.2.5** Current Land Use

42

The Upper and Lower Cobbs Ponds AOC is currently utilized for recreational purposes, as it contains abundant fish and wildlife. OHARNG and facility workers use the ponds for fishing. Currently, due to the fact that the investigative activities are still ongoing at the AOC under the restoration program,
fishing in the ponds is catch and release. The pond bank areas are lightly vegetated and covered with
grass. A picnic area, pavilion, and playground are located approximately 100 ft from the edge of the
eastern bank of Upper Cobbs Pond.

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- 6 7

8

2.3 POTENTIAL HUMAN RECEPTORS AND ECOLOGICAL RESOURCES AT UPPER AND LOWER COBBS PONDS AOC

- 9 The following sections discuss potential human receptors and ecological resources at the Upper and 10 Lower Cobbs Ponds AOC.
- 11

12 2.3.1 Human Receptors

13

The Upper and Lower Cobbs Ponds AOC is located in the east-central portion of the facility and is currently used for recreational purposes (Figure 1-2). Future OHARNG Land Use for the AOC (which includes the ponds and their associated sediment and banks) includes recreational use, dam maintenance, wildlife management activities, and training purposes. Surrounding adjacent areas may be used for recreational activities as well as maneuver and multipurpose military training operations. This RI Report evaluates the Resident Receptor (Adult and Child) (formerly called the Resident Farmer) to assess Unrestricted (Residential) Land Use.

21

22 2.3.2 Ecological Resources

23

Camp Ravenna has a diverse range of vegetation and habitat resources. Habitats present within the
 facility include large tracts of closed-canopy hardwood forest, scrub/shrub open areas, grasslands,
 wetlands, open-water ponds and lakes, and semi-improved administration areas (OHARNG 2014).

27

An abundance of wildlife is present on the facility: 35 species of land mammals, 214 species of birds, 41 species of fish, and 34 species of amphibians and reptiles have been identified. The ponds support a variety of aquatic animals (e.g., fish, turtles, and frogs) and semi-aquatic wildlife, such as waterfowl (e.g., ducks and geese) and wading birds (e.g., great blue heron).

32

The northern long-eared bat (*Myotis septentrionalis*; federally threatened) exists at Camp Ravenna. There are no other federally listed species and no critical habitat occurs (OHARNG 2014). Ohio statelisted plant and animal species have been identified through confirmed sightings and/or biological inventories at the facility and are presented in Table 2-1. Currently, the AOC is surrounded by forest, grassland, and wetland habitat. Additional information specific to ecological resources at the Upper and Lower Cobbs Ponds AOC is included in Section 7.3.

- 39
- 40 41

2.4 CO-LOCATED OR PROXIMATE SITES

- The following subsections summarize sites that are co-located or proximate to the Upper and Lower Cobbs Ponds AOC, but are addressed separately.
- 44

1 2.4.1 Facility-wide Sewers

| 2 | |
|----------|--|
| 3 | There are no facility-wide sewers within or adjacent to the AOC boundary. |
| 4 | |
| 5 | 2.4.2 Facility-wide Groundwater |
| 6 | |
| 7 | As part of the Installation Restoration Program, the Army implements the Facility-wide Groundwater |
| 8 9 | Monitoring Program (FWGWMP) in accordance with previous agreements made with Ohio EPA. The FWGWMP was initiated in 2005 and involves quarterly and semi-annual sampling of selected |
| 10 | wells within the former RVAAP. From 2008–2009 for the FWGWMP, groundwater samples were |
| 11 | collected from each of the six wells (ULCPmw-001 through ULCPmw-006) associated with the |
| 12 | Upper and Lower Cobbs Ponds AOC. These wells are shown on Figure 2-2. |
| 13 | |
| 14 15 | In January 2011, additional groundwater data from the wells presented below were collected under the EWCWAR as indicated in Attachment P of the English wide Crowndwater Monitoring Program |
| 15 16 | the FWGWMP, as indicated in Attachment B of the Facility-wide Groundwater Monitoring Program Report on the January 2011 Sampling Event (EQM 2011): |
| 17 | |
| 18 | • ULCPmw-002 – Bis(2-ethylhexyl)phthalate was detected at a concentration above the |
| 19 20 | cleanup goal (CUG), |
| 20 | • ULCPmw-005 – There were indications that barium concentrations were increasing, and |
| 21 | • ULCPmw-006 – Manganese and naphthalene were detected at a concentration above the |
| 22 23 | CUG. |
| 23 24 | Most chemical concentrations in groundwater were below the maximum contaminant level (MCL) or |
| 25 | regional screening level (RSL) [target risk (TR) of 1E-05, hazard quotient (HQ) of 1]. The exceptions |
| 26 | were (EQM 2015): |
| 27 | |
| 28 | • Arsenic exceeded the MCL of 10 μ g/L in each sample collected at ULCPmw-005 with a |
| 29 | maximum concentration of 65 μ g/L. All samples collected from ULCPmw-006 were below |
| 30 | the MCL except the sample collected in October 2009 that had a concentration of 18.7 μ g/L. |
| 31 32 | • Manganese in ULCPmw-006 exceeded the RSL of 1E-05 (430 µg/L) in all samples collected with a maximum concentration of 3,260J µg/L from the January 2009 sample. |
| 33 | • Bis(2-ethylhexyl)phthalate in ULCPmw-002 exceeded the MCL (6 µg/L) in one sample with |
| 34 | a concentration of 16 μ g/L. The subsequent two samples had concentrations below the MCL |
| 35 | and RSL of 1E-05 (5.6 µg/L). |
| 36 | • Naphthalene in ULCPmw-006 exceeded the RSL of 1E-05 (0.17 μ g/L) in one sample in the |
| 37 | final quarterly sample collected in January 2009; however, the detected concentration was |
| 38 | below the reporting limit (J-flagged). All other samples from this well had non-detectable |
| 39 | concentrations of naphthalene. |
| 40 | • 2,6-DNT exceeded the RSL (TR of 1E-05, HQ of 1, 0.048 µg/L) in ULCPmw-003 from the |
| 41 | first two samples collected in 2008. These samples had detected concentrations below the |
| 42 | reporting limit (J-flagged). All subsequent samples from this well did not have detectable |
| 43 | concentrations of 2,6-DNT. |
| 44 | |

Additional groundwater level monitoring was performed in May 2014 at the six monitoring wells at the Upper and Lower Cobbs Ponds AOC; however, no samples were collected (EQM 2015). Facilitywide groundwater is currently at the RI phase of the CERCLA process. Any future decisions or actions respective to groundwater at the Upper and Lower Cobbs Ponds AOC will be addressed as part of that facility-wide AOC.

6 7

2.4.3 Munitions Response Sites

8 9

There is no munitions response site within or adjacent to the AOC boundary identified as part of the Military Munitions Response Program.

10 11

12 2.4.4 Compliance Restoration Sites

13

14 There are no compliance restoration sites, such as former or existing underground storage tanks, 15 within or adjacent to the AOC boundary.

- 16
- 17

Table 2–1. Federal- and State-listed Species List

| | CAMP RAVENNA JOINT MILITARY TH | RAIN | ING CENTER RARE SPECIES LIST |
|------|---|-------|---|
| | Decemb | er 20 |)14 |
| | | | |
| I. S | pecies confirmed to be on Camp Ravenna property | by b | iological inventories and confirmed sightings. |
| | A. Federal Threatened | | |
| 1. N | Northern long-eared bat, Myotis septentrionalis | | |
| | B. State Endangered | | |
| 1. | American bittern, Botaurus lentiginosus | 8. | Tufted Moisture-loving Moss, Philonotis |
| | (migrant) | | Fontana var. caespitosa |
| 2. | Northern harrier, Circus cyaneus | 9. | Appalachian quillwort, Isoetes engelmannii |
| 3. | Sandhill Crane, Grus Canadensis (probable | 10. | Handsome sedge, Carex formosa |
| | nester) | 11. | Narrow-necked Pohl's Moss, Pohlia elongata |
| 4. | Black bear, Ursus americanus | | var. elongate |
| 5. | Mountain Brook Lamprey, Ichthyomyzon | 12. | Philadelphia panic-grass, Panicum |
| | greeleyi | | philadelphicum |
| 6. | Brush-tipped emerald, Somatochlora walshii | 13. | Variegated scouring-rush, Equisetum |
| 7. | Graceful Underwing, Catocala gracilis | | variegatum |
| | C. State Threatened | | |
| 1. | Barn owl, Tyto alba | 6. | Northern long-eared bat, Myotis septentrionalis |
| 2. | Least Bittern, Ixobrychus exilis | 7. | Hobblebush, Viburnum alnifolium |
| 3. | Trumpeter swan, Cygnus buccinators (migrant) | 8. | Simple willow-herb, Epilobium strictum |
| 4. | Bobcat, Felis rufus | 9. | Lurking leskea, Plagiothecium latebricola |
| 5. | Caddis fly, Psilotreta indecisa | 10. | Strict blue-eyed grass, Sisyrinchium montanum |
| | D. State Potentially Threatened Plants | | |
| 1. | Arborvitae, Thuja occidentalis | 6. | Sharp-glumed manna-grass, Glyceria acutifolia |
| 2. | False hop sedge, Carex lupiliformis | 7. | Straw sedge, Carex straminea |
| 3. | Greenwhite sedge, Carex albolutescens | 8. | Water avens, Geum rivale |
| 4. | Long Beech Fern, Phegopteris connectilis | 9. | Woodland Horsetail, Equisetum sylvaticum |
| | (Thelypteris phegopteris) | 10. | Shining ladies'-tresses, Spiranthes lucida |
| 5. | Pale sedge, Carex pallescens | | |

18

| Decer | nber 2014 |
|--|---|
| E. State Species of Concern | |
| 1. Big brown bat, <i>Eptesicus fuscus</i> | 18. Common moorhen, Gallinula chloropus |
| 2. Deer mouse, <i>Peromyscus maniculatus</i> | 19. Great egret, Ardea alba (migrant) |
| 3. Eastern red bat, <i>Lasiurus borealis</i> | 20. Sora, Porzana carolina |
| 4. Hoary bat, <i>Lasiurus cinereus</i> | 21. Virginia Rail, Rallus limicola |
| 5. Little brown bat, <i>Myotis lucifugus</i> | 22. Yellow-bellied Sapsucker, Sphyrapicus varius |
| 6. Pygmy shrew, Sorex hovi | 23. Creek heelsplitter, Lasmigona compressa |
| 7. Southern bog lemming, <i>Svnaptomys cooperi</i> | 24. Eastern box turtle, Terrapene carolina |
| 8. Star-nosed mole, Condylura cristata | 25. Four-toed Salamander, Hemidacrylium scutatur |
| 9. Tri-colored bat, Perimyotis subflavus | 26. Eastern garter snake, Thamnophis sirtalis |
| 10. Woodland jumping mouse, Napaeozapu. | s 27. Smooth green snake, <i>Opheodrys vernalis</i> |
| insignis | 28. Eastern sand darter, Ammocrypta pellucida |
| 11. Sharp-shinned hawk, Accipiter striatus | 29. Mayfly, Stenonema ithica |
| 12. Marsh wren, Cistothorus palustris | 30. Moth, Apamea mixta |
| 13. Henslow's sparrow, Ammodramus henslowii | 31. Moth, Brachylomia algens |
| 14. Cerulean warbler, Dendroica cerulean | 32. Scurfy quaker, Homorthodes furfurata |
| 15. Prothonotary warbler, Protonotaria citrea | 33. Sedge wren, Cistothorus platensis |
| 16. Bobolink, Dolichonyx oryzivorus | |
| 17. Northern bobwhite, Colinus virginianus | |
| F. State Special Interest | |
| 1. American black duck, Anas rubripes | 13. Purple finch, Carpodacus purpureus |
| 2. Canada warbler, Wilsonia Canadensis | 14. Red-breasted nuthatch, Sitta Canadensis |
| 3. Dark-eyed junco, Junco hyemalis (migrant) | 15. Golden-crowned kinglet, Regulus satrapa |
| 4. Hermit thrush, Catharus guttatus (migrant) | 16. Blackburnian warbler, Dendroica fusca |
| 5. Least flycatcher, <i>Empidonax minimus</i> | 17. Gadwall, Anas strepera |
| 6. Magnolia warbler, <i>Dendroica magnolia</i> | 18. Green-winged teal, Anas crecca |
| 7. Northern waterthrush, <i>Seiurus noveboracensis</i> | 19. Northern shoveler, Anas clypeata |
| 8. Winter wren, <i>Troglodytes</i> | 20. Redhead duck, Aytya Americana |
| 9. Back-throated blue warbler, Dendroice | |
| caerulescens | 22. Wilson's snipe, Gallinago delicata |
| 10. Brown creeper, Certhia Americana | 23. Subflava sedge borer, Capsula subflava |
| 11. Mourning warbler, Oporornis Philadelphia | |
| 12. Pine siskit, Carduelis pinus | |

Note: The Integrated Natural Resources Plan (OHARNG 2014) indicated that no federally listed species are known to reside at Camp Ravenna, and no critical habitat occurs. However, Table 2-1 reflects that the northern long-eared bat exists at Camp Ravenna and is federally threatened (USFWS 2016) and state threatened (ODNR 2016).

1

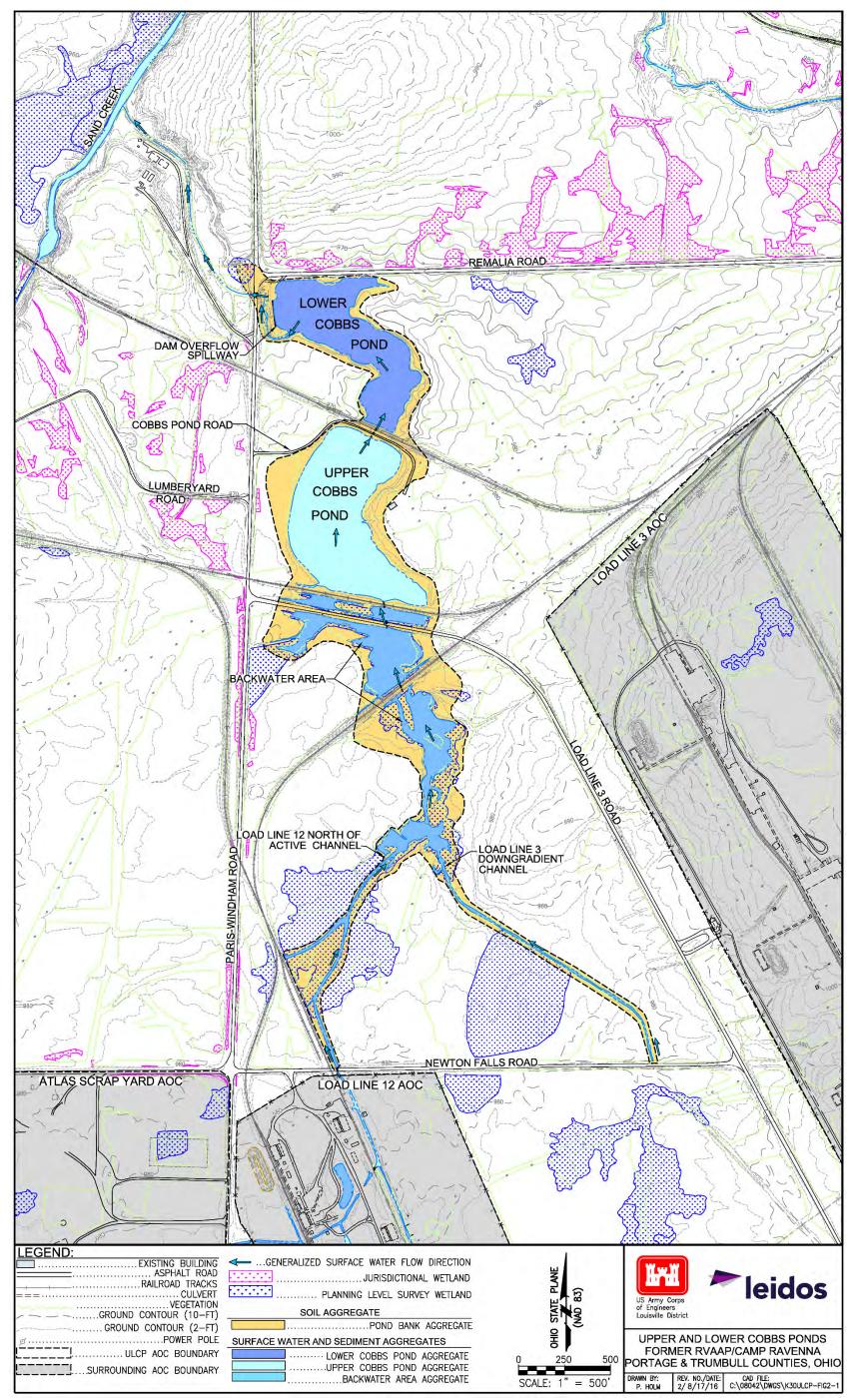


Figure 2-1. Surface Water Conveyance at Upper and Lower Cobbs Ponds

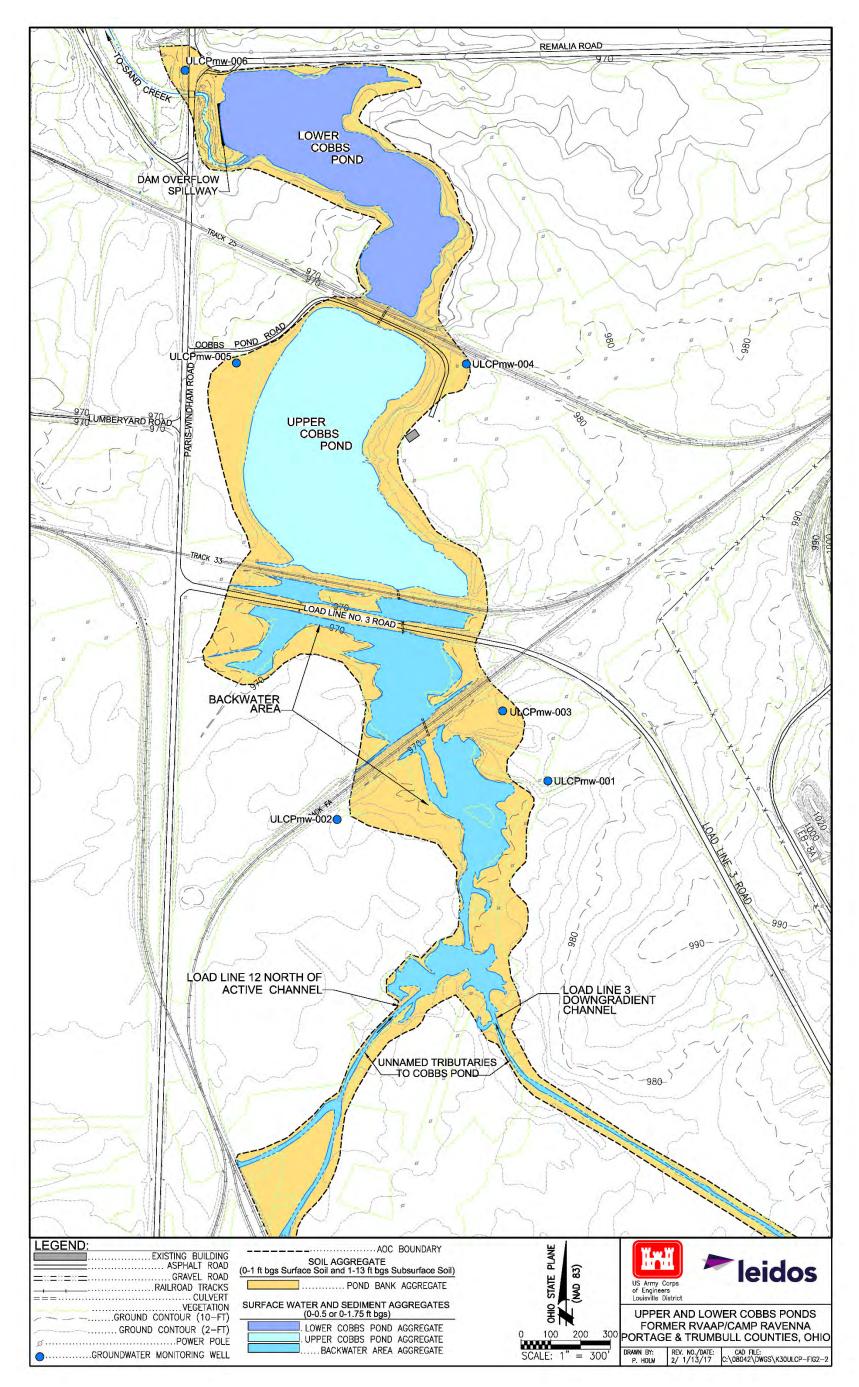


Figure 2–2. Upper and Lower Cobbs Ponds AOC Aggregates

1 3.0 ENVIRONMENTAL SETTING

2

This section describes the physical features, topography, geology, hydrogeology, and environmental characteristics of Camp Ravenna at the Upper and Lower Cobbs Ponds AOC that are factors in identifying the potential contaminant transport pathways, receptor populations, and exposure scenarios to evaluate human health and ecological risk.

7 8

3.1 CAMP RAVENNA PHYSIOGRAPHIC SETTING

9

10 Camp Ravenna is located within the southern New York Section of the Appalachian Plateaus 11 physiographic province (USGS 1968). This province is characterized by elevated uplands underlain 12 primarily by Mississippian and Pennsylvanian age bedrock units that are horizontal or gently dipping. 13 The province is characterized by its rolling topography, with incised streams having dendritic 14 drainage patterns. The southern New York Section has been modified by glaciation, which rounded 15 ridges, filled major valleys, and blanketed many areas with glacially-derived unconsolidated deposits (e.g., sand, gravel, and finer-grained outwash deposits). As a result of glacial activity in this section, 16 17 old stream drainage patterns were disrupted in many locales, and extensive wetland areas developed.

- 18
- 19 20

3.2 SURFACE FEATURES AND AOC TOPOGRAPHY

The topography of Camp Ravenna is gently undulating, with an overall decrease in ground elevation from a topographic high of approximately 1,220 ft above mean sea level (amsl) in the far western portion of the facility to low areas at approximately 930 ft amsl in the far eastern portion of the facility.

25

USACE mapped the facility topography in February 1998 using a 2-ft contour interval with an accuracy of 0.02 ft. USACE based the topographic information on aerial photographs taken during the spring of 1997. The USACE survey is the basis for the topographical information illustrated in figures included in this report.

30

31 The Upper and Lower Cobbs Ponds AOC is in the east-central portion of Camp Ravenna, east of 32 Paris-Windham Road and south of Remalia Road (Figures 1-2 and 2-1). No fences or perimeter 33 boundaries exist at the AOC. The topography at the AOC is dominated by the ponds and their 34 associated drainage conveyances (Figure 3-1). The pond banks are characterized by mild to steep 35 contours, surrounded by relatively flat land surface. The ground elevation varies from 960-980 ft 36 amsl, with a slight topographic high located near the southeastern portion of the AOC and a 37 topographic low west of the dam overflow spillway in the northwest portion of the AOC. Perennial 38 surface water is present within Upper Cobbs Pond, Lower Cobbs Pond, and the Backwater Area 39 (Figure 3-1). Surface water flows through the Backwater Area to Upper Cobbs Pond then to Lower 40 Cobbs Pond, ultimately exiting the AOC through a dam overflow spillway located near the intersection of Remalia and Paris-Windham Roads. Surface water then enters an unnamed tributary 41 42 that discharges into Sand Creek, northwest of the AOC.

1 3.3 SOIL AND GEOLOGY

3 3.3.1 Regional Geology

5 The regional geology at Camp Ravenna consists of horizontal to gently dipping bedrock strata of 6 Mississippian and Pennsylvanian age, overlain by varying thicknesses of unconsolidated glacial 7 deposits. The bedrock and unconsolidated geology at Camp Ravenna and the geology specific to the 8 Upper and Lower Cobbs Ponds AOC are presented in the following subsections.

9

2

4

3.3.2 Soil and Glacial Deposits

10 11

Bedrock at Camp Ravenna is overlain by deposits of the Wisconsin-age Lavery Till in the western portion of the facility and the younger Hiram Till and associated outwash deposits in the eastern twothirds of the facility (Figure 3-2). Unconsolidated glacial deposits vary considerably in their character and thickness across Camp Ravenna, from zero in some of the eastern portions of the facility to an estimated 150 ft in the south-central portion.

17

18 Thin coverings of glacial material have been completely removed as a consequence of human 19 activities at locations such as Ramsdell Quarry. Bedrock is present at or near the ground surface in 20 locations such as Load Line 1 and the Erie Burning Grounds (USACE 2001a). Where this glacial 21 material is still present, its distribution and character indicate its origin as ground moraine. These tills 22 consist of laterally discontinuous assemblages of yellow-brown, brown, and gray silty clays to clayey 23 silts, with sand and rock fragments. Lacustrine sediment from bodies of glacial-age standing water 24 has also been encountered in the form of deposits of uniform light gray silt greater than 50-ft thick in 25 some areas (USACE 2001a).

26

Soil at Camp Ravenna is generally derived from the Wisconsin-age silty clay glacial till. Distributions of soil types are discussed and mapped in the *Soil Survey of Portage County, Ohio*, which describes soil as nearly level to gently sloping and poor to moderately well drained (USDA 1978). Much of the native soil at Camp Ravenna was disturbed during construction activities in former production and operational areas of the facility.

32

The Sharon Member of the Pennsylvanian Pottsville Formation is the primary bedrock beneath Camp Ravenna. In the western half of the facility, the upper members of the Pottsville Formation, including the Connoquenessing Sandstone (also known as the Massillon Sandstone), Mercer Shale, and uppermost Homewood Sandstone, have been found. The regional dip of the Pottsville Formation measured in the western portion of Camp Ravenna is between 5–11.5 ft per mile to the south.

38

40

39 **3.3.3** Geologic Setting of Upper and Lower Cobbs Ponds

41 Bedrock was not encountered at the Upper and Lower Cobbs Pond AOC during the PBA08 RI or 42 previous characterization activities. The bedrock formation underlying the unconsolidated deposits at 43 the Upper and Lower Cobbs Pond AOC, is the Pennsylvanian age Pottsville Formation, Sharon 44 Sandstone Member, informally referred to as the Sharon Conglomerate (Figure 3-3) (Winslow et al. 1 1966). The Sharon Sandstone Member, the lowest unit of the Pottsville Formation, is a highly porous, 2 loosely cemented, permeable, cross-bedded, frequently fractured and weathered orthoquartzite 3 sandstone, which is locally conglomeratic. The Sharon Conglomerate exhibits locally occurring thin 4 shale lenses in the upper portion of the unit. Upper members of the Pottsville Formation are not 5 present at the AOC.

6

7 The soil types observed surrounding the ponds and Backwater Area are the Mahoning silt loam (2– 8 6% slopes) and the Trumbull silt loam (0-2% slopes). Mahoning silt loam is a gently sloping, poorly 9 drained soil formed in silty clay loam or clay loam glacial till, generally where bedrock is greater than 10 6 ft bgs. The Mahoning silt loam has low permeability, with rapid runoff and seasonal wetness and is present primarily around 85% of the water at the AOC (USDA 2010). The Trumbull silt loam covers 11 12 the remaining 15% of the pond banks at the AOC, primarily in the southeastern portion within the 13 Backwater Area. The Trumbull silt loam is poorly drained soil formed in silty clay till, generally 14 where bedrock is greater than 6 ft bgs. The Trumbull silt loam is typically formed in depressions with 15 a moderate water capacity with groundwater existing near ground surface (USDA 2010).

16

17 As observed in PBA08 RI soil borings, the composition of unconsolidated deposits at the AOC 18 generally consist of yellowish-brown to gray, medium dense clay to sand-rich silt tills with 19 interbedded sands scattered throughout. Bedrock was not encountered during PBA08 RI investigation 20 activities at the AOC. Groundwater was observed from 1–10.3 ft bgs in the PBA08 RI soil borings. 21 PBA08 RI boring logs containing geologic descriptions of unconsolidated deposits at the Upper and 22 Lower Cobbs Pond AOC are included in Appendix A. Geologic descriptions and geotechnical 23 analyses of subsurface soil samples collected during the PBA08 RI are generally consistent with the 24 conclusions from the Phase II RI.

25

Two geotechnical samples were collected from the AOC during the PBA08 RI. Samples collected from CPCsb-033 at 4–5.4 ft bgs and 8–9.7 ft bgs were classified as clayey silts with trace sand. A summary of geotechnical analysis, including porosity, density, and moisture content, is presented in Section 5.3.5.

- 30
- 31 3.4 HYDROGEOLOGY
- 32

33 3.4.1 Regional Hydrogeology

34

Sand and gravel aquifers are present in the buried-valley and outwash deposits in Portage County, as described in the *Phase I Remedial Investigation Report for High-Priority Areas of Concern* (USACE 1998) (herein referred to as the Phase I RI Report). Generally, these saturated zones are too thin and localized to provide large quantities of water for industrial or public water supplies; however, yields are sufficient for residential water supplies. Lateral continuity of these aquifers is unknown. Recharge of these units is derived from surface water infiltration of precipitation and surface streams. Specific groundwater recharge and discharge areas at Camp Ravenna have not been delineated.

42

The thickness of the unconsolidated interval at Camp Ravenna ranges from thin to absent in the eastern and northeastern portion of Camp Ravenna, to an estimated 150 ft in the central portion of the 1 facility. The groundwater table occurs within the unconsolidated zone in many areas of the facility. 2 Because of the heterogeneous nature of the unconsolidated glacial material, groundwater flow 3 patterns are difficult to determine with a high degree of accuracy. Vertical recharge from precipitation 4 likely occurs via infiltration along root zones, desiccation cracks, and partings within the soil column. Laterally, most groundwater flow likely follows topographic contours and stream drainage patterns, 5 with preferential flow along pathways (e.g., sand seams, channel deposits, or other stratigraphic 6 7 discontinuities) having higher permeabilities than surrounding clay or silt-rich material. Figure 3-4 8 illustrates facility-wide potentiometric surface data in the unconsolidated interval from the January 9 2010 contemporaneous measurement event (EQM 2010a).

10

11 Within bedrock units at Camp Ravenna, the principle water-bearing aquifer is the Sharon Sandstone/Conglomerate. Depending on the existence and depth of overburden, the Sharon 12 13 Sandstone/Conglomerate ranges from an unconfined to a leaky artesian aquifer. Water yields from 14 on-site water supply wells completed in the Sharon Sandstone/Conglomerate ranged from 30-400 15 gallons per minute (gpm) (USATHAMA 1978). Well yields of 5-200 gpm were reported for on-site 16 bedrock wells completed in the Sharon Sandstone/Conglomerate (Kammer 1982). Other local 17 bedrock units capable of producing water include the Homewood Sandstone, which is generally 18 thinner and only capable of well yields less than 10 gpm, and the Connoquenessing Sandstone. Wells completed in the Connoquenessing Sandstone in Portage County have yields ranging from 5-100 19 20 gpm but are typically less productive than the Sharon Sandstone/Conglomerate due to lower 21 permeabilities (Winslow et al. 1966).

22

23 Figure 3-5 shows the potentiometric surface within bedrock strata at Camp Ravenna in January 2010 24 (EQM 2010a). The bedrock potentiometric map shows a more uniform and regional eastward flow 25 direction than the unconsolidated zone that is not as affected by local surface topography. Due to the 26 lack of well data in the western portion of Camp Ravenna, general flow patterns are difficult to 27 discern. For much of the eastern half of Camp Ravenna, bedrock potentiometric elevations are higher 28 than the overlying unconsolidated potentiometric elevations, indicating an upward hydraulic gradient. 29 This evidence suggests there is a confining layer that separates the two aquifers. In the far eastern 30 area, the two potentiometric surfaces are at approximately the same elevation, suggesting hydraulic 31 communication between the two aquifers is occurring.

32

3.4.2 Upper and Lower Cobbs Ponds Hydrologic/Hydrogeologic Setting

33 34

Six groundwater monitoring wells were installed at the Upper and Lower Cobbs Ponds AOC during the Phase II RI (MKM 2005). Initial depths to groundwater encountered during well installation varied from 5.75–39.75 ft bgs. Monitoring wells at the AOC ranged in completion from 18–27 ft bgs. All monitoring wells have groundwater elevations collected under the FWGWMP. All monitoring wells at the Upper and Lower Cobbs Ponds AOC were screened in the unconsolidated overburden.

40

41 The potentiometric surface of the AOC from the January 2010 monitoring event is shown in Figure 3-

The estimated groundwater flow directions reflect the January 2010 facility-wide potentiometric
 data presented in the *Facility-wide Groundwater Monitoring Program Report on the January 2010*

45 data presented in the Faculty-while Groundwater Montioning Frogram Report on the January 2010

1 amsl. The potentiometric surface shows the groundwater flow pattern to the northwest toward Sand

2 Creek. The average hydraulic gradient for the AOC is approximately 0.0046 ft/ft.

3

Results of rising and falling head slug tests performed at the six monitoring wells during the Phase II
RI indicate an average hydraulic conductivity of 5.64E-06 cm/s (MKM 2005). Table 3-1 presents the
hydraulic conductivity result for each well at the Upper and Lower Cobbs Ponds AOC.

7 8

3.4.3 Surface Water

9

10 The following sections describe the regional and AOC-specific surface water.

11

12 3.4.3.1 <u>Regional Surface Water</u>

13

14 Camp Ravenna resides within the Mahoning River watershed, which is part of the Ohio River basin. 15 The west branch of the Mahoning River is the main surface stream in the area. The west branch flows 16 adjacent to the west end of the facility, generally in a north to south direction, before flowing into the 17 Michael J. Kirwan Reservoir, located to the south of State Route 5 (Figure 1-1). The west branch 18 flows out of the reservoir and parallels the southern Camp Ravenna boundary before joining the 19 Mahoning River east of Camp Ravenna. The western and northern portions of Camp Ravenna display 20 low hills and a dendritic surface drainage pattern. The eastern and southern portions are characterized 21 by an undulating to moderately level surface, with less dissection of the surface drainage. The facility 22 is marked with marshy areas and flowing and intermittent streams whose headwaters are located in 23 the upland areas of the facility.

24

25 The three primary watercourses that drain Camp Ravenna are (Figure 1-2):

26 27

28

• South Fork Eagle Creek,

- Sand Creek, and
- Hinkley Creek.
- 29 30

31 These watercourses have many associated tributaries. Sand Creek, with a drainage area of 13.9 square 32 miles (36 km²), generally flows in a northeast direction to its confluence with South Fork Eagle 33 Creek. In turn, South Fork Eagle Creek continues in a northerly direction for 2.7 miles to its 34 confluence with Eagle Creek. The drainage area of South Fork Eagle Creek is 26.2 square miles, 35 including the area drained by Sand Creek. Hinkley Creek originates just southeast of the intersection 36 between State Routes 88 and 303 to the north of the facility. Hinkley Creek, with a drainage area of 37 11 square miles, flows in a southerly direction through the facility and converges with the west 38 branch of the Mahoning River (USACE 2001a).

39

40 Previous jurisdictional wetland delineations have surveyed approximately 5,680 acres (or 26% of the

41 Camp Ravenna land). Approximately 715 acres of jurisdictional wetlands have been delineated within

- 42 the 5,680 acres surveyed, which comprises approximately 13% of the total surveyed area. Wetland
- 43 areas at Camp Ravenna include seasonal wetlands, wet fields, and forested wetlands. Many of the

1 wetland areas are the result of natural drainage or beaver activity; however, some wetland areas are

2 associated with anthropogenic settling ponds and drainage areas.

3

4 Approximately 30 ponds are scattered throughout the facility. Many were constructed within natural drainageways to function as settling ponds or basins for process effluent and runoff. Others are 5 natural in origin, resulting from glacial action or beaver activity. Water bodies at Camp Ravenna 6 7 support aquatic vegetation and biota as described in Section 2.3.2. Storm water runoff is controlled 8 primarily by natural drainage, except in former operations areas where an extensive storm sewer 9 network helps to direct runoff to drainage ditches and settling ponds. In addition, the storm sewer 10 system was one of the primary drainage mechanisms for process effluent while production facilities 11 were operational.

12

13 14

3.4.3.2 <u>Upper and Lower Cobbs Ponds Surface Water</u>

Surface water at the AOC is present perennially within the ponds and Backwater Area (Figure 3-1). The flow of surface water begins with the drainage channels from Load Line 3 and Load Line 12 that enter the AOC from the south and flows through the Backwater Area. Surface water flows north through Upper Cobbs Pond to Lower Cobbs Pond.

19

Flow from the Backwater Area into Upper Cobbs Pond and, subsequently, into Lower Cobbs Pond is through culverts beneath railroad trackbeds and roads that cross the AOC. Lower Cobbs Pond surface water discharges into a dam overflow spillway adjacent to the intersection of Remalia and Paris-Windham Roads, entering an unnamed tributary to Sand Creek, which is the primary migration pathway for contamination to exit the AOC.

25

Several large planning wetlands exist within the AOC boundary (OHARNG 2008), generally along the Pond Banks, which may receive overland surface water flow or runoff from ponds when water levels within the AOC are elevated.

29

30 3.5 CLIMATE

31

32 The general climate of the Camp Ravenna area is continental and is characterized by moderately 33 warm and humid summers, reasonably cold and cloudy winters, and wide variations in precipitation 34 from year to year. The climate data presented below for the Camp Ravenna area were obtained from 35 available National Weather Service records for the 30-year period of record from 1981–2010 at the 36 Youngstown Regional Airport, Ohio (http://www.weather.gov/climate/xmacis.php?wfo=cle). Wind 37 speed data for Youngstown, Ohio, are from the National Climatic Data Center 38 (http://www1.ncdc.noaa.gov/pub/data/ccd-data/wndspd14.txt) for the available 30-year period of 39 record from 1984–2014.

40

Average annual rainfall at Camp Ravenna area is 38.86 inches, with the highest monthly average
occurring in July (4.31 inches) and the lowest monthly average occurring in February (2.15 inches).
Average annual snowfall totals approximately 62.9 inches, with the highest monthly average
occurring in January (17.1 inches). Due to the influence of lake-effect snowfall events associated with

1 Lake Erie (located approximately 35 miles to the northwest of Camp Ravenna), snowfall totals vary

2 widely throughout northeastern Ohio.

3

4 The average annual daily temperature in the Camp Ravenna area is 49.3°F, with an average daily high 5 temperature of 70.9°F and an average daily low temperature of 26.1°F. The record high temperature 6 of 100°F occurred in July 1988, and the record low temperature of -22°F occurred in January 1994. 7 The prevailing wind direction at Camp Ravenna is from the southwest, with the highest average wind 8 speed occurring in January (10.3 miles per hour) and the lowest average wind speed occurring in 9 August (6.5 miles per hour). Thunderstorms occur on approximately 35 days per year and are most 10 abundant from April through August. Camp Ravenna is susceptible to tornadoes; minor structural 11 damage to several buildings on facility property occurred as the result of a tornado in 1985.

12 13

Table 3–1. Hydraulic Conductivities Measured During the Phase II Remedial Investigation

| Monitoring Well | Screened Interval | | Hydraulic Conductivity |
|-----------------|-------------------|---------------------------------------|------------------------|
| ID | (ft bgs) | Geologic Material Adjacent to Screen | (cm/s) |
| ULCPmw-001 | 5.5-15.5 | Silty sand, clayey silt | 1.98E-06 |
| ULCPmw-002 | 5.5–15.5 | Clayey silt with silty sand interbeds | 6.01E-06 |
| ULCPmw-003 | 8-18 | Clayey silt | 1.38E-05 |
| ULCPmw-004 | 9.5–19.5 | Silty sand, clayey silt | 8.38E-07 |
| ULCPmw-005 | 29.5–39.5 | Clayey silt; sandy silt | 1.16E-06 |
| ULCPmw-006 | 8–18 | Clayey silt | 1.01E-05 |

Source = Phase II Remedial Investigation Report for Upper and Lower Cobbs Ponds (MKM 2005).

bgs = Below ground surface.

cm/s = Centimeters per second.

ft= Feet. ID = Identification. THIS PAGE INTENTIONALLY LEFT BLANK.

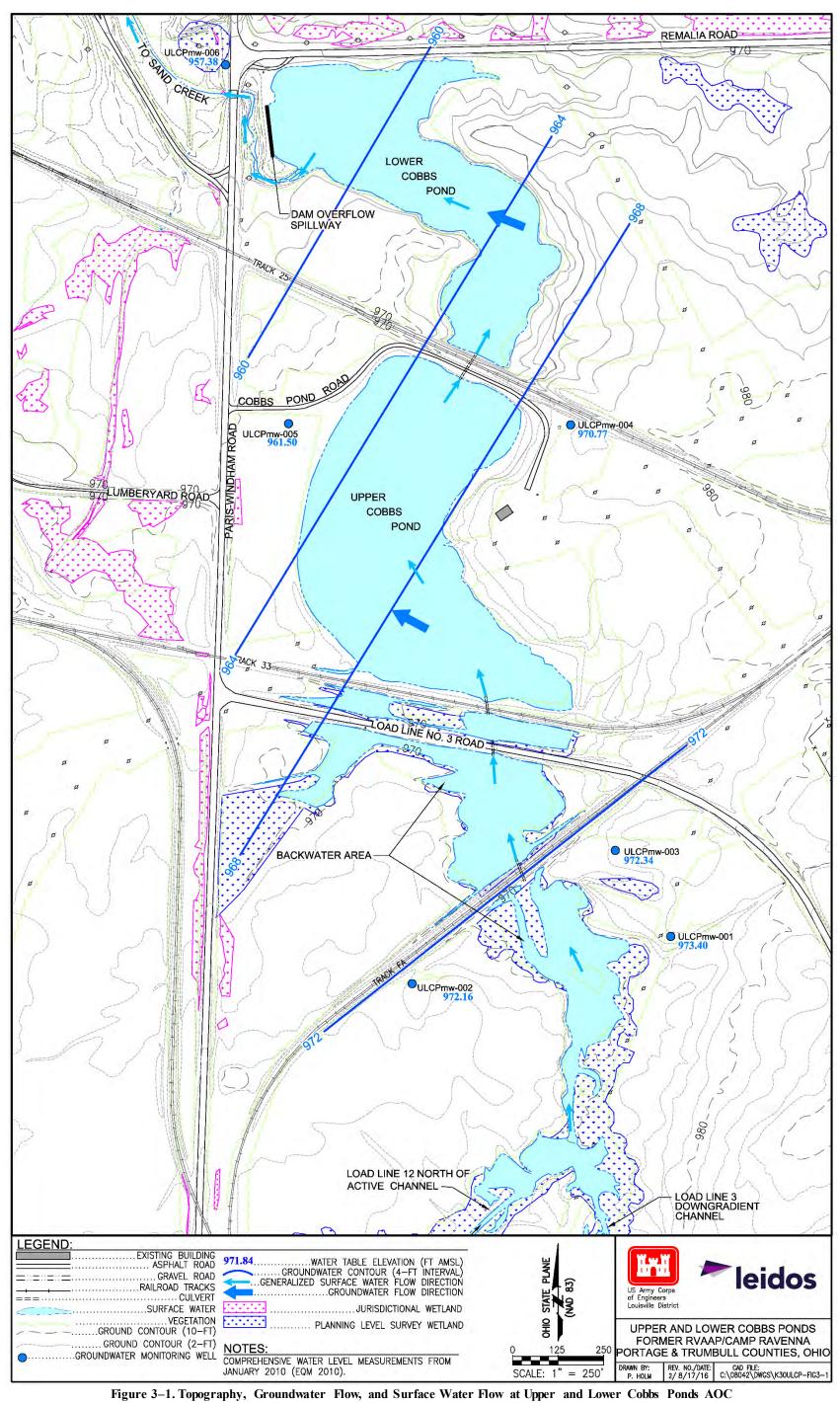


Figure 3-1. Topography, Groundwater Flow, and Surface Water Flow at Upper and Lower Cobbs Ponds AOC

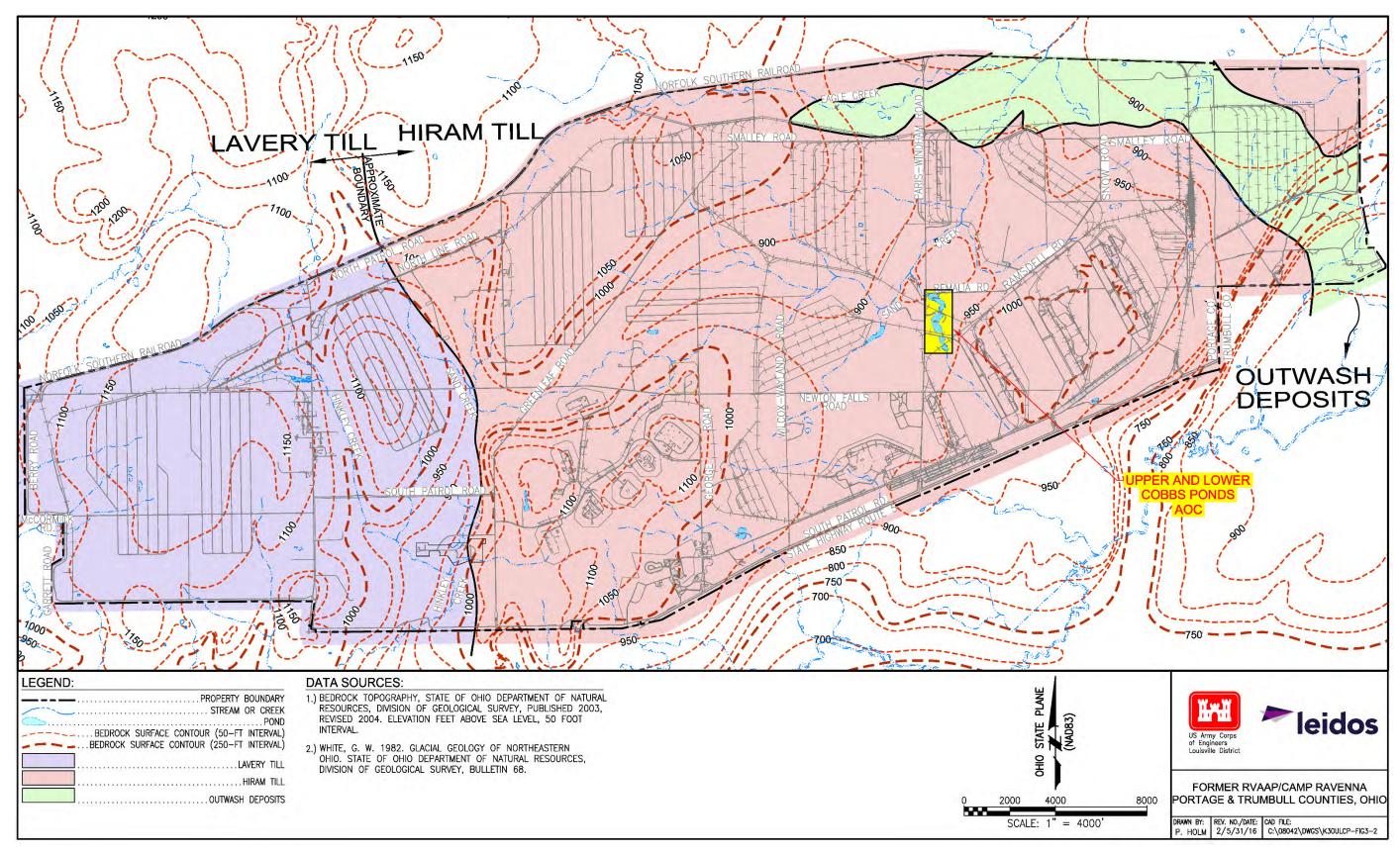


Figure 3-2. Geologic Map of Unconsolidated Deposits on Camp Ravenna

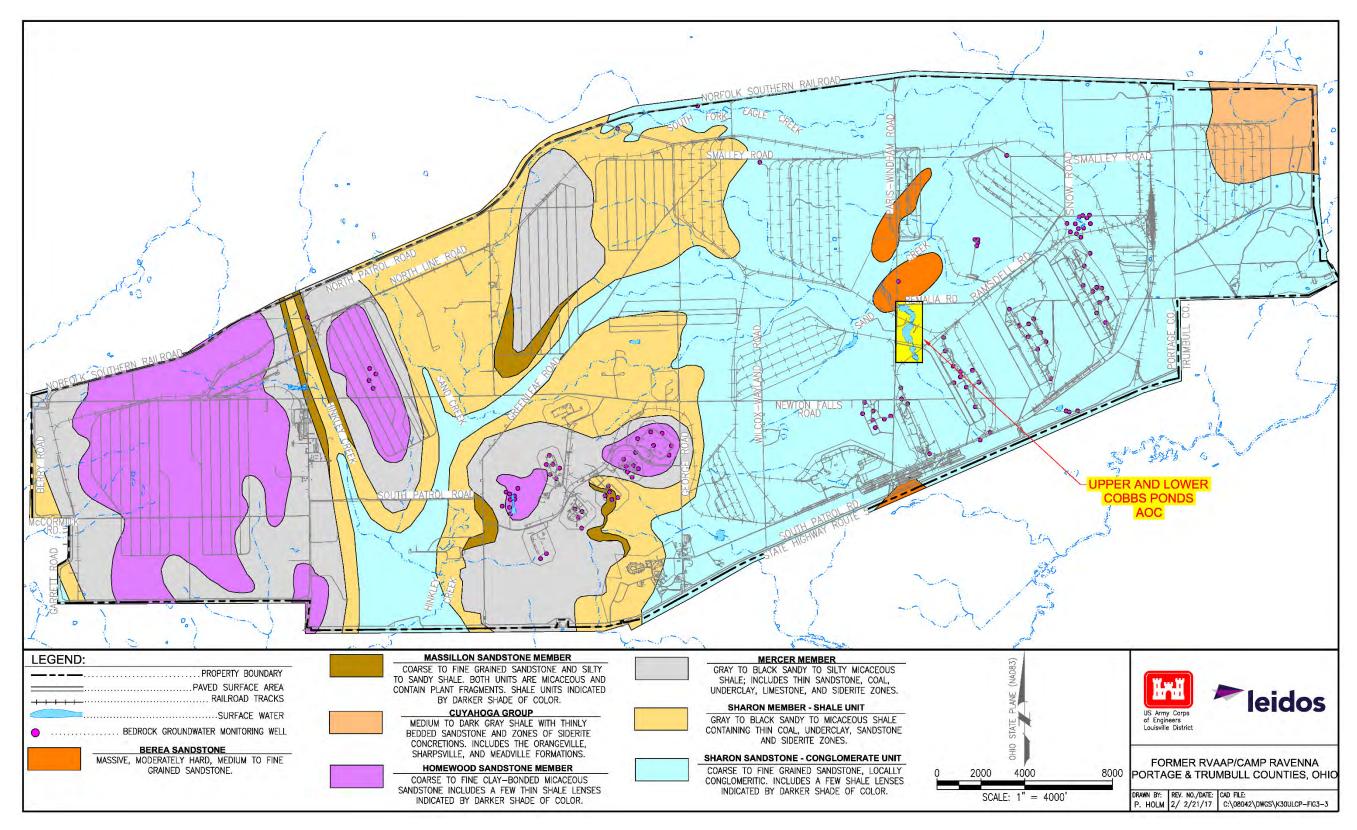


Figure 3-3. Geologic Bedrock Map and Stratigraphic Description of Units on Camp Ravenna

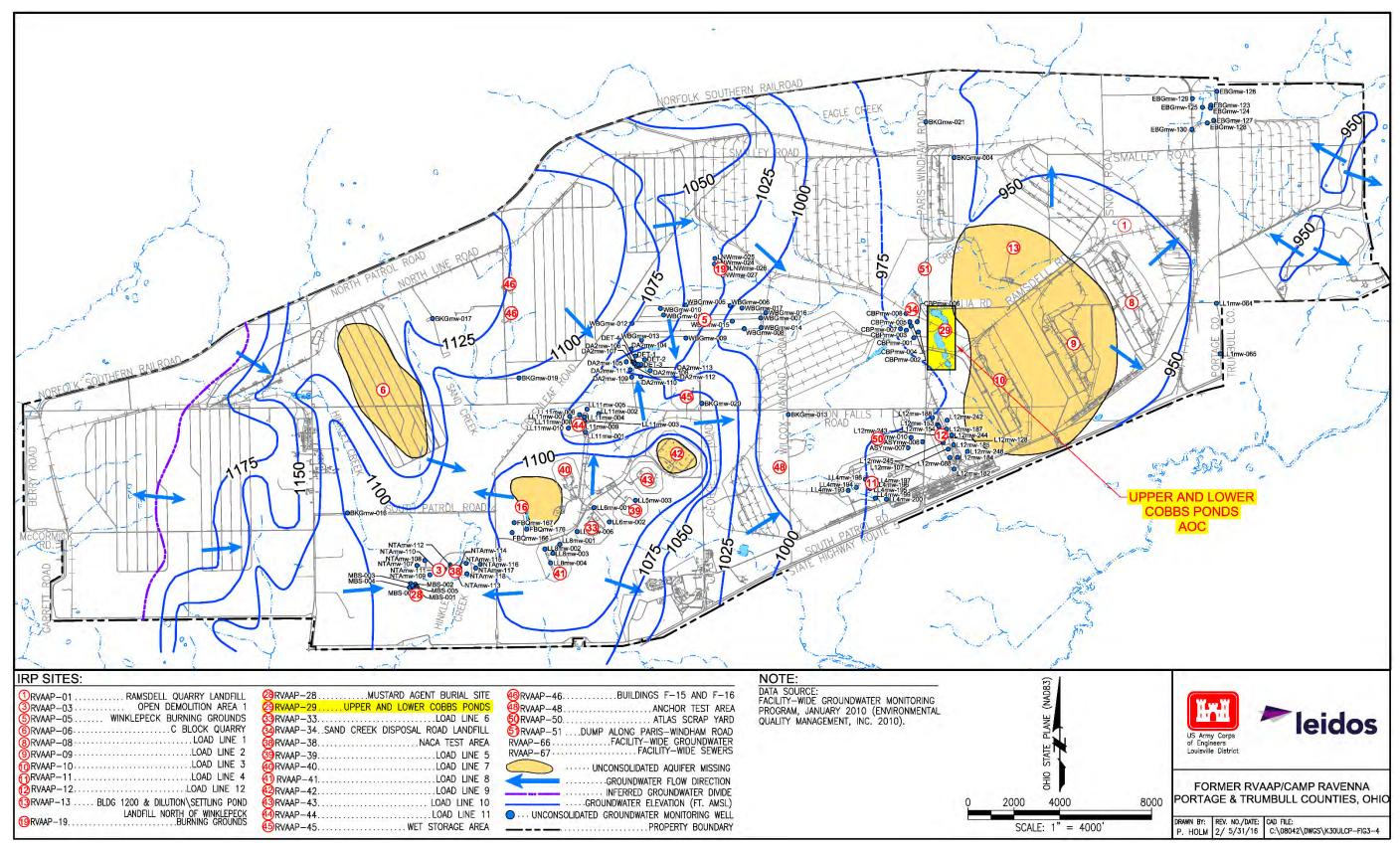


Figure 3-4. Potentiometric Surface of Unconsolidated Aquifer at Camp Ravenna

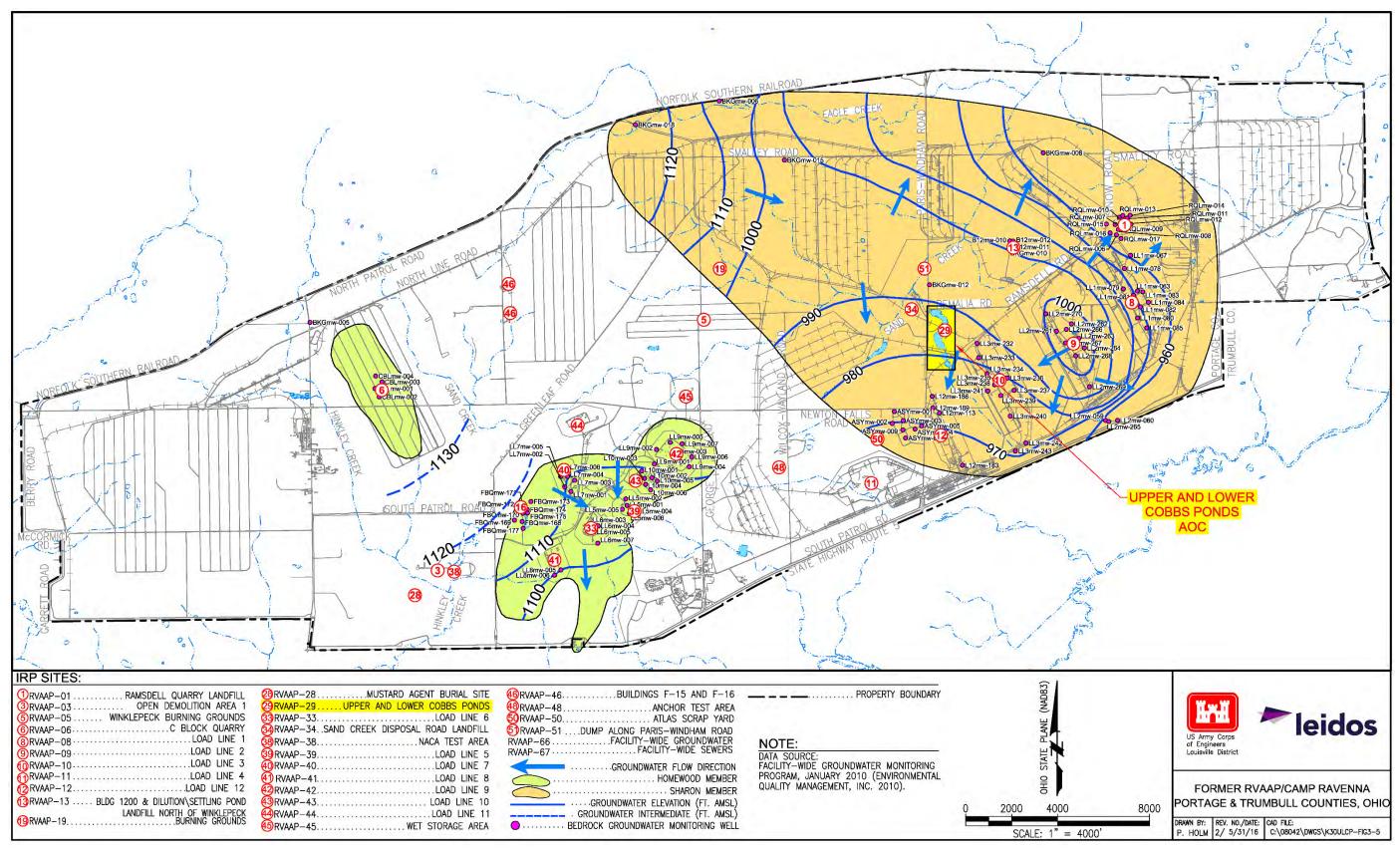


Figure 3-5. Potentiometric Surface of Bedrock Aquifers at Camp Ravenna

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4.0 SITE ASSESSMENTS, INVESTIGATIONS, AND DATA 2 ASSEMBLY

3 4

5

This section summarizes all previous site assessments and investigations conducted at the Upper and Lower Cobbs Ponds AOC. These previous activities included assessments to prioritize the AOC and investigations that collected data used in support of the RI.

6 7 8

9

4.1 PREVIOUS ASSESSMENTS AND EVALUATIONS

10 This section summarizes previous assessments and evaluations conducted at the Upper and Lower 11 Cobbs Ponds AOC. These activities were generally performed as an initial evaluation and/or 12 prioritization assessment of the AOC. The data collected as part of these prioritization assessments 13 and evaluations are not used in the nature and extent, fate and transport, HHRA, or ERA due to their 14 age and lack of data quality documentation.

15 16

4.1.1 Installation Assessment of Ravenna Army Ammunition Plant

17

18 The *Installation Assessment of Ravenna Army Ammunition Plant* incorporated a review of historical 19 operational information and available environmental data to assess the potential for contaminant 20 releases from operational facilities. No sampling was performed at Upper and Lower Cobbs Ponds as 21 part of the assessment. The assessment identified the following conditions at RVAAP, applicable to 22 Upper and Lower Cobbs Ponds (USATHAMA 1978):

- 23
- Load Line 3 was used for TNT and composition B melt-pour activities, and Load Line 12 was
 used for ammonium nitrate production. Both load lines were also used for demilitarization
 activities. Pink water and washout of residue, dusts, and spills at Load Lines 3 and 12 were
 discharged into Upper Cobbs Pond through surface drainage channels.
- 28 From 1965–1967, Hercules Alcor, Inc. leased Building FF-19 in Load Line 12 for producing 29 aluminum chloride. A fish kill occurred on November 15, 1966 at Lower Cobbs Pond as a 30 result of improper handling of aluminum chloride during these manufacturing operations. The 31 bulk of the aluminum chloride at Load Line 12 was removed and disposed in Ramsdell 32 Ouarry. The pond was settled and drained, and the contaminants were moved to Ramsdell 33 Ouarry. Contaminated metals were flashed at a burning ground to a five times condition 34 (USATHAMA 1978 and 1982, RVAAP 2000). The Army terminated the lease early on 35 December 3, 1967 due to environmental concerns related to air emissions and wastewater 36 discharges to Upper and Lower Cobbs Ponds.
- Surface waters exiting the facility were not required to be monitored for nitrobodies and
 heavy metals.
- In August 1949, the ammonium nitrate fertilizer operations at Load Line 12 were investigated to determine whether high nitrate concentrations observed in Warren, Ohio were related to Load Line 12 waste disposal procedures. Warren, Ohio used the Mahoning River as a water supply source. The Mahoning River received influent from upgradient waters, including Upper and Lower Cobbs Ponds, which received drainage from Load Line 12. Improvements

| 1 2 | | in the operations and housekeeping at RVAAP in 1949 reduced nitrate concentrations to an acceptable concentration (less than 10 parts per million) in on-site surface water and |
|--|---|---|
| 3 | | downstream. |
| 4 | • | Analysis of the well water indicated potable quality. |
| 5 | • | No environmental stress was identified at RVAAP. |
| 6 | | |
| 7 8 | 4.1.2 | Soil and Sediment Analysis |
| 9 | The M | logul Corporation performed surface soil and sediment analysis at selected RVAAP ponds and |
| 10 | stream | s for the explosives TNT and RDX in May 1982. A description of field activities and data |
| 11 | summa | ary are provided in the Soil and Sediment Analysis Performed for Ravenna Arsenal (Mogul |
| 12 | 1982). | Results relevant to Upper and Lower Cobbs Ponds are as follows: |
| 13 | | |
| 14 15 | • | One composite sample collected at Upper Cobbs Pond had non-detectable concentrations of TNT. RDX had a concentration of $1.16 \mu\text{g/mL}$. |
| 16 17 | • | One composite sample collected from Lower Cobbs Pond had non-detectable concentrations of TNT and RDX. |
| 18 | • | One grab sample collected at Load Line 12 Stream #1 had a TNT concentration of 0.17 |
| 19 | | μg/mL and an RDX concentration of 1.12 μg/mL. |
| 20 21 22 | • | Samples at upstream locations from Upper Cobbs Pond, including Load Line 12 Stream #2, Load Line 12 Stream #3, and the Load Line 12 pond had non-detectable concentrations of TNT and RDX. |
| | | |
| 23 | | |
| 23 24 | 4.1.3 | RVAAP Preliminary Assessment |
| 23 24 25 | 4.1.3 | RVAAP Preliminary Assessment |
| 24 | | RVAAP Preliminary Assessment reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP |
| 24 25 | The Pr | |
| 24 25 26 | The <i>Pr</i> history | eliminary Assessment for the Characterization of Areas of Contamination researched RVAAP |
| 24 25 26 27 | The <i>Pr</i> history historic | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional |
| 24 25 26 27 28 | The <i>Pr</i> history historic report | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The |
| 24 25 26 27 28 29 | The <i>Pr</i> history historic report | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through |
| 24 25 26 27 28 29 30 | The <i>Pr</i> history historic report evalua | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through |
| 24 25 26 27 28 29 30 31 | The <i>Pr</i> history historic report evalua | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. |
| 24 25 26 27 28 29 30 31 32 | The <i>Pr</i> history historic report evaluation. | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: |
| 24 25 26 27 28 29 30 31 32 33 34 35 | The <i>Pr</i> history historic report evaluation. | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water |
| 24 25 26 27 28 29 30 31 32 33 34 | The <i>Pr</i> history historic report evaluation. | <i>reliminary Assessment for the Characterization of Areas of Contamination</i> researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. |
| 24 25 26 27 28 29 30 31 32 33 34 35 | The <i>Pr</i> history historic report evaluation. | reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. The ponds have been used for recreational purposes and support abundant fish and wildlife. |
| 24 25 26 27 28 29 30 31 32 33 34 35 36 37 38 | The <i>Pr</i> history historic report evalua The as Ponds | reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. The ponds have been used for recreational purposes and support abundant fish and wildlife. A ponded area (Backwater Area) was created by beavers and exists south (upstream) of |
| 24 25 26 27 28 29 30 31 32 33 34 35 36 37 | The <i>Pr</i> history historic report evalua The as Ponds | reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. The ponds have been used for recreational purposes and support abundant fish and wildlife. |
| 24 25 26 27 28 29 30 31 32 33 34 35 36 37 38 39 40 41 | The <i>Pr</i> history historic report evalua The as Ponds | reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. The ponds have been used for recreational purposes and support abundant fish and wildlife. A ponded area (Backwater Area) was created by beavers and exists south (upstream) of Upper Cobbs Pond. This area did not exist during plant operations. This Backwater Area supports abundant fish and wildlife. A fish kill occurred at Lower Cobbs Pond in 1966 due to improper handling of aluminum |
| 24 25 26 27 28 29 30 31 32 33 34 35 36 37 38 39 40 41 42 | The <i>Pr</i> history historic report evalua The as Ponds | reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. seessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. The ponds have been used for recreational purposes and support abundant fish and wildlife. A ponded area (Backwater Area) was created by beavers and exists south (upstream) of Upper Cobbs Pond. This area did not exist during plant operations. This Backwater Area supports abundant fish and wildlife. A fish kill occurred at Lower Cobbs Pond in 1966 due to improper handling of aluminum chloride at Load Line 12. |
| 24 25 26 27 28 29 30 31 32 33 34 35 36 37 38 39 40 41 | The <i>Pr</i> history historic report evalua The as Ponds | reliminary Assessment for the Characterization of Areas of Contamination researched RVAAP , process operations, and historical data to identify AOCs (USACE 1996a). No additional cal documents were available and minimal service history was available for the AOC. The provided preliminary assessment scoring, subsequent prioritization of AOCs through tion of exposure pathways, and a relative risk site evaluation model. ssessment also recognized the following conditions relevant to the Upper and Lower Cobbs AOC: Upper and Lower Cobbs Ponds consists of two unlined ponds that received pink water effluent and storm water runoff from Load Lines 3 and 12. The ponds have been used for recreational purposes and support abundant fish and wildlife. A ponded area (Backwater Area) was created by beavers and exists south (upstream) of Upper Cobbs Pond. This area did not exist during plant operations. This Backwater Area supports abundant fish and wildlife. A fish kill occurred at Lower Cobbs Pond in 1966 due to improper handling of aluminum |

- The total volume of wastewater received by Upper and Lower Cobbs Ponds during the
 AOC's operational period is unknown.
- The greatest potential for release of contaminants to groundwater was identified as
 wastewater effluent discharge to unlined earthen settling ponds (i.e., Upper and Lower Cobbs
 Ponds).
- The greatest potential for historical off-site contaminant migration during load line operations
 was identified as surface water.
 - The greatest potential for current off-site contaminant migration was identified as groundwater and surface water.
- 9 10

8

Based on qualitative assessment of the potential hazards, release mechanisms, and environmental
 conditions, the Upper and Lower Cobbs Ponds AOC was ranked as a high-priority AOC.

13

14 4.2 REMEDIAL INVESTIGATIONS

15

16 This section summarizes previous investigations [including the 2003 Facility-wide Biological and 17 Water Quality Study (FWBWQS)] conducted at the Upper and Lower Cobbs Ponds AOC. These 18 investigations collected data of sufficient provenance and quality to be used to support the evaluations 19 in this RI, including the nature and extent, fate and transport, HHRA, and/or ERA.

20

The previous Phase I RI Report (USACE 1998) and *Phase II Remedial Investigation Report for Upper and Lower Cobbs Ponds* (MKM 2005) (herein referred to as the Phase II RI Report) presented SRCs and/or COPCs based on data evaluation protocols in use at the time the investigations were completed. The data and information is used in this report; however, an updated screening process and the addition of new data and information may result in a different list of SRCs and/or COPCs.

26

27 References to "RVAAP full-suite analytes" generally include analyses of target analyte list (TAL)
28 metals, explosives, propellants (nitrocellulose and nitroguanidine), semi-volatile organic compounds
29 (SVOCs), volatile organic compounds (VOCs), polychlorinated biphenyls (PCBs), and pesticides.

30 31

32

4.2.1 Phase I Remedial Investigation

In 1996, sampling was conducted at the Upper and Lower Cobbs Ponds AOC in accordance with the
 RVAAP Facility-wide Sampling and Analysis Plan (FWSAP; USACE 1996b) and *Phase I RI SAP Addendum for High-Priority AOCs* (USACE 1996c). The Phase I RI was performed to accomplish
 the following tasks:

37 38

39

- Collect environmental samples from potentially impacted media to confirm if contamination is present and is being released to the environment,
- 40 Determine the nature of the COPCs, and
 - Provide a human health risk screening evaluation.
- 41 42

Results of this Phase I RI are presented in the Phase I RI Report (USACE 1998) and are summarizedin the following subsections.

1 4.2.1.1 Field Activities

The following field activities were conducted during the Phase I RI in August 1996 at the Upper and
Lower Cobbs Ponds AOC to assess the potential impacts from former operations:

5 6

7

8

9

2

- Collected 10 composite sediment samples from the ponds and drainage conveyances (5 samples at Upper Cobbs Pond, 3 samples at Lower Cobbs Pond, and 2 samples downstream of Lower Cobbs Pond);
- Installed, developed, and sampled three temporary well points; and
- Surveyed sampling locations.
- 10 11

12 Temporary well points were installed, developed, and sampled; however, these activities are pertinent 13 to other co-located AOCs associated with Upper and Lower Cobbs Ponds (e.g., facility-wide 14 groundwater) and are not discussed in this section.

15

Sediment samples were analyzed for TAL metals and explosives, with one sample submitted for
analysis for RVAAP full-suite analytes (10%). All groundwater samples were submitted for analysis
of RVAAP full-suite analytes.

19

Figure 4-1 presents the locations sampled under the Phase I RI. Table 4-1 presents the sample locations, associated operations, and suite of chemicals analyzed as part of the Phase I RI. Table 4-2 presents the results of the analytes detected from samples collected during the Phase I RI.

23

24 Analytical laboratory procedures were completed in accordance with applicable professional 25 standards, USEPA requirements, government regulations and guidelines, and specific project goals 26 and requirements. Samples were analyzed as specified by the FWSAP current at the time of the 27 investigation and the Phase IRI SAP Addendum for High-Priority AOCs (USACE 1996c). The data 28 quality objectives (DQOs) were established for the Phase I RI and complied with USEPA Region 5 29 guidance. The requisite number of quality assurance (QA)/quality control (QC) samples was obtained during the investigation. The data validation determined that the data met the completeness 30 31 requirements for the project (100% complete), was usable, and that it satisfied the DQOs for the 32 project.

33

34 4.2.1.2 <u>Nature and Extent of Contamination</u>

35

Five TAL metals were detected in sediment at concentrations exceeding background concentrations and identified as SRCs. Concentrations of inorganic chemicals in drainage sediment tended to be higher towards the center of the ponds. In addition, a single detection of the explosive nitrobenzene was observed in sediment at the outfall of Upper Cobbs Pond (location CPCsd-005) at a concentration of 0.38 mg/kg. Eight non-nutrient TAL metals, cyanide, and the organic compound 4-chloro-3methyl-phenol were identified as SRCs in groundwater samples.

| 1 | 4.2.1.3 | Risk Evaluation |
|----------------------|------------|--|
| 2 | 1 | |
| 3 | | valuation for the Upper and Lower Cobbs Ponds AOC was included in the Phase I RI. The |
| 4 | | luation compared chemical concentrations detected in AOC samples to RVAAP criteria in |
| 5 | | t that time, which included facility-wide background concentrations for inorganic chemicals, |
| 6 | | Region 4 sediment criteria, and state of Ohio or federal MCLs for drinking water. The results |
| 7 | of the ris | sk evaluation identified chromium in sediment and manganese in groundwater as COPCs. |
| 8 | 4214 | |
| 9 10 | 4.2.1.4 | Conclusions and Recommendations |
| 10 11 | The Dhe | ase I RI Report provided the following conclusions applicable to the Upper and Lower Cobbs |
| 12 | Ponds A | |
| 12 | I Unus A | |
| | _ | The DI abjectives were to abarratorize the native and extent of natural contention in |
| 14 | | The RI objectives were to characterize the nature and extent of potential contamination in |
| 15 16 | | sediment at the site. The objectives of the RI were met through the field investigation conducted in 1996. Field investigation included sampling sediment and groundwater. |
| | | |
| 17 | | The Phase I RI re-categorized the Upper and Lower Cobbs Ponds AOC as a "medium priority" AOC, as there was one detection of explosives and concentrations of inorganic |
| 18 19 | | chemicals were slightly above background concentrations (at the time of the investigation) |
| 19 20 | | |
| | | primarily in the center of the ponds. |
| 21 | | The concentrations of inorganic chemicals in the drainage downstream of the AOC were generally below background concentrations (at the time of the investigation), indicating |
| 22 23 | | contaminants were not migrating from the ponds. |
| 23 24 | | containing were not migrating from the polids. |
| 2 4 25 | Recomm | nendations of the Phase I RI include: |
| 25 26 | Reconn | includations of the 1 mase 1 K1 include. |
| 20 27 | • | Conduct a Phase II RI to determine extent of soil and sediment contamination at high- and |
| 28 | | medium-priority AOCs as defined by the Phase I RI. |
| 28 29 | | Collect site-specific hydrogeologic data to determine the vertical and lateral extent of |
| 29 30 | | potential groundwater contamination in the unconsolidated zone. |
| 31 | | Perform chemical fate and transport modeling to identify contaminant migration potential via |
| 32 | | groundwater and surface water within the Upper and Lower Cobbs Ponds AOC. |
| 32 33 | | Complete a baseline human health risk assessment (BHHRA) for all environmental media. |
| 33 34 | | Complete a baseline numan nearin risk assessment (Brittika) for all environmental media. |
| 34 35 | • | Complete an EKA for all environmental media. |
| | 422 | |
| 36 27 | 4.2.2 | Phase II Remedial Investigation |
| 37 38 | In Juky | September 2001, sampling was conducted at the Upper and Lower Cobbs Ponds AOC in |
| 38 39 | • | nce with the Sampling and Analysis Plan Addendum for the Phase II Remedial Investigation |
| 40 | | <i>Ipper and Lower Cobbs Pond (AOC 29)</i> (MKM 2001; herein referred to as the Phase II RI |
| 10 | | pper una Lower Cooss I on a (100 27) (mixin 2001, herein referred to as the I hase if Ki |

41 SAP Addendum).

- 1 The Phase II RI was performed to accomplish the following:
 - Define the vertical and horizontal extent of contamination at Upper and Lower Cobbs Ponds and their surroundings,
 - Gather sufficient data to conduct a BHHRA and a screening ecological risk assessment (SERA),
 - Define potential transport pathways and receptor populations, and
 - Provide sufficient data for selecting remedial action alternatives.
- 8 9

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10 The primary focus of the Phase II RI is to evaluate sediment, surface water, and groundwater. Soil 11 samples were collected from the pond bank soils (0–1 ft bgs) associated with banks on the pond to 12 determine if fluctuations in water levels in the ponds affected the bank soil. The data from the RI 13 sampling was incorporated into an evaluation of risk for the AOC, followed by recommendations for 14 remedial efforts, as necessary. Results of this Phase II RI are presented in the Phase II RI Report 15 (MKM 2005) and are summarized in the following subsections.

16 17

18

4.2.2.1 Field Activities

19 The following field activities were conducted during the Phase II RI in July–September 2001 at the 20 AOC [as documented in the Phase II RI Report (MKM 2005)] to assess the potential impacts from 21 former operations:

22 23

24

- Conducted unexploded ordnance-avoidance screening before initiating field activities,
- Collected 10 discrete surface soil (0–1 ft bgs) samples from Pond Bank,
- Field-screened Pond Bank soil samples for RDX and TNT,
- Collected 7 surface water samples,
- Collected 29 sediment samples, and
- Surveyed sampling locations.
- 29

The following activities also took place; however, these activities are pertinent to other co-located AOCs associated with the Upper and Lower Cobbs Ponds AOC (e.g., facility-wide groundwater) and are not discussed in this section:

- 33 34
 - Installed, developed, and sampled six monitoring wells;
- Conducted in-situ permeability testing (slug tests);
- Collected geotechnical samples from three monitoring well borings; and
- Surveyed monitoring well locations.
- 38
- Soil, sediment, and surface water samples were analyzed for TAL metals and explosives.
 Additionally, 10% of samples were analyzed for RVAAP full-suite analytes. Groundwater samples
- 41 were submitted for RVAAP full-suite analysis.
- 42

1 Figure 4-1 presents the locations sampled under the Phase II RI. Table 4-3 presents the sample

2 locations, associated operations, and suite of chemicals analyzed as part of the Phase II RI. Tables 4-4

3 through 4-6 present the results of the analytes detected from samples collected during the Phase II RI.

4

5 Analytical laboratory procedures were completed in accordance with applicable professional standards, USEPA requirements, government regulations and guidelines, and specific project goals 6 7 and requirements. Samples were analyzed as specified by the FWSAP current at the time of the 8 investigation (USACE 2001a) and the Phase II RI SAP Addendum (MKM 2001). DQOs were 9 established for the Phase II RI and complied with USEPA Region 5 guidance. The requisite number 10 of QA/QC samples was obtained during the investigation. The data validation determined that the data met the completeness requirements for the project (100% complete), were usable, and that they 11 12 satisfied the DQOs for the project.

- 13
- 14 15

4.2.2.2 Nature and Extent of Contamination

16 A summary of the nature and extent of contamination for each medium sampled during the Phase II 17 RI is presented below.

18

19 Pond Bank Soil (0-1 ft bgs)

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- Arsenic was detected in three samples exceeding the background and preliminary remediation • goals (PRGs) and in eight samples exceeding PRGs. Chromium was detected in three samples exceeding background.
- 24 • No explosives, VOCs, SVOCs, pesticides, PCBs, or nitrates/nitrites were detected above 25 detection limits.
 - Nitrates/nitrites were detected in one sample (ULCPss-001).
- 27 The propellants detected exceed background because the background value for propellants is • 28 zero. However, none of the concentrations exceeded PRGs.
- 30 Se dime nt
- 31
- 32 Arsenic was detected in 3 samples exceeding the background and PRGs and in 27 samples • 33 exceeding PRGs. Chromium was detected in nine samples exceeding background and PRGs. 34
 - No pesticides were detected above detection limits. •
- 35 No background values were established for VOCs, SVOCs, PCBs, or total organic carbon in 36 sediment.
- 37 • Aroclor-1254 was detected in four samples (ULCPsd-006, ULCPsd-011, ULCPsd-020, and 38 ULCPsd-020 duplicate) but at concentrations below PRGs.
- 39 • Sediment samples were analyzed for hexavalent chromium separately (as well as total 40 chromium). Four samples (ULCPsd-010, ULCPsd-018, ULCPsd-021, and ULCPsd-022) 41 contained concentrations of hexavalent chromium below the Region 9 PRG. At each of these 42 locations, the total chromium concentration was also below the Region 9 PRG.

| Arsenic was detected in three samples exceeding PRGs and in three samples exceeding the background and PRGs. Manganese was detected in three samples exceeding background an PRGs and in two samples exceeding the background. No explosives, cyanide, SVOCs, pesticides, PCBs, or nitrates were detected above detection limits. No background values were established for VOCs, sulfides, or sulfates in surface water. Metals and cyanide samples were not filtered; therefore, results represent total concentrations Chloromethane, acetone, and toluene were detected in sample ULCPSW-002 and in duplicat at concentrations below Region 9 PRGs. Groundwater Arsenic was detected in three samples exceeding PRGs and in one sample exceeding th background and PRGs. Manganese was detected in two samples exceeding background an PRGs. No explosives, propellants, cyanide, VOCs, PCBs, or nitrates/nitrites were detected above detection limits. Bis(2-ethylhexyl)phthalate was detected in the duplicate of sample ULCPGW-006 but not in the original sample. No background values were established for SVOCs in groundwater. The pesticide endrin was detected in one sample (ULCPGW-003) at a concentration below the Region 9 PRG. No background values were established for pesticides in groundwater. |
|--|
| 4 background and PRGs. Manganese was detected in three samples exceeding background an PRGs and in two samples exceeding the background. 6 No explosives, cyanide, SVOCs, pesticides, PCBs, or nitrates were detected above detection limits. 8 No background values were established for VOCs, sulfides, or sulfates in surface water. 9 Metals and cyanide samples were not filtered; therefore, results represent total concentrations. 10 Chloromethane, acetone, and toluene were detected in sample ULCPSW-002 and in duplicat at concentrations below Region 9 PRGs. 13 Groundwater 14 15 Arsenic was detected in three samples exceeding PRGs and in one sample exceeding th background and PRGs. Manganese was detected in two samples exceeding background an PRGs. 18 No explosives, propellants, cyanide, VOCs, PCBs, or nitrates/nitrites were detected above detection limits. 20 Bis(2-ethylhexyl)phthalate was detected in the duplicate of sample ULCPGW-006 but not if the original sample. No background values were established for SVOCs in groundwater. 21 The pesticide endrin was detected in one sample (ULCPGW-003) at a concentration below the Region 9 PRG. No background values were established for pesticides in groundwater. |
| PRGs and in two samples exceeding the background. No explosives, cyanide, SVOCs, pesticides, PCBs, or nitrates were detected above detection limits. No background values were established for VOCs, sulfides, or sulfates in surface water. Metals and cyanide samples were not filtered; therefore, results represent total concentrations. Chloromethane, acetone, and toluene were detected in sample ULCPSW-002 and in duplicat at concentrations below Region 9 PRGs. Groundwater Arsenic was detected in three samples exceeding PRGs and in one sample exceeding th background and PRGs. Manganese was detected in two samples exceeding background an PRGs. No explosives, propellants, cyanide, VOCs, PCBs, or nitrates/nitrites were detected above detection limits. Bis(2-ethylhexyl)phthalate was detected in the duplicate of sample ULCPGW-006 but not in the original sample. No background values were established for SVOCs in groundwater. The pesticide endrin was detected in one sample (ULCPGW-003) at a concentration below the Region 9 PRG. No background values were established for pesticides in groundwater. |
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| detection limits. Bis(2-ethylhexyl)phthalate was detected in the duplicate of sample ULCPGW-006 but not if the original sample. No background values were established for SVOCs in groundwater. The pesticide endrin was detected in one sample (ULCPGW-003) at a concentration below the Region 9 PRG. No background values were established for pesticides in groundwater. |
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| the original sample. No background values were established for SVOCs in groundwater. The pesticide endrin was detected in one sample (ULCPGW-003) at a concentration below the Region 9 PRG. No background values were established for pesticides in groundwater. |
| The pesticide endrin was detected in one sample (ULCPGW-003) at a concentration below the Region 9 PRG. No background values were established for pesticides in groundwater. |
| 23 the Region 9 PRG. No background values were established for pesticides in groundwater. |
| |
| |
| 24 |
| 25 Figure 4-4 presents locations that exceed current screening criteria. |
| 26 27 4222 Baseline Human Haakk Disk Assessment |
| 4.2.2.3 <u>Baseline Human Health Risk Assessment</u> 28 |
| A BHHRA was included in the Phase II RI Report (MKM 2005) for the Upper and Lower Cobb |
| 30 Ponds AOC. The BHHRA utilized the screening process in effect at that time, based on the protoco |
| 31 established in the RVAAP Facility-wide Human Health Risk Assessor Manual – Amendment |
| 32 (USACE 2005b) (herein referred to as the FWHHRAM), in order to identify potential contaminant |
| 33 of concern. Risks were evaluated for the receptors available at that time [National Guard Trainee (th |
| 34 primary receptor of concern), National Guard Resident/Trainer, National Guard Fire and Dust Contro |
| 35 Specialist, Hunter/Trapper/Fisher, Security Guard/Maintenance Worker, and Resident Farmer (Adu |
| 36 and Child)]. COPCs were selected and toxicological and exposure factors were applied to evaluat |
| 37 risk. The COPCs identified in the quantitative BHHRA are summarized in Table 4-10. |
| 38 |
| 39 The BHHRA indicated that several significant uncertainties associated with the risk assessment wer |
| 40 identified and must be considered carefully when making any risk management decisions, includin |
| 41 the fact that the nature and extent of the contamination in soil, sediment, surface water, an |
| 42 groundwater at the Upper and Lower Cobbs Ponds AOC had not been completely defined. Therefore 43 the results of the BHHRA are preliminary and subject to change based on the additional data and |
| 43 the results of the BritikA are preliminary and subject to change based on the additional data an 44 information presented in the Phase III RI Report. |

1 4.2.2.4 Screening Ecological Risk Assessment

2

The SERA compared chemical concentrations detected in Upper and Lower Cobbs Ponds

The SERA compared chemical concentrations detected in Upper and Lower Cobbs Ponds environmental media to RVAAP facility-wide background concentrations for inorganic chemicals and ecological screening values (ESVs). The SERA followed screening methodology guidance presented in the 2003 RVAAP Facility-wide Ecological Risk Work Plan (USACE 2003a) (herein

7 referred to as the FWERWP) and *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA

- 8 2003). Chemicals were retained if they did not have screening values.
- 9

10 A SERA was performed to assess whether adverse ecological impacts are present at the Upper and 11 Lower Cobbs Ponds AOC as a result of SRCs detected. The SERA concluded no unacceptable ecological risks were identified. No COPCs were retained for direct toxicity to terrestrial plants, soil 12 13 invertebrates, sediment invertebrates (benthos), and aquatic organisms. Refining conservative 14 assumptions used during food chain modeling led to determining that the risks were acceptable. 15 Refinement included using less conservative bioavailability, less conservative dietary benchmarks 16 [lowest observed adverse effect levels (LOAELs)], and more realistic average exposure scenarios (not 17 maximum) and considering aquatic habitat quality in the ponds, backwater area, and site drainage channels. 18

- 19
- 20 21

4.2.2.5 <u>Conclusions and Recommendations</u>

The Phase II RI Report (MKM 2005) provided the following conclusions applicable to the Upper andLower Cobbs Ponds AOC.

24

The RI objectives were to characterize the nature and extent of potential contamination at the site. The objectives of the RI were met through the field investigations conducted in 2001.
 Field investigation included sampling the pond bank soil, surface water, groundwater, and sediment. Six soil borings were selected to complete as groundwater monitoring wells. In addition to collecting samples to meet RI objectives, aquifer testing was conducted at each well to characterize subsurface lithologic conditions.

- The results of the Phase II RI defined the lateral and vertical extent of contamination, including explosives, propellants, inorganic chemicals, and SVOCs.
- Collection of additional data was recommended to support human health risk management
 decisions and future remedial actions, particularly with respect to elevated inorganic chemical
 concentrations above background levels.
- 36

31

32

- No additional action was recommended for ecological risk management.
- 37

38 4.2.3 2003 Facility-wide Biological and Water Quality Study

39

The FWBWQS collected samples at locations throughout the former RVAAP to evaluate sediment
and surface water from ponds and streams with potential repositories of contamination (USACE
2005a).

- 1 This evaluation included the following specific objectives.
- Establish biological conditions in select streams on the former RVAAP by evaluating fish and
 macroinvertebrate communities, and assessing physical habitat conditions.
 - Measure and evaluate surface water and sediment chemical quality in select streams within the former RVAAP. General COCs included munitions and explosives of concern, semi-volatile organic compounds (SVOCs), and inorganic constituents.
- Determine the aquatic life use attainment status of streams with regard to warmwater habitat
 or other applicable aquatic life use designations codified in the Ohio Water Quality Standards
 (WQS).
 - Determine if chemical contamination within the streams is adversely affecting biological communities.
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• Complete an ecological assessment report summarizing the sediment, surface water, and aquatic biological results.

15

16 The data most relevant to the Upper and Lower Cobbs Ponds AOC were collected from two stations 17 within the Upper and Lower Cobbs Ponds (i.e., stations Lowcobb Pond and Uppercobb Pond), and a 18 sampling station (S-8) along a tributary to Sand Creek, which is fed by drainage from Upper and 19 Lower Cobbs Ponds, which in turn have drainages from Load Lines 3 and 12 (Figure 4-1). Station S-20 8 is adjacent to the former Sand Creek Sewage Treatment Plant. All surface water that exits the Upper 21 and Lower Cobbs Ponds AOC enters the tributary to Sand Creek and passes through the sampling 22 station S-8. One multi-increment sediment sample in each pond, one sample at downstream location 23 S-8, and two surface water samples from each location were collected for the FWBWQS (USACE 24 2005a) in June and August/September 2003. Sediment and surface water samples were analyzed for 25 TAL metals, explosives, pesticides, PCBs, SVOCs, ammonia, and phosphorus. Sediment collected 26 from Lower Cobbs Pond reflected slightly contaminated conditions. Three metals (cadmium, copper, 27 zinc) were reported at levels above the threshold effect concentration (TEC) [and above the sediment reference values (SRVs)]. In addition, chromium (121 mg/kg) was measured above the probable 28 29 effect concentration (PEC), a level above which harmful effects potentially could occur to aquatic 30 biota.

31

4.2.4 PBA08 Remedial Investigation – February to April 2010 33

34 In November 2008, Science Applications International Corporation (SAIC) scientists performed a site 35 walk of the Upper and Lower Cobbs Ponds AOC. The site walk was conducted to develop the 36 Performance-based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan 37 Addendum No. 1 (USACE 2009) (herein referred to as the PBA08 SAP), which supplemented the 38 Phase I and II RIs and completed the RI phase of the CERCLA process. No significant physical 39 changes occurred at the Upper and Lower Cobbs Ponds AOC between the 1996 Phase I RI sampling 40 and the development of the PBA08 SAP. The PBA08 SAP considered prior investigations and 41 changes in AOC conditions during development of the DQOs and sampling scheme for completing 42 the Upper and Lower Cobbs Ponds AOC RI. Section 4.4.4 discusses the suitability and use of 43 samples collected to support the RI, with respect to changes in AOC conditions. The PBA08 SAP was 44 reviewed and approved by representatives of the Army and Ohio EPA in January 2010.

1 As part of the PBA08 RI DQOs, an initial screening approach was used to help focus the 2 investigation on specific chemicals and areas to be further evaluated by assessing the nature and 3 extent of contamination observed in historical samples (Section 3.2.2 of the PBA08 SAP). Decision 4 flowcharts for PBA08 RI surface and subsurface sampling are presented in Figures 4-2 and 4-3, respectively. The screening approach presented in the PBA08 SAP compared sample results from 5 previous investigations at the Upper and Lower Cobbs Ponds AOC to chemical-specific facility-wide 6 7 cleanup goals (FWCUGs) at the 1E-06 cancer risk level and non-carcinogenic risk HQ of 0.1, as 8 presented in the FWHHRAM (USACE 2005b). The most protective FWCUGs for the three potential 9 receptors are referred to as "screening criteria." Previous results were also compared to FWCUGs at the higher TR of 1E-05, HQ of 1 to facilitate identifying potential source areas that may require 10 additional sampling to refine the extent of contamination. Table 4-11 lists the chemicals with detected 11 12 concentrations that exceeded screening criteria at the time of the PBA08 SAP in historical soil 13 samples.

14

15 In February–April 2010, the PBA08 RI was implemented by collecting surface and subsurface soil 16 and surface water using discrete sampling techniques, and sediment using the multi-aliquot composite 17 and discrete sampling techniques. The results of the PBA08 RI sampling, combined with the results 18 of the Phase I and II RIs were used to evaluate the nature and extent of contamination, assess 19 potential future impacts to groundwater, conduct HHRAs and ERAs, and evaluate the need for 20 remedial alternatives.

21

22 No groundwater samples were collected during the PBA08 RI, as the current conditions of 23 groundwater will be evaluated as an individual AOC for the entire facility (designated as RVAAP-66) 24 and addressed in a separate RI/FS Report.

25

26 A sample log for each sample and lithologic soil description for each soil boring collected during the 27 PBA08 RI is included in Appendix A. The DQOs, field activities, sampling methodologies, QA/QC, 28 and management of analytical data for the PBA08 RI are further expanded upon in Appendix I.

29

31

30 4.2.4.1 **Surface Soil Sampling Rationale**

32 Soil samples were collected during the PBA08 RI to assess contaminant occurrence and distribution 33 in surface soil. The PBA08 RI samples were designed to delineate the extent of areas previously 34 identified as having the greatest likelihood of contamination (e.g., along the shoreline/pond banks and 35 drainage ditches). Table 4-12 presents the specific rationale for each surface soil sample collected 36 during the PBA08 RI. Table 4-13 presents the results of the analytes detected from surface soil 37 samples collected during the PBA08 RI. All PBA08 RI sample locations are presented on Figure 4-4.

38

39 Eight surface soil samples were collected at the Upper and Lower Cobbs Ponds AOC during the 40 PBA08 RI to further delineate surface soil above screening criteria, as presented in Table 4-12, and to

41 completely characterize the AOC (Figure 4-4). Of the eight surface soil samples collected, five

42 surface soil samples were collected to delineate contamination along the shoreline and in drainage

- 43 ditches where historical screening criteria exceedances were observed. Three discrete samples were
- 44 collected in areas not previously investigated at the AOC. These samples were collected to further

1 delineate the lateral extent of surface soil contamination. All surface soil samples collected during the 2 PBA08 RI were collected from 0-1 ft bgs in accordance with the bucket hand auger method described 3 in Section 4.5.2.1.1 of the FWSAP (USACE 2001a). An updated version of the FWSAP was 4 developed in February 2011 and approved by Ohio EPA; however, the PBA08 RI was implemented 5 prior to approval of this updated version. Discrete surface soil samples were analyzed for TAL metals, explosives, and SVOCs. One sample (10% of the total number of samples collected) was 6 7 analyzed for RVAAP full-suite analytes. One QC field duplicate and one QA split sample were 8 collected to satisfy the QA/QC requirement of 10% of the total samples collected.

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- 10 11

4.2.4.2 <u>Subsurface Soil Sampling Rationale and Methods</u>

- 12 The PBA08 RI used discrete samples from soil borings to characterize subsurface soil. Subsurface 13 soil sampling was conducted according to the decision rules approved in the PBA08 SAP. The 14 subsurface soil borings were located based on three objectives:
- 15
- 16 17
- 1. Borings at locations where previous surface soil sampling results exceeded screening criteria and vertical delineation was warranted.
- 18 19
- and vertical delineation was warranted.2. Borings at locations where previous surface soil sampling results only slightly exceeded screening criteria to confirm that contaminant concentrations did not increase with depth.
- 20
- 21

3. Borings at locations not previously sampled to fully characterize surface and subsurface soil.

- 22 Subsurface soil was characterized by placing borings in various areas, including areas with previous 23 surface soil results greater than the screening criteria, areas with previous results only slightly greater 24 than the screening criteria, and areas not previously sampled. Four locations were near five historical 25 locations where chemicals exceeded the screening criteria. One location was near Lower Cobbs Pond 26 in an area not previously investigated to verify the absence of subsurface soil contamination. In all 27 cases, soil samples were collected from the subsurface borings to further define the vertical extent of 28 contamination in subsurface soil at the AOC (Figure 4-4). Table 4-14 presents the specific rationale 29 for each subsurface soil sample collected for the PBA08 RI. Results of detected analytes are 30 presented in Table 4-15.
- 31

To assess the depths of exposure of the Resident Receptor, each soil boring was sampled at 0–1, 1–4, 4–7, and 7–13 ft bgs. The deep sample interval was archived on site, while the 4–7 ft bgs sample was analyzed under an expedited five-day turnaround time. As specified in the PBA08 SAP, the deep sample interval would be analyzed for the following reasons:

- 36
- 37 38
- One chemical had a concentration that exceeded screening criteria in the 4–7 ft bgs sample, or
- 39 40

• To ensure at least 10% of all subsurface samples from 7–13 ft bgs were submitted for laboratory analysis to adequately characterize subsurface soil to 13 ft bgs.

41

Each interval was composited and homogenized in a stainless steel bowl, with the exception of VOC
samples. The sample collected from 7–13 ft bgs (where achievable) was archived on site, while the 4–
7 ft bgs sample was analyzed under an expedited five day turnaround time. No 7-13 ft bgs samples

1 were analyzed due to preliminary screening criteria exceedances within the 4-7 ft bgs sample interval,

2 but one sample (CPCSB-032-5116-SO) was analyzed to ensure adequate characterization of 7–13 ft

bgs. All samples were analyzed for TAL metals, explosives, and SVOCs; three samples were analyzed for RVAAP full-suite analytes to satisfy the PBA08 SAP sample requirements of a

- 5 minimum of 10% frequency for full-suite analysis.
- 6

Two QC field duplicates and two QA split samples were collected to satisfy the QA/QC sample
requirements of 10% frequency for subsurface soil samples.

9

10 Two geotechnical samples were collected from one boring location to provide soil data for fate and transport modeling. A pilot boring was installed with a Geoprobe® to a depth of 13 ft bgs to 11 12 lithologically characterize the soil and determine the appropriate geotechnical sample intervals 13 (Appendix A). The geotechnical sample location was offset from the pilot boring and drilled with 14 hollow stem auger attachments. Geotechnical samples were collected through the hollow stem augers 15 directly into Shelby tubes. Shelby tubes were collected from 4–5.4 ft bgs and 8–9.7 ft bgs, directly 16 above the saturated zone observed in the pilot boring. Shelby tubes were sealed with wax, capped, and submitted for laboratory geotechnical analysis for porosity, bulk density, moisture content, total 17 organic carbon, grain size fraction analysis, and permeability. Laboratory analytical results for 18 19 geotechnical samples are presented in Section 5.3.5 and Appendix D.

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- 21 22

4.2.4.3 <u>Surface Water and Sediment Rationale and Methods</u>

Surface water and sediment samples were collected to characterize current conditions and assess
 potential exit pathways from the AOC (Figure 4-4). Five co-located surface water and sediment
 samples were collected during the PBA08 RI.

26

27 The surface water grab samples were collected by the handheld bottle method in accordance with 28 Section 4.3 of the PBA08 SAP and analyzed for RVAAP full-suite analytes. Additionally, water quality parameters for temperature, pH, conductivity, dissolved oxygen, and turbidity were collected 29 30 using calibrated water quality meters (Hanna Instrument Models 9828 and 98703). Ten composite and discrete sediment samples were collected during the PBA08 RI. Six composite samples were 31 32 collected as multi-aliquot composite sediment samples in accordance with Section 4.2 of the PBA08 33 SAP. Each sample consisted of a multi-aliquot composite with 10 aliquots selected randomly within a 34 5-ft radius of the identified sample location. Each aliquot was collected by means of a stainless steel 35 push probe or Ponar sediment sampler lowered from the side of a boat to a maximum depth of 0.5 ft 36 bgs. The aliquots were homogenized in a stainless steel bowl and transferred to the appropriate 37 labeled sample container. CPCSD-044-5022-SD was analyzed for TAL metals, explosives, and 38 SVOCs, and CPCSD-049-5032-SD (chromium speciation sample) was analyzed for total and 39 hexavalent chromium. All other composite sediment samples were analyzed for RVAAP full-suite 40 analytes (and their associated QA/QC samples). For the sample location designated for VOC analysis, 41 one discrete sample collected from 0.5 ft bgs was collected within the 5-ft sampling radius of the 42 discrete sample location and placed directly in the appropriate, labeled sample container.

43

1 Four discrete subsurface sediment samples were collected at a maximum depth of 2 ft bgs from a 2 stainless steel split core barrel attached to the end of a slide hammer in accordance with Section 4.2 of 3 the PBA08 SAP. Sediment from 0.5-2 ft bgs was homogenized in a stainless steel bowl, transferred to 4 the appropriate, labeled sample container, and analyzed for RVAAP full-suite analytes. No QC field duplicate or QA split sediment samples were collected. Table 4-16 presents the specific rationale for 5 the surface water and sediment samples collected for the PBA08 RI. The locations of these samples 6 7 are presented in Figure 4-4 and the results are presented in Tables 4-17 and 4-18 for surface water and 8 sediment, respectively.

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- 10

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4.2.4.4 Changes from the Work Plan

Significant changes to the work plan are documented in field change requests (Appendix B). Changes made in the field based on site conditions are not documented on field change requests but on the field sampling logs (Appendix A). These changes are presented on Table 4-19. New coordinates for all station locations can be found on the field sampling logs.

16 17

4.3 FACILITY-WIDE BACKGROUND EVALUATION

18

19 Facility-wide background values for inorganic constituents in soil, sediment, surface water, and 20 groundwater were developed in 1998, as documented in the *Phase II Remedial Investigation Report* 21 *for the Winklepeck Burning Grounds* (USACE 2001b). These background values are currently being 22 reassessed, but the background valued developed in 1998 are used throughout this report.

23

24 These facility-wide background values developed in 1998 were employed in the data reduction and 25 screening process described in Section 4.4.2 and the remainder of the evaluations in this RI (e.g., 26 nature and extent, fate and transport). Background locations were selected using aerial photographs 27 and site visits from areas believed to be unaffected by RVAAP activities. Soil, sediment, surface 28 water, and groundwater samples were collected from those locations to determine the range of 29 background concentrations that could be expected in these media. Results from the site-specific 30 background data collection were used to determine if detected metals and potential anthropogenic 31 compounds [such as polycyclic aromatic hydrocarbons (PAHs)] are site-related, naturally occurring, 32 or from non-RVAAP-related anthropogenic sources.

33

A total of 14 wells were installed in established background locations to collect filtered and unfiltered samples from the bedrock and unconsolidated zones. These samples were analyzed for TAL metals and cyanide for determining background concentrations.

37

Soil samples were collected from each of the background monitoring well locations from three intervals: 0–1, 1–3, and greater than 3 ft bgs. Because boring locations were changed during sampling based on the lithological requirements for well screen intervals, all depth intervals for soil were not sampled for each boring. Background surface soil samples were analyzed for TAL metals, cyanide, SVOCs, pesticides, PCBs, and VOCs. Surface water samples were analyzed for TAL metals and cvanide.

44

1 Seven stream locations upstream of RVAAP activities along Hinkley, Sand, and Eagle Creeks were

2 sampled for sediment and surface water to characterize background conditions. Background sediment

3 samples were analyzed for TAL metals, cyanide, SVOCs, pesticides, PCBs, and VOCs. Surface water

- 4 samples were analyzed for TAL metals and cyanide.
- 5

6 Using the sampling results, an evaluation of outliers, data assessment, and statistical analyses were 7 performed to determine background concentrations for each medium. For surface soil samples, PAHs, 8 in addition to metals, were elevated in four samples. PAHs are related to combustion products and 9 could indicate human disturbance at the locations where they were detected. Visits to the sampling 10 locations and a review of aerial photography showing the area prior to the establishment of RVAAP 11 indicated that these sampling locations were near homes or farms and could have been influenced by 12 activities associated with those structures. During the finalization of background concentrations at the 13 former RVAAP, the Army and Ohio EPA agreed that formal background concentrations would only be applicable for inorganics. All organic analytes (e.g., PAHs, VOCs, explosives) were classified as 14 15 anthropogenic and potentially related to RVAAP operations; therefore, no background values were 16 established for these classes of compounds. The final, approved facility-wide background 17 concentrations for inorganics are presented in Table 4-20.

18 19

20

4.4 DATA EVALUATION METHOD

Data evaluation methods for the Upper and Lower Cobbs Ponds AOC are consistent with those established in the FWCUG Report. These methods were specified in the PBA08 SAP (USACE 2009). The processes used to evaluate the analytical data involved three general steps: (1) defining data aggregates; (2) conducting data verification, reduction, and screening; and (3) presenting data.

25 26

27

4.4.1 Definition of Aggregates

The Upper and Lower Cobbs Ponds AOC data were aggregated in three ways for evaluating contaminant nature and extent and completing the HHRA and ERA. The initial basic aggregation of data was by environmental medium: surface soil, subsurface soil, sediment, and surface water. For each medium-specific aggregate, an evaluation was conducted to determine if further aggregation was warranted with respect to AOC characteristics, ecological habitat, and potential future remedial strategy and Land Use (e.g., spatial aggregates). Data for soil were further aggregated based on depth and sample type for consistency with guidance established in the FWHHRAM and FWCUG Report.

35

36 Data aggregates for evaluating the nature and extent of contamination at the Upper and Lower Cobbs37 Ponds AOC are as follows:

38

Surface Soil (0-1 ft bgs). Using the above data aggregation criteria, surface soil within the geographic area of Upper Cobbs Pond, Lower Cobbs Pond, and the Backwater Area were evaluated as an AOC-wide spatial aggregate, designated as the Pond Bank aggregate. The Pond Bank is defined as the perimeter of the natural shorelines throughout the entirety of the AOC. It extends from the Backwater Area to the south at the confluence of the channels exiting Load Lines 3 and 12 and north to the intersection of Remalia Road and Paris-

1 Windham Road. The areas encompassed within the Pond Bank are dry the majority of the 2 time, although they may be intermittently wet in association with heavy rainfall events when 3 they may receive runoff from the surrounding areas or when water levels within the AOC are 4 elevated.

5 6 • Subsurface Soil (greater than 1 ft bgs). This medium was classified as an AOC-wide spatial aggregate on the same basis as surface soil (i.e., Pond Bank).

- Sediment. Sediment samples were divided into three spatial aggregates for this report:
 Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. All three of these aggregates
 are perennially inundated water bodies. The Backwater Area is in the southern portion of the
 complex, extending from the confluence of the drainage channels exiting Load Lines 3 and
 12 and reaching north to Track 33. Upper Cobbs Pond is the portion of the water body north
 of Track 33 to Cobbs Pond Road. Lower Cobbs Pond is the portion north of the Cobbs Pond
 Road, bounded by Remalia Road on the north and Paris-Windham Road to the west.
- Surface Water. This medium is subdivided into the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond spatial aggregates on the same basis as sediment. Similar to sediment, surface water was divided into three spatial aggregates for this report: Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. The direction of flow within the AOC is from the south (Backwater Area) towards the dam overflow spillway at the north (Lower Cobbs Pond).
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4.4.2 Data Verification, Reduction, and Screening

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4.4.2.1 Data Verification

Data verification was performed on 38 surface soil, subsurface soil, sediment, and surface water samples (including QC duplicates) collected during the PBA08 RI in February–April 2010. Historical data were verified and completed as presented in the historical reports. Analytical results were reported by the laboratory in electronic format and loaded into the Ravenna Environmental Information Management System (REIMS) database. Data verification was performed to ensure all requested data were received and complete. Data qualifiers were assigned to each result based on the laboratory QA review and verification criteria.

31

32 Results were qualified as follows:

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- "U" not detected;
- "UJ" not detected, reporting limit estimated;
- "J" indicates the analyte was positively identified, but the associated numerical value is
 an approximate concentration of the analyte in the sample; and
- 38 39

•

"R" result not usable.

In addition to assigning qualifiers, the verification process also selected the appropriate result to use when re-analyses or dilutions were performed. Where laboratory surrogate recovery data or laboratory QC samples were outside of analytical method specifications, the verification chemist determined whether laboratory re-analysis should be used in place of an original reported result. If the laboratory reported results for both diluted and undiluted samples, diluted sample results were used for those analytes that exceeded the calibration range of the undiluted sample. A complete discussion of verification process results is contained in the data QC summary report (Appendix C). The data QC summary report also includes a summary table of the assigned data qualifiers and an accompanying rationale. Independent, third-party validation of 10% of the RI data, and 100% of the USACE QA laboratory data, was performed by a subcontractor to the USACE Louisville District.

6

7 4.4.2.2 Data Reduction

8

9 Calculating data summary statistics was the initial step in the data reduction process to identify SRCs. 10 Eligible historic and current AOC data were extracted from the database. Results from OC splits and 11 field duplicates, as well as rejected results, were excluded from the data screening process. All 12 analytes having at least one detected value were included in the data reduction process. As stated in 13 Section 5.4.7 of the FWSAP, "The duplicate is submitted as 'blind' to the laboratory and is used to 14 determine whether the field sampling technique is reproducible and to check the accuracy of reported 15 laboratory results." Therefore, duplicates are not used in the data screening process. All analytes 16 having at least one detected value were included in the data reduction process.

17

Summary statistics calculated for each data aggregate included the minimum, maximum, and average (mean) values and the proportion of detected results to the total number of samples collected. For calculating mean values, non-detected results were addressed by using one-half of the reported detection limit as a surrogate value when calculating the mean result for each compound (USEPA 1989). Non-detected results with elevated detection limits (i.e., more than five times the contract-required detection limit) were excluded from the summary statistics in order to avoid skewing the mean value calculations.

25

26 4.4.2.3 Data Screening

27

28 After reduction, the data were screened to identify SRCs using the processes outlined below. Additional screening of identified SRCs against applicable criteria (e.g., USEPA RSLs, FWCUGs, 29 30 and ESVs) was conducted (1) in the fate and transport evaluation (Section 6.0) to identify CMCOPCs, 31 (2) in the HHRA to identify human health COPCs and COCs (Section 7.2), and (3) in the ERA to 32 evaluate COPECs (Section 7.3). Figure 4-5 illustrates the screening process to identify SRCs and 33 COPCs at the Upper and Lower Cobbs Ponds AOC in accordance with the FWCUG Report. All 34 chemicals that were not eliminated during the screening steps were retained as SRCs. The steps 35 involved in the SRC screening are summarized below:

- 36
- 37 38
- **Data quality assessment**. Review the usability of the RI data set with respect to established DQOs as discussed in Section I.4.5 of Appendix I.
- Background screening. Compare the maximum detected concentrations (MDCs) of
 inorganic chemicals to background concentrations. If background concentrations are
 exceeded, the respective inorganic chemicals are retained as SRCs. No background
 concentrations were established for organic chemicals at the Upper and Lower Cobbs Ponds
 AOC. As such, all detected organic chemicals were retained as SRCs.

Screening of essential human nutrients. Evaluate chemicals that are considered essential 1 2 nutrients (e.g., calcium, chloride, iodine, iron, magnesium, potassium, phosphorous, and 3 sodium) that are an integral part of the human food supply and are often added to foods as 4 supplements. USEPA recommends these chemicals not be evaluated unless they are grossly elevated relative to background concentrations or would exhibit toxicity at the observed 5 concentrations at an AOC (USEPA 1989). Recommended daily allowance (RDA) and 6 7 recommended daily intake (RDI) values are available for all of these chemicals (Table 4-21). 8 Screening values are calculated for receptors ingesting 100 mg of soil per day or 1 L of 9 groundwater per day to meet their RDA/RDI. In the case of calcium, magnesium, 10 phosphorous, potassium, and sodium, a receptor ingesting 100 mg of soil per day would receive less than the RDA/RDI value, even if the soil consisted of the pure mineral (i.e., soil 11 12 concentrations at 1,000,000 mg/kg). Essential nutrients detected at or below their RDA/RDI-13 based SLs are eliminated as SRCs.

- 14 Frequency of detection/WOE screening: The FWCUG Report and the Final (Revised) • 15 USACE RVAAP Position Paper for the Application and Use of Facility-wide Human Health Cleanup Goals (USACE 2012a) (hereafter referred to as the FWCUG Position Paper) 16 17 establish the protocol for frequency of detection and WOE screening. These guidance 18 documents denote that analytes (with exception of explosives and propellants) detected in 19 less than 5% of the discrete samples are screened out from further consideration, if the 20 sample population is 20 or more samples and evidence exists that the analyte is not AOC-21 related. Chemicals that were never detected in a given medium are eliminated as SRCs. For 22 chemicals with at least 20 samples and a frequency of detection of less than 5%, a WOE 23 approach is used to determine if the chemical is AOC-related. The WOE evaluates magnitude 24 and location (clustering) of detected results and if the distribution of detected results indicates 25 a potential source of the chemical. If the detected results for a chemical show (1) no 26 clustering, (2) concentrations were not substantially elevated relative to detection limit, and 27 (3) the chemical did not have an evident source, the results are considered spurious, and the 28 chemical is eliminated from further consideration. This screen is applied to all organic 29 chemicals and inorganic chemicals (with the exception of explosives and propellants); all 30 detected explosives and propellants are considered SRCs regardless of frequency of 31 detection. Frequency of detection/WOE screening only applied to the Pond Bank surface soil 32 data set, as it was comprised of 20 or more samples.
- 33

34 4.4.3 Data Presentation

35

36 Data summary statistics and screening results for SRCs in surface soil, subsurface soil, sediment, and 37 surface water at the Upper and Lower Cobbs Ponds AOC are presented below for each media and 38 spatial aggregate. Analytical results for SRCs are presented in Table 4-22 for surface soil, Table 4-23 39 for subsurface soil, Tables 4-24 through 4-31 for sediment, and Tables 4-32 through 4-34 for surface 40 water.

41

The complete laboratory analytical data packages are included in Appendix D. In order to maximize efficiency for laboratory reporting and data management activities, all of the samples received at the laboratory on a given day were reported in a single data package. Therefore, results may be present in the data packages in Appendix D that are associated with different AOCs. All samples for the Upper and Lower Cobbs Ponds AOC have sample identifications beginning with "CPC." Each table in Appendix D presents the results for each sampling location for a specific medium aggregate (i.e., surface soil, subsurface soil, sediment, and surface water), spatial aggregate (i.e., Pond Bank, Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond), and class of analyte (e.g., explosives, inorganic chemicals, SVOCs, and VOCs).

7

8 The tables in Appendix D present the analytical results for samples collected during the 1996 Phase I 9 RI, 2001 Phase II RI, 2003 FWBWQS, and PBA08 RI. Sample locations from these investigations 10 are presented on Figure 4-6. Analytical results are grouped by media (e.g., surface soil, subsurface 11 soil) and class of analyte (e.g., explosives, inorganic chemicals) for ease of reference.

12

14

13 4.4.4 Data Evaluation

All quality-assured sample data were further evaluated to determine suitability for use in the various key RI data screens and evaluations (i.e., nature and extent, fate and transport, risk assessment).
Evaluating data suitability for use in the PBA08 RI involved considering representativeness with respect to current AOC conditions. Table 4-35 presents the designated use for all available Upper and Lower Cobbs Ponds samples.

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21 4.4.4.1 <u>Soil</u>

22

23 Surface and subsurface soil samples at the Upper and Lower Cobbs Ponds AOC were collected using 24 discrete sample methods during the 2001 Phase II RI and the PBA08 RI. Samples from 2001 (Phase 25 II RI) were evaluated to determine if conditions had changed substantively between earlier 26 characterization efforts and the PBA08 RI activities. No AOC disturbance activities occurred at the 27 Upper and Lower Cobbs Ponds AOC between the Phase I RI sampling in 1996 and the PBA08 RI 28 sampling in 2010. However, since more recent sample data were available from the PBA08 RI, the 29 Phase II samples were utilized for nature and extent purposes only. The full Phase II RI and PBA08 30 RI data sets were incorporated into the SRC screening process, and more recent sample data were 31 carried forward into the risk assessment.

32

33 4.4.4.2 Sediment and Surface Water

34

35 The only medium sampled in the 1996 Phase I RI was sediment. Sediment throughout the AOC was 36 resampled in the 2001 Phase II RI and also under the 2010 PBA08 RI activities. The Phase II RI, in 37 addition to sediment, also incorporated surface water sample collection. PBA08 RI activities 38 consisted of sediment and surface water sampling. Under the PBA08 RI sampling activities, sediment 39 samples were collected from 0-0.5 ft bgs and 0.5-2 ft bgs at a subset of locations in order to 40 characterize vertical extent. The PBA08 RI samples collected from 0.5-2 ft bgs were screened as 41 subsurface sediment. The surface sediment screening data set includes both the PBA08 RI samples 42 from 0-0.5 ft bgs and historical sediment samples.

43

1 Samples from the 1996 Phase I RI and 2001 Phase II RI data sets were evaluated to determine if 2 conditions had changed substantively between earlier characterization efforts and the 2010 PBA08 RI 3 activities. No disturbance activities have occurred at the AOC between Phase I RI and PBA08 RI 4 activities. For sediment and surface water data, if a PBA08 RI sample was obtained from a historical 5 (1996 Phase I RI or 2001 Phase II RI) sample location, the PBA08 RI result was considered to represent current conditions and was screened for SRCs. All historical sediment and surface water 6 7 samples superseded in the SRC screen by PBA08 RI data were used only for evaluating contaminant 8 nature and extent (e.g., temporal trends) and contaminant transport.

9

10 Off AOC sediment and surface water samples collected at location S-8 (FSW-012 and FSW-052) 11 during the 2003 FWBWQS were used for nature and extent comparison only. Samples collected at 12 locations Uppercobb Pond (FSW-030) and Lowcobb Pond (FSW-031) were used in the risk 13 assessments included in this Phase III RI.

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- 15

| , | Table 4–1 | . Phase I | RI Sa | ampling | Locations |
|---|-----------|-----------|-------|---------|-----------|

| Phase I RI Sample | Sample Depth | | Potential Sources or Areas for |
|-------------------|--------------|---|--|
| Location | (ft bgs) | Analytes | Investigation |
| CPCsd-001(P) | 0-1 | Metals, Explosives | Upper Cobbs Pond – southern portion near culvert draining Backwater Area |
| CPCsd-002(P) | 0–1 | Metals, Explosives | Upper Cobbs Pond – southwestern portion |
| CPCsd-003(P) | 0-1 | Metals, Explosives | Upper Cobbs Pond – south-central |
| CPCsd-004(P) | 0–1 | Metals, Explosives | Upper Cobbs Pond – north-central |
| CPCsd-005(P) | 0–1 | Metals, Explosives | Upper Cobbs Pond – northern portion near culvert draining to Lower Cobbs Pond |
| CPCsd-006(P) | 0-1 | Metals, Explosives | Lower Cobbs Pond – southwestern portion near culvert draining Upper Cobbs Pond |
| | 0-1 | Metals, Explosives | |
| CPCsd-007(P) | 0-1 | Metals, Explosives (QA/QC sample collected) | Lower Cobbs Pond – north-central portion |
| CPCsd-008(P) | 0-1 | Metals, Explosives | Lower Cobbs Pond – northwestern portion |
| CPCsd-009(P) | 0–0.5 | Metals, Explosives | Lower Cobbs Pond – northwestern portion adjacent to access road |
| CPCsd-010(P) | 0–0.5 | Metals, Explosives | Drainage Ditch West of Lower Cobbs Pond (near ULCPmw-006) |

bgs = Below ground surface.

Comp B = Composition B (a combination of TNT and RDX).

DNT = Dinitrotoluene.

ft = Feet.

HMX = Octahy dro-1, 3, 5, 7-tetranitro-1, 3, 5, 7-tetrazocine.

QA = Quality assurance.

QC = Quality control.

RDX = Hexahy dro-1,3,5-trinitro-1,3,5-triazine.

RI = Remedial investigation.

TNT = 2,4,6-Trinitrotoluene.

| Area | | Lower Cobbs Pond | Upper Cobbs Pond |
|------------------------|------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|--------------------------|
| Station | | CPCsd-006(p) | CPCsd-007(p) | CPCsd-007(p) | CPCsd-008(p) | CPCsd-009(p) | CPCsd-010(p) | CPCsd-001(p) | CPCsd-002(p) | CPCsd-003(p) | CPCsd-004(p) | CPCsd-005(p) |
| Sample ID | | CPCsd-006(P)- 0655-SD | CPCsd-007(P)- 0657-FD | CPCsd-007(P)- 0656-SD | CPCsd-008(P)- 0659-SD | CPCsd-009(P)- 0660-SD | CPCsd-010(P)- 0661-SD | CPCsd-001(P)- 0650-SD | CPCsd-002(P)- 0651-SD | CPCsd-003(P)- 0652-SD | CPCsd-004(P)- 0653-SD | CPCsd-005(P)- 0654-SD |
| Date | | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | TAL Metals, | RVAAP Full- | RVAAP Full- | TAL Metals, |
| Analyte | Criteria | Explosives | suite analytes | suite analytes | Explosives |
| | | · · · · | Ú Ú | J. v | Metals (mg | | . | · · · | . | | | <u> </u> |
| Aluminum | 13900 | 12500 | 9930 | 8500 | 9610 | 4470 | 4310 | 5830 | 12000 | 9040 | 10300 | 9520 |
| Antimony | 0 | NR | 1.3J* | 1.9* | NR |
| Arsenic | 19.5 | 16.4 | 10.5 | 14.4 | 8.5 | 8.2 | 9.4 | 4.9 | 23.4* | 5.9 | 5 | 7.7 |
| Barium | 123 | 100 | 82 | 79.1 | 77.3 | 40.8 | 41.4 | 43.4 | 66.1 | 52.3 | 46.7 | 71.1 |
| Beryllium | 0.38 | NR | 0.74* | 0.73* | NR |
| Cadmium | 0 | 1J* | 1.2J* | 1.4* | 1.4* | <0.2U | <0.13U | 0.54J* | 0.75J* | 0.54J* | <0.29U | 0.54J* |
| Calcium | 5510 | NR | 2110 | 2320 | NR |
| Chromium | 18.1 | 86.2* | 147* | 329* | 63.2* | 8.3 | 16.2 | 7.9 | 40.9* | 54.3* | 30.8* | 38.9* |
| Cobalt | 9.1 | NR | 10.9* | 11.1* | NR |
| Copper | 27.6 | NR | 143* | 316* | NR |
| Iron | 28200 | NR | 19400 | 19600 | NR |
| Lead | 27.4 | 37.8* | 31.3* | 34.2* | 33.3* | 10.1 | 8.8 | 12.6 | 31.6* | 17.9 | 14.8 | 19.1 |
| Magnesium | 2760 | NR | 1960 | 1640 | NR |
| Manganese | 1950 | 407 | 354 | 345 | 517 | 425 | 816 | 312J | 483J | 239J | 171J | 375J |
| Mercury | 0.059 | 0.15* | 0.09* | 0.11* | <0.09U | <0.05U | <0.05U | <0.06U | 0.12* | 0.08* | 0.23* | 0.11* |
| Nickel | 17.7 | NR | 22.4* | 20.9* | NR |
| Potassium | 1950 | NR | 842J | 683J | NR |
| Selenium | 1.7 | 2.1* | 2.2* | 2.2* | 2.4* | 1.2 | 1.3 | 1.1 | 2.9* | 1.8* | 1.4 | 1.7 |
| Silver | 0 | <0.42U | 1.1J* | 1.7J* | 1.4J* | <0.27U | <0.27U | <0.31U | 1.3J* | 1.4J* | 0.38J* | 0.98J* |
| Sodium | 112 | NR | 482J* | 458J* | NR |
| Thallium | 0.89 | NR | 2.4* | 2.3* | NR |
| Vanadium | 26.1 | NR | 16.3 | 15.1 | NR |
| Zinc | 532 | 222 | 202 | 254 | 182 | 46.2 | 49.5 | 68.3 | 167 | 133 | 90.8 | 127 |
| | | | T | 1 | Explosives (| 0 0, | 1 | T | 1 | T | T | |
| Nitrobenzene | None | <0.26U | 0.38* | <0.26U | 0.38J* |
| | | | 0.0071 | | SVOCs (m | | | | | | | |
| Benz(a)anthracene | None | NR | 0.22J* | 0.21J* | NR |
| Benzo(a)pyrene | None | NR | 0.26J* | 0.26J* | NR |
| Benzo(b)fluoranthene | None | NR | 0.37J* | 0.56J* | NR |
| Benzo(ghi)perylene | None | NR | 0.23J* | 0.2J* | NR |
| Benzo(k)fluoranthene | None | NR | 0.32J* | <1.9U | NR |
| Chrysene | None | NR | 0.29J* | 0.27J* | NR |
| Fluoranthene | None | NR | 0.43J* | 0.38J* | NR |
| Indeno(1,2,3-cd)pyrene | None | NR | 0.22J* | 0.19J* | NR |
| Pyrene | None | NR | 0.29J* | 0.27J* | NR |

Table 4-2. Analytes Detected in Phase I RI Sediment Samples

Table 4–2. Analytes Detected in Phase I RI Sediment Samples (continued)

| | | Lower Cobbs | Lower Cobbs | Lower Cobbs | Lower Cobbs | Lower Cobbs | Lower Cobbs | Upper Cobbs |
|---------------------|------------|---------------|----------------|--------------------|---------------|---------------|---------------|---------------|---------------|---------------|---------------|---------------|
| Area | | Pond | Pond | Pond | Pond | Pond | Pond | Pond | Pond | Pond | Pond | Pond |
| Station | | CPCsd-006(p) | CPCsd-007(p) | CPCsd-007(p) | CPCsd-008(p) | CPCsd-009(p) | CPCsd-010(p) | CPCsd-001(p) | CPCsd-002(p) | CPCsd-003(p) | CPCsd-004(p) | CPCsd-005(p) |
| | | CPCsd-006(P)- | CPCsd-007(P)- | CPCsd-007(P)- | CPCsd-008(P)- | CPCsd-009(P)- | CPCsd-010(P)- | CPCsd-001(P)- | CPCsd-002(P)- | CPCsd-003(P)- | CPCsd-004(P)- | CPCsd-005(P)- |
| Sample ID | | 0655-SD | 0657-FD | 0656-SD | 0659-SD | 0660-SD | 0661-SD | 0650-SD | 0651-SD | 0652-SD | 0653-SD | 0654-SD |
| Date | | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 | 08/19/96 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | TAL Metals, | RVAAP Full- | RVAAP Full- | TAL Metals, |
| Analyte | Criteria | Explosives | suite analytes | suite analytes | Explosives |
| | | | | | VOCs (mg | r/kg) | | | | | | |
| Acetone | None | NR | 0.28J* | 0.33J* | NR |
| Methylene chloride | None | NR | 0.035* | <0.029U | NR |

^aOnly detected site-related contaminants are presented in the table. ^bBackground concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

^bBackground concentrations are published in th ft = Feet. ID = Identification. J = Estimated value less than reporting limits. mg/kg = Milligrams per kilogram. NR = Not reported/not analyzed. RI = Remedial investigation. RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

VOC = Volatile organic compound.

TAL = Target analyte list.U = Not detected.

* = Result exceeds background criteria or no background criteria was available.

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation |
|-------------------------------|--------------------------|--|--|
| ULCPss-001 | 0-1 | Metals, Explosives, Anions | Pond Bank – Backwater Area |
| ULCPss-002 | 0-1 | Metals, Explosives, Anions | Pond Bank –Backwater Area near Track FA |
| ULCPss-003 | 0–1 | Metals, Explosives, Anions | Pond Bank – Northeastern bank of Backwater Area |
| ULCPss-004 | 0-1 | Metals, Explosives, Anions | Pond Bank – Northwestern bank of Backwater Area |
| ULCPss-005 | 0-1 | Metals, Explosives, Anions | Pond Bank – Southeastern bank of Upper Cobbs Pond |
| ULCPss-006 | 0-1 | Metals, Explosives, Anions | Pond Bank – Southwestern bank of Upper Cobbs Pond |
| | 0-1 | Full suite | |
| ULCPss-007 | 0-1 | Full suite (QA/QC sample collected) | Pond Bank – Northwestern bank of Upper Cobbs Pond |
| ULCPss-008 | 0–1 | Metals, Explosives, Anions | Pond Bank – Northeastern bank of Upper Cobbs Pond |
| ULCPss-009 | 0–1 | Metals, Explosives, Anions | Pond Bank – Southwestern bank of Lower Cobbs Pond |
| ULCPss-010 | 0–1 | Metals, Explosives, Anions | Pond Bank – Northern bank of Lower Cobbs Pond |
| ULCPsd-010 | 0–0.5 | Metals, Explosives, Anions | Pond Bank –Backwater Area along Track FA |
| | 0-1 | Metals, Explosives, Anions | Backwater Area at Confluence of |
| ULCPsd-004 | 0–1 | Metals, Explosives, Anions (QA/QC sample collected) | Southern Tributaries from Load Lines 3 and 12 |
| ULCPsd-005 | 0-1 | Metals, Explosives, Anions | Backwater Area Downgradient of Southern Tributaries from Load Lines 3 and 12 |
| ULCPsd-006 | 0-1 | Full suite | Backwater Area Downgradient of Southern Tributaries from Load Lines 3 and 12 |
| ULCPsd-007 | 0-0.8 | Metals, Explosives, Anions | Backwater Area – South-central portion, South of Track FA |
| ULCPsd-008 | 0–0.5 | Metals, Explosives, Anions | Backwater Area – Northeastern Portion, South of Track FA |
| ULCPsd-009 | 0-0.9 | Metals, Explosives, Anions | Backwater Area – Northwestern Portion, Just South of Track FA Culvert |
| ULCPsd-011 | 0–0.8 | Full suite | Backwater Area Adjacent to Track FA culvert |
| ULCPsd-012 | 0–1.8 | Metals, Explosives, Anions | Backwater Area Northwest of Track FA Culvert |
| | 0-1.8 | Metals, Explosives, Anions | Backwater Area – Northern Portion |
| ULCPsd-013 | 0-1.8 | Metals, Explosives, Anions | South of Load Line 3 Road |
| | 0-1.0 | (QA/QC sample collected) | |
| ULCPsd-014 | 0-0.5 | Metals, Explosives, Anions | Backwater Area – Northwestern Portion, South of Load Line 3 Road |
| ULCPsd-015 | 0-0.5 | Metals, Explosives, Anions | Backwater Area between Load Line 3 |
| ULCPsd-015 | 0-1 | VOCs, SVOCs | Road and Track 33 culverts |
| ULCPsd-016 | 0–1.5 | Metals, Explosives, Anions | Backwater Area between Load Line 3 Road and Track 33, East of Culverts |
| ULCPsd-017 | 0-1 | Metals, Explosives, Anions | Upper Cobbs Pond – Eastern Portion |

Table 4–3. Phase II RI Sampling Locations

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation |
|-------------------------------|--------------------------|--|--|
| ULCPsd-018 | 0–0.5 | Metals, Explosives, Anions | Upper Cobbs Pond – Southwestern Portion |
| ULCPsd-019 | 0–0.5 | Metals, Explosives, Anions | Upper Cobbs Pond – Northwestern Portion |
| ULCPsd-020 | 0-0.8 | Full suite | Llan - Califa Daniel - Narthann Darthan |
| ULCPsd-020 | 0–0.8 | Full suite (QA/QC sample collected) | Upper Cobbs Pond – Northern Portion near Track 25 Culvert |
| ULCPsd-021 | 0-1 | Metals, Explosives, Anions | Lower Cobbs Pond – Southern Portion |
| ULCPsd-021 | 0-1 | VOCs, SVOCs | Lower Cobbs Pond – Southern Portion |
| ULCPsd-022 | 0–0.8 | Metals, Explosives, Anions | Lower Cobbs Pond – East-central Portion |
| ULCPsd-023 | 0–0.8 | Metals, Explosives, Anions | Lower Cobbs Pond – West-central Portion |
| ULCPsd-024 | 0-0.5 | Metals, Explosives, Anions | Lower Cobbs Pond – outlet channel |
| ULCPsd-025 | 0-0.5 | Metals, Explosives, Anions | Downstream of Lower Cobbs Pond, near Dam Overflow Spillway |
| ULCPsd-026 | 0-1 | Metals, Explosives, Anions | Lower Cobbs Pond – Northern Portion |
| ULCPsd-027 | 0-0.5 | Metals, Explosives, Anions | Headwaters of the Load Line 3 Tributary |
| ULCPsw-001 | NA | Metals, Explosives, Anions | Backwater Area Downgradient of Southern Tributaries from Load Lines 3 and 12 |
| ULCPsw-002 | NA | Metals, Explosives, Anions | Backwater Area Northwest of Track |
| ULCPsw-002 | NA | Metals, Explosives, Anions (QA/QC sample collected) | FA Culvert |
| ULCPsw-003 | NA | Metals, Explosives, Anions | Upper Cobbs Pond – Eastern Portion |
| ULCPsw-004 | NA | Metals, Explosives, Anions | Upper Cobbs Pond – Northwestern Portion |
| ULCPsw-005 | NA | Metals, Explosives, Anions | Lower Cobbs Pond – Southern Portion |
| ULCPsw-006 | NA | Metals, Explosives, Anions | Lower Cobbs Pond – Outlet Channel |
| ULCPsw-007 | NA | Metals, Explosives, Anions | Downstream of Lower Cobbs Pond, near Dam Overflow Spillway |

Table 4-3. Phase II RI Sampling Locations (continued)

Full suite = Target analyte list metals, explosives, propellants, VOCs, SVOCs, polychlorinated biphenyls, pesticides and cyanide.

bgs = Below ground surface.

Comp B = Composition B (a combination of TNT and RDX).

DNT = Dinitrotoluene.

ft = Feet.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

NA = Not applicable.

QA = Quality assurance.

QC= Quality control.

RDX = Hexahydro-1,3,5-trinitro-1,3,5-triazine.

RI = Remedial investigation.

SVOC = Semi-volatile organic compound.

TNT = 2,4,6-Trinitrotoluene. VOC = Volatile organic compound.

Upper and Lower Cobbs Ponds

| A | | Dand Dank | Dand Dank | Dand Dank | Dend Denk | Dand Dank | Pond Bank | David David | Pond Bank | Dend Denk | Dend Denk | Dend Denk | Pond Bank |
|---|------------|-------------|-------------|-------------|-------------|---------------|-------------|-------------|--------------------|--------------------|-------------|-------------|------------|
| Area | 4 | Pond Bank | | Pond Bank | | Pond Bank | Pond Bank | Pond Bank | |
| Station | _ | ULCPsd-010 | ULCPss-001 | ULCPss-002 | ULCPss-003 | ULCPss-004 | ULCPss-005 | ULCPss-006 | ULCPss-007 | ULCPss-007 | ULCPss-008 | ULCPss-009 | ULCPss-01 |
| ~ | | ULCPsd-010- | ULCPss-001- | ULCPss-002- | ULCPss-003- | ULCPss-004- | ULCPss-005- | ULCPss-006- | ULCPss-007- | ULCPss-007- | ULCPss-008- | ULCPss-009- | ULCPss-01 |
| Sample ID | _ | 0001-SD | 0001-SO | 0001-SO | 0001-SO | 0001-SO | 0001-SO | 0001-SO | 0001-FD | 0001-SO | 0001-SO | 0001-SO | 0001-SO |
| Date | | 08/21/01 | 07/17/01 | 07/17/01 | 07/18/01 | 07/17/01 | 07/18/01 | 07/19/01 | 07/18/01 | 07/18/01 | 07/18/01 | 07/19/01 | 07/19/01 |
| Depth (ft) | | 0.0 - 0.5 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | | Misc., TAL | Misc., TAL | Misc., TAL | | | Misc., TAL | Misc., TAL | Misc., TA |
| • | Background | Metals, | Metals, | Metals, | Metals, | Metals, | Metals, | Metals, | RVAAP Full- | RVAAP Full- | Metals, | Metals, | Metals, |
| Analyte (mg/kg) | Criteria | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | suite analytes | suite analytes | Explosives | Explosives | Explosives |
| | • | <u> </u> | - | · • | · • | Metals (mg | y/kg) | <u> </u> | · · · · · | · · · · · · | - | - | <u> </u> |
| Aluminum | 17700 | 6450 | 11000 | 11100 | 10200 | 18000* | 19200* | 10700 | 12300 | 12900 | 13300 | 11200 | 16200 |
| Antimony | 0.96 | 0.25J | <0.19U | <0.22U | 0.34 | 0.46 | 0.3 | 0.28 | <0.21U | <0.19U | <0.23U | <0.18U | 0.39 |
| Arsenic | 15.4 | 28.4* | 8.5 | 16.9* | 5.3 | 4.2 | 16.3* | 5.2 | 16.7* | 13.7 | 13.5 | 9.3 | 11.2 |
| Barium | 88.4 | 65 | 90.4* | 60.9 | 75.9 | 64.8 | 61.7 | 61.6 | 52.1 | 49.3 | 74.4 | 35.2 | 57.9 |
| Beryllium | 0.88 | 0.6 | 0.56 | 0.58 | 0.48 | 0.4 | 0.86 | 0.39 | 0.43 | 0.43 | 0.7 | 0.39 | 0.61 |
| Cadmium | 0 | 0.49* | <0.1U | 0.15* | 0.16* | 0.11* | <0.11U | 0.11* | <0.11U | <0.1U | 0.13* | <0.093U | <0.12U |
| Calcium | 15800 | 1530 | 1190 | 4040 | 538 | 512 | 1050 | 876 | 1290 | 1320 | 1870 | 195 | 572 |
| Chromium | 17.4 | 9.6 | 13.6 | 17.3 | 11.3 | 14.2 | 24.8* | 11.6 | 15.7 | 16 | 22.3* | 12.4 | 23.7 * |
| Chromium, hexavalent | | 1* | NR | NR | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Cobalt | 10.4 | 9.2 | 10.7* | 10.6* | 6.2 | 6.7 | 11.6* | 6.1 | 6.2 | 6.1 | 12* | 5.6 | 8.4 |
| Copper | 17.7 | 12.2 | 9.5 | 18.7* | 4.2 | 6.4 | 23.9* | 8.3 | 13.6 | 13.8 | 22.9* | 11.3 | 21 * |
| Cyanide | 0 | 1.2* | <0.28U | 0.45* | <0.24U | 0.31* | <0.16U | 0.42* | <0.27U | <0.22U | <0.27U | <0.23U | <0.26U |
| Iron | 23100 | 21800 | 15300 | 24000* | 13800 | 16900 | 44000* | 13400 | 22200 | 24400* | 30800* | 18400 | 34100 * |
| Lead | 26.1 | 30.8* | 22.6 | 18.1 | 19.1 | 25 | 19.6 | 18.6 | 12.3 | 20.4 | 22.5 | 14.5 | 20.1 |
| Magnesium | 3030 | 1340 | 1880 | 3630* | 1600 | 2540 | 4310* | 1550 | 2300 | 2380 | 4080* | 1770 | 4080* |
| Manganese | 1450 | 453 | 241 | 397 | 186 | 107 | 231 | 125 | 115 | 129 | 267 | 126 | 208 |
| Mercury | 0.036 | 0.025J | 0.066* | 0.029 | 0.041* | 0.046* | 0.03 | 0.035 | 0.032 | 0.032 | 0.029 | 0.033 | 0.027 |
| Nickel | 21.1 | 14.8 | 12.9 | 24.7* | 9.7 | 10.7 | 26.6* | 8.6 | 13.9 | 13.9 | 27.2* | 13.9 | 23.9 * |
| Potassium | 927 | 704 | 959* | 1490* | 727 | 1290* | 2220* | 568 | 1550* | 1510* | 1750* | 696 | 1610* |
| Selenium | 1.4 | <0.18U | 0.83 | <0.59U | 0.85 | 0.48 | 1.2 | 0.86 | <0.52U | 1.4* | 0.72 | 0.67 | <0.57U |
| Silver | 0 | <0.19U | 0.45* | 0.41* | <0.14U | 0.32* | <0.22U | <0.18U | <0.21U | <0.19U | <0.24U | <0.18U | <0.24U |
| Sodium | 123 | 42.9 | 136* | 83.4 | 104 | 63.5 | 143* | 95.1 | 123 | 116 | 159* | 55.6 | 84.5 |
| Thallium | 0 | <0.22U | 0.25* | <0.22U | <0.21U | <0.22U | <0.2U | <0.44U | <1.1U | <0.19U | <1.1U | <0.36U | <0.52U |
| Vanadium | 31.1 | 14.4 | 18.9 | 20.5 | 17.8 | 35.3* | 28.3 | 21 | 22.7 | 23.1 | 22.4 | 17.7 | 27 |
| Zinc | 61.8 | 121* | 81.1* | 72.1* | 63.1* | 50.9 | 72.7* | 52.3 | 50.4 | 47.7 | 77* | 53 | 68.3 * |
| | | | | • | • | Anions (mg | | | • | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | None | <0.96U | 1.4* | <0.97U | <0.9U | <0.92U | <0.89U | <0.96U | <0.9U | <0.89U | <1U | <0.75U | <0.96U |
| | 1 | 1 | r | | | Explosives (n | | 1 | | 1 | r | r | |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | 0.66* | 1.5* | NR | NR | NR |

 Table 4-4. Analytes Detected in Phase II RI Surface Soil Samples

^aOnly detected site-related contaminants are presented in the table. ^bBackground concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

ft = Feet. ID = Identification.

ID = Identification. J = Estimated value less than reporting limits. mg/kg = Milligrams per kilogram. NR = Not reported/not analyzed. RI = Remedial investigation. RVAAP = Ravenna Army Ammunition Plant. TAL = Target analyte list. UL = Not detected

U = Not detected.

* = Result exceeds background criteria or no background criteria was available.

| A | | Declaration Arres | Dealers |
|--|------------|------------------------------|------------------------------|------------------------------|------------------------------|------------------------------|------------------------------|--------------|
| Area Station | | Backwater Area ULCPsd-004 | Backwater Area ULCPsd-004 | Backwater Area ULCPsd-005 | Backwater Area ULCPsd-006 | Backwater Area ULCPsd-007 | Backwater Area ULCPsd-008 | Backw ULC |
| Station | | ULCPsd-004-0001- | ULCPsd-004-0001- | ULCPsd-005-0001- | ULCPsd-006-0001- | ULCPsd-007-0001- | ULCPsd-008-0001- | ULCPsd |
| Sample ID | | FD | SD | SD | SD | SD | SD | ULCESU |
| Date | | 08/22/01 | 08/22/01 | 08/22/01 | 08/22/01 | 08/21/01 | 08/21/01 | 08/ |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.8 | 0.0 - 0.5 | 0.0 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., T |
| Analyte (mg/kg) | Criteria | Explosives | Explosives | Explosives | analytes | Explosives | Explosives | Exp |
| | | * | L* | | Metals (mg/kg) | • • | | * |
| Aluminum | 13900 | 14700* | 15900* | 13000 | 12800 | 13200 | 9010 | 16200* |
| Antimony | 0 | 0.47J* | <0.38U | 0.85* | 0.65* | 0.76* | <0.28U | 1.4* |
| Arsenic | 19.5 | 15.5 | 11.6 | 6.5 | 10.5 | 14.9 | 4.4 | 13.9 |
| Barium | 123 | 122 | 132* | 147* | 103 | 113 | 101 | 104 |
| Beryllium | 0.38 | 0.97* | 1.1* | 0.96* | 0.81* | 1* | 0.61J* | 1* |
| Cadmium | 0 | 2* | 1.4* | 1.9* | 1.2* | 2.2* | 0.57* | 0.97* |
| Calcium | 5510 | 3290 | 2880 | 2420 | 3540 | 3830 | 1350 | 2490 |
| Chromium | 18.1 | 24.1* | 25.1* | 19.3* | 21.1* | 41.4* | 12.6 | 56.5* |
| Chromium, hexavalent | None | <1U | <1U | <1U | <1U | <1U | <1U | <1U |
| Cobalt | 9.1 | 23.3* | 17.8* | 13.9* | 17.1* | 20.1* | 10.8* | 20.7* |
| Copper | 27.6 | 36.1* | 36.7* | 24.6 | 21.9 | 61.1* | 11.7 | 50.5* |
| Cyanide | 0 | <0.46U | <0.49U | <0.49U | <0.38U | 0.12* | 0.1* | 0.55* |
| Iron | 28200 | 27300 | 27100 | 24800 | 23600 | 25400 | 14400 | 26700 |
| Lead | 27.4 | 40.8* | 35* | 42.6* | 32.5* | 40* | 19.2 | 57.9* |
| Magnesium | 2760 | 2290 | 2400 | 2090 | 2240 | 2030 | 1530 | 2570 |
| Manganese | 1950 | 516 | 462 | 348 | 711 | 964 | 343 | 721 |
| Mercury | 0.059 | 0.098* | 0.1* | 0.11* | 0.068J* | 0.089J* | 0.042J | 0.085J* |
| Nickel | 17.7 | 26.3* | 24.9* | 24.4* | 21.6* | 29.4* | 15.8 | 24.8* |
| Potassium | 1950 | 1590 | 1640 | 1230 | 1430 | 1150 | 846 | 1290 |
| Selenium | 1.7 | 0.47 | 0.39 | 0.35J | <0.34U | 0.44J | <0.24U | <0.34U |
| Silver | 0 | 10.9* | 10.9* | 1.4* | 5.6* | 23* | 1.1* | 21.7* |
| Sodium | 112 | 156J* | 145J* | 98.8J | 228* | 139* | 107 | 239* |
| Thallium | 0.89 | <0.4U | <0.38U | <0.39U | <0.4U | <0.46U | <0.28U | <0.4U |
| Vanadium | 26.1 | 22.9 | 24.8 | 21.8 | 19.2 | 15.9 | 15.2 | 25.7 |
| Zinc | 532 | 298 | 234 | 222 | 199 | 418 | 114 | 212 |
| | | T | | 1 | Anions (mg/kg) | T | T | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N | N) None | <1.5U | <1.4U | <1.5U | <1.5U | <1.8U | <1.1U | <1.8U |
| | | | | 1 | Explosives (mg/kg) | | | |
| 2,4,6-Trinitrotoluene | None | <0.033U | <0.034U | 0.32* | 0.18* | <0.033U | <0.034U | 0.17* |
| 2,6-Dinitrotoluene | None | <0.047U | <0.047U | 0.16J* | <0.047U | <0.047U | <0.048U | <0.048U |
| Nitrocellulose | None | NR | NR | NR | 3* | NR | NR | NR |
| Anthracene | None | NR | NR | NR | <0.17U | NR | NR | NR |
| Benz(a)anthracene | None | NR | NR | NR | <0.13U | NR | NR | NR |
| Benzo(a)pyrene | None | NR | NR | NR | 0.23J* | NR | NR | NR |
| Benzo(b)fluoranthene | None | NR | NR | NR | <0.26U | NR | NR | NR |
| Benzo(ghi)perylene | None | NR | NR | NR | <0.36U | NR | NR | NR |
| Benzo(k)fluoranthene | None | NR | NR | NR | <0.27U | NR | NR | NR |
| Chrysene | None | NR | NR | NR | 0.24J* | NR | NR | NR |
| Fluoranthene | None | NR | NR | NR | <0.22U | NR | NR | NR |
| Indeno(1,2,3-cd)pyrene | None | NR | NR | NR | <0.26U | NR | NR | NR |
| Phenanthrene | None | NR | NR | NR | <0.16U | NR | NR | NR |
| Desus a s | I NI | | 1 N 10 | | 1 <0.2411 | 1 5 10 | | ND |

NR

Table 4–5. Analytes Detected in Phase II RI Sediment Samples

NR

None

NR

Pyrene

NR

NR

NR

<0.34U

| Declaration Arres | Deckson Arres | Declaration Arres |
|-------------------------|-----------------------------------|------------------------|
| Backwater Area | Backwater Area | Backwater Area |
| ULCPsd-009 | ULCPsd-011 | ULCPsd-012 |
| ULCPsd-009-0001- SD | ULCPsd-011-0001- SD | ULCPsd-012-0001- SD |
| 08/21/01 | 08/20/01 | 08/20/01 |
| 0.0 - 0.9 | 0.0 - 0.8 | 0.0 - 1.8 |
| Misc., TAL Metals, | RVAAP Full-suite | Misc., TAL Metals, |
| Explosives | analytes | Explosives |
| | | |
| 16200* | 12200 | 10000 |
| 1.4* | 0.24J* | 0.66* |
| 13.9 | 13.8 | 6.7 |
| 104 | 69.6 | 95.4 |
| 1* | 0.63* | 0.63* |
| 0.97* | 0.11J* | 0.52* |
| 2490 | 1180 | 1450 |
| 56.5* | 30.4* | 13.9 |
| <1U | <1U | <1U |
| 20.7* | 11.5* | 12* |
| 50.5* | 23.7 | 14 |
| 0.55* | <0.36U | <0.45U |
| 26700 | 23500 | 15600 |
| 57.9* | 21.7 | 19.7 |
| 2570 | 2560 | 1700 |
| 721 | 296 | 570 |
| 0.085J* | 0.04J | 0.068* |
| 24.8* | 21.8* | 16.3 |
| 1290 | 1370 | 731 |
| <0.34U | <0.18U | 0.26J |
| 21.7* | <0.21U | 1.8* |
| 239* | 91.4J | 87.1J |
| <0.4U | <0.21U | <0.28U |
| 25.7 | 20.4 | 15.4 |
| 212 | 79.1 | 101 |
| .1.017 | .411 | .1.017 |
| <1.8U | <1U | <1.2U |
| A 17* | 0.049J* | <0.024U |
| 0.17* <0.048U | <pre>0.049J^ <0.047U</pre> | <0.034U <0.048U |
| <u><0.0480</u> NR | 2.5* | <0.0480 NR |
| NR | <0.11U | NR |
| NR | <0.077U | NR |
| NR | <0.0770 <0.084U | NR |
| NR | <0.0840 <0.16U | NR |
| NR | <0.100 <0.22U | NR |
| NR | <0.17U | NR |
| NR | <0.058U | NR |
| NR | <0.0580 <0.14U | NR |
| NR | <0.140 | NR |
| NR | <0.100 <0.1U | NR |
| NR | <0.10 <0.21U | NR |
| 1 111 | -0.210 | 1111 |

| Area | | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area |
|---------------------|------------|--------------------|--------------------|--------------------|-------------------------|--------------------|--------------------|--------------------|-------------------------|--------------------|
| Station | | ULCPsd-004 | ULCPsd-004 | ULCPsd-005 | ULCPsd-006 | ULCPsd-007 | ULCPsd-008 | ULCPsd-009 | ULCPsd-011 | ULCPsd-012 |
| | | ULCPsd-004-0001- | ULCPsd-004-0001- | ULCPsd-005-0001- | ULCPsd-006-0001- | ULCPsd-007-0001- | ULCPsd-008-0001- | ULCPsd-009-0001- | ULCPsd-011-0001- | ULCPsd-012-0001- |
| Sample ID | | FD | SD | SD | SD | SD | SD | SD | SD | SD |
| Date | | 08/22/01 | 08/22/01 | 08/22/01 | 08/22/01 | 08/21/01 | 08/21/01 | 08/21/01 | 08/20/01 | 08/20/01 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.8 | 0.0 - 0.5 | 0.0 - 0.9 | 0.0 - 0.8 | 0.0 - 1.8 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | Misc., TAL Metals, |
| Analyte (mg/kg) | Criteria | Explosives | Explosives | Explosives | analytes | Explosives | Explosives | Explosives | analytes | Explosives |
| | | | | | Pesticides/PCBs | | | | | |
| PCB-1254 | None | NR | NR | NR | 0.047* | NR | NR | NR | 0.046* | NR |
| | | | | | VOCs | | • | | | · |
| 2-Butanone | None | NR | NR | NR | <0.01U | NR | NR | NR | <0.006U | NR |
| Acetone | None | NR | NR | NR | 0.13* | NR | NR | NR | 0.02* | NR |
| Methylene chloride | None | NR | NR | NR | 0.048* | NR | NR | NR | <0.003U | NR |

 Table 4–5. Analytes Detected in Phase II RI Sediment Samples (continued)

| | | 1 | | Table 4–5. Analytes De | | | , | F | 1 | |
|---|------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|------------------------|
| Area | | Backwater Area | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond |
| Station | | ULCPsd-013 | ULCPsd-013 | ULCPsd-014 | ULCPsd-015 | ULCPsd-015 | ULCPsd-016 | ULCPsd-021 | ULCPsd-021 | ULCPsd-022 |
| Sample ID | | ULCPsd-013-0001- FD | ULCPsd-013-0001- SD | ULCPsd-014-0001- SD | ULCPsd-015-0001- SD | ULCPsd-015-0002- SD | ULCPsd-016-0001- SD | ULCPsd-021-0001- SD | ULCPsd-021-0002- SD | ULCPsd-022-0001- SD |
| Date | | 08/20/01 | 08/20/01 | 07/25/01 | 08/17/01 | 09/27/01 | 08/17/01 | 08/15/01 | 09/27/01 | 08/15/01 |
| Depth (ft) | 1 | 0.0 - 1.8 | 0.0 - 1.8 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 1.0 | 0.0 - 1.5 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.8 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | | Misc., TAL Metals, | Misc., TAL Metals, | | Misc., TAL Metals, |
| Analyte (mg/kg) | Criteria | Explosives | Explosives | Explosives | Explosives | SVOCs, VOCs | Explosives | Explosives | SVOCs, VOCs | Explosives |
| | | | | | Metals | | | | | |
| Aluminum | 13900 | 13900 | 18600* | 11500 | 10300 | NR | 8000 | 11900 | NR | 14800* |
| Antimony | 0 | <0.3U | <0.4U | 0.25J* | 0.31J* | NR | 0.3J* | <0.37U | NR | <0.31U |
| Arsenic | 19.5 | 4.8 | 12.3 | 3.5 | 11.3 | NR | 11.1 | 20.2* | NR | 5.1 |
| Barium | 123 | 105 | 142* | 58.3 | 57.3 | NR | 53.8 | 77.3 | NR | 96.7 |
| Beryllium | 0.38 | 0.72* | 1* | 0.41J* | 0.94* | NR | 0.34J | 0.81* | NR | 0.66* |
| Cadmium | 0 | 0.45* | 0.67* | 0.57* | 0.36* | NR | 0.23J* | 1* | NR | 0.43* |
| Calcium | 5510 | 1940 | 3000 | 2130 | 15400* | NR | 1620 | 2470 | NR | 1480 |
| Chromium | 18.1 | 21* | 27* | 15 | 19.3* | NR | 10.1 | 122* | NR | 33.6* |
| Chromium, hexavalent | None | <1U | <1U | <1U | <1U | NR | <1U | 5.7* | NR | 5* |
| Cobalt | 9.1 | 12.2* | 14.2* | 5.4 | 7.3 | NR | 6.1 | 10.6* | NR | 10.7* |
| Copper | 27.6 | 19.6 | 26.9 | 12.7 | 25.2 | NR | 10.9 | 115* | NR | 31.8* |
| Cyanide | 0 | <0.46U | <0.59U | <0.28U | <0.38U | NR | <0.42U | <0.48U | NR | <0.31U |
| Iron | 28200 | 21000 | 27100 | 12700 | 15000 | NR | 12800 | 22200 | NR | 20000 |
| Lead | 27.4 | 20.1 | 32.7* | 21 | 14.2 | NR | 20.5 | 35.4* | NR | 18.7 |
| Magnesium | 2760 | 2180 | 2870* | 1710 | 3830* | NR | 1170 | 2220 | NR | 2590 |
| Manganese | 1950 | 383 | 480 | 179 | 563 | NR | 197 | 367 | NR | 307 |
| Mercury | 0.059 | 0.046J | 0.073J* | 0.045J | 0.035J | NR | 0.031J | 0.093* | NR | 0.041J |
| Nickel | 17.7 | 19.1* | 24.8* | 12.8 | 12.2 | NR | 9.4 | 22* | NR | 19.6* |
| Potassium | 1950 | 1250 | 1460 | 1080 | 767 | NR | 682 | 1190 | NR | 1520 |
| Selenium | 1.7 | 0.29J | 0.37J | 0.95 | <0.18U | NR | <0.21U | 0.45 | NR | <0.27U |
| Silver | 0 | 2.3* | 3.1* | 0.54J* | 2* | NR | 0.51J* | 2.4* | NR | <0.27U |
| Sodium | 112 | 119J* | 173J* | 55.2 | 237* | NR | 114J* | 148J* | NR | 190* |
| Thallium | 0.89 | <0.3U | <0.4U | <0.21U | <0.21U | NR | <0.25U | <0.37U | NR | <0.31U |
| Vanadium | 26.1 | 23.7 | 30* | 23 | 11.7 | NR | 15.8 | 18.9 | NR | 26.6* |
| Zinc | 532 | 121 | 170 | 74.8 | 117 | NR | 79.2 | 275 | NR | 108 |
| | 1 | | | · | Anions | I | 1 | 1 | I | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | None | <1.2U | <1.5U | <1U | <1.1U | NR | 2.2* | <1.7U | NR | <1.1U |
| 2 4 <i>c</i> m : 1 | 3.7 | 0.02.411 | 0.00714 | -0.02211 | Explosives | | -0.02 MI | 0.02477 |) ID | -0.02.4T.I |
| 2,4,6-Trinitrotoluene | None | <0.034U | 0.086J* | <0.033U | <0.034U | NR | <0.034U | <0.034U | NR | <0.034U |
| 2,6-Dinitrotoluene | None | <0.047U | <0.048U | <0.047U | <0.048U | NR | <0.047U | <0.047U | NR | <0.047U |
| Nitrocellulose | None | NR |
| Anthracene | None | NR | NR | NR | NR | 0.22J* | NR | NR | <0.18U | NR |
| Benz(a)anthracene | None | NR | NR | NR | NR | 0.83J* | NR | NR | <0.13U | NR |
| Benzo(a)pyrene | None | NR | NR | NR | NR | 0.89J* | NR | NR | <0.14U | NR |
| Benzo(b)fluoranthene | None | NR | NR | NR | NR | 0.86J* | NR | NR | <0.26U | NR |
| Benzo(ghi)perylene | None | NR | NR | NR | NR | 0.49J* | NR | NR | <0.37U | NR |
| Benzo(k)fluoranthene | None | NR | NR | NR | NR | 0.8J* | NR | NR | <0.28U | NR |
| Chrysene | None | NR | NR | NR | NR | 0.93J* | NR | NR | <0.096U | NR |
| Fluoranthene | None | NR | NR | NR | NR | 1.8* | NR | NR | <0.23U | NR |
| Indeno(1,2,3-cd)pyrene | None | NR | NR | NR | NR | 0.48J* | NR | NR | <0.27U | NR |
| Phenanthrene | None | NR | NR | NR | NR | 0.91J* | NR | NR | <0.17U | NR |
| Pyrene | None | NR | NR | NR | NR | 1.9* | NR | NR | <0.34U | NR |

| Table 4–5. Analytes | Detected i | in Phase II l | RI Sediment | Samples | (continued) |
|---------------------|------------|---------------|--------------------|---------|-------------|
| | | | | | |

| Area | | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond |
|---------------------|------------|--------------------|--------------------|--------------------|--------------------|------------------|--------------------|--------------------|------------------|--------------------|
| Station | | ULCPsd-013 | ULCPsd-013 | ULCPsd-014 | ULCPsd-015 | ULCPsd-015 | ULCPsd-016 | ULCPsd-021 | ULCPsd-021 | ULCPsd-022 |
| | | ULCPsd-013-0001- | ULCPsd-013-0001- | ULCPsd-014-0001- | ULCPsd-015-0001- | ULCPsd-015-0002- | ULCPsd-016-0001- | ULCPsd-021-0001- | ULCPsd-021-0002- | ULCPsd-022-0001- |
| Sample ID | | FD | SD | SD | SD | SD | SD | SD | SD | SD |
| Date | | 08/20/01 | 08/20/01 | 07/25/01 | 08/17/01 | 09/27/01 | 08/17/01 | 08/15/01 | 09/27/01 | 08/15/01 |
| Depth (ft) | | 0.0 - 1.8 | 0.0 - 1.8 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 1.0 | 0.0 - 1.5 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.8 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | | Misc., TAL Metals, | Misc., TAL Metals, | | Misc., TAL Metals, |
| Analyte (mg/kg) | Criteria | Explosives | Explosives | Explosives | Explosives | SVOCs, VOCs | Explosives | Explosives | SVOCs, VOCs | Explosives |
| | | | | | Pesticides/PCBs | | | | | |
| PCB-1254 | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | VOCs | | | | | |
| 2-Butanone | None | NR | NR | NR | NR | 0.035* | NR | NR | 0.025* | NR |
| Acetone | None | NR | NR | NR | NR | 0.14* | NR | NR | 0.076* | NR |
| Methylene chloride | None | NR | NR | NR | NR | <0.005U | NR | NR | <0.004U | NR |

 Table 4–5. Analytes Detected in Phase II RI Sediment Samples (continued)

| | | | | | | | , | | | |
|-----------------------------|------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|------------------------------|-----------------------|
| Area | - | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond |
| Station | - | ULCPsd-023 | ULCPsd-024 | ULCPsd-025 | ULCPsd-026 | ULCPsd-017 | ULCPsd-018 | ULCPsd-019 | ULCPsd-020 | ULCPsd-020 |
| | | ULCPsd-023-0001- | ULCPsd-024-0001- | ULCPsd-025-0001- | ULCPsd-026-0001- | ULCPsd-017-0001- | ULCPsd-018-0001- | ULCPsd-019-0001- | ULCPsd-020-0001- | ULCPsd-020-0001- |
| Sample ID | - | SD 08/15/01 | SD 09/14/01 | SD 07/25/01 | SD 09/14/01 | SD 09/17/01 | SD | SD | FD | SD |
| Date | | 0.0 - 0.8 | 08/14/01 0.0 - 0.5 | 0.0 - 0.5 | 08/14/01 0.0 - 1.0 | 08/17/01 0.0 - 1.0 | 08/16/01 0.0 - 0.5 | 08/16/01 0.0 - 0.5 | 08/16/01 0.0 - 0.8 | 08/16/01 0.0 - 0.8 |
| Depth (ft) | Deele successed | | | | | | | | | |
| Parameters Analyzed | Background Criteria | Misc., TAL Metals, Explosives | RVAAP Full-suite analytes | RVAAP Full-suite |
| Analyte (mg/kg) | Criteria | Explosives | Exprosives | Explosives | Metals | Explosives | Explosives | Explosives | analytes | analytes |
| Aluminum | 13900 | 12900 | 11600 | 10000 | 16400* | 21400* | 15100* | 14000* | 13000 | 13400 |
| Antimony | 13900 | <0.46U | <0.22U | <0.22U | <0.2U | 1.3* | 0.21J* | 0.26* | <0.2U | 0.23J* |
| Arsenic | 19.5 | 10.9 | 8.5 | 17.6 | 34.3* | 17.5 | 5.1 | 8.8 | 11 | 11.2 |
| Barium | 19.5 | 80.2 | 79.4 | 70.8 | 51.3 | 17.5 147* | 58.7 | 84 | 102 | 83.2 |
| Beryllium | 0.38 | 0.76* | 0.63* | 0.67* | 0.55 * | 1.3* | 0.44 * | 0.8* | 0.69* | 0.67* |
| Cadmium | 0.58 | 0.98* | <0.1U | 0.18* | <0.1U | 2.2* | <0.094U | <0.11U | 0.09 [±] | 0.07# 0.15J* |
| Calcium | 5510 | 2370 | 7360* | 2050 | 699 | 5440 | <0.0940 771 | 1290 | 2310 | 1950 |
| Chromium | 18.1 | 150* | 15.8 | 11.9 | 20 * | 87.5* | 19.9* | 27.8* | 50 * | 43.6* |
| Chromium, hexavalent | None | <10 | 13.8 <1U | <1U | <1U | <1U | 19.9* | <1U | <1U | 43.0" <1U |
| Cobalt | 9.1 | 12.1* | 9.3* | 21.3* | 9.3* | 19.3* | 5.9 | 12* | 11* | 10.9* |
| | 27.6 | 149* | 14.2 | 12.4 | 14.5 | 19.5 | 14.1 | 29.6* | 36* | 31.7* |
| Copper Cyanide | 0 | <0.67U | 0.4* | <0.21U | <0.23U | <1.2U | <0.23U | <0.26U | <0.27U | <0.28U |
| Iron | 28200 | 24500 | 22700 | 32200* | 24600 | 36400* | 18300 | 30600* | 23200 | 24400 |
| Lead | 27.4 | 43.8* | 60.4 * | 12.6 | 16.3 | 40.9* | 13.1 | 14.5 | 16.9 | 17.5 |
| Magnesium | 2760 | 2520 | 2510 | 3240* | 2790* | 3550* | 2370 | 3690* | 2820* | 3040* |
| Manganese | 1950 | 423 | 711 | 966 | 217 | 705 | 138 | 310 | 358 | 326 |
| Mercury | 0.059 | 0.078J* | 0.033J | 0.016J | 0.022J | 0.11J* | 0.029J | 0.014J | 0.023J | 0.028J |
| Nickel | 17.7 | 24.4* | 15.6 | 19.3* | 16.6 | 35.4* | 13.1 | 30.7* | 22.6* | 23.1* |
| Potassium | 1950 | 1320 | 1490 | 1650 | 1860 | 2100* | 1060 | 2070* | 1640 | 1750 |
| Selenium | 1.7 | <0.39U | <0.19U | 1.5 | <0.17U | 0.91 | 0.22 | <0.15U | <0.17U | <0.18U |
| Silver | 0 | 2.1* | <0.1) C | 0.23J* | <0.170 <0.2U | 11* | <0.18U | <0.21U | 0.77* | 0.55* |
| Sodium | 112 | 138J* | 87J | 57.4 | 146* | 340* | 101 | 82.7J | 94.1J | 114* |
| Thallium | 0.89 | <0.46U | <0.22U | <0.22U | <0.2U | <0.73U | <0.2U | 0.24 | <0.2U | <0.22U |
| Vanadium | 26.1 | 21.8 | 22.8 | 19.4 | 32.2* | 31.2* | 33.4* | 23.4 | 24 | 23.3 |
| Zinc | 532 | 251 | 73.9 | 44.6 | 58.3 | 479 | 71.3 | 89.7 | 115 | 99.9 |
| | 552 | 201 | 15.7 | | Anions | 777 | /1.5 | 0).1 | 115 |)).) |
| Nitrate/Nitrite (NO3/NO2-N) | None | <1.9U | <0.93U | <0.87U | <0.91U | <3.2U | <0.86U | <0.9U | <1.2U | <1.1U |
| | 1.0110 | 100 | | 0.070 | Explosives | 0.20 | 0.000 | 0.50 | | |
| 2,4,6-Trinitrotoluene | None | <0.034U | <0.033U | <0.034U | <0.034U | 0.087J* | <0.034U | <0.034U | <0.033U | <0.033U |
| 2,6-Dinitrotoluene | None | <0.037U | <0.046U | <0.03 TC | <0.048U | <0.048U | <0.0310 | <0.047U | <0.047U | <0.047U |
| Nitrocellulose | None | NR | 2.1* | 2* |
| | 1 | 1 | 1 | 1 | SVOCs | 1 | 1 | 1 | 1 | 1 |
| Anthracene | None | NR | <0.13U | <0.12U |
| Benz(a)anthracene | None | NR | <0.096U | <0.09U |
| Benzo(a)pyrene | None | NR | <0.11U | <0.098U |
| Benzo(b)fluoranthene | None | NR | <0.2U | <0.18U |
| Benzo(ghi)perylene | None | NR | <0.28U | <0.26U |
| Benzo(k)fluoranthene | None | NR | <0.21U | <0.19U |
| Chrysene | None | NR | <0.073U | <0.068U |
| Fluoranthene | None | NR | <0.17U | <0.16U |
| Indeno(1,2,3-cd)pyrene | None | NR | <0.2U | <0.19U |
| Phenanthrene | None | NR | <0.13U | <0.12U |
| Pyrene | None | NR | <0.26U | <0.24U |

 Table 4–5. Analytes Detected in Phase II RI Sediment Samples (continued)

| Area | | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond |
|---------------------|------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|-------------------------|-------------------------|
| Station | | ULCPsd-023 | ULCPsd-024 | ULCPsd-025 | ULCPsd-026 | ULCPsd-017 | ULCPsd-018 | ULCPsd-019 | ULCPsd-020 | ULCPsd-020 |
| | | ULCPsd-023-0001- | ULCPsd-024-0001- | ULCPsd-025-0001- | ULCPsd-026-0001- | ULCPsd-017-0001- | ULCPsd-018-0001- | ULCPsd-019-0001- | ULCPsd-020-0001- | ULCPsd-020-0001- |
| Sample ID | | SD | FD | SD |
| Date | | 08/15/01 | 08/14/01 | 07/25/01 | 08/14/01 | 08/17/01 | 08/16/01 | 08/16/01 | 08/16/01 | 08/16/01 |
| Depth (ft) | | 0.0 - 0.8 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 0.8 | 0.0 - 0.8 |
| Parameters Analyzed | Background | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite |
| Analyte (mg/kg) | Criteria | Explosives | analytes | analytes |
| | | | | | Pesticides/PCBs | | | | | |
| PCB-1254 | None | NR | 0.031* | 0.032* |
| | | | | | VOCs | | | | | |
| 2-Butanone | None | NR | <0.008U | 0.022* |
| Acetone | None | NR | 0.011* | 0.088* |
| Methylene chloride | None | NR | ND | NR | NR | NR | ND | NR | <0.003U | <0.003U |

Table 4-5. Analytes Detected in Phase II RI Sediment Samples (continued)

^aOnly detected site-related contaminants are presented in the table. ^bBackground concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

ft = Feet.

ID = Identification.

J = Estimated value less than reporting limits.

J = Estimated value less than reporting limits. mg/kg = Milligrams per kilogram. NR = Not reported/not analyzed. PCB = Polychlorinated biphenyl. RI = Remedial Investigation RVAAP = Ravenna Army Ammunition Plant. SVOC = Semi-volatile organic compound. TAL = Target analyte list. VOC = Volatile organic compound.

VOC = Volatile organic compound.

U = Not detected.

* = Result exceeds background criteria or no background criteria was available.

| Area | | Backwater Area | Backwater Area | Backwater Area | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond |
|--|------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|-----------------------|
| Station | | ULCPsw-001 | ULCPsw-002 | ULCPsw-002 | ULCPsw-005 | ULCPsw-006 | ULCPsw-007 | ULCPsw-003 | ULCPsw-004 |
| Sample ID | | ULCPsw-001-0001-SW | ULCPsw-002-0001-FD | ULCPsw-002-0001-SW | ULCPsw-005-0001-SW | ULCPsw-006-0001-SW | ULCPsw-007-0001-SW | ULCPsw-003-0001-SW | ULCPsw-004-0001-SW |
| Date | | 07/26/01 | 07/25/01 | 07/25/01 | 07/24/01 | 07/24/01 | 07/24/01 | 07/24/01 | 07/24/01 |
| | | | RVAAP Full-suite | RVAAP Full-suite | | | | | |
| | | TAL Metals, Explosives, | analytes + Sulfate and | analytes + Sulfate and | TAL Metals, Explosives, | TAL Metals, Explosives, | TAL Metals, Explosives, | TAL Metals, Explosives, | TAL Metals, Explosive |
| Parameters Analyzed | | Sulfate, Sulfide | Sulfide | Sulfide | Sulfate, Sulfide | Sulfate, Sulfide | Sulfate, Sulfide | Sulfate, Sulfide | Sulfate, Sulfide |
| Filtered? | Background | | | | | | | | |
| Analyte (mg/L) | Criteria | Total | Total |
| (ing, 2) | 0.1.01.10 | | | | Metals (mg/L) | | | | 1000 |
| Aluminum | 3.37 | 1.45 | 0.456 | 2.52 | 0.114J | 0.161J | 0.377 | 0.276 | 0.107J |
| Arsenic | 0.0032 | 0.0104* | 0.0038* | 0.0043* | <0.0018U | 0.0024 | 0.002 | <0.0018U | 0.0025 |
| Barium | 0.0475 | 0.141* | 0.159* | 0.382* | 0.0138 | 0.0118 | 0.0228 | 0.0226 | 0.0148 |
| Calcium | 41.4 | 38.3 | 44.2* | 67.7* | 19.8 | 20.6 | 38.7 | 22.3 | 21.8 |
| Chromium | 0 | 0.004J* | 0.0031J* | 0.0067J* | <0.003U | <0.003U | <0.003U | <0.003U | <0.003U |
| Cobalt | 0 | 0.0107* | <0.002U | 0.0054* | <0.002U | <0.002U | <0.002U | <0.002U | <0.002U |
| Copper | 0.0079 | <0.0028U | <0.0028U | 0.0088J* | <0.0028U | <0.0028U | <0.0028U | <0.0028U | <0.0028U |
| Iron | 2.56 | 12.4* | 4.53* | 16.3* | 0.348 | 0.294 | 0.443 | 1.18 | 0.419 |
| Magnesium | 10.8 | 6.67 | 8.97 | 11.2* | 4.05 | 3.98 | 6.71 | 4.41 | 4.21 |
| Manganese | 0.391 | 8.59* | 7.36* | 15.8* | 0.157 | 0.124 | 0.654* | 0.528* | 0.171 |
| Nickel | 0 | 0.0042J* | <0.0026U | 0.0067J* | <0.0026U | <0.0026U | <0.0026U | <0.0026U | <0.0026U |
| Potassium | 3.17 | 5.82* | 10.1* | 15.7* | 0.497J | 0.751 | 1.51 | 0.894 | 0.921 |
| Selenium | 0 | <0.0049U | 0.0075* | 0.0104* | <0.0049U | <0.0049U | <0.0049U | <0.0049U | <0.0049U |
| Silver | 0 | <0.0013U | <0.0013U | 0.0017J* | <0.0013U | <0.0013U | <0.0013U | <0.0013U | <0.0013U |
| Sodium | 21.3 | 7.55 | 9.99 | 12.9 | 3.61 | 3.28 | 4.4 | 3.68 | 3.46 |
| Vanadium | 0 | 0.005* | <0.0023U | 0.0049J* | <0.0023U | 0.003J* | 0.0024J* | <0.0023U | <0.0023U |
| Zinc | 0.042 | 0.0312 | 0.0329 | 0.109* | 0.0027J | 0.0139 | 0.0125 | 0.0115 | 0.0056J |
| ~ 10 | | | | | Anions (mg/L) | | | | |
| Sulfate | None | 29* | <2.5U | <2.5U | 14* | 16* | 19* | 13* | 13* |
| Sulfide | None | <0.5U | 1.3* | 2.3* | <0.5U | <0.5U | 0.7* | 0.7* | 0.7* |
| • | N | | 0.005* | 0.002* | VOCs (mg/L) | | | | |
| Acetone | None | NR | 0.005* | 0.003* | NR | NR | NR | NR | NR |
| Chloromethane | None | NR | 0.001* | 0.0007* | NR | NR | NR | NR | NR |
| Toluene ^a Only detected site-related | None | NR | 0.001* | 0.001* | NR | NR | NR | NR | NR |

Table 4-6. Analytes Detected in Phase II RI Surface Water Samples

^aOnly detected site-related contaminants are presented in the table. ^bBackground concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

ID = Identification.

J = Estimated value less than reporting limits.

J = Estimated Value less than reporting limits.
mg/L = Milligrams per liter.
NR = Not reported/not analyzed.
RI = Remedial investigation.
RVAAP = Ravenna Army Ammunition Plant.
TAL = Target analyte list.
U = Not detected.

VOC = Volatile organic compound. * = Result exceeds background criteria or no background criteria was available.

Table 4–7. Facility-wide Biological and Water Quality Study Sampling Locations

| FWBWQS Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|--|---|---|-------------------------------------|--|
| S-8/FSW-SD-012 | 0–0.5 | Metals, Explosives, Pesticides/PCBs, SVOCs, Ammonia, Phosphorus | - Tributary to Sand Creek, | Station S-8 is adjacent to the former Sand Creek Sewage Treatment Plant. Station S-8 is along a | | |
| S-8/FSW-SW- 012 | NA | Metals, Explosives, Pesticides/PCBs, SVOCs, Ammonia, and Phosphorus | downstream from Lower Cobbs Pond outlet | tributary to Sand Creek, which is fed by drainage from Upper and Lower Cobbs Ponds, which in turn have drainages from Load Lines 3 and 12 (1941–1971). All surface water that exits the Upper and Lower Cobbs Ponds AOC enters the tributary to Sand Creek and passes through the sampling | | |
| S-8/FSW-SW- 052 | NA | Metals, Explosives, SVOCs, Ammonia, and Phosphorus | | station S-8. | | |
| Uppercobb Pond/ FSW-SD-030 | 0–0.3 | Metals, Explosives, Pesticides/PCBs, SVOCs, Ammonia, Phosphorus | | 1941–1971 - Upper and Lower Cobbs Ponds received effluent from Load Lines 3 and 12 sawdust filtration units, wash water, storm water runoff, and surface water runoff. Rinsate from demilitarization operations at Load Lines 3 and 12 was initially allowed to flow out of buildings and | Fish kill from | TNT, RDX, Comp B, HMX, |
| Uppercobb Pond/ FSW-SW-030 | NA | Metals, Explosives, Pesticides/PCBs, SVOCs, Ammonia, and Phosphorus | Upper Cobbs Pond | directly onto the ground or to drainage ditches which ultimately discharged to Upper Cobbs Pond and Lower Cobbs Pond. From 1951–1961, explosives melt-out and demilitarization activities occurred at Load Line 12. On November 15, 1966, a fish kill occurred at Lower Cobbs Pond as a | mishandling aluminum chloride at | DNT, nitrocellulose, nitrate, aluminum chloride, arsenic, cadmium, chromium, lead, |
| Uppercobb Pond/ FSW-SW-070 | NA | Metals, Explosives, SVOCs, Ammonia, and Phosphorus | | result of improper handling of aluminum chloride during these manufacturing operations at Load Line 12. | Load Line 12. | mercury |
| Lowcobb Pond/ FSW-SD-031 | 0–0.3 | Metals, Explosives, Pesticides/PCBs, SVOCs, Ammonia, Phosphorus | | 1941–1971 - Upper and Lower Cobbs Ponds received effluent from Load Lines 3 and 12 sawdust filtration units, wash water, storm water runoff, and surface water runoff. Rinsate from | | |
| Lowcobb Pond/ FSW-SW-031 | NA | Metals, Explosives, Pesticides/PCBs, SVOCs, Ammonia, and Phosphorus | Lower Cobbs Pond | demilitarization operations at Load Lines 3 and 12 was initially allowed to flow out of buildings and directly onto the ground or to drainage ditches which ultimately discharged to Upper Cobbs Pond and Lower Cobbs Pond. From 1951–1961, explosives melt-out and demilitarization activities | | |
| Lowcobb Pond/ FSW-SW-071 | NA | Metals, Explosives, SVOCs, Ammonia, and Phosphorus | | occurred at Load Line 12. On November 15, 1966, a fish kill occurred at Lower Cobbs Pond as a result of improper handling of aluminum chloride during these manufacturing operations at Load Line 12. | | |

AOC = Area of concern. bgs = Below ground surface. Comp B = Composition B (a combination of TNT and RDX). DNT = Dinitrotoluene.

ft = Feet.

ft = Feet. FWBWQS = Facility-wide Biological and Water Quality Study HMX = Octahy dro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine. NA = Not applicable. PCB = Polychlorinated biphenyl. RDX = Hexahy dro-1,3,5-trinitro-1,3,5-triazine. RI = Remedial investigation. TNT = 2,4,6-Trinitrotoluene. SVOC = Somi volatile organic compound.

SVOC = Semi-volatile organic compound.

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| A | | Lower Cobbs | 05 400 | Upper Cobbs |
|------------------------------------|------------|-----------------------|---------------------|------------------|
| Area | | Pond | Off AOC | Pond |
| Station | | Lowcobb Pond | S-8 | Uppercobb Pond |
| Sample ID | | FSW-SD-031-0000 | FSW-SD-012- 0000 | FSW-SD-030-0000 |
| Date | | 06/23/03 | 06/24/03 | 06/23/03 |
| Depth (ft) | | 0.0 - 0.3 | 0.0 - 0.5 | 0.0 - 0.3 |
| Parameters Analyzed | | | TAL Metals, | TAL Metals, |
| | _ | TAL Metals, | Explosives, | Explosives, |
| | | Explosives, | Pesticides/PCBs, | Pesticides/PCBs, |
| | | Pesticides/PCBs, | SVOCs, | SVOCs, |
| | Background | SVOCs, Ammonia, | Ammonia, | Ammonia, |
| Analyte | Criteria | Phosphorus | Phosphorus | Phosphorus |
| | | Metals (mg/kg) | | |
| Aluminum | 13900 | 15400J* | 6110J | 9030J |
| Antimony | 0 | 1.3J* | 0.51J* | 1J* |
| Arsenic | 19.5 | 10.2J | 12.6 | 6.5J |
| Barium | 123 | 116 | 50.6 | 78.9 |
| Beryllium | 0.38 | 0.83* | 0.53* | 0.48* |
| Cadmium | 0 | 1.2* | 0.12J* | 0.6* |
| Calcium | 5510 | 2550 | 1430J | 1960 |
| Chromium | 18.1 | 121* | 15.1 | 18.6* |
| Cobalt | 9.1 | 12* | 9.5J* | 7.7 |
| Copper | 27.6 | 120R | 13.3J | 23R |
| Iron | 28200 | 26000R | 19800 | 16200J |
| Lead | 27.4 | 34.2* | 12.2J | 17.8 |
| Magnesium | 2760 | 2600 | 1580J | 1530 |
| Manganese | 1950 | 602 | 1090 | 383 |
| Mercury | 0.059 | 0.1J* | 0.046J | 0.048J |
| Nickel | 17.7 | 24.6J* | 24.2* | 14.4J |
| Potassium | 1950 | 1640J | 1260R | 858J |
| Selenium | 1.7 | 1.6J | 1.1 | 1.1J |
| Silver | 0 | 1.5* | 0.19* | 1.1* |
| Sodium | 112 | 203* | 85.6 | 106 |
| Thallium | 0.89 | <3.6U | 0.69J | <2.8U |
| Vanadium | 26.1 | 24.5 | 11.5 | 15.6 |
| Zinc | 532 | 259 | 97.3 | 153 |
| | | Anions (mg/kg) | | - |
| Nitrate/Nitrite (NO ₃ / | | | | |
| NO ₂ -N) | None | 8.5* | <1.8U | 5.3* |
| | | Miscellaneous (mg/kg) | | |
| Ammonia | None | 61* | 10* | 69* |
| Total Phosphorus as P | None | 360* | 500* | 270* |

Table 4-8. Analytes Detected in Facility-wide Biological and Water Quality Study Sediment Samples

| | 1 | I GU | | |
|----------------------------|------------|------------------|------------------|------------------|
| | | Lower Cobbs | | Upper Cobbs |
| Area | | Pond | Off AOC | Pond |
| Station | | Lowcobb Pond | S-8 | Uppercobb Pond |
| | | | FSW-SD-012- | |
| Sample ID | | FSW-SD-031-0000 | 0000 | FSW-SD-030-0000 |
| Date | | 06/23/03 | 06/24/03 | 06/23/03 |
| Depth (ft) | | 0.0 - 0.3 | 0.0 - 0.5 | 0.0 - 0.3 |
| Parameters Analyzed | | | TAL Metals, | TAL Metals, |
| | | TAL Metals, | Explosives, | Explosives, |
| | | Explosives, | Pesticides/PCBs, | Pesticides/PCBs, |
| | | Pesticides/PCBs, | SVOCs, | SVOCs, |
| | Background | SVOCs, Ammonia, | Ammonia, | Ammonia, |
| Analyte | Criteria | Phosphorus | Phosphorus | Phosphorus |
| | | SVOCs (mg/kg) | | |
| Bis(2-ethylhexyl)phthalate | None | <1.3U | <0.42U | 0.84J* |
| Butyl benzyl phthalate | None | <1.3U | <0.42U | 0.16J* |
| Di-n-butyl phthalate | None | 1.035J* | <0.42U | 2.7* |

Table 4-8. Analytes Detected in Facility-wide Biological and Water Quality Study Sediment Samples (continued)

^aOnly detected site-related contaminants are presented in the table.

^bBackground concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

AOC = Area of concern.

ft = Feet.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

R = Rejected result.

SVOC = Semi-volatile organic compound. TAL = Target analyte list.

U = Not detected.

* = Result exceeds background criteria or no background criteria was available.

| Table 4–9. Analytes D | Detected in Facility-wide | Biological and Water | Ouality Study | Surface Water Samples |
|-----------------------|---------------------------|-----------------------------|----------------------|-----------------------|
| | | | | ····· |

| Area | | Lower Cobbs Pond | Lower Cobbs Pond | Off AOC | Off AOC | Upper Cobbs Pond | Upper Cobbs Pond |
|---|---------------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------------|
| Station | | Lowcobb Pond | Lowcobb Pond | S-8 | S-8 | Uppercobb Pond | Uppercobb Pond |
| Sample ID | | FSW-SW-031-0000 | FSW-SW-071-0000 | FSW-SW-012-0000 | FSW-SW-052-0000 | FSW-SW-030-0000 | FSW-SW-070-0000 |
| Date | | 06/23/03 | 08/05/03 | 06/24/03 | 09/16/03 | 06/23/03 | 08/05/03 |
| Parameters Analyzed | | TAL Metals, Explosives, |
| | - | Pesticides/PCBs, SVOCs, | SVOCs, Ammonia, and | Pesticides/PCBs, SVOCs, | SVOCs, Ammonia, and | Pesticides/PCBs, SVOCs, | SVOCs, Ammonia, and |
| Filtered? | | Ammonia, and Phosphorus | Phosphorus | Ammonia, and Phosphorus | Phosphorus | Ammonia, and Phosphorus | Phosphorus |
| Analyte (mg/L) | Background Criteria | Total | Total | Total | Total | Total | Total |
| | | • | Ме | etals (mg/kg) | · | | • |
| Aluminum | 3.37 | 0.226 | 0.125 | 0.311 | 0.0678J | 0.119 | 0.0916 |
| Antimony | 0 | 0.0034J* | 0.0022J* | <0.0019UJ | <0.007UJ | <0.007UJ | <0.007U |
| Barium | 0.0475 | 0.0135 | 0.013 | 0.0163 | 0.0156 | 0.0162 | 0.0138 |
| Calcium | 41.4 | 20.1 | 10.9 | 22.8 | 22.3 | 24.3 | 13.2 |
| Chromium | 0 | 0.0011J* | 0.00053J* | 0.001J* | 0.0018* | 0.001J* | <0.0016U |
| Cobalt | 0 | <0.0016U | <0.0016U | 0.00052J* | <0.0016U | <0.0016U | <0.0016U |
| Copper | 0.0079 | 0.0029J | 0.0023J | 0.00098J | <0.0032U | 0.0051 | 0.0026J |
| Iron | 2.56 | 1.35 | 0.899 | 1.38 | 0.303 | 1.3 | 0.873 |
| Lead | 0 | <0.008UJ | 0.0029J* | <0.0022U | <0.008U | 0.0033J* | 0.0022J* |
| Magnesium | 10.8 | 3.3 | 1.55 | 3.95 | 3.67 | 3.6 | 1.75 |
| Manganese | 0.391 | 0.19 | 0.182 | 0.216 | 0.03 | 0.188 | 0.0772 |
| Nickel | 0 | <0.0044U | <0.0044U | 0.002J* | <0.0044U | <0.0044U | <0.0044U |
| Potassium | 3.17 | 1.58 | 1.61 | 1.78 | 1.75 | 1.57 | 1.44 |
| Selenium | 0 | <0.007UJ | <0.007U | 0.0043J* | <0.007U | <0.007UJ | <0.007U |
| Sodium | 21.3 | 2.04 | 1.37 | 2.71J | 2.11 | 2.07 | 1.46 |
| Vanadium | 0 | 0.00082J* | 0.00071J* | 0.00086J* | 0.00054J* | <0.002U | <0.002U |
| Zinc | 0.042 | 0.0042J | 0.0044J | 0.0063J | <0.012UJ | 0.008J | 0.0041J |
| | | | | ions (mg/kg) | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | None | 0.13* | 0.1* | 0.17* | 0.07* | 0.1* | 0.18* |
| | | | | laneous (mg/kg) | | | |
| Total Phosphorus as P | None | 0.12* | 0.1* | 0.2* | 0.17* | 0.24* | 0.06* |
| | | | - | osives (mg/kg) | | | |
| 4-Amino-2,6-Dinitrotoluene | None | <0.00026U | 0.00029J* | <0.00026UJ | <0.00026UJ | <0.00026U | 0.00024J* |
| | - | | | OCs (mg/kg) | | | |
| Bis(2-ethylhexyl)phthalate | None | 0.0029J* | 0.0022J* | 0.0017J* | <0.011UJ | <0.011U | 0.002J* |
| Di-n-butyl phthalate | None | 0.012* | 0.0023J* | <0.011U | 0.00445J* | 0.0046J* | 0.0018J* |

^aOnly detected site-related contaminants are presented in the table. ^bBackground concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

AOC = Area of concern.

ID = Identification.

ID = Identification. J = Estimated value less than reporting limits. mg/kg = Milligrams per kilogram. mg/L = Milligrams per liter. PCB = Polychlorinated biphenyl. SVOC = Semi-volatile organic compound. TAL = Target analyte list. U = Not detected. LU = Non detected.

UJ = Non-detectable concentration and reporting limit estimated. VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

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| Pond Bank Soil | Sediment | Surface Water | Groundwater |
|----------------|----------------------|---------------|----------------------------|
| Aluminum | Aluminum | Arsenic | Arsenic |
| Arsenic | Arsenic | Barium | Manganese |
| Chromium | Cadmium | Manganese | Bis(2-ethylhexyl)phthalate |
| Vanadium | Silver | Vanadium | 4,4'-DDD |
| | Vanadium | | |
| | Benz(a)anthracene | | |
| | Benzo(a)pyrene | | |
| | Benzo(b)fluoranthene | | |

Table 4-10. Human Health COPCs per the Phase II RI

Source: Phase II Remedial Investigation Report for Upper and Lower Cobbs Ponds (MKM 2005).

COPC = Chemical of potential concern.

DDD = Dichlorodiphenyldichloroethane

RI = Remedial investigation.

Table 4–11. Chemicals Detected at Concentrations above Screening Criteria in Previous Investigations

| Surface Soil | Subsurface Soil | Sediment | Surface Water |
|----------------------------|----------------------|----------------------|------------------|
| Aluminum | Media not previously | Aluminum | Arsenic |
| Arsenic | sampled | Arsenic | Manganese |
| Chromium | | Cadmium | |
| Cobalt | | Chromium | |
| Thallium | | Chromium, | |
| Benzo(a)pyrene | | hexavalent | |
| Dibenz(a,h)anthracene | | Copper | |
| N-Nitroso-di-n-propylamine | | Silver | |
| | | Thallium | |
| | | Benz(a)anthracene | |
| | | Benzo(a)pyrene | |
| | | Benzo(b)fluoranthene | |
| | | Indeno(1,2,3- | |
| | | cd)pyrene | |

Source: Phase I Remedial Investigation Report for High Priority Areas of Concern (USACE 1998) and Phase II Remedial Investigation Report for Upper and Lower Cobbs Ponds (MKM 2005).

This table was generated using data from the following sources:

Screening criteria are the smaller of the FWCUGs for the Resident Farmer Adult, Resident Farmer Child, and National Guard Trainee based on hazard quotient =0.1 or target risk =1E-06

RI = Remedial investigation.

| | | | | Ana | alyses Perfor | med | |
|----------------------|---|--|--------|------------|---------------|---------------------|------|
| PBA08 RI Location | Targeted Area | Purpose | Metals | Explosives | VOCs | Pesticides/ PCBs | SVOC |
| CPCss-036 | Pond Bank – northeastern bank of Lower Cobbs Pond | Characterize an area not previously sampled. | Y | Y | Ν | N | Y |
| CPUSS-057 | Pond Bank – southeastern | Delineate lateral extent of previously identified surface contamination. | Y | Y | Ν | Ν | Y |
| CI C33-037 | bank of Lower Cobbs Pond | QA/QC | Y | Y | Ν | N | Y |
| | | QA/QC | Y | Y | Ν | N N N | Y |
| CPCss-038 | Pond Bank –eastern bank of Upper Cobbs Pond | Characterize an area not previously sampled. | Y | Y | Ν | N | Y |
| CPCss-039 | Pond Bank –southeastern bank of Upper Cobbs Pond | Delineate lateral extent of previously identified surface contamination. Analyzed for RVAAP full-suite analytes. | Y | Y | Y | Y | Y |
| CPCss-040 | Pond Bank – eastern bank south of Track FA | Delineate lateral extent of previously identified surface contamination. | Y | Y | Ν | N | Y |
| CPCss-041 | Pond Bank – western bank of Upper Cobbs Pond | Delineate lateral extent of previously identified surface contamination. | Y | Y | Ν | N | Y |
| CPCss-042 | Pond Bank – northwestern bank of Upper Cobbs Pond | Delineate lateral extent of previously identified surface contamination. | Y | Y | Ν | Ν | Y |
| CPCss-043 | Pond Bank – south/southwestern bank of Lower Cobbs Pond | Characterize an area not previously sampled. | Y | Y | Ν | N | Y |

Table 4–12. PBA08 RI Surface Soil Samples and Rationales

PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation.

PCB = Polychlorinated biphenyl.

QA = Quality assurance. QC = Quality control.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

VOC = Volatile organic compound.

| Area | | Pond Bank | Pond Bank | Pond Bank | Pond Bank | Pond Bank | Pond Bank | Pond Bank | Pond Bank | Pond Bank | Pond Bank |
|----------------------------|--------------|-------------------|----------------------|----------------------------|-------------------|-------------------------|-------------------|-------------------|-------------------|-------------------|-------------------------|
| Station | - | CPCsb-030 | CPCsb-031 | CPCsb-032 | CPCsb-034 | CPCsb-035 | CPCss-036 | CPCss-037 | CPCss-037 | CPCss-038 | CPCss-039 |
| | - | CPCsb-030-5105- | CPCsb-031-5109- | CPCsb-032-5113- | CPCsb-034-5119- | CPCsb-035-5123- | CPCss-036-5014- | CPCss-037-6041- | CPCss-037-5015- | CPCss-038-5016- | CPCss-039-5017- |
| Sample ID | | SO | so | SO | so | SO | so | FD | SO | SO | SO |
| Date | - | 03/29/10 | 03/24/10 | 03/24/10 | 03/29/10 | 03/29/10 | 02/23/10 | 02/23/10 | 02/23/10 | 02/23/10 | 02/23/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | RVAAP Full-suite | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | RVAAP Full-suite |
| Analyte | Criteria | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVOCs | analytes | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVOCs | analytes |
| | - | - | - | | | (mg/kg) | - | - | - | | - |
| Aluminum | 17700 | 9230 | 13400J | 12200J | 9920 | 10500 | 14200 | 10800 | 11900 | 13200 | 8790 |
| Antimony | 0.96 | <0.67UJ | 0.12J | 0.086J | <0.72UJ | 0.088J | <0.66R | <0.74R | <0.76R | <0.65R | <0.63R |
| Arsenic | 15.4 | 6.5 | 15.1J | 13.1J | 19.7* | 8.7 | 6.8J | 6J | 6.7J | 11.2J | 11J |
| Barium | 88.4 | 75.7 | 80.7 | 52.4 | 67.7 | 62.9 | 61.8J | 64.5J | 57.6J | 43.3J | 38.7J |
| Beryllium | 0.88 | 0.57 | 0.65 | 0.58 | 0.65 | 0.52 | 0.36 | 0.51 | 0.58 | 0.41 | 0.58 |
| Cadmium | 0 | 0.053J* | 0.17J* | 0.095J* | 0.19J* | 0.2J* | 0.058J* | 0.2J* | 0.26J* | 0.058J* | 0.1J* |
| Calcium | 15800 | 1130J | 1680J | 1360J | 2400J | 3810J | 804J | 689J | 725J | 279J | 1500J |
| Chromium | 17.4 | 10.8 | 20.8* | 17.4* | 13.9 | 12.5 | 15.9J | 12.6J | 14.6J | 17.2J | 11.7J |
| Cobalt | 10.4 | 8.1 | 13.3* | 16.4* | 12.1* | 6.5 | 4.5 | 8.8 | 8.1 | 7.6 | 10.9* |
| Copper | 17.7 | 8.7 | 20* | 21.3* | 17.9* | 10.4 | 12.3J | 10.2J | 14J | 14.4J | 19.9J* |
| Iron | 23100 | 15200 | 31200* | 30200* | 31700* | 21000 | 20600 | 18000 | 19300 | 25700* | 18700 |
| Lead | 26.1 | 10.6 | 19.9J | 15.7J | 15.9 | 30.1* | 11.1 | 12.7 | 14.2 | 12 | 11.8 |
| Magnesium | 3030 1450 | 1880 80.7 | 3540J* 330 | 3290J* 237 | 2520 667 | 2200 506 | 2200J 78.6 | 2080J 238 | 2370J 242 | 2930J 200 | 1980J 186 |
| Manganese | | 0.023J | | <0.12U | | | | 238 0.063J* | 0.045J* | 0.052J* | 0.023J |
| Mercury Nickel | 0.036 | 12.8 | 0.021J 24.6J* | <u><0.120</u> 25.7J* | <0.14U 20.5 | 0.034J 11.9 | <0.13U 12.3J | 14.3J | 16.5J | | 17.6J |
| | 21.1 927 | 508 | 24.0J* 1040J* | 25.7J* 1090J* | 877 | 11.9 1190* | 12.3J 571J | 14.5J 560J | 788J | 18J 785J | 612J |
| Potassium Selenium | 1.4 | 0.79J | 1.4* | 1.2 | 8// 1.2J | 1190 1J | 0.72 | 0.76 | 0.92 | 0.63J | 0.89 |
| Silver | 0 | 0.795 0.046J* | 0.032J* | <0.022UJ | <0.014UJ | 1J 0.038J* | <0.028UJ | <0.043UJ | 0.92 0.062J* | <0.034UJ | <0.021UJ |
| Sodium | 123 | 38.4J | 43.3J | 47J | 108J | 40.7J | 42.7J | <0.04303 66.8J | 62.8J | 51J | 48.8J |
| Thallium | 0 | 0.12J* | 0.18J * | 0.18J* | 0.14J* | 0.14J* | 42.73 0.19J* | 0.18J* | 02.8J | 0.18J* | 0.11J * |
| Vanadium | 31.1 | 15.2 | 26.5 | 20.3 | 19 | 20 | 26.3J | 18.4J | 18.7J | 21.5J | 14.1J |
| Zinc | 61.8 | 49.7J | 61.3 | 61.3 | 96.4J* | 61.9J* | 54.1J | 48.3J | 60.8J | 49.1J | 54.6J |
| | 01.0 | 19.75 | 01.5 | 01.5 | Explosive | | 5 1.15 | 10.55 | 00.03 | 17.15 | 5 1.05 |
| Nitrocellulose | None | NR | NR | NR | NR | 1.5J* | NR | NR | NR | NR | 1.2J* |
| | 1.0110 | | 1.11 | | SVOCs | | 1.11 | | 1.11 | 1.11 | 1120 |
| 2-Methylnaphthalene | None | <0.44U | <0.44U | <0.41U | <0.48U | 0.0091J* | <0.43U | <0.49U | <0.5U | <0.43U | <0.41U |
| Acenaphthylene | None | <0.067U | <0.067U | <0.062U | <0.072U | 0.012J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Benz(a)anthracene | None | <0.067U | <0.067U | <0.062U | 0.017J* | 0.035J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Benzenemethanol | None | <0.44U | <0.44U | <0.41U | <0.48U | <0.43U | <0.43U | <0.49U | <0.5U | 0.44* | <0.41U |
| Benzo(a)pyrene | None | <0.067U | <0.067U | <0.062U | 0.014J* | 0.042J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Benzo(b)fluoranthene | None | <0.067U | <0.067U | <0.062U | 0.021J* | 0.059J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Benzo(ghi)perylene | None | <0.067U | <0.067U | <0.062U | <0.072U | 0.031J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Benzo(k)fluoranthene | None | <0.067U | <0.067U | <0.062U | <0.072U | 0.028J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Benzoic acid | None | <1.1U | <1.1U | <0.99U | <1.2U | <1.1U | <1.1U | <1.2U | <1.2U | 17* | <1U |
| Bis(2-ethylhexyl)phthalate | None | <0.44U | <0.44U | <0.41U | <0.48U | <0.43U | <0.43U | <0.49U | <0.5U | <0.43U | 0.23J* |
| Chrysene | None | <0.067U | <0.067U | <0.062U | 0.014J* | 0.048J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Di-n-butyl phthalate | None | 0.021J* | <0.44U | <0.41U | 0.027J* | <0.43U | <0.43U | <0.49U | <0.5U | <0.43U | <0.41U |
| Fluoranthene | None | <0.067U | <0.067U | <0.062U | 0.032J* | 0.093* | 0.012J* | 0.021J* | 0.023J* | <0.065U | <0.063U |
| Indeno(1,2,3-cd)pyrene | None | <0.067U | <0.067U | <0.062U | <0.072U | 0.026J* | <0.066U | <0.074U | <0.076U | <0.065U | <0.063U |
| Phenanthrene | None | <0.067U | <0.067U | <0.062U | <0.072U | 0.05J* | <0.066U | 0.01J* | <0.076U | <0.065U | <0.063U |
| Pyrene | None | <0.067U | <0.067U | <0.062U | 0.025J* | 0.072* | 0.011J* | 0.016J* | 0.016J* | <0.065U | <0.063U |
| | 1 | | | | PCBs (| | | | | | |
| beta-BHC | None | NR | NR | NR | NR | 0.0035J* | NR | NR | NR | NR | <0.0044U |

Table 4–13. Analytes Detected in PBA08 RI Surface Soil Samples

| Area | | Pond Bank | Pond Bank | Pond Bank | Pond Bank |
|----------------------------------|------------|---------------------|--------------------|------------------------------|--------------------|
| Station | | CPCss-040 | CPCss-041 | CPCss-042 | CPCss-043 |
| Sample ID | | CPCss-040-5018-SO | CPCss-041-5019-SO | CPCss-042-5020-SO | CPCss-043-5021-5 |
| Date | | 02/23/10 | 02/23/10 | 02/23/10 | 02/23/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, |
| Analyte | Criteria | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVO |
| | | Metals | | | |
| Aluminum | 17700 | 10900 | 12400 | 9340 | 9890 |
| Antimony | 0.96 | <0.67R | <0.66R | <0.69R | 0.23J |
| Arsenic | 15.4 | 6.2J | 11.3J | 7.1J | 10.8J |
| Barium | 88.4 | 62.9J | 47.3J | 57.3J | 59.6J |
| Beryllium | 0.88 | 0.37 | 0.45 | 0.37 | 0.48J |
| Cadmium | 0 | 0.1J* | 0.18J* | 0.17J* | 0.2J* |
| Calcium | 15800 | 290J | 975J | 737J | 2090J |
| Chromium | 17.4 | 12.2J | 15.3J | 12.5J | 18.7J* |
| Cobalt | 10.4 | 8.1 | 7.9 | 6.5 | 10.4 |
| Copper | 17.7 | 5.4J | 11.9J | 9.6J | 20.2J* |
| Iron | 23100 | 16700 | 23800* | 16900 | 24100* |
| Lead | 26.1 | 13.8 | 17.4 | 31.3* | 14.5 |
| Magnesium | 3030 | 1810J | 2350J | 1970J | 3230J* |
| Manganese | 1450 | 567 | 468 | 326 | 374 |
| Mercury | 0.036 | 0.052J* | 0.047J* | 0.072J* | 0.074J* |
| Nickel | 21.1 | 10.8J | 13.9J | 12.4J | 23.2J* |
| Potassium | 927 | 671J | 1090J* | 786J | 840J |
| Selenium | 1.4 | 0.68 | 0.78 | 0.67J | 0.96J |
| Silver | 0 | <0.04UJ | <0.04UJ | 0.075 0.057J* | 0.085J* |
| Sodium | 123 | 47.6J | 44.9J | 46.5J | 70.7J |
| Thallium | 0 | 0.2J* | 0.16J* | 0.16J* | 0.13J* |
| Vanadium | 31.1 | 17.6J | 21.6J | 18.4J | 17J |
| Zinc | 61.8 | 51.8J | 57.8J | 52.8J | 63.5J* |
| Zine | 01.0 | Explosive | | 52.05 | 03.30 |
| Nitrocellulose | None | NR | NR | NR | NR |
| Wildeenulose | None | SVOCs | | TVIX | NIC |
| 2-Methylnaphthalene | None | <0.44U | <0.44U | <0.45U | <3UJ |
| Acenaphthylene | None | <0.067U | <0.066U | <0.450 <0.069U | <0.45UJ |
| Benz(a)anthracene | None | <0.0070 | <0.066U | 0.066J* | <0.45UJ |
| Benzenemethanol | None | <0.44U | <0.000C | <0.45U | <3UJ |
| Benzo(a)pyrene | None | <0.440 | <0.066U | 0.057J* | <0.45UJ |
| Benzo(b)fluoranthene | None | <0.007U | <0.066U | <0.069U | <0.45UJ |
| Benzo(ghi)perylene | None | <0.007U | <0.066U | <0.069U | <0.45UJ |
| Benzo(k)fluoranthene | None | <0.007U | <0.066U | <0.069U | <0.45UJ |
| Benzoic acid | None | <0.0070 | <0.0000 | <0.0090 | <0.4505 <7.2U |
| Bis(2-ethylhexyl)phthalate | | <0.44U | <0.44U | <0.45U | <7.20 <3UJ |
| | None | <0.440 <0.067U | <0.440 <0.066U | <0.450 <0.069U | <30J <0.45UJ |
| Chrysene Din hutrul nhthalata | None | | <0.066U <0.44U | <0.0690 <0.45U | <0.45UJ <3UJ |
| Di-n-butyl phthalate | None | <0.44U | <0.44U 0.033J* | <0.450 0.11 * | |
| Fluoranthene | None | 0.018J* | | 0.11 [★] <0.069U | <0.45UJ <0.45UJ |
| Indeno(1,2,3-cd)pyrene | None | <0.067U 0.0095J* | <0.066U 0.017J* | <0.069U 0.093* | |
| Phenanthrene | None | | | | <0.45UJ |
| Pyrene | None | 0.016J* | 0.028J* | 0.079* | <0.45UJ |

Table 4-13. Analytes Detected in PBA08 RI Surface Soil Samples (continued)

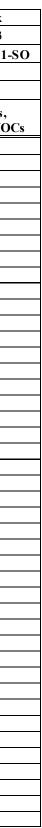


Table 4–13. Analytes Detected in PBA08 RI Surface Soil Samples (continued)

| Area | | Pond Bank | Pond Bank | Pond Bank | Pond Bank |
|---------------------|------------|-------------------|-------------------|-------------------|-----------------|
| Station | | CPCss-040 | CPCss-041 | CPCss-042 | CPCss-043 |
| Sample ID | | CPCss-040-5018-SO | CPCss-041-5019-SO | CPCss-042-5020-SO | CPCss-043-5021- |
| Date | | 02/23/10 | 02/23/10 | 02/23/10 | 02/23/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, |
| Analyte | Criteria | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVOCs | Explosives, SVO |
| | | PCBs (| mg/kg) | | |
| beta-BHC | None | NR | NR | NR | NR |

^aOnly detected site-related contaminants are presented in the table.

^bBackground concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

BHC = Hexachlorocyclohexane.

ft = Feet.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation. PCB = Polychlorinated biphenyl. R = Rejected result.

RI = Remedial investigation.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

TAL = Target analyte list.

U = Not detected.

UJ = Non-detectable concentration and reporting limit estimated.

* = Result exceeds background criteria or no background criteria was available.



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| PBA08 RI Location | Comments/Rationale | Sample Type | Depth (ft bgs) | Analyses Performed Metals | Explosives | VOCs | Pesticides/ PCBs | SVOC |
|----------------------|---|-----------------------|-------------------|---------------------------------|------------|------|---------------------|------|
| | | Discrete | 0-1 | Y | Y | N | N | Y |
| | Confirm absence of contamination in an | NS | 1-4 | N | N | N | N | N |
| CPCsb-030 | area not previously sampled. | NS | 4–7 | N | N | N | N | Ν |
| | Groundwater encountered at 1 ft bgs. | NS | 7–13 | N | N | N | N | Ν |
| _ | | Discrete | 0–1 | Y | Y | Ν | N | Y |
| CDC 1 021 | Delineate vertical extent of previously | NS | 1–4 | Ν | Ν | Ν | N | Ν |
| CPCsb-031 | identified surface contamination. | NS | 4–7 | N | Ν | Ν | N | Ν |
| | Groundwater encountered at 1.3 ft bgs. | NS | 7–13 | N | Ν | Ν | N | Ν |
| | Delineate vertical extent of previously identified surface contamination. Groundwater encountered at 10 ft bgs. | Discrete | 0-1 | Y | Y | N | N | Y |
| | | Discrete | 1–4 | Y | Y | Ν | Ν | Y |
| CPCsb-032 | | Discrete | 4–7 | Y | Y | Ν | Ν | Y |
| CrCs0-052 | | Discrete ^a | 7–10 | Y | Y | Ν | Ν | Y |
| | QA/QC. | Discrete | 1–4 | Y | Y | Ν | Ν | Y |
| | QA/QC. | Discrete | 1–4 | Y | Y | Ν | Ν | Y |
| CPCsb-033 | Geotechnical. | Discrete | 4-5.4 | N | N | N | N | Ν |
| CI CSU-055 | Geotechnical. | Discrete | 8-9.7 | N | N | N | N | N |
| | Delineate vertical extent of previously | Discrete | 0-1 | Y | Y | N | N | Y |
| CPCsb-034 | identified surface contamination. | Discrete | 1–2 | Y | Y | N | N | Y |
| CI C30-054 | Groundwater encountered at 2 ft bgs. | NS | 4–7 | N | N | N | N | Ν |
| | | NS | 7–13 | N | N | N N | | N |
| | Delineate vertical extent of previously | Discrete | 0-1 | Y | Y | Y | Y | Y |
| | identified surface contamination. | Discrete | 1–4 | Y | Y | Y | Y | Y |
| CPCsb-035 | Analyzed for RVAAP full-suite | Discrete | 4–7 | Y | Y | Y | Y | Y |
| CI C30 033 | analytes. | NA | 7–13 | N | N | N | N | Ν |
| | QA/QC. Analyzed for RVAAP full-suite | Discrete | 4–7 | Y | Y | Y | Y | Y |
| | analytes. | Discrete | 4–7 | Y | Y | Y | Y | Y |

Table 4–14. PBA08 RI Subsurface Soil Rationale and Analyses

bgs = Below ground surface.

ft = Feet.

NA = Sample not analyzed by the laboratory based on preliminary screening criteria of the 4-7 ft bgs sample interval.

NS = Not sampled due to refusal.

PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation.

PCB = Polychlorinated biphenyl.

QA = Quality assurance.

QC = Quality control.

RVAAP = Ravenna Army and Ammunition Plant.

SVOC = Semi-volatile organic compound.

VOC = Volatile organic compound.

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| A | | David David | Dand Dank | Dend Denk | David David | David David | Dand Dank | Den d Denle | Dand Dank |
|----------------------------|------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------------|-------------------------|-------------------------|
| Area | | Pond Bank | Pond Bank | Pond Bank |
| Station | | CPCsb-032 | CPCsb-032 | CPCsb-032 | CPCsb-032 | CPCsb-034 | CPCsb-035 | CPCsb-035 | CPCsb-035 |
| Sample ID | | CPCsb-032-6073-FD | CPCsb-032-5114-SO | CPCsb-032-5115-SO | CPCsb-032-5116-SO | CPCsb-034-5120-SO | CPCsb-035-5124-SO | CPCsb-035-6072-FD | CPCsb-035-5125-SO |
| Date | | 03/24/10 | 03/24/10 | 03/24/10 | 03/24/10 | 03/29/10 | 03/29/10 | 03/29/10 | 03/29/10 |
| Depth (ft) | | 1.0 - 4.0 | 1.0 - 4.0 | 4.0 - 7.0 | 7.0 - 10.0 | 1.0 - 4.0 | 1.0 - 4.0 | 4.0 - 7.0 | 4.0 - 7.0 |
| Parameters Analyzed | Background | TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite |
| Analyte | Criteria | Explosives, SVOCs | analytes | analytes | analytes |
| | | | | Metals | s (mg/kg) | | | | |
| Aluminum | 19500 | 8690J | 9420J | 11000 | 11000J | 7120 | 10100 | 14800 | 10900 |
| Antimony | 0.96 | <0.6UJ | 0.13J | <0.6R | <0.58UJ | <0.64UJ | 0.084J | 0.079J | 0.086J |
| Arsenic | 19.8 | 11.1J | 17.9J | 14.3 | 15.9J | 14.2 | 9.7 | 18.6 | 15.4 |
| Barium | 124 | 68.8 | 63.2 | 49.1J | 38.1 | 40.7 | 62.7 | 77.5 | 61 |
| Beryllium | 0.88 | 0.58J | 0.57 | 0.61 | 0.49 | 0.41 | 0.47 | 0.87 | 0.58 |
| Cadmium | 0 | 0.069J* | 0.046J* | 0.035J* | 0.032J* | 0.068J* | 0.18J* | 0.078J* | 0.053J* |
| Calcium | 35500 | 1850J | 1530J | 5820 | 8240J | 1080J | 6870J | 3420Ј | 2340J |
| Chromium | 27.2 | 13.6 | 15.9 | 17.5 | 18.5 | 10.5 | 12.8 | 22.4 | 15.5 |
| Cobalt | 23.2 | 6.3 | 9.4 | 12.4 | 13.3 | 9.7 | 6.6 | 16.5 | 14.1 |
| Copper | 32.3 | 19.4 | 22.4 | 17.1 | 20 | 20.4 | 12.8 | 27.5 | 23 |
| Iron | 35200 | 22000 | 29900 | 29600 | 33600 | 25600 | 21300 | 37600* | 27300 |
| Lead | 19.1 | 11.3J | 11.9J | 11.8 | 11.2J | 10.9 | 23.3* | 15.5 | 13.4 |
| Magnesium | 8790 | 2650J | 2830J | 5460J | 7120J | 2110 | 2640 | 5250 | 3570 |
| Manganese | 3030 | 131 | 181 | 307 | 343 | 243 | 426 | 438 | 541 |
| Mercury | 0.044 | <0.12U | <0.12U | <0.12U | <0.12U | <0.13U | 0.022J | <0.12U | <0.12U |
| Nickel | 60.7 | 21.5J | 26.5J | 27.7J | 31.7J | 19.8 | 14.6 | 35.9 | 26.1 |
| Potassium | 3350 | 959J | 1050J | 1890J | 2190Ј | 856 | 1210 | 2050 | 1420 |
| Selenium | 1.5 | 1.6* | 1.6* | 0.88J | 1.1 | 1.2J | 1J | 1.5J | 1.1J |
| Silver | 0 | <0.016UJ | <0.0078UJ | 0.029J* | <0.024UJ | <0.0098UJ | <0.026UJ | 0.033J* | 0.013J* |
| Sodium | 145 | 46.6J | 46.6J | 78.4J | 110J | 40.8J | 44.5J | 58.8J | 46.9J |
| Thallium | 0.91 | 0.12J | 0.14J | 0.17J | 0.16J | 0.093J | 0.13J | 0.19J | 0.17J |
| Vanadium | 37.6 | 16.3 | 17.3 | 16.3 | 17.2 | 12.1 | 17.8 | 24.9 | 17.9 |
| Zinc | 93.3 | 55.1 | 59.7 | 61.1 | 62.5 | 57.2J | 61.9J | 86J | 65.4J |
| | | | | | ves (mg/kg) | | | | |
| Nitrocellulose | None | NR | NR | NR | NR | NR | 1.6J* | <6.2U | <6.2U |
| | | | | | s (mg/kg) | - | | | |
| 2-Methylnaphthalene | None | <0.4U | <0.38U | <0.4U | 0.014J* | <0.42U | 0.0091J* | <0.41U | <0.41U |
| Benz(a)anthracene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.047J* | <0.062U | <0.062U |
| Benzo(a)pyrene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.066* | <0.062U | <0.062U |
| Benzo(b)fluoranthene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.08* | <0.062U | <0.062U |
| Benzo(ghi)perylene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.06J* | <0.062U | <0.062U |
| Benzo(k)fluoranthene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.03J* | <0.062U | <0.062U |
| Bis(2-ethylhexyl)phthalate | None | <0.4U | <0.38U | <0.4U | <0.38U | <0.42U | <0.41U | 0.024J* | <0.41U |
| Chrysene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.063* | <0.062U | <0.062U |
| Di-n-butyl phthalate | None | <0.4U | <0.38U | 0.019J* | <0.38U | <0.42U | <0.41U | 0.025J* | <0.41UJ |
| Dibenz(a,h)anthracene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.021J* | <0.062U | <0.062U |
| Fluoranthene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.11* | <0.062U | <0.062U |
| Indeno(1,2,3-cd)pyrene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.037J* | <0.062U | <0.062U |
| Phenanthrene | None | <0.06U | <0.058U | <0.06U | 0.014J* | <0.064U | 0.051J* | <0.062U | <0.062U |
| Pyrene | None | <0.06U | <0.058U | <0.06U | <0.058U | <0.064U | 0.088* | <0.062U | <0.062U |

 Table 4–15. Analytes Detected in PBA08 RI Subsurface Soil Samples

^aOnly detected site-related contaminants are presented in the table.

^bBackground concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

ft = Feet.

ID = Identification.

J = Estimated value less than reporting limits. mg/kg = M illigrams per kilogram.

NR = Not reported/not analyzed.

PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation. R = Rejected result.RVAAP = Ravenna Army Ammunition Plant. SVOC = Semi-volatile organic compound. TAL = Target analyte list.

U = Non-detectable concentration.

< = Less than.

UJ = Non-detectable concentration and reporting limit estimated. * = Result exceeds background concentration.

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| | | | | | | Analy | ses Perfo | rmed | |
|----------------------|------------------------------|--|----------------|-------------------|--------|------------|-----------|---------------------|------|
| PBA08 RI Location | Targeted Area | Comments/Rationale | Sample Type | Depth (ft bgs) | Metals | Explosives | VOCs | Pesticides/ PCBs | svoc |
| CPCsd-044 | Outlet of | | Composite | 00.5 | Y | Y | N | Ν | Y |
| CPCsw-044 | Lower Cobbs Pond | Characterize off-AOC migration. | Grab | NA | Y | Y | Y | Y | Y |
| CPCsd-045 | T C 11 | Confirm presence of | Composite | 0–0.5 | Y | Y | Y | Y | Y |
| CPCsd-045 | Lower Cobbs Pond | contamination in previously | Discrete | 0.5–2 | Y | Y | Y | Y | Y |
| CPCsw-045 | Pond | sampled area. | Grab | NA | Y | Y | Y | Y | Y |
| CPCsd-046 | Upper Cobbs | Confirm presence of | Composite | 0–0.5 | Y | Y | Y | Y | Y |
| CPCsw-046 | Pond | contamination in previously sampled area. | Discrete | 0.5–2 | Y | Y | Y | Y | Y |
| CPCsd-047 | | Confirm presence of | Composite | 0–0.5 | Y | Y | Y | Y | Y |
| CPCsd-04/ | Inlet efflueren | contamination in previously | Discrete | 0.5–2 | Y | Y | Y | Y | Y |
| | Inlet of Upper Cobbs Pond | sampled area. | Grab | NA | Y | Y | Y | Y | Y |
| CPCsw-047 | Cobbs Pond | 04/05 | Grab | NA | Y | Y | Y | Y | Y |
| | | QA/QC | Grab | NA | Y | Y | Y | Y | Y |
| CPCsd-048 | Backwater | Confirm presence of | Composite | 0–0.5 | Y | Y | Y | Y | Y |
| CPCsd-048 | | contamination in previously | Discrete | 0.5–2 | Y | Y | Y | Y | Y |
| CPCsw-048 | Area | sampled area. | Grab | NA | Y | Y | Y | Y | Y |
| CPCsd-049 | Upper Cobbs Pond | Chromium speciation - Previous Cr result represents maximum Cr concentration | Composite | 0–0.5 | Cr | Ν | N | Ν | Ν |

Table 4–16. PBA08 RI Surface Water and Sediment Samples and Rationales

AOC = Area of concern.

bgs = Below ground surface.

Cr = Chromium.

ft = Feet.

NA = Not applicable.

PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation.

PCB = Polychlorinated biphenyl.

QA = Quality assurance.

QC = Quality control.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

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| Area | | Backwater Area | Backwater Area | Backwater Area | Lower Cobbs Pond | Lower Cobbs Pond | Upper Cobbs Pond |
|----------------------------|------------|-------------------|-------------------------|-------------------|-------------------------|-------------------------|-------------------|
| Station | | CPCsw-047 | CPCsw-047 | CPCsw-048 | CPCsw-044 | CPCsw-045 | CPCsw-046 |
| Sample ID | | CPCsw-047-6045-FD | CPCsw-047-5030-SW | CPCsw-048-5031-SW | CPCsw-044-5027-SW | CPCsw-045-5028-SW | CPCsw-046-5029-SW |
| Date | | 04/01/10 | 04/01/10 | 04/01/10 | 03/29/10 | 04/01/10 | 03/25/10 |
| Parameters Analyzed | | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite |
| Filtered? | Background | analytes | analytes | analytes | analytes | analytes | analytes |
| Analyte (mg/L) | Criteria | Total | Total | Total | Total | Total | Total |
| | · | • | Met | als (mg/kg) | • | • | • |
| Aluminum | 3.37 | 0.254 | 0.357 | 0.26 | 0.285 | 0.236 | 0.332 |
| Antimony | 0 | 0.00095J* | 0.00096J* | 0.001J* | 0.00073J* | 0.00088J* | 0.00094J* |
| Arsenic | 0.0032 | 0.00086J | 0.001J | 0.0009J | 0.00078J | 0.00092J | 0.00078J |
| Barium | 0.0475 | 0.0161 | 0.0216 | 0.0198 | 0.0121 | 0.0117 | 0.0168 |
| Beryllium | 0 | 0.00008J* | 0.00006J* | 0.00005J* | <0.001U | 0.00003J* | <0.001U |
| Cadmium | 0 | 0.00004J* | 0.00006J* | 0.00006J* | <0.002U | 0.00004J* | <0.002U |
| Calcium | 41.4 | 22.3 | 21.6 | 22.1 | 14.7 | 18 | 16.2 |
| Chromium | 0 | <0.005U | 0.00062J* | 0.00056J* | <0.00072UJ | <0.005U | 0.00059J* |
| Cobalt | 0 | 0.00027J* | 0.00039J* | 0.00041J* | <0.00018UJ | 0.00015J* | <0.00024UJ |
| Copper | 0.0079 | 0.0018J | 0.0021J | 0.0019J | 0.0016J | 0.0016J | 0.0019J |
| Iron | 2.56 | 0.814 | 0.995 | 0.846 | 0.814 | 0.683 | 0.818 |
| Lead | 0 | 0.0003J* | 0.00047J* | 0.00032J* | 0.00032J* | 0.00029J* | 0.00039J* |
| Magnesium | 10.8 | 3.34 | 3.27 | 3.25 | 2.71 | 2.88 | 2.85 |
| Manganese | 0.391 | 0.102 | 0.136 | 0.132 | 0.0915 | 0.107 | 0.115 |
| Nickel | 0 | 0.0017J* | 0.0019J* | 0.002J* | 0.0016J* | 0.0014J* | 0.0019J* |
| Potassium | 3.17 | 1.42 | 1.49 | 1.39 | 1.26 | 1.39 | 1.49 |
| Selenium | 0 | 0.0002J* | 0.00023J* | 0.00031J* | <0.005U | 0.00032J* | 0.00024J* |
| Sodium | 21.3 | 1.66 | 1.6 | 1.68 | 1.44 | 1.49 | 1.41 |
| Thallium | 0 | <0.002U | 0.00046J* | <0.002U | <0.002U | 0.00035J* | <0.002U |
| Vanadium | 0 | 0.00054J* | 0.00051J* | <0.01U | 0.00063J* | <0.01U | 0.00089J* |
| Zinc | 0.042 | <0.04U | 0.0109J | 0.0101J | <0.04U | <0.04U | <0.04U |
| | | · | Explo | sives (mg/kg) | | - | · |
| 4-Amino-2,6-Dinitrotoluene | None | 0.00004J* | 0.00004J* | 0.00007J* | <0.00015U | <0.00015U | 0.00007J* |
| | | | | Cs (mg/kg) | | | |
| Benzenemethanol | None | <0.01U | <0.01U | <0.01U | <0.01U | 0.0049J* | <0.01U |
| Bis(2-ethylhexyl)phthalate | None | <0.01U | <0.01UJ | <0.01UJ | <0.01U | <0.01UJ | 0.0019J* |
| Butyl benzyl phthalate | None | <0.01U | <0.01U | <0.01U | <0.01U | 0.0018J* | <0.01U |
| | | | | Bs (mg/kg) | | | |
| oeta-BHC | None | 0.00002J* | <0.0001U | <0.00005U | <0.00005UJ | <0.00005U | <0.00005U |

Table 4–17. Analytes Detected in PBA08 RI Surface Water Samples

^a Only detected site-related contaminants are presented in the table.

^b Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram. PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation. RVAAP = Ravenna Army Ammunition Plant. SVOC = Semi-volatile organic compound.

U = Non-detectable concentration.

UJ = Non-detectable concentration and reporting limit estimated. * = **Result exceeds background concentration.**

< = Less than.

| Table 4–18. Analytes | Detected in PBA08 | 3 RI Sediment Samples |
|----------------------|--------------------------|------------------------------|
| | | |

| | 1 | | | | | | 1 | 1 | 1 | | |
|----------------------------|------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------|-------------------------|-------------------------|-------------------------|-------------------------|-----------------|
| Area | _ | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Upper Cobbs Pond | | |
| Station | | CPCsd-047 | CPCsd-047 | CPCsd-048 | CPCsd-048 | CPCsd-044 | CPCsd-045 | CPCsd-045 | CPCsd-046 | CPCsd-046 | CPCsd-049 |
| | | CPCsd-047-5025- | CPCsd-047-5785- | CPCsd-048-5026- | CPCsd-048-5786- | CPCsd-044-5022- | CPCsd-045-5023- | CPCsd-045-5783- | CPCsd-046-5024- | CPCsd-046-5784- | CPCsd-049-5032- |
| Sample ID | | SD | SD | SD | SD | SD | SD | SD | SD | SD | SD |
| Date | _ | 04/01/10 | 04/01/10 | 04/01/10 | 04/01/10 | 03/29/10 | 04/01/10 | 04/01/10 | 03/25/10 | 03/25/10 | 03/25/10 |
| Depth (ft) | _ | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 |
| Parameters Analyzed | Background | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | |
| Analyte | Criteria | analytes | analytes | analytes | analytes | Explosives, SVOCs | analytes | analytes | analytes | analytes | TAL Metals |
| | 1 | | ſ | ſ | | s (mg/kg) | 1 | 1 | 1 | Γ | 1 |
| Aluminum | 13900 | 17300J* | 9150J | 6620J | 9120J | 8920 | 16200J* | 13400J | 18400J* | 8500J | NR |
| Antimony | 0 | 2.1J* | 0.15J* | 0.45J* | 0.17J* | <0.66UJ | 1.4J* | 0.15J* | 1.9J* | 0.19J* | NR |
| Arsenic | 19.5 | 16.7 | 6.4 | 5.8 | 4.9 | 8.8 | 14 | 7.7 | 18.8J | 5.5J | NR |
| Barium | 123 | 151* | 73.9 | 65.8 | 84.2 | 73.5 | 149* | 114 | 157* | 82.3 | NR |
| Beryllium | 0.38 | 1* | 0.59* | 0.52* | 0.58* | 0.42* | 0.93* | 0.68* | 1* | 0.45* | NR |
| Cadmium | 0 | 2.3* | 0.24J* | 0.61* | 0.28J* | 0.1J* | 1.5* | 0.44* | 2.3* | 0.28* | NR |
| Calcium | 5510 | 10100* | 1010 | 1720 | 3550 | 1260J | 3660 | 2110 | 4440J | 1650J | NR |
| Chromium | 18.1 | 24.3J* | 16.6J | 9.9J | 13.1J | 16 | 103J* | 21.3J* | 91.6* | 15.5 | 34.3* |
| Cobalt | 9.1 | 22.8* | 8.9 | 10.6* | 10.4* | 8.6 | 13.9* | 10.5* | 18.3* | 10.1* | NR |
| Copper | 27.6 | 62.3J* | 14J | 11.9J | 12.1J | 12.8 | 103J* | 18J | 104* | 12.6 | NR |
| Iron | 28200 | 36100* | 16600 | 15400 | 16500 | 15800 | 28100 | 22500 | 41800* | 17200 | NR |
| Lead | 27.4 | 40.8* | 14.3 | 15.9 | 14.2 | 12.3 | 35.9* | 16.8 | 52.4J* | 13.5J | NR |
| Magnesium | 2760 | 3200* | 1710 | 1340 | 1950 | 2140 | 2840* | 2780* | 3090J* | 1770Ј | NR |
| Manganese | 1950 | 999 | 279 | 445 | 332 | 287 | 734 | 596 | 712 | 916 | NR |
| Mercury | 0.059 | <0.64U | <0.14U | 0.036J | 0.031J | <0.13U | 0.082J* | 0.054J | 0.15J* | <0.14U | NR |
| Nickel | 17.7 | 35.5* | 13.6 | 12.7 | 15.3 | 15.3 | 30.2* | 21.4* | 41.1J* | 14.8J | NR |
| Potassium | 1950 | 1740 | 655 | 541 | 724 | 673 | 1440 | 989 | 1680J | 534J | NR |
| Selenium | 1.7 | 2.7J* | 0.77J | 0.79J | 0.86J | 1.1J | 2.2J* | 1J | 2.9J* | 0.97 | NR |
| Silver | 0 | 6.2* | 0.25J* | 1.8* | 0.71J* | 0.052J* | 1.7J* | 0.075J* | 7.3* | 0.09J* | NR |
| Sodium | 112 | 178J* | 56.9J | 73J | 85.5J | 47.3J | 142J* | 86.9J | 142J* | 58.4J | NR |
| Thallium | 0.89 | <1.3U | 0.15J | 0.16J | 0.15J | 0.13J | 0.41J | 0.21J | 0.37J | 0.13J | NR |
| Vanadium | 26.1 | 26.1 | 16.4 | 12.9 | 16.7 | 15.3 | 23 | 21 | 27.6* | 15 | NR |
| Zinc | 532 | 384 | 65.5 | 104 | 73.7 | 53.1J | 267 | 91.8 | 458 | 63.9 | NR |
| | | | Γ | Γ | | ves (mg/kg) | 1 | 1 | 1 | 1 | |
| 1,3-Dinitrobenzene | None | <0.25U | <0.25U | <0.26U | <0.25U | <0.25U | <0.26U | <0.24U | 0.036J* | <0.25U | NR |
| 2,4,6-Trinitrotoluene | None | <0.25U | <0.25U | 0.088J* | <0.25U | <0.25U | <0.26U | <0.24U | 0.15J* | <0.25U | NR |
| 4-Amino-2,6- | | | | | | | | | | | |
| Dinitrotoluene | None | <0.25U | <0.25U | <0.26U | <0.25U | <0.25U | <0.26U | <0.24U | 0.12J* | <0.25U | NR |
| HMX | None | <0.25U | <0.25U | <0.26U | <0.25U | 0.017J* | <0.26U | 0.015J* | 0.083J* | <0.25U | NR |
| Nitrocellulose | None | 10.4J* | <7.2U | 3.1J* | <7.6U | NR | 7.8J* | 1.9J* | 5.7J* | <7.1U | NR |
| Tetryl | None | 0.024J* | <0.25U | <0.26U | <0.25U | <0.25U | 0.022J* | <0.24U | 0.019J* | <0.25U | NR |
| | | | ſ | ſ | | s (mg/kg) | 1 | 1 | 1 | Γ | 1 |
| 2-Methylnaphthalene | None | <2.1U | <0.48U | <0.58U | <0.5U | <0.44U | <1.9U | 0.025J* | <2.1UJ | <0.47U | NR |
| Acenaphthene | None | <0.32U | <0.072U | <0.088U | <0.076U | <0.066U | <0.29U | 0.0099J* | <0.32U | <0.071U | NR |
| Acenaphthylene | None | <0.32U | <0.072U | <0.088U | <0.076U | <0.066U | <0.29U | 0.091* | <0.32U | 0.02J* | NR |
| Anthracene | None | 0.074J* | <0.072U | <0.088U | <0.076U | <0.066U | <0.29U | 0.07J* | <0.32U | 0.01J* | NR |
| Benz(a)anthracene | None | 0.38* | 0.027J* | 0.058J* | 0.015J* | 0.021J* | 0.045J* | 0.34* | 0.078J* | 0.05J* | NR |
| Benzo(a)pyrene | None | 0.39* | 0.032J* | 0.063J* | 0.012J* | 0.02J* | 0.052J* | 0.47* | 0.12J* | 0.063J* | NR |
| Benzo(b)fluoranthene | None | 0.52* | 0.064J* | 0.1* | 0.025J* | 0.038J* | 0.075J* | 0.8* | 0.19J* | 0.11* | NR |
| Benzo(ghi)perylene | None | 0.27J* | 0.029J* | 0.05J* | 0.011J* | 0.014J* | 0.045J* | 0.47* | 0.11J* | 0.061J* | NR |
| Benzo(k)fluoranthene | None | 0.26J* | 0.027J* | 0.044J* | <0.076U | 0.013J* | <0.29U | 0.21* | 0.067J* | 0.036J* | NR |
| Bis(2-ethylhexyl)phthalate | None | 0.16J* | <0.48U | <0.58U | <0.5U | <0.44U | <1.9U | <0.48U | <2.1UJ | <0.47U | NR |
| Chrysene | None | 0.43* | 0.048J* | 0.06J* | 0.014J* | 0.027J* | 0.052J* | 0.4* | 0.11J* | 0.062J* | NR |

| Area | | Backwater Area | Backwater Area | Backwater Area | Backwater Area | Lower Cobbs Pond | Lower Cobbs Pond | Lower Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond | Upper Cobbs Pond |
|------------------------|------------|-------------------------|-------------------------|-------------------------|-------------------------|-------------------|-------------------------|-------------------------|-------------------------|-------------------------|------------------|
| Station | | CPCsd-047 | CPCsd-047 | CPCsd-048 | CPCsd-048 | CPCsd-044 | CPCsd-045 | CPCsd-045 | CPCsd-046 | CPCsd-046 | CPCsd-049 |
| | - | CPCsd-047-5025- | CPCsd-047-5785- | CPCsd-048-5026- | CPCsd-048-5786- | CPCsd-044-5022- | CPCsd-045-5023- | CPCsd-045-5783- | CPCsd-046-5024- | CPCsd-046-5784- | CPCsd-049-5032- |
| Sample ID | | SD | SD | SD | SD | SD | SD | SD | SD | SD | SD |
| Date | | 04/01/10 | 04/01/10 | 04/01/10 | 04/01/10 | 03/29/10 | 04/01/10 | 04/01/10 | 03/25/10 | 03/25/10 | 03/25/10 |
| Depth (ft) | | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 | 0.5 - 2.0 | 0.0 - 0.5 |
| Parameters Analyzed | Background | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | |
| Analyte | Criteria | analytes | analytes | analytes | analytes | Explosives, SVOCs | analytes | analytes | analytes | analytes | TAL Metals |
| Di-n-butyl phthalate | None | <2.1U | <0.48U | <0.58U | 0.034J* | 0.022J* | <1.9U | 0.023J* | <2.1UJ | <0.47U | NR |
| Dibenz(a,h)anthracene | None | 0.066J* | <0.072U | <0.088U | <0.076U | <0.066U | <0.29U | 0.12* | <0.32U | <0.071U | NR |
| Fluoranthene | None | 0.96* | 0.07J* | 0.11* | 0.022J* | 0.042J* | 0.095J* | 0.48* | 0.22J* | 0.087* | NR |
| Fluorene | None | 0.053J* | <0.072U | <0.088U | <0.076U | <0.066U | <0.29U | 0.02J* | <0.32U | <0.071U | NR |
| Indeno(1,2,3-cd)pyrene | None | 0.23J* | 0.026J* | 0.04J* | <0.076U | 0.014J* | 0.042J* | 0.43* | 0.09J* | 0.054J* | NR |
| Naphthalene | None | <0.32U | <0.072U | <0.088U | <0.076U | <0.066U | <0.29U | 0.038J* | <0.32U | <0.071U | NR |
| Phenanthrene | None | 0.35* | 0.027J* | 0.034J* | 0.01J* | 0.014J* | <0.29U | 0.15* | 0.07J* | 0.024J* | NR |
| Pyrene | None | 0.76* | 0.053J* | 0.084J* | 0.019J* | 0.032J* | 0.083J* | 0.41* | 0.18J* | 0.069J* | NR |
| | | | | | PCBs | (mg/kg) | | | | | |
| delta-BHC | None | <0.13UJ | <0.029U | <0.035U | <0.03UJ | NR | <0.12U | <0.029U | <0.13UJ | 0.0018J* | NR |
| | | | | | VOCs | (mg/kg) | | | | | |
| 2-Butanone | None | 0.055J* | 0.013J* | 0.03J* | 0.01J* | NR | 0.047J* | 0.012J* | 0.033J* | 0.0072J* | NR |
| Acetone | None | 0.19* | <0.055U | 0.12* | <0.051U | NR | 0.17* | <0.052U | 0.091J* | <0.028UJ | NR |
| Carbon disulfide | None | <0.032U | <0.0072U | <0.0088U | <0.0076U | NR | 0.0033J* | <0.0072U | <0.032UJ | <0.0071UJ | NR |

Table 4–18. Analytes Detected in PBA08 RI Sediment Samples (continued)

^a Only detected site-related contaminants are presented in the table. ^b Background concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b). BHC = Hexachlorocy clohexane. ft = Feet.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

ID = Identification.

J = Estimated value less than reporting limits. mg/kg = M illigrams per kilogram.

NR = Not reported/not analyzed.

PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

RVAAP = Ravenna Army Ammunition Plant.

SVOC= Semi-volatile organic compound.

TAL = Target analyte list.U = Not detected.

UJ = Not detected, reporting limit estimated.

VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

< = Less than.

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| Location | Affected Sample | Date Sampled | Change/Rationale |
|-----------|-------------------|-----------------|--|
| | CPCSB-032-5113-SO | 3/24/2010 | |
| | CPCSB-032-5114-SO | 3/24/2010 | Leastion married due to standing |
| CPCsb-032 | CPCSB-032-5115-SO | 3/24/2010 | Location moved due to standing water. |
| | CPCSB-032-5116-SO | 3/24/2010 | water. |
| | CPCSB-032-6073-FD | 3/24/2010 | |
| CPCsb-033 | CPCSB-033-5117-SO | 3/29/2010 | Location moved due to standing |
| CPCSD-055 | CPCSB-033-5118-SO | 3/29/2010 | water. |
| CDC-1 024 | CPCSB-034-5119-SO | 3/29/2010 | Location moved due to standing |
| CPCsb-034 | CPCSB-034-5120-SO | 3/29/2010 | water. |
| | CPCSW-048-5031-SW | 4/1/2010 | Location moved approximately |
| CPCsw-048 | CPCSD-048-5786-SD | 4/1/2010 | 30 ft to the south due to access |
| | CPCSD-048-5026-FD | 4/1/2010 | issues. |

Table 4–19. Changes from the PBA08 SAP

ft = Feet. PBA08 SAP = Performance-Based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1 (USACE 2009).

| Surface Soil | | Subsurface soil | Sediment | Surface Water | | -Unconsolidated ng/L) | Groundwater | -Bedrock (mg/L) |
|--------------|---------|-----------------|----------|---------------|----------|--------------------------|-------------|-----------------|
| Chemical | (mg/kg) | (mg/kg) | (mg/kg) | (mg/L) | Filtered | Unfiltered | Filtered | Unfiltered |
| Aluminum | 17700 | 19500 | 13900 | 3.37 | NA | 48 | NA | 9.41 |
| Antimony | 0.96 | 0.96 | 0 | 0 | 0 | 0.0043 | 0 | 0 |
| Arsenic | 15.4 | 19.8 | 19.5 | 0.0032 | 0.0117 | 0.215 | 0 | 0.0191 |
| Barium | 88.4 | 124 | 123 | 0.0475 | 0.0821 | 0.327 | 0.256 | 0.241 |
| Beryllium | 0.88 | 0.88 | 0.38 | 0 | 0 | 0 | 0 | 0 |
| Cadmium | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calcium | 15800 | 35500 | 5510 | 41.4 | 115 | 194 | 53.1 | 48.2 |
| Chromium | 17.4 | 27.2 | 18.1 | 0 | 0.0073 | 0.0852 | 0 | 0.0195 |
| Cobalt | 10.4 | 23.2 | 9.1 | 0 | 0 | 0.0463 | 0 | 0 |
| Copper | 17.7 | 32.3 | 27.6 | 0.0079 | 0 | 0.289 | 0 | 0.017 |
| Cyanide | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Iron | 23100 | 35200 | 28200 | 2.56 | 0.279 | 195 | 1.43 | 21.5 |
| Lead | 26.1 | 19.1 | 27.4 | 0 | 0 | 0.183 | 0 | 0.023 |
| Magnesium | 3030 | 8790 | 2760 | 10.8 | 43.3 | 58.4 | 15 | 13.7 |
| Manganese | 1450 | 3030 | 1950 | 0.391 | 1.02 | 2.86 | 1.34 | 1.26 |
| Mercury | 0.036 | 0.044 | 0.059 | 0 | 0 | 0.00025 | 0 | 0 |
| Nickel | 21.1 | 60.7 | 17.7 | 0 | 0 | 0.117 | 0.0834 | 0.0853 |
| Potassium | 927 | 3350 | 1950 | 3.17 | 2.89 | 7.48 | 5.77 | 6.06 |
| Selenium | 1.4 | 1.5 | 1.7 | 0 | 0 | 0.0057 | 0 | 0 |
| Silver | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sodium | 123 | 145 | 112 | 21.3 | 45.7 | 44.7 | 51.4 | 49.7 |
| Thallium | 0 | 0.91 | 0.89 | 0 | 0 | 0.0024 | 0 | 0 |
| Vanadium | 31.1 | 37.6 | 26.1 | 0 | 0 | 0.0981 | 0 | 0.0155 |
| Zinc | 61.8 | 93.3 | 532 | 0.042 | 0.0609 | 0.888 | 0.0523 | 0.193 |

Table 4–20. RVAAP Background Concentrations

Background concentrations were developed in 1998 and published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b). These background values are currently being reassessed, but the background valued developed in 1998 are used throughout this report.

mg/kg = M illigrams per kilogram.

mg/L = M illigrams per liter.

NA = Not available. Aluminum results were rejected in validation.

RVAAP = Ravenna Army Ammunition Plant.

| Essential Human Nutrient | USDA RDA/RDI ^a Value |
|--------------------------|---------------------------------|
| Calcium | 1000 mg/d |
| Chloride ^b | 3400 mg/d |
| Iodine | 150 μg/d |
| Iron | 8 mg/d |
| Magnesium | 400 mg/d |
| Potassium ^b | 4700 mg/d |
| Phosphorous | 700 mg/d |
| Sodium ^b | 2300 mg/d |

 Table 4–21. Recommended Dietary Allowances/Reference Daily Intake Values

Source= Values were obtained from http://fnic.nal.usda.gov charts.

^a Dietary reference intakes vary by gender and age, values present are for life stage group: Males 19-30 years.

^bAdequate intake value.

 $\mu g/d = M$ icrograms per day.

mg/d = M illigram per day.

RDA= Recommended dietary allowance.

RDI= Reference daily intake.

USDA = U.S. Department of Agriculture.

| | | | Minimum | Maaia | | Dealermand | | |
|---|------------|-----------------------|----------|-------------------|-------------------|-------------------------------------|------|--------------------|
| Analyte (mg/kg) | CAS Number | Freq of Detect | Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | SRC Justification |
| Analyte (mg/kg) | CAS Number | Freq of Detect | Metals | | Kesuit | Criteria | SKC: | SIC JUSTIICATION |
| Aluminum | 7429-90-5 | 24/24 | 6450 | 19200 | 11900 | 17700 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 10/17 | 0.086 | 0.46 | 0.22 | 0.96 | No | Below background |
| Arsenic | 7440-38-2 | 24/24 | 4.2 | 28.4 | 11.1 | 15.4 | Yes | Exceeds background |
| Barium | 7440-39-3 | 24/24 | 35.2 | 90.4 | 61 | 88.4 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 24/24 | 0.36 | 0.86 | 0.524 | 0.88 | No | Below background |
| Cadmium | 7440-41-7 | <u>24/24</u> 19/24 | 0.30 | 0.80 | 0.324 | 0.88 | Yes | Exceeds background |
| Calcium | 7440-43-9 | 24/24 | 195 | 4040 | 1310 | 15800 | No | Essential Nutrient |
| | | 24/24 24/24 | <u> </u> | | | | | |
| Chromium | 7440-47-3 | | | 24.8 | 15.4 | 17.4 | Yes | Exceeds background |
| Chromium, hexavalent | 18540-29-9 | 1/1 | 1 | 1 | 1 | 10.4 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 24/24 | 4.5 | 16.4 | 8.9 | 10.4 | Yes | Exceeds background |
| Copper | 7440-50-8 | 24/24 | 4.2 | 23.9 | 14.1 | 17.7 | Yes | Exceeds background |
| Cyanide | 57-12-5 | 4/11 | 0.31 | 1.2 | 0.292 | 0 | Yes | Exceeds background |
| Iron | 7439-89-6 | 24/24 | 13400 | 44000 | 23000 | 23100 | No | Essential Nutrient |
| Lead | 7439-92-1 | 24/24 | 10.6 | 31.3 | 18.7 | 26.1 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 24/24 | 1340 | 4310 | 2560 | 3030 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 24/24 | 78.6 | 667 | 281 | 1450 | No | Below background |
| Mercury | 7439-97-6 | 21/24 | 0.021 | 0.074 | 0.043 | 0.036 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 24/24 | 8.6 | 27.2 | 17 | 21.1 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 24/24 | 508 | 2220 | 1020 | 927 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 21/24 | 0.48 | 1.4 | 0.813 | 1.4 | Yes | Exceeds background |
| Silver | 7440-22-4 | 9/24 | 0.032 | 0.45 | 0.0996 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 24/24 | 38.4 | 159 | 74 | 123 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 14/24 | 0.11 | 0.25 | 0.173 | 0 | Yes | Exceeds background |
| Vanadium | 7440-62-2 | 24/24 | 14.1 | 35.3 | 20.9 | 31.1 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 24/24 | 47.7 | 121 | 63.9 | 61.8 | Yes | Exceeds background |
| | • | • | Anions | | | - | • | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 1/11 | 1.4 | 1.4 | 0.545 | None | Yes | Exceeds background |
| | | | Explosiv | es | | | | |
| Nitrocellulose | 9004-70-0 | 3/3 | 1.2 | 1.5 | 1.4 | None | Yes | Detected organic |

| Analyte (mg/kg) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | SRC Justification | | | | |
|----------------------------|------------|----------------|-------------------|-------------------|-------------------|-------------------------------------|------|-------------------|--|--|--|--|
| | SVOCs | | | | | | | | | | | |
| 2-Methylnaphthalene | 91-57-6 | 1/14 | 0.0091 | 0.0091 | 0.293 | None | Yes | Detected organic | | | | |
| Acenaphthylene | 208-96-8 | 1/14 | 0.012 | 0.012 | 0.0459 | None | Yes | Detected organic | | | | |
| Benz(a)anthracene | 56-55-3 | 3/14 | 0.017 | 0.066 | 0.0484 | None | Yes | Detected organic | | | | |
| Benzenemethanol | 100-51-6 | 1/13 | 0.44 | 0.44 | 0.337 | None | Yes | Detected organic | | | | |
| Benzo(a)pyrene | 50-32-8 | 3/14 | 0.014 | 0.057 | 0.0482 | None | Yes | Detected organic | | | | |
| Benzo(b)fluoranthene | 205-99-2 | 2/14 | 0.021 | 0.059 | 0.0506 | None | Yes | Detected organic | | | | |
| Benzo(ghi)perylene | 191-24-2 | 1/14 | 0.031 | 0.031 | 0.0519 | None | Yes | Detected organic | | | | |
| Benzo(k)fluoranthene | 207-08-9 | 1/14 | 0.028 | 0.028 | 0.0499 | None | Yes | Detected organic | | | | |
| Benzoic acid | 65-85-0 | 1/13 | 17 | 17 | 2.05 | None | Yes | Detected organic | | | | |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | 1/14 | 0.23 | 0.23 | 0.304 | None | Yes | Detected organic | | | | |
| Chrysene | 218-01-9 | 2/14 | 0.014 | 0.048 | 0.0462 | None | Yes | Detected organic | | | | |
| Di-n-butyl phthalate | 84-74-2 | 2/14 | 0.021 | 0.027 | 0.27 | None | Yes | Detected organic | | | | |
| Fluoranthene | 206-44-0 | 7/14 | 0.012 | 0.11 | 0.0549 | None | Yes | Detected organic | | | | |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/14 | 0.026 | 0.026 | 0.0494 | None | Yes | Detected organic | | | | |
| Phenanthrene | 85-01-8 | 4/14 | 0.0095 | 0.093 | 0.0506 | None | Yes | Detected organic | | | | |
| Pyrene | 129-00-0 | 7/14 | 0.011 | 0.079 | 0.0517 | None | Yes | Detected organic | | | | |
| | | | Pesticides/P | CBs | | | | | | | | |
| beta-BHC | 319-85-7 | 1/3 | 0.0035 | 0.0035 | 0.00193 | None | Yes | Detected organic | | | | |

Table 4-22. SRC Screening for Discrete Shallow Surface Soil (0-1 ft bgs) Samples at Cobbs Ponds (continued)

^a Background concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b). Site-related contaminant screening tables include all available and appropriate data as presented in Section 4.4.4.

bgs = Below ground surface.

BHC = Hexachlorocyclohexane.

CAS = Chemical Abstract Service.

ft = Feet.

1

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound

| | CAS | Freq of | Minimum | Maximum | Average | Background | | | | | | | |
|-----------------|-----------|---------|----------|---------------|---------|-----------------------|------|--------------------|--|--|--|--|--|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification | | | | | |
| | Metals | | | | | | | | | | | | |
| Aluminum | 7429-90-5 | 6/6 | 7120 | 11000 | 9920 | 19500 | No | Below background | | | | | |
| Antimony | 7440-36-0 | 3/5 | 0.084 | 0.13 | 0.182 | 0.96 | No | Below background | | | | | |
| Arsenic | 7440-38-2 | 6/6 | 9.7 | 17.9 | 14.6 | 19.8 | No | Below background | | | | | |
| Barium | 7440-39-3 | 6/6 | 38.1 | 63.2 | 52.5 | 124 | No | Below background | | | | | |
| Beryllium | 7440-41-7 | 6/6 | 0.41 | 0.61 | 0.522 | 0.88 | No | Below background | | | | | |
| Cadmium | 7440-43-9 | 6/6 | 0.032 | 0.18 | 0.069 | 0 | Yes | Exceeds background | | | | | |
| Calcium | 7440-70-2 | 6/6 | 1080 | 8240 | 4310 | 35500 | No | Essential Nutrient | | | | | |
| Chromium | 7440-47-3 | 6/6 | 10.5 | 18.5 | 15.1 | 27.2 | No | Below background | | | | | |
| Cobalt | 7440-48-4 | 6/6 | 6.6 | 14.1 | 10.9 | 23.2 | No | Below background | | | | | |
| Copper | 7440-50-8 | 6/6 | 12.8 | 23 | 19.3 | 32.3 | No | Below background | | | | | |
| Iron | 7439-89-6 | 6/6 | 21300 | 33600 | 27900 | 35200 | No | Essential Nutrient | | | | | |
| Lead | 7439-92-1 | 6/6 | 10.9 | 23.3 | 13.8 | 19.1 | Yes | Exceeds background | | | | | |
| Magnesium | 7439-95-4 | 6/6 | 2110 | 7120 | 3960 | 8790 | No | Essential Nutrient | | | | | |
| Manganese | 7439-96-5 | 6/6 | 181 | 541 | 340 | 3030 | No | Below background | | | | | |
| Mercury | 7439-97-6 | 1/6 | 0.022 | 0.022 | 0.0545 | 0.044 | No | Below background | | | | | |
| Nickel | 7440-02-0 | 6/6 | 14.6 | 31.7 | 24.4 | 60.7 | No | Below background | | | | | |
| Potassium | 7440-09-7 | 6/6 | 856 | 2190 | 1440 | 3350 | No | Essential Nutrient | | | | | |
| Selenium | 7782-49-2 | 6/6 | 0.88 | 1.6 | 1.15 | 1.5 | Yes | Exceeds background | | | | | |
| Silver | 7440-22-4 | 2/6 | 0.013 | 0.029 | 0.0126 | 0 | Yes | Exceeds background | | | | | |
| Sodium | 7440-23-5 | 6/6 | 40.8 | 110 | 61.2 | 145 | No | Essential Nutrient | | | | | |
| Thallium | 7440-28-0 | 6/6 | 0.093 | 0.17 | 0.144 | 0.91 | No | Below background | | | | | |
| Vanadium | 7440-62-2 | 6/6 | 12.1 | 17.9 | 16.4 | 37.6 | No | Below background | | | | | |
| Zinc | 7440-66-6 | 6/6 | 57.2 | 65.4 | 61.3 | 93.3 | No | Below background | | | | | |
| | | | Explosiv | es/Propellant | 'S | | | | | | | | |
| Nitrocellulose | 9004-70-0 | 1/2 | 1.6 | 1.6 | 2.35 | None | Yes | Detected organic | | | | | |

Table 4-23. SRC Screening for Discrete Subsurface Soil (1-13 ft bgs) Samples at Cobbs Ponds

| | CAS | Freq of | Minimum | Maximum | Average | Background | | | | | |
|--|----------|---------|---------|---------|---------|-----------------------|------|-------------------------|--|--|--|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification | | | |
| | SVOCs | | | | | | | | | | |
| 2-Methylnaphthalene 91-57-6 2/6 0.0091 0.014 0.138 None Yes Detected organic | | | | | | | | | | | |
| Benz(a)anthracene | 56-55-3 | 1/6 | 0.047 | 0.047 | 0.033 | None | Yes | Detected organic | | | |
| Benzo(a)pyrene | 50-32-8 | 1/6 | 0.066 | 0.066 | 0.0362 | None | Yes | Detected organic | | | |
| Benzo(b)fluoranthene | 205-99-2 | 1/6 | 0.08 | 0.08 | 0.0385 | None | Yes | Detected organic | | | |
| Benzo(ghi)perylene | 191-24-2 | 1/6 | 0.06 | 0.06 | 0.0352 | None | Yes | Detected organic | | | |
| Benzo(k)fluoranthene | 207-08-9 | 1/6 | 0.03 | 0.03 | 0.0302 | None | Yes | Detected organic | | | |
| Chrysene | 218-01-9 | 1/6 | 0.063 | 0.063 | 0.0357 | None | Yes | Detected organic | | | |
| Di-n-butyl phthalate | 84-74-2 | 1/6 | 0.019 | 0.019 | 0.17 | None | Yes | Detected organic | | | |
| Dibenz(a,h)anthracene | 53-70-3 | 1/6 | 0.021 | 0.021 | 0.0287 | None | Yes | Detected organic | | | |
| Fluoranthene | 206-44-0 | 1/6 | 0.11 | 0.11 | 0.0435 | None | Yes | Detected organic | | | |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/6 | 0.037 | 0.037 | 0.0313 | None | Yes | Detected organic | | | |
| Phenanthrene | 85-01-8 | 2/6 | 0.014 | 0.051 | 0.0312 | None | Yes | Detected organic | | | |
| Pyrene | 129-00-0 | 1/6 | 0.088 | 0.088 | 0.0398 | None | Yes | Detected organic | | | |

Table 4-23. SRC Screening for Discrete Subsurface Soil (1-13 ft bgs) Samples at Cobbs Ponds (continued)

^a Background concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b). SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

bgs = Below ground surface.

CAS = Chemical Abstract Service.

ft = Feet.

1

mg/kg = Milligrams per kilogram.

SRC = Site-related contaminant.

SVOC = Semi-volatile organic compound.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|---|------------|---------|---------|-----------|---------|-----------------------|------|--------------------|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | 20000 | | Metals | 100 000 | 01100110 | SHOT | |
| Aluminum | 7429-90-5 | 14/14 | 6620 | 18600 | 12500 | 13900 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 11/14 | 0.24 | 2.1 | 0.607 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 14/14 | 3.5 | 16.7 | 10.2 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 14/14 | 53.8 | 151 | 99.5 | 123 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 14/14 | 0.34 | 1.1 | 0.782 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 14/14 | 0.11 | 2.3 | 0.972 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 14/14 | 1180 | 15400 | 3790 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 14/14 | 9.9 | 56.5 | 23.3 | 18.1 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 14/14 | 5.4 | 22.8 | 13.6 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 14/14 | 10.9 | 62.3 | 28.2 | 27.6 | Yes | Exceeds background |
| Cyanide | 57-12-5 | 3/12 | 0.1 | 0.55 | 0.224 | 0 | Yes | Exceeds background |
| Iron | 7439-89-6 | 14/14 | 12700 | 36100 | 21400 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 14/14 | 14.2 | 57.9 | 29.6 | 27.4 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 14/14 | 1170 | 3830 | 2230 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 14/14 | 179 | 999 | 520 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 13/14 | 0.031 | 0.11 | 0.0816 | 0.059 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 14/14 | 9.4 | 35.5 | 20.5 | 17.7 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 14/14 | 541 | 1740 | 1140 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 8/14 | 0.26 | 2.7 | 0.5 | 1.7 | Yes | Exceeds background |
| Silver | 7440-22-4 | 13/14 | 0.51 | 23 | 5.7 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 14/14 | 55.2 | 239 | 140 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 1/14 | 0.16 | 0.16 | 0.196 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 14/14 | 11.7 | 30 | 19.9 | 26.1 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 14/14 | 74.8 | 418 | 179 | 532 | No | Below background |
| | | | | Anions | • | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 1/12 | 2.2 | 2.2 | 0.804 | None | Yes | Exceeds background |
| · · · · · · · · · · · · · · · · · · · | | | E. | xplosives | | | • | |
| 2,4,6-Trinitrotoluene | 118-96-7 | 6/14 | 0.049 | 0.32 | 0.0811 | None | Yes | Detected organic |
| 2,6-Dinitrotoluene | 606-20-2 | 1/14 | 0.16 | 0.16 | 0.0483 | None | Yes | Detected organic |
| Nitrocellulose | 9004-70-0 | 4/4 | 2.5 | 10.4 | 4.75 | None | Yes | Detected organic |
| Tetryl | 479-45-8 | 1/14 | 0.024 | 0.024 | 0.0294 | None | Yes | Detected organic |

| Table 4–24. SRC Screening f | or Discrete Sediment | Samples at the Backwater Area |
|-----------------------------|----------------------|-------------------------------|
| | | |

| Analyte (mg/kg) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | SRC Justification | | | | |
|----------------------------|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------------------|------|-------------------|--|--|--|--|
| Analyte (ing/kg) | Tumber | Detect | | SVOCs | Kesuit | CITICITA | SKC: | SIC Justification | | | | |
| | | | | | | | | | | | | |
| Anthracene | 120-12-7 | 2/5 | 0.074 | 0.22 | 0.0956 | None | Yes | Detected organic | | | | |
| Benz(a)anthracene | 56-55-3 | 3/5 | 0.058 | 0.83 | 0.274 | None | Yes | Detected organic | | | | |
| Benzo(a)pyrene | 50-32-8 | 4/5 | 0.063 | 0.89 | 0.323 | None | Yes | Detected organic | | | | |
| Benzo(b)fluoran the ne | 205-99-2 | 3/5 | 0.1 | 0.86 | 0.338 | None | Yes | Detected organic | | | | |
| Benzo(ghi)perylene | 191-24-2 | 3/5 | 0.05 | 0.49 | 0.22 | None | Yes | Detected organic | | | | |
| Benzo(k)fluoranthene | 207-08-9 | 3/5 | 0.044 | 0.8 | 0.265 | None | Yes | Detected organic | | | | |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | 1/5 | 0.16 | 0.16 | 0.165 | None | Yes | Detected organic | | | | |
| Chrysene | 218-01-9 | 4/5 | 0.06 | 0.93 | 0.338 | None | Yes | Detected organic | | | | |
| Dibenz(a,h)anthr acene | 53-70-3 | 1/5 | 0.066 | 0.066 | 0.095 | None | Yes | Detected organic | | | | |
| Fluoranthene | 206-44-0 | 3/5 | 0.11 | 1.8 | 0.61 | None | Yes | Detected organic | | | | |
| Fluorene | 86-73-7 | 1/5 | 0.053 | 0.053 | 0.0844 | None | Yes | Detected organic | | | | |
| Indeno(1,2,3-c d)pyrene | 193-39-5 | 3/5 | 0.04 | 0.48 | 0.192 | None | Yes | Detected organic | | | | |
| Phenanthrene | 85-01-8 | 3/5 | 0.034 | 0.91 | 0.285 | None | Yes | Detected organic | | | | |
| Pyrene | 129-00-0 | 3/5 | 0.084 | 1.9 | 0.604 | None | Yes | Detected organic | | | | |
| | | | Pesti | icides/PCBs | • | | | | | | | |
| PCB-1254 | 11097-69-1 | 2/4 | 0.046 | 0.047 | 0.0568 | None | Yes | Detected organic | | | | |
| | | | | VOCs | | | | - | | | | |
| 2-Butanone | 78-93-3 | 3/5 | 0.03 | 0.055 | 0.0256 | None | Yes | Detected organic | | | | |
| Acetone | 67-64-1 | 5/5 | 0.02 | 0.19 | 0.12 | None | Yes | Detected organic | | | | |
| Methylene chloride | 75-09-2 | 1/5 | 0.048 | 0.048 | 0.0145 | None | Yes | Detected organic | | | | |

Table 4–24. SRC Screening for Discrete Sediment Samples at the Backwater Area (continued)

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

| Analyte (mg/kg) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteriaª | SRC? | SRC Justification |
|-----------------|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------|------|--------------------|
| | | | 1 | Metals | | | | |
| Aluminum | 7429-90-5 | 2/2 | 9120 | 9150 | 9140 | 13900 | No | Below background |
| Antimony | 7440-36-0 | 2/2 | 0.15 | 0.17 | 0.16 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 2/2 | 4.9 | 6.4 | 5.65 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 2/2 | 73.9 | 84.2 | 79.1 | 123 | No | Below background |
| Beryllium | 7440-41-7 | 2/2 | 0.58 | 0.59 | 0.585 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 2/2 | 0.24 | 0.28 | 0.26 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 2/2 | 1010 | 3550 | 2280 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 2/2 | 13.1 | 16.6 | 14.9 | 18.1 | No | Below background |
| Cobalt | 7440-48-4 | 2/2 | 8.9 | 10.4 | 9.65 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 2/2 | 12.1 | 14 | 13.1 | 27.6 | No | Below background |
| Iron | 7439-89-6 | 2/2 | 16500 | 16600 | 16600 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 2/2 | 14.2 | 14.3 | 14.3 | 27.4 | No | Below background |
| Magnesium | 7439-95-4 | 2/2 | 1710 | 1950 | 1830 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 2/2 | 279 | 332 | 306 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 1/2 | 0.031 | 0.031 | 0.0505 | 0.059 | No | Below background |
| Nickel | 7440-02-0 | 2/2 | 13.6 | 15.3 | 14.5 | 17.7 | No | Below background |
| Potassium | 7440-09-7 | 2/2 | 655 | 724 | 690 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 2/2 | 0.77 | 0.86 | 0.815 | 1.7 | No | Below background |
| Silver | 7440-22-4 | 2/2 | 0.25 | 0.71 | 0.48 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 2/2 | 56.9 | 85.5 | 71.2 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 2/2 | 0.15 | 0.15 | 0.15 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 2/2 | 16.4 | 16.7 | 16.6 | 26.1 | No | Below background |
| Zinc | 7440-66-6 | 2/2 | 65.5 | 73.7 | 69.6 | 532 | No | Below background |

Table 4-25. SRC Screening for Discrete Subsurface Sediment Samples at the Backwater Area

| | CAS | Freq of | Minimum | Maximum | Average | Background | | | | | |
|---|----------|---------|---------|---------|---------|-----------------------|------|-------------------|--|--|--|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification | | | |
| SVOCs | | | | | | | | | | | |
| Benz(a)anthracene56-55-32/20.0150.0270.021NoneYesDetected organic | | | | | | | | | | | |
| Benzo(a)pyrene | 50-32-8 | 2/2 | 0.012 | 0.032 | 0.022 | None | Yes | Detected organic | | | |
| Benzo(b)fluoranthene | 205-99-2 | 2/2 | 0.025 | 0.064 | 0.0445 | None | Yes | Detected organic | | | |
| Benzo(ghi)perylene | 191-24-2 | 2/2 | 0.011 | 0.029 | 0.02 | None | Yes | Detected organic | | | |
| Benzo(k)fluoranthene | 207-08-9 | 1/2 | 0.027 | 0.027 | 0.0325 | None | Yes | Detected organic | | | |
| Chrysene | 218-01-9 | 2/2 | 0.014 | 0.048 | 0.031 | None | Yes | Detected organic | | | |
| Di-n-butyl phthalate | 84-74-2 | 1/2 | 0.034 | 0.034 | 0.137 | None | Yes | Detected organic | | | |
| Fluoranthene | 206-44-0 | 2/2 | 0.022 | 0.07 | 0.046 | None | Yes | Detected organic | | | |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/2 | 0.026 | 0.026 | 0.032 | None | Yes | Detected organic | | | |
| Phenanthrene | 85-01-8 | 2/2 | 0.01 | 0.027 | 0.0185 | None | Yes | Detected organic | | | |
| Pyrene | 129-00-0 | 2/2 | 0.019 | 0.053 | 0.036 | None | Yes | Detected organic | | | |
| VOCs | | | | | | | | | | | |
| 2-Butanone | 78-93-3 | 2/2 | 0.01 | 0.013 | 0.0115 | None | Yes | Detected organic | | | |

 Table 4–25. SRC Screening for Discrete Subsurface Sediment Samples at the Backwater Area (continued)

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

mg/kg = M illigrams per kilogram.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

| | CAS | Freq of | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | |
|---|------------|---------|-------------------|-------------------|-------------------|-------------------------------------|-----------|--------------------|
| Analyte | Number | Detect | (mg/kg) | (mg/kg) | (mg/kg) | (mg/kg) | (ves/no) | SRC Justification |
| Tuarye | Tumor | Dettet | (ing/kg) | Metals | (ing/kg) | (mg/ Kg) | (903/110) | Sice sustilication |
| Aluminum | 7429-90-5 | 1/1 | 9030 | 9030 | 9030 | 13900 | No | Below background |
| Antimony | 7440-36-0 | 1/1 | 1 | 1 | 1 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 1/1 | 6.5 | 6.5 | 6.5 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 1/1 | 78.9 | 78.9 | 78.9 | 123 | No | Below background |
| Beryllium | 7440-41-7 | 1/1 | 0.48 | 0.48 | 0.48 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 1/1 | 0.6 | 0.6 | 0.6 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 1/1 | 1960 | 1960 | 1960 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 1/1 | 18.6 | 18.6 | 18.6 | 18.1 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 1/1 | 7.7 | 7.7 | 7.7 | 9.1 | No | Below background |
| Iron | 7439-89-6 | 1/1 | 16200 | 16200 | 16200 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 1/1 | 17.8 | 17.8 | 17.8 | 27.4 | No | Below background |
| Magnesium | 7439-95-4 | 1/1 | 1530 | 1530 | 1530 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 1/1 | 383 | 383 | 383 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 1/1 | 0.048 | 0.048 | 0.048 | 0.059 | No | Below background |
| Nickel | 7440-02-0 | 1/1 | 14.4 | 14.4 | 14.4 | 17.7 | No | Below background |
| Potassium | 7440-09-7 | 1/1 | 858 | 858 | 858 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 1/1 | 1.1 | 1.1 | 1.1 | 1.7 | No | Below background |
| Silver | 7440-22-4 | 1/1 | 1.1 | 1.1 | 1.1 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 1/1 | 106 | 106 | 106 | 112 | No | Essential Nutrient |
| Vanadium | 7440-62-2 | 1/1 | 15.6 | 15.6 | 15.6 | 26.1 | No | Below background |
| Zinc | 7440-66-6 | 1/1 | 153 | 153 | 153 | 532 | No | Below background |
| | • | | | Anions | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 1/1 | 5.3 | 5.3 | 5.3 | None | Yes | Exceeds background |
| | | | 1 | Miscellaneous | | | | |
| Ammonia | 7664-41-7 | 1/1 | 69 | 69 | 69 | None | Yes | Exceeds background |
| Total Phosphorus as P | 7723-14-0 | 1/1 | 270 | 270 | 270 | None | Yes | Exceeds background |
| | | | | SVOCs | | | | |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | 1/1 | 0.84 | 0.84 | 0.84 | None | Yes | Detected organic |
| Butyl benzyl phthalate | 85-68-7 | 1/1 | 0.16 | 0.16 | 0.16 | None | Yes | Detected organic |
| Di-n-butyl phthalate | 84-74-2 | 1/1 | 2.7 | 2.7 | 2.7 | None | Yes | Detected organic |

Table 4-26. SRC Screening for ISM Sediment Samples at the Upper Cobbs Pond

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

SRC = Site-related contaminant.

ISM = Incremental sampling methodology. mg/kg = Milligrams per kilogram. SVOC= Semi-volatile organic compound.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|----------------------------|------------|---------|---------|-----------|---------|-----------------------|------|--------------------|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 5/ 5 | 13400 | 21400 | 16500 | 13900 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 5/5 | 0.21 | 1.9 | 0.78 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 5/5 | 5.1 | 18.8 | 12.3 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 5/ 5 | 58.7 | 157 | 106 | 123 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 5/5 | 0.44 | 1.3 | 0.842 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 3/5 | 0.15 | 2.3 | 0.95 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 5/5 | 771 | 5440 | 2780 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 6/6 | 19.9 | 91.6 | 50.8 | 18.1 | Yes | Exceeds background |
| Chromium, hexavalent | 18540-29-9 | 1/5 | 10.6 | 10.6 | 2.73 | | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 5/ 5 | 5.9 | 19.3 | 13.3 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 5/5 | 14.1 | 106 | 57.1 | 27.6 | Yes | Exceeds background |
| Iron | 7439-89-6 | 5/5 | 18300 | 41800 | 30300 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 5/5 | 13.1 | 52.4 | 27.7 | 27.4 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 5/5 | 2370 | 3690 | 3150 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 5/5 | 138 | 712 | 438 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 5/ 5 | 0.014 | 0.15 | 0.0662 | 0.059 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 5/5 | 13.1 | 41.1 | 28.7 | 17.7 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 5/5 | 1060 | 2100 | 1730 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 3/5 | 0.22 | 2.9 | 0.839 | 1.7 | Yes | Exceeds background |
| Silver | 7440-22-4 | 3/5 | 0.55 | 11 | 3.81 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 5/5 | 82.7 | 340 | 156 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 2/5 | 0.24 | 0.37 | 0.237 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 5/5 | 23.3 | 33.4 | 27.8 | 26.1 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 5/5 | 71.3 | 479 | 240 | 532 | No | Below background |
| | | | Ex | cplosives | | | | |
| 1,3-Dinitrobenzene | 99-65-0 | 1/5 | 0.036 | 0.036 | 0.0144 | None | Yes | Detected organic |
| 2,4,6-Trinitrotoluene | 118-96-7 | 2/5 | 0.087 | 0.15 | 0.0575 | None | Yes | Detected organic |
| 4-Amino-2,6-Dinitrotoluene | 19406-51-0 | 1/1 | 0.12 | 0.12 | 0.12 | None | Yes | Detected organic |
| HMX | 2691-41-0 | 1/5 | 0.083 | 0.083 | 0.0606 | None | Yes | Detected organic |
| Nitrocellulose | 9004-70-0 | 2/2 | 2 | 5.7 | 3.85 | None | Yes | Detected organic |
| Tetryl | 479-45-8 | 1/5 | 0.019 | 0.019 | 0.021 | None | Yes | Detected organic |

Table 4–27. SRC Screening for Discrete Sediment Samples at the Upper Cobbs Pond

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|-------------------------|------------|---------|---------|------------|---------|-----------------------|------|-------------------|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | | Å | SVOCs | | | | |
| Benz(a)anthracene | 56-55-3 | 1/2 | 0.078 | 0.078 | 0.0615 | None | Yes | Detected organic |
| Benzo(a)pyrene | 50-32-8 | 1/2 | 0.12 | 0.12 | 0.0845 | None | Yes | Detected organic |
| Benzo(b)fluoranthene | 205-99-2 | 1/2 | 0.19 | 0.19 | 0.14 | None | Yes | Detected organic |
| Benzo(ghi)perylene | 191-24-2 | 1/2 | 0.11 | 0.11 | 0.12 | None | Yes | Detected organic |
| Benzo(k)fluoranthene | 207-08-9 | 1/2 | 0.067 | 0.067 | 0.081 | None | Yes | Detected organic |
| Chrysene | 218-01-9 | 1/2 | 0.11 | 0.11 | 0.072 | None | Yes | Detected organic |
| Fluoranthene | 206-44-0 | 1/2 | 0.22 | 0.22 | 0.15 | None | Yes | Detected organic |
| Indeno(1,2,3-c d)pyrene | 193-39-5 | 1/2 | 0.09 | 0.09 | 0.0925 | None | Yes | Detected organic |
| Phenanthrene | 85-01-8 | 1/2 | 0.07 | 0.07 | 0.065 | None | Yes | Detected organic |
| Pyrene | 129-00-0 | 1/2 | 0.18 | 0.18 | 0.15 | None | Yes | Detected organic |
| | | | Pesti | cides/PCBs | | | | |
| PCB-1254 | 11097-69-1 | 1/2 | 0.032 | 0.032 | 0.0685 | None | Yes | Detected organic |
| | | | | VOCs | | | | |
| 2-Butanone | 78-93-3 | 2/2 | 0.022 | 0.033 | 0.0275 | None | Yes | Detected organic |
| Acetone | 67-64-1 | 2/2 | 0.088 | 0.091 | 0.0895 | None | Yes | Detected organic |

Table 4–27. SRC Screening for Discrete Sediment Samples at the Upper Cobbs Pond (continued)

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

| Analyte (mg/kg) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteriaª | SRC? | SRC Justification |
|-----------------|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------|------|--------------------|
| | | | 1 | Metals | | | | |
| Aluminum | 7429-90-5 | 1/1 | 8500 | 8500 | 8500 | 13900 | No | Below background |
| Antimony | 7440-36-0 | 1/1 | 0.19 | 0.19 | 0.19 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 1/1 | 5.5 | 5.5 | 5.5 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 1/1 | 82.3 | 82.3 | 82.3 | 123 | No | Below background |
| Beryllium | 7440-41-7 | 1/1 | 0.45 | 0.45 | 0.45 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 1/1 | 0.28 | 0.28 | 0.28 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 1/1 | 1650 | 1650 | 1650 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 1/1 | 15.5 | 15.5 | 15.5 | 18.1 | No | Below background |
| Cobalt | 7440-48-4 | 1/1 | 10.1 | 10.1 | 10.1 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 1/1 | 12.6 | 12.6 | 12.6 | 27.6 | No | Below background |
| Iron | 7439-89-6 | 1/1 | 17200 | 17200 | 17200 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 1/1 | 13.5 | 13.5 | 13.5 | 27.4 | No | Below background |
| Magnesium | 7439-95-4 | 1/1 | 1770 | 1770 | 1770 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 1/1 | 916 | 916 | 916 | 1950 | No | Below background |
| Nickel | 7440-02-0 | 1/1 | 14.8 | 14.8 | 14.8 | 17.7 | No | Below background |
| Potassium | 7440-09-7 | 1/1 | 534 | 534 | 534 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 1/1 | 0.97 | 0.97 | 0.97 | 1.7 | No | Below background |
| Silver | 7440-22-4 | 1/1 | 0.09 | 0.09 | 0.09 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 1/1 | 58.4 | 58.4 | 58.4 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 1/1 | 0.13 | 0.13 | 0.13 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 1/1 | 15 | 15 | 15 | 26.1 | No | Below background |
| Zinc | 7440-66-6 | 1/1 | 63.9 | 63.9 | 63.9 | 532 | No | Below background |

 Table 4–28. SRC Screening for Discrete Subsurface Sediment Samples at the Upper Cobbs Pond

| | CAS | Freq of | Minimum | Maximum | Average | Background | | | | |
|------------------------|----------|---------|---------|------------|---------|-----------------------|------|-------------------------|--|--|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification | | |
| | | | | SVOCs | | | | | | |
| Acenaphthylene | 208-96-8 | 1/1 | 0.02 | 0.02 | 0.02 | None | Yes | Detected organic | | |
| Anthracene | 120-12-7 | 1/1 | 0.01 | 0.01 | 0.01 | None | Yes | Detected organic | | |
| Benz(a)anthracene | 56-55-3 | 1/1 | 0.05 | 0.05 | 0.05 | None | Yes | Detected organic | | |
| Benzo(a)pyrene | 50-32-8 | 1/1 | 0.063 | 0.063 | 0.063 | None | Yes | Detected organic | | |
| Benzo(b)fluoran the ne | 205-99-2 | 1/1 | 0.11 | 0.11 | 0.11 | None | Yes | Detected organic | | |
| Benzo(ghi)perylene | 191-24-2 | 1/1 | 0.061 | 0.061 | 0.061 | None | Yes | Detected organic | | |
| Benzo(k)fluoranthene | 207-08-9 | 1/1 | 0.036 | 0.036 | 0.036 | None | Yes | Detected organic | | |
| Chrysene | 218-01-9 | 1/1 | 0.062 | 0.062 | 0.062 | None | Yes | Detected organic | | |
| Fluoranthene | 206-44-0 | 1/1 | 0.087 | 0.087 | 0.087 | None | Yes | Detected organic | | |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/1 | 0.054 | 0.054 | 0.054 | None | Yes | Detected organic | | |
| Phenanthrene | 85-01-8 | 1/1 | 0.024 | 0.024 | 0.024 | None | Yes | Detected organic | | |
| Pyrene | 129-00-0 | 1/1 | 0.069 | 0.069 | 0.069 | None | Yes | Detected organic | | |
| | | | Pestie | cides/PCBs | | | | | | |
| delta-BHC | 319-86-8 | 1/1 | 0.0018 | 0.0018 | 0.0018 | None | Yes | Detected organic | | |
| VOCs | | | | | | | | | | |
| 2-Butanone | 78-93-3 | 1/1 | 0.0072 | 0.0072 | 0.0072 | None | Yes | Detected organic | | |

Table 4–28. SRC Screening for Discrete Subsurface Sediment Samples at the Upper Cobbs Pond (continued)

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

BHC = Hexachlorocy clohexane.

CAS = Chemical Abstract Service.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

Bold indicates analyte identified as an SRC.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | | | | | |
|---|------------|---------|---------|-------------|---------|------------|------|--------------------|--|--|--|--|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteriaª | SRC? | SRC Justification | | | | |
| | | | | Metals | • | | | · | | | | |
| Aluminum 7429-90-5 1/1 15400 15400 13900 Yes Exceeds background | | | | | | | | | | | | |
| Antimony | 7440-36-0 | 1/1 | 1.3 | 1.3 | 1.3 | 0 | Yes | Exceeds background | | | | |
| Arsenic | 7440-38-2 | 1/1 | 10.2 | 10.2 | 10.2 | 19.5 | No | Below background | | | | |
| Barium | 7440-39-3 | 1/1 | 116 | 116 | 116 | 123 | No | Below background | | | | |
| Beryllium | 7440-41-7 | 1/1 | 0.83 | 0.83 | 0.83 | 0.38 | Yes | Exceeds background | | | | |
| Cadmium | 7440-43-9 | 1/1 | 1.2 | 1.2 | 1.2 | 0 | Yes | Exceeds background | | | | |
| Calcium | 7440-70-2 | 1/1 | 2550 | 2550 | 2550 | 5510 | No | Essential Nutrient | | | | |
| Chromium | 7440-47-3 | 1/1 | 121 | 121 | 121 | 18.1 | Yes | Exceeds background | | | | |
| Cobalt | 7440-48-4 | 1/1 | 12 | 12 | 12 | 9.1 | Yes | Exceeds background | | | | |
| Lead | 7439-92-1 | 1/1 | 34.2 | 34.2 | 34.2 | 27.4 | Yes | Exceeds background | | | | |
| Magnesium | 7439-95-4 | 1/1 | 2600 | 2600 | 2600 | 2760 | No | Essential Nutrient | | | | |
| Manganese | 7439-96-5 | 1/1 | 602 | 602 | 602 | 1950 | No | Below background | | | | |
| Mercury | 7439-97-6 | 1/1 | 0.1 | 0.1 | 0.1 | 0.059 | Yes | Exceeds background | | | | |
| Nickel | 7440-02-0 | 1/1 | 24.6 | 24.6 | 24.6 | 17.7 | Yes | Exceeds background | | | | |
| Potassium | 7440-09-7 | 1/1 | 1640 | 1640 | 1640 | 1950 | No | Essential Nutrient | | | | |
| Selenium | 7782-49-2 | 1/1 | 1.6 | 1.6 | 1.6 | 1.7 | No | Below background | | | | |
| Silver | 7440-22-4 | 1/1 | 1.5 | 1.5 | 1.5 | 0 | Yes | Exceeds background | | | | |
| Sodium | 7440-23-5 | 1/1 | 203 | 203 | 203 | 112 | No | Essential Nutrient | | | | |
| Vanadium | 7440-62-2 | 1/1 | 24.5 | 24.5 | 24.5 | 26.1 | No | Below background | | | | |
| Zinc | 7440-66-6 | 1/1 | 259 | 259 | 259 | 532 | No | Below background | | | | |
| | | | | Anions | | | | | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 1/1 | 8.5 | 8.5 | 8.5 | None | Yes | Exceeds background | | | | |
| | | | Mis | scellaneous | | | | | | | | |
| Ammonia | 7664-41-7 | 1/1 | 61 | 61 | 61 | None | Yes | Exceeds background | | | | |
| Total Phosphorus as P | 7723-14-0 | 1/1 | 360 | 360 | 360 | None | Yes | Exceeds background | | | | |
| | | | | SVOCs | | | | | | | | |
| Di-n-butyl phthalate | 84-74-2 | 1/1 | 1.035 | 1.035 | 1.04 | None | Yes | Detected organic | | | | |

Table 4–29. SRC Screening for ISM Sediment Samples at Lower Cobbs Pond

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

ISM = Incremental sampling methodology.

mg/kg = Milligrams per kilogram. SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

Bold indicates analyte identified as an SRC.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|----------------------|------------|---------|---------|-----------|---------|-----------------------|------|--------------------|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 8/8 | 8920 | 16400 | 12800 | 13900 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 1/8 | 1.4 | 1.4 | 0.328 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 8/8 | 5.1 | 34.3 | 14.9 | 19.5 | Yes | Exceeds background |
| Barium | 7440-39-3 | 8/8 | 51.3 | 149 | 84.8 | 123 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 8/8 | 0.42 | 0.93 | 0.679 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 6/8 | 0.1 | 1.5 | 0.536 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 8/8 | 699 | 7360 | 2670 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 8/8 | 11.9 | 150 | 59 | 18.1 | Yes | Exceeds background |
| Chromium, hexavalent | 18540-29-9 | 2/6 | 5 | 5.7 | 2.12 | | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 8/8 | 8.6 | 21.3 | 12 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 8/8 | 12.4 | 149 | 56.6 | 27.6 | Yes | Exceeds background |
| Cyanide | 57-12-5 | 1/6 | 0.4 | 0.4 | 0.225 | 0 | Yes | Exceeds background |
| Iron | 7439-89-6 | 8/8 | 15800 | 32200 | 23800 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 8/8 | 12.3 | 60.4 | 29.4 | 27.4 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 8/8 | 2140 | 3240 | 2610 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 8/8 | 217 | 966 | 502 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 7/8 | 0.016 | 0.093 | 0.0538 | 0.059 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 8/8 | 15.3 | 30.2 | 20.4 | 17.7 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 8/8 | 673 | 1860 | 1390 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 4/8 | 0.45 | 2.2 | 0.72 | 1.7 | Yes | Exceeds background |
| Silver | 7440-22-4 | 5/8 | 0.052 | 2.4 | 0.852 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 8/8 | 47.3 | 190 | 119 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 2/8 | 0.13 | 0.41 | 0.179 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 8/8 | 15.3 | 32.2 | 22.5 | 26.1 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 8/8 | 44.6 | 275 | 141 | 532 | No | Below background |
| | | | E. | xplosives | | | | |
| HMX | 2691-41-0 | 1/8 | 0.017 | 0.017 | 0.0596 | None | Yes | Detected organic |
| Nitrocellulose | 9004-70-0 | 1/1 | 7.8 | 7.8 | 7.8 | None | Yes | Detected organic |
| Tetryl | 479-45-8 | 1/8 | 0.022 | 0.022 | 0.0344 | None | Yes | Detected organic |

Table 4–30. SRC Screening for Discrete Sediment Samples at Lower Cobbs Pond

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|-------------------------|----------|---------|---------|---------|---------|-----------------------|------|-------------------|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | | | SVOCs | | | | |
| Benz(a)anthracene | 56-55-3 | 2/3 | 0.021 | 0.045 | 0.0437 | None | Yes | Detected organic |
| Benzo(a)pyrene | 50-32-8 | 2/3 | 0.02 | 0.052 | 0.0473 | None | Yes | Detected organic |
| Benzo(b)fluoran the ne | 205-99-2 | 2/3 | 0.038 | 0.075 | 0.081 | None | Yes | Detected organic |
| Benzo(ghi)perylene | 191-24-2 | 2/3 | 0.014 | 0.045 | 0.0813 | None | Yes | Detected organic |
| Benzo(k)fluoranthene | 207-08-9 | 1/3 | 0.013 | 0.013 | 0.0993 | None | Yes | Detected organic |
| Chrysene | 218-01-9 | 2/3 | 0.027 | 0.052 | 0.0423 | None | Yes | Detected organic |
| Di-n-butyl phthalate | 84-74-2 | 1/3 | 0.022 | 0.022 | 0.352 | None | Yes | Detected organic |
| Fluoranthene | 206-44-0 | 2/3 | 0.042 | 0.095 | 0.084 | None | Yes | Detected organic |
| Indeno(1,2,3-c d)pyrene | 193-39-5 | 2/3 | 0.014 | 0.042 | 0.0637 | None | Yes | Detected organic |
| Phenanthrene | 85-01-8 | 1/3 | 0.014 | 0.014 | 0.0813 | None | Yes | Detected organic |
| Pyrene | 129-00-0 | 2/3 | 0.032 | 0.083 | 0.095 | None | Yes | Detected organic |
| | | | | VOCs | | | | |
| 2-Butanone | 78-93-3 | 2/2 | 0.025 | 0.047 | 0.036 | None | Yes | Detected organic |
| Acetone | 67-64-1 | 2/2 | 0.076 | 0.17 | 0.123 | None | Yes | Detected organic |
| Carbon disulfide | 75-15-0 | 1/2 | 0.0033 | 0.0033 | 0.0029 | None | Yes | Detected organic |

Table 4-30. SRC Screening for Discrete Sediment Samples at Lower Cobbs Pond (continued)

SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

mg/kg = Milligrams per kilogram.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|-----------------|-----------|---------|---------|----------|---------|-----------------------|------|--------------------|
| Analyte (mg/kg) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | | 1 | Metals | | | | |
| Aluminum | 7429-90-5 | 1/1 | 13400 | 13400 | 13400 | 13900 | No | Below background |
| Antimony | 7440-36-0 | 1/1 | 0.15 | 0.15 | 0.15 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 1/1 | 7.7 | 7.7 | 7.7 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 1/1 | 114 | 114 | 114 | 123 | No | Below background |
| Beryllium | 7440-41-7 | 1/1 | 0.68 | 0.68 | 0.68 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 1/1 | 0.44 | 0.44 | 0.44 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 1/1 | 2110 | 2110 | 2110 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 1/1 | 21.3 | 21.3 | 21.3 | 18.1 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 1/1 | 10.5 | 10.5 | 10.5 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 1/1 | 18 | 18 | 18 | 27.6 | No | Below background |
| Iron | 7439-89-6 | 1/1 | 22500 | 22500 | 22500 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 1/1 | 16.8 | 16.8 | 16.8 | 27.4 | No | Below background |
| Magnesium | 7439-95-4 | 1/1 | 2780 | 2780 | 2780 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 1/1 | 596 | 596 | 596 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 1/1 | 0.054 | 0.054 | 0.054 | 0.059 | No | Below background |
| Nickel | 7440-02-0 | 1/1 | 21.4 | 21.4 | 21.4 | 17.7 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 1/1 | 989 | 989 | 989 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 1/1 | 1 | 1 | 1 | 1.7 | No | Below background |
| Silver | 7440-22-4 | 1/1 | 0.075 | 0.075 | 0.075 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 1/1 | 86.9 | 86.9 | 86.9 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 1/1 | 0.21 | 0.21 | 0.21 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 1/1 | 21 | 21 | 21 | 26.1 | No | Below background |
| Zinc | 7440-66-6 | 1/1 | 91.8 | 91.8 | 91.8 | 532 | No | Below background |
| | | | Ex | plosives | | | | |
| HMX | 2691-41-0 | 1/1 | 0.015 | 0.015 | 0.015 | None | Yes | Detected organic |
| Nitrocellulose | 9004-70-0 | 1/1 | 1.9 | 1.9 | 1.9 | None | Yes | Detected organic |

 Table 4–31. SRC Screening for Discrete Subsurface Sediment Samples at the Lower Cobbs Pond

| Analyte (mg/kg) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | SRC Justification |
|-------------------------|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------------------|------|-------------------|
| | 1 (unioci | Dettett | | SVOCs | nesure | enterna | sner | Site ousuiteuton |
| 2-Methylnaphthalene | 91-57-6 | 1/1 | 0.025 | 0.025 | 0.025 | None | Yes | Detected organic |
| Acenaphthene | 83-32-9 | 1/1 | 0.0099 | 0.0099 | 0.0099 | None | Yes | Detected organic |
| Acenaphthylene | 208-96-8 | 1/1 | 0.091 | 0.091 | 0.091 | None | Yes | Detected organic |
| Anthracene | 120-12-7 | 1/1 | 0.07 | 0.07 | 0.07 | None | Yes | Detected organic |
| Benz(a)anthracene | 56-55-3 | 1/1 | 0.34 | 0.34 | 0.34 | None | Yes | Detected organic |
| Benzo(a)pyrene | 50-32-8 | 1/1 | 0.47 | 0.47 | 0.47 | None | Yes | Detected organic |
| Benzo(b)fluoranthene | 205-99-2 | 1/1 | 0.8 | 0.8 | 0.8 | None | Yes | Detected organic |
| Benzo(ghi)perylene | 191-24-2 | 1/1 | 0.47 | 0.47 | 0.47 | None | Yes | Detected organic |
| Benzo(k)fluoranthene | 207-08-9 | 1/1 | 0.21 | 0.21 | 0.21 | None | Yes | Detected organic |
| Chrysene | 218-01-9 | 1/1 | 0.4 | 0.4 | 0.4 | None | Yes | Detected organic |
| Di-n-butyl phthalate | 84-74-2 | 1/1 | 0.023 | 0.023 | 0.023 | None | Yes | Detected organic |
| Dibenz(a,h)an thr acene | 53-70-3 | 1/1 | 0.12 | 0.12 | 0.12 | None | Yes | Detected organic |
| Fluoranthene | 206-44-0 | 1/1 | 0.48 | 0.48 | 0.48 | None | Yes | Detected organic |
| Fluorene | 86-73-7 | 1/1 | 0.02 | 0.02 | 0.02 | None | Yes | Detected organic |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/1 | 0.43 | 0.43 | 0.43 | None | Yes | Detected organic |
| Naphthalene | 91-20-3 | 1/1 | 0.038 | 0.038 | 0.038 | None | Yes | Detected organic |
| Phenanthrene | 85-01-8 | 1/1 | 0.15 | 0.15 | 0.15 | None | Yes | Detected organic |
| Pyrene | 129-00-0 | 1/1 | 0.41 | 0.41 | 0.41 | None | Yes | Detected organic |
| | | | | VOCs | | | | |
| 2-Butanone | 78-93-3 | 1/1 | 0.012 | 0.012 | 0.012 | None | Yes | Detected organic |

Table 4–31. SRC Screening for Discrete Subsurface Sediment Samples at the Lower Cobbs Pond (continued)

^a Background concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b). SRC screening tables include all available and appropriate data as presented in Section 4.4.4.

CAS = Chemical Abstract Service.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

mg/kg = Milligrams per kilogram.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound.

VOC = Volatile organic compound.

Bold indicates analyte identified as an SRC.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|----------------------------|------------|---------|----------|------------|-----------|-----------------------|------|--------------------|
| Analyte (mg/L) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| Timityte (iig, 2) | 1 (unioci | Dettett | Dettett | Metals | itesuit | Criteriu | snet | Site ousuiteuton |
| Aluminum | 7429-90-5 | 4/4 | 0.26 | 2.52 | 1.15 | 3.37 | No | Below background |
| Antimony | 7440-36-0 | 2/4 | 0.00096 | 0.001 | 0.00119 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 4/4 | 0.0009 | 0.0104 | 0.00415 | 0.0032 | Yes | Exceeds background |
| Barium | 7440-39-3 | 4/4 | 0.0198 | 0.382 | 0.141 | 0.0475 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 2/4 | 0.000053 | 0.000064 | 0.000102 | 0 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 2/4 | 0.000057 | 0.000062 | 0.00011 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 4/4 | 21.6 | 67.7 | 37.4 | 41.4 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 4/4 | 0.00056 | 0.0067 | 0.00297 | 0 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 4/4 | 0.00039 | 0.0107 | 0.00423 | 0 | Yes | Exceeds background |
| Copper | 7440-50-8 | 3/4 | 0.0019 | 0.0088 | 0.00355 | 0.0079 | Yes | Exceeds background |
| Iron | 7439-89-6 | 4/4 | 0.846 | 16.3 | 7.64 | 2.56 | No | Essential Nutrient |
| Lead | 7439-92-1 | 2/4 | 0.00032 | 0.00047 | 0.000698 | 0 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 4/4 | 3.25 | 11.2 | 6.1 | 10.8 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 4/4 | 0.132 | 15.8 | 6.16 | 0.391 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 4/4 | 0.0019 | 0.0067 | 0.0037 | 0 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 4/4 | 1.39 | 15.7 | 6.1 | 3.17 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 3/4 | 0.00023 | 0.0104 | 0.00335 | 0 | Yes | Exceeds background |
| Silver | 7440-22-4 | 1/4 | 0.0017 | 0.0017 | 0.000603 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 4/4 | 1.6 | 12.9 | 5.93 | 21.3 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 1/4 | 0.00046 | 0.00046 | 0.000865 | 0 | Yes | Exceeds background |
| Vanadium | 7440-62-2 | 3/4 | 0.00051 | 0.005 | 0.00385 | 0 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 4/4 | 0.0101 | 0.109 | 0.0403 | 0.042 | Yes | Exceeds background |
| | | | | Anions | | | | |
| Sulfate | 14808-79-8 | 1/2 | 29 | 29 | 15.1 | None | Yes | Exceeds background |
| Sulfide | 18496-25-8 | 1/2 | 2.3 | 2.3 | 1.28 | None | Yes | Exceeds background |
| | | | E | Explosives | | | | |
| 4-Amino-2,6-Dinitrotoluene | 19406-51-0 | 2/2 | 0.000043 | 0.00007 | 0.0000565 | None+ | Yes | Detected organic |

Table 4–32. SRC Screening for Surface Water at the Backwater Area (continued)

| Analyte (mg/L) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | SRC Justification |
|----------------|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------------------|------|-------------------|
| | | | | VOCs | | | | |
| Acetone | 67-64-1 | 1/3 | 0.003 | 0.003 | 0.00433 | None | Yes | Detected organic |
| Chloromethane | 74-87-3 | 1/3 | 0.0007 | 0.0007 | 0.000567 | None | Yes | Detected organic |
| Toluene | 108-88-3 | 1/3 | 0.001 | 0.001 | 0.000667 | None | Yes | Detected organic |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

CAS = Chemical Abstract Service.

mg/L = M illigrams per liter.

SRC = Site-related contaminant.

VOC = Volatile organic compound.

| | CAS | E f | | · | | | | |
|---|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------------------|------|--------------------|
| Analyte (mg/L) | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | SRC Justification |
| Metals | Tumber | Dettet | Dettet | Dutt | ittsuit | Criteria | SRC. | SICE 9 usuiteation |
| Aluminum | 7429-90-5 | 5/5 | 0.0916 | 0.332 | 0.185 | 3.37 | No | Below background |
| Antimony | 7440-36-0 | 1/5 | 0.00094 | 0.00094 | 0.00215 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 2/ 5 | 0.00078 | 0.0025 | 0.00464 | 0.0032 | No | Below background |
| Barium | 7440-39-3 | 5/ 5 | 0.0138 | 0.0226 | 0.0168 | 0.0475 | No | Below background |
| Calcium | 7440-70-2 | 5/5 | 13.2 | 24.3 | 19.6 | 41.4 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 2/5 | 0.00059 | 0.001 | 0.00108 | 0 | Yes | Exceeds background |
| Copper | 7440-50-8 | 3/ 5 | 0.0019 | 0.0051 | 0.00248 | 0.0079 | No | Below background |
| Iron | 7439-89-6 | 5/ 5 | 0.419 | 1.3 | 0.918 | 2.56 | No | Essential Nutrient |
| Lead | 7439-92-1 | 3/5 | 0.00039 | 0.0033 | 0.00158 | 0 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 5/5 | 1.75 | 4.41 | 3.36 | 10.8 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 5/5 | 0.0772 | 0.528 | 0.216 | 0.391 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 1/5 | 0.0019 | 0.0019 | 0.00178 | 0 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 5/5 | 0.894 | 1.57 | 1.26 | 3.17 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 1/5 | 0.00024 | 0.00024 | 0.00243 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 5/5 | 1.41 | 3.68 | 2.42 | 21.3 | No | Essential Nutrient |
| Vanadium | 7440-62-2 | 1/5 | 0.00089 | 0.00089 | 0.00104 | 0 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 4/5 | 0.0041 | 0.0115 | 0.00984 | 0.042 | No | Below background |
| | | | | Anions | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 2/4 | 0.1 | 0.18 | 0.087 | None | Yes | Exceeds background |
| Sulfate | 14808-79-8 | 2/2 | 13 | 13 | 13 | None | Yes | Exceeds background |
| Sulfide | 18496-25-8 | 2/2 | 0.7 | 0.7 | 0.7 | None | Yes | Exceeds background |
| | | | Mis | scellaneous | | | | |
| Total Phosphorus as P | 7723-14-0 | 2/2 | 0.06 | 0.24 | 0.15 | None | No | Essential Nutrient |
| | | | | xplosives | | | | |
| 4-Amino-2,6-Dinitrotoluene | 19406-51-0 | 2/3 | 0.000072 | 0.00024 | 0.000147 | None | Yes | Detected organic |
| | | | | SVOCs | | | | |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | 2/3 | 0.0019 | 0.002 | 0.00313 | None | Yes | Detected organic |
| Di-n-butyl phthalate | 84-74-2 | 2/3 | 0.0018 | 0.0046 | 0.0038 | None | Yes | Detected organic |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b). Bold indicates analyte identified as an SRC.

CAS = Chemical Abstract Service.

1

mg/L = M illigrams per liter. SRC = Site-related contaminant.

SVOC = Semi-volatile organic compound.

| | CAS | Freq of | Minimum | Maximum | Average | Background | | | |
|---|------------|---------|----------|-----------|----------|-----------------------|------|-------------------------|--|
| Analyte (mg/L) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification | |
| Metals | | | | | | | | | |
| Aluminum | 7429-90-5 | 7/7 | 0.114 | 0.377 | 0.218 | 3.37 | No | Below background | |
| Antimony | 7440-36-0 | 4/7 | 0.00073 | 0.0034 | 0.00163 | 0 | Yes | Exceeds background | |
| Arsenic | 7440-38-2 | 4/7 | 0.00078 | 0.0024 | 0.00371 | 0.0032 | No | Below background | |
| Barium | 7440-39-3 | 7/7 | 0.0117 | 0.0228 | 0.0141 | 0.0475 | No | Below background | |
| Beryllium | 7440-41-7 | 1/7 | 0.000034 | 0.000034 | 0.000196 | 0 | Yes | Exceeds background | |
| Cadmium | 7440-43-9 | 1/7 | 0.000043 | 0.000043 | 0.000389 | 0 | Yes | Exceeds background | |
| Calcium | 7440-70-2 | 7/7 | 10.9 | 38.7 | 20.4 | 41.4 | No | Essential Nutrient | |
| Chromium | 7440-47-3 | 2/7 | 0.00053 | 0.0011 | 0.00128 | 0 | Yes | Exceeds background | |
| Cobalt | 7440-48-4 | 1/7 | 0.00015 | 0.00015 | 0.000691 | 0 | Yes | Exceeds background | |
| Copper | 7440-50-8 | 4/7 | 0.0016 | 0.0029 | 0.0018 | 0.0079 | No | Below background | |
| Iron | 7439-89-6 | 7/7 | 0.294 | 1.35 | 0.69 | 2.56 | No | Essential Nutrient | |
| Lead | 7439-92-1 | 3/7 | 0.00029 | 0.0029 | 0.0015 | 0 | Yes | Exceeds background | |
| Magnesium | 7439-95-4 | 7/7 | 1.55 | 6.71 | 3.6 | 10.8 | No | Essential Nutrient | |
| Manganese | 7439-96-5 | 7/7 | 0.0915 | 0.654 | 0.215 | 0.391 | Yes | Exceeds background | |
| Nickel | 7440-02-0 | 2/7 | 0.0014 | 0.0016 | 0.00161 | 0 | Yes | Exceeds background | |
| Potassium | 7440-09-7 | 7/7 | 0.497 | 1.61 | 1.23 | 3.17 | No | Essential Nutrient | |
| Selenium | 7782-49-2 | 1/7 | 0.00032 | 0.00032 | 0.00245 | 0 | Yes | Exceeds background | |
| Sodium | 7440-23-5 | 7/7 | 1.37 | 4.4 | 2.52 | 21.3 | No | Essential Nutrient | |
| Thallium | 7440-28-0 | 1/7 | 0.00035 | 0.00035 | 0.00348 | 0 | Yes | Exceeds background | |
| Vanadium | 7440-62-2 | 5/7 | 0.00063 | 0.003 | 0.00196 | 0 | Yes | Exceeds background | |
| Zinc | 7440-66-6 | 5/7 | 0.0027 | 0.0139 | 0.0111 | 0.042 | No | Below background | |
| | | | A | nions | | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 2/5 | 0.1 | 0.13 | 0.0664 | None | Yes | Exceeds background | |
| Sulfate | 14808-79-8 | 3/3 | 14 | 19 | 16.3 | None | Yes | Exceeds background | |
| Sulfide | 18496-25-8 | 1/3 | 0.7 | 0.7 | 0.4 | None | Yes | Exceeds background | |
| | | | | ellaneous | | | | | |
| Total Phosphorus as P | 7723-14-0 | 2/2 | 0.1 | 0.12 | 0.11 | None | No | Essential Nutrient | |
| | | | | olosives | | | | | |
| 4-Amino-2,6-Dinitrotoluene | 19406-51-0 | 1/4 | 0.00029 | 0.00029 | 0.000143 | None | Yes | Detected organic | |

| Table 4–34. SRC Screening for Surface Water at Lower Cob | os Pond |
|--|---------|
|--|---------|

| | CAS | Freq of | Minimum | Maximum | Average | Background | | |
|----------------------------|----------|---------|---------|---------|---------|-----------------------|------|-------------------|
| Analyte (mg/L) | Number | Detect | Detect | Detect | Result | Criteria ^a | SRC? | SRC Justification |
| | | | S | VOCs | | | | |
| Benzenemethanol | 100-51-6 | 1/2 | 0.0049 | 0.0049 | 0.00495 | None | Yes | Detected organic |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | 2/4 | 0.0022 | 0.0029 | 0.00378 | None | Yes | Detected organic |
| Butyl benzyl phthalate | 85-68-7 | 1/4 | 0.0018 | 0.0018 | 0.00433 | None | Yes | Detected organic |
| Di-n-butyl phthalate | 84-74-2 | 2/4 | 0.0023 | 0.012 | 0.00608 | None | Yes | Detected organic |

Table 4-34. SRC Screening for Surface Water at Lower Cobbs Pond (continued)

CAS = Chemical Abstract Service.

mg/L = M illigrams per liter.

SRC = Site-related contaminant.

SVOC = Semi-volatile organic compound.

| Sample ID | Туре | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|----------------------|------|----------|---------------|----------------|-------|-----|-----|--------|-----|---|
| Sample ID | Туре | Date | (11) | Samping Event | Sedin | | ræi | IIIINA | LNA | Comments |
| CPCsd-001(P)-0650-SD | D | 08/19/96 | 0–1 | Phase I RI | | X | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-002(P)-0651-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-003(P)-0652-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-004(P)-0653-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-005(P)-0654-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-006(P)-0655-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-007(P)-0656-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-007(P)-0657-FD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-008(P)-0659-SD | D | 08/19/96 | 0–1 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-009(P)-0660-SD | D | 08/19/96 | 0–0.5 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-010(P)-0661-SD | D | 08/19/96 | 0–0.5 | Phase I RI | | Х | | | | Sample not used for HHRA and ERA because more recent sample was used. |
| CPCsd-044-5022-SD | D | 03/29/10 | 0–0.5 | PBA08 RI | | Х | Х | Х | X | |
| CPCsd-045-5023-SD | D | 04/01/10 | 0–0.5 | PBA08 RI | | Х | Х | Х | X | |
| CPCsd-045-5783-SD | D | 04/01/10 | 0.5–2 | PBA08 RI | | Х | Х | Х | | |
| CPCsd-046-5024-SD | D | 03/25/10 | 0–0.5 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsd-046-5784-SD | D | 03/25/10 | 0.5–2 | PBA08 RI | | Х | Х | Х | | |
| CPCsd-047-5025-SD | D | 04/01/10 | 0-0.5 | PBA08 RI | | Х | Х | Х | X | |
| CPCsd-047-5785-SD | D | 04/01/10 | 0.5–2 | PBA08 RI | | Х | Х | Х | | |
| CPCsd-048-5026-SD | D | 04/01/10 | 0–0.5 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsd-048-5786-SD | D | 04/01/10 | 0.5–2 | PBA08 RI | | Х | Х | Х | | |
| CPCsd-049-5032-SD | D | 03/25/10 | 0–0.5 | PBA08 RI | | Х | Х | Х | Х | |
| FSW-SD-012-0000 | ISM | 06/24/03 | 0–0.5 | FWBWQS 2003 | | Х | | | | Off AOC sample used for nature and extent comparison only. |

Table 4–35. Data Summary and Designated Use for RI

| | T | D (| Depth | | 00 | NOF | | | | |
|--------------------|------|----------|-------|----------------|----|-----|-----|------|-----|--|
| Sample ID | Туре | Date | (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
| FSW-SD-030-0000 | ISM | 06/23/03 | 0-0.3 | FWBWQS 2003 | | X | X | X | X | |
| FSW-SD-031-0000 | ISM | 06/23/03 | 0-0.3 | FWBWQS 2003 | | Х | Х | Х | Х | |
| ULCPsd-001-0001-SD | D | 08/23/01 | 0–0.5 | Phase II RI | | | | | | Do not use sample because it was included in the Load Line 12 evaluation. |
| ULCPsd-002-0001-SD | D | 08/23/01 | 0.3–1 | Phase II RI | | | | | | Do not use sample because it was included in the Load Line 12 evaluation. |
| ULCPsd-003-0001-SD | D | 08/22/01 | 0–0.8 | Phase II RI | | | | | | Do not use sample because it was included in the Load Line 12 evaluation. |
| ULCPsd-004-0001-FD | D | 08/22/01 | 0-1 | Phase II RI | Х | | | | | Field duplicate. |
| ULCPsd-004-0001-SD | D | 08/22/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-005-0001-SD | D | 08/22/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-006-0001-SD | D | 08/22/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-007-0001-SD | D | 08/21/01 | 0-0.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-008-0001-SD | D | 08/21/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-009-0001-SD | D | 08/21/01 | 0-0.9 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-011-0001-SD | D | 08/20/01 | 0-0.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-012-0001-SD | D | 08/20/01 | 0-1.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-013-0001-FD | D | 08/20/01 | 0-1.8 | Phase II RI | Х | | | | | Field duplicate. |
| ULCPsd-013-0001-SD | D | 08/20/01 | 0-1.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-014-0001-SD | D | 07/25/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-015-0001-SD | D | 08/17/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-015-0002-SD | D | 09/27/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-016-0001-SD | D | 08/17/01 | 0-1.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-017-0001-SD | D | 08/17/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-018-0001-SD | D | 08/16/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-019-0001-SD | D | 08/16/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-020-0001-FD | D | 08/16/01 | 0–0.8 | Phase II RI | Х | | | | | Field duplicate. |
| ULCPsd-020-0001-SD | D | 08/16/01 | 0–0.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-021-0001-SD | D | 08/15/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-021-0002-SD | D | 09/27/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-022-0001-SD | D | 08/15/01 | 0-0.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-023-0001-SD | D | 08/15/01 | 0-0.8 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-024-0001-SD | D | 08/14/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-025-0001-SD | D | 07/25/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPsd-026-0001-SD | D | 08/14/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |

Table 4-35. Data Summary and Designated Use for RI (continued)

| Sample ID | Туре | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|--------------------|------|----------|---------------|----------------|----|-----|-----|------|-----|--|
| ULCPsd-027-0001-SD | D | 07/26/01 | 0–0.5 | Phase II RI | | | | | | Do not use sample because location was included in the Load Line 3 evaluation. |
| Soil | | | | | | | | | | |
| CPCsb-030-5105-SO | D | 03/29/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsb-031-5109-SO | D | 03/24/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsb-032-5113-SO | D | 03/24/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsb-032-5114-SO | D | 03/24/10 | 1–4 | PBA08 RI | | Х | Х | Х | | |
| CPCsb-032-5115-SO | D | 03/24/10 | 4–7 | PBA08 RI | | Х | Х | Х | | |
| CPCsb-032-5116-SO | D | 03/24/10 | 7–10 | PBA08 RI | | Х | Х | Х | | |
| CPCsb-032-6073-FD | D | 03/24/10 | 1–4 | PBA08 RI | Х | | | | | Field duplicate. |
| CPCsb-034-5119-SO | D | 03/29/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsb-034-5120-SO | D | 03/29/10 | 1–4 | PBA08 RI | | Х | Х | Х | | |
| CPCsb-035-5123-SO | D | 03/29/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCsb-035-5124-SO | D | 03/29/10 | 1–4 | PBA08 RI | | Х | Х | Х | | |
| CPCsb-035-5125-SO | D | 03/29/10 | 4–7 | PBA08 RI | | Х | Х | Х | | |
| CPCsb-035-6072-FD | D | 03/29/10 | 4–7 | PBA08 RI | Х | | | | | Field duplicate. |
| CPCss-036-5014-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-037-5015-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-037-6041-FD | D | 02/23/10 | 0-1 | PBA08 RI | Х | | | | | Field duplicate. |
| CPCss-038-5016-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-039-5017-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-040-5018-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-041-5019-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-042-5020-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| CPCss-043-5021-SO | D | 02/23/10 | 0-1 | PBA08 RI | | Х | Х | Х | Х | |
| ULCPsd-010-0001-SD | D | 08/21/01 | 0-0.5 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-001-0001-SO | D | 07/17/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-002-0001-SO | D | 07/17/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-003-0001-SO | D | 07/18/01 | 0-1 | Phase II RI | | Х | Х | Х | X | |
| ULCPss-004-0001-SO | D | 07/17/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-005-0001-SO | D | 07/18/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-006-0001-SO | D | 07/19/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-007-0001-FD | D | 07/18/01 | 0-1 | Phase II RI | Х | | | | | Field duplicate. |
| ULCPss-007-0001-SO | D | 07/18/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| ULCPss-008-0001-SO | D | 07/18/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |

| Table 4-35 | . Data Summary | and | Designated | Use f | for RI | (continued) |
|------------|----------------|-----|------------|-------|--------|-------------|
| | | | | | | |

Upper and Lower Cobbs Ponds

| Courselo ID | T | Dete | Depth | | | NPE | Б е Т | | | Comments |
|--------------------|------|----------|-------|----------------|--------|-------|-------|------|-----|--|
| Sample ID | Туре | Date | (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
| ULCPss-009-0001-SO | D | 07/19/01 | 0-1 | Phase II RI | | Х | X | Х | Х | |
| ULCPss-010-0001-SO | D | 07/19/01 | 0-1 | Phase II RI | | Х | Х | Х | Х | |
| | | | | S | urface | Water | | | | |
| CPCsw-044-5027-SW | D | 03/29/10 | | PBA08 RI | | Х | Х | Х | Х | |
| CPCsw-045-5028-SW | D | 04/01/10 | | PBA08 RI | | Х | Х | Х | Х | |
| CPCsw-046-5029-SW | D | 03/25/10 | | PBA08 RI | | Х | Х | Х | Х | |
| CPCsw-047-5030-SW | D | 04/01/10 | | PBA08 RI | | Х | Х | Х | Х | |
| CPCsw-047-6045-FD | D | 04/01/10 | | PBA08 RI | Х | | | | | Field duplicate. |
| CPCsw-048-5031-SW | D | 04/01/10 | | PBA08 RI | | Х | Х | Х | Х | |
| FSW-SW-012-0000 | D | 06/24/03 | 0–1 | FWBWQS 2003 | | Х | | | | Off-AOC sample used for nature and extent comparison only. |
| FSW-SW-030-0000 | D | 06/23/03 | 0–1 | FWBWQS 2003 | | Х | Х | Х | Х | |
| FSW-SW-031-0000 | D | 06/23/03 | 0-1 | FWBWQS 2003 | | Х | Х | Х | Х | |
| FSW-SW-052-0000 | D | 09/16/03 | 0–1 | FWBWQS 2003 | | Х | | | | Off-AOC sample used for nature and extent comparison only. |
| FSW-SW-070-0000 | D | 08/05/03 | 0–1 | FWBWQS 2003 | | Х | Х | Х | Х | |
| FSW-SW-071-0000 | D | 08/05/03 | 0–1 | FWBWQS 2003 | | Х | Х | Х | Х | |
| ULCPsw-001-0001-SW | D | 07/26/01 | | Phase II RI | | Х | Х | Х | Х | |
| ULCPsw-002-0001-FD | D | 07/25/01 | | Phase II RI | Х | | | | | Field duplicate. |
| ULCPsw-002-0001-SW | D | 07/25/01 | | Phase II RI | | Х | Х | Х | Х | |
| ULCPsw-003-0001-SW | D | 07/24/01 | | Phase II RI | | Х | Х | Х | Х | |
| ULCPsw-004-0001-SW | D | 07/24/01 | | Phase II RI | | Х | Х | Х | Х | |
| ULCPsw-005-0001-SW | D | 07/24/01 | | Phase II RI | | Х | Х | Х | Х | |
| ULCPsw-006-0001-SW | D | 07/24/01 | | Phase II RI | | Х | Х | Х | Х | |
| ULCPsw-007-0001-SW | D | 07/24/01 | | Phase II RI | | Х | Х | Х | Х | |

Table 4–35. Data Summary and Designated Use for RI (continued)

AOC = Area of concern.

D = Discrete.

ERA = Ecological risk assessment.

ft = Feet.

F&T = Fate and transport.

FWBWQS = Facility-wide Biological Water Quality Study.

HHRA = Human health risk assessment.

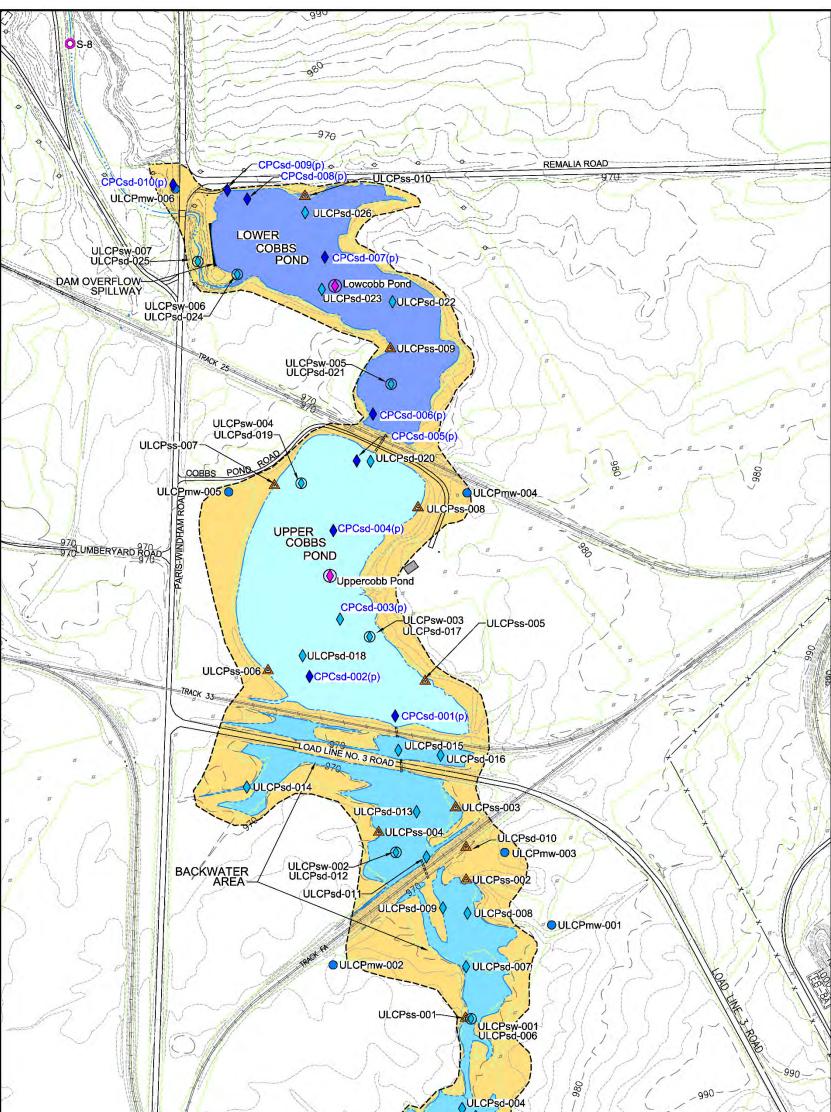
ID = Identification.

N&E = Nature and extent.

PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

QC = Quality control.

RI = Remedial investigation.



| | LOAD LINE 3 DOWNGRADIENT CHANNEL MED TRIBUTARIES COBBS POND 980- |
|---|--|
| LEGEND: ASPHALT ROADAOC BOUNDARY ASPHALT ROADSOIL AGGREGATE RAILROAD TRACKS RAILROAD TRACKS SURFACE WATER AND SEDIMENT AGGREGATES | PHASE II RI SAMPLE LOCATIONS ♦ ULCP*SEDIMENT © ULCP*AND SEDIMENT ▲ ULCP*SURFACE SOIL |
| VEGETATION | FACILITY-WIDE BIOLOGICAL & WATER QUALITY STUDY SAMPLE LOCATIONS UPPER AND LOWER COBBS PONDS FORMER RVAAP/CAMP RAVENNA PORTAGE & TRUMBULL COUNTIES, OHI Image: Colored |

Figure 4-1. Historical Sample Locations at Upper and Lower Cobbs Ponds

Upper and Lower Cobbs Ponds

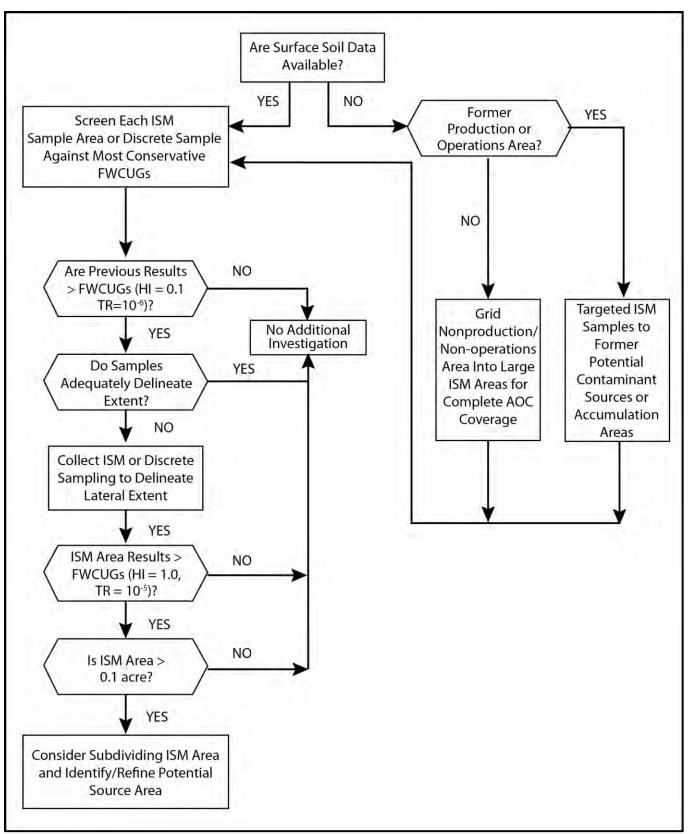


Figure 4–2. PBA08 RI Surface Soil Sampling

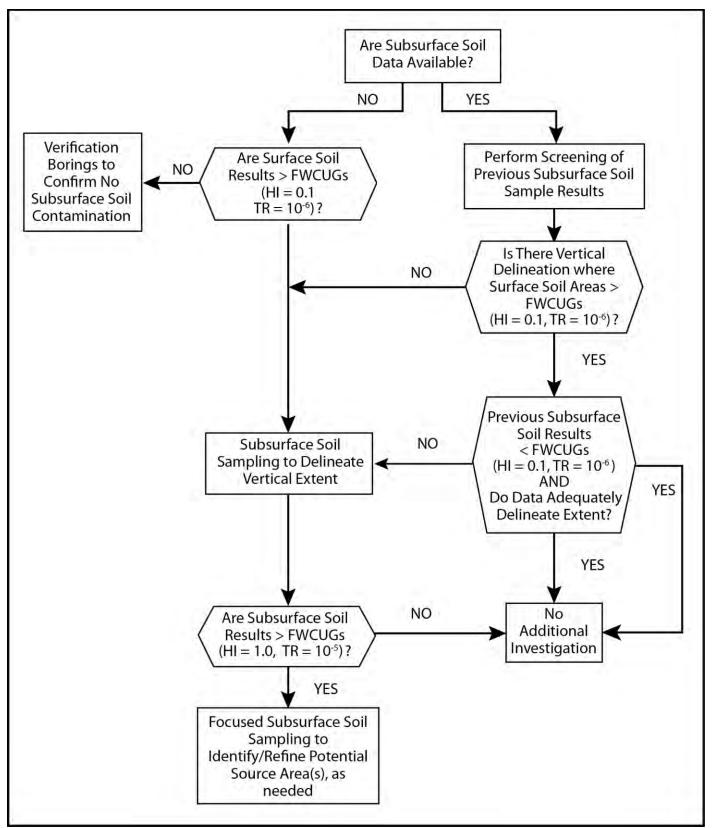


Figure 4–3. PBA08 RI Subsurface Soil Sampling

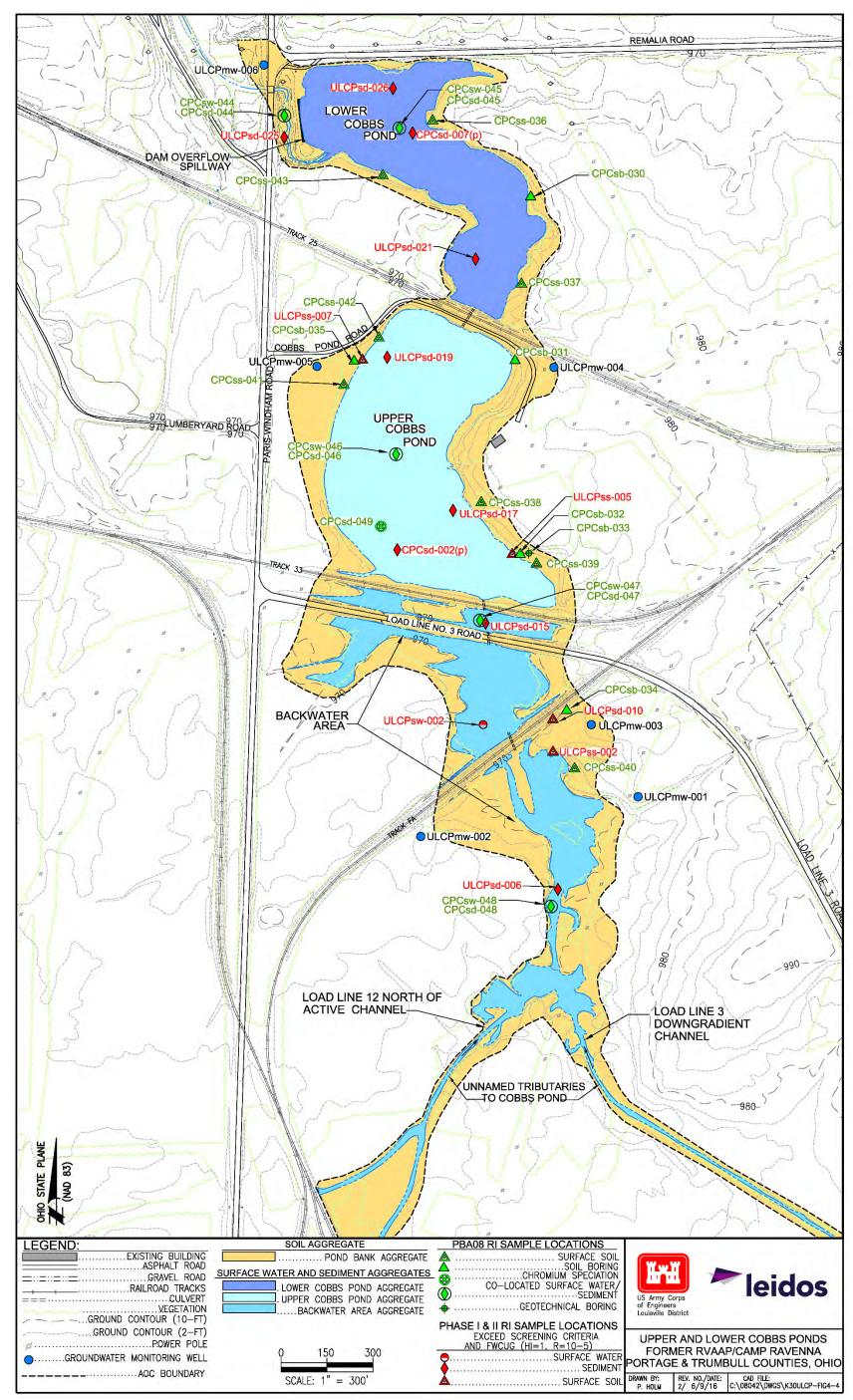


Figure 4-4. PBA08 RI Sample Locations at Upper and Lower Cobbs Ponds

Upper and Lower Cobbs Ponds

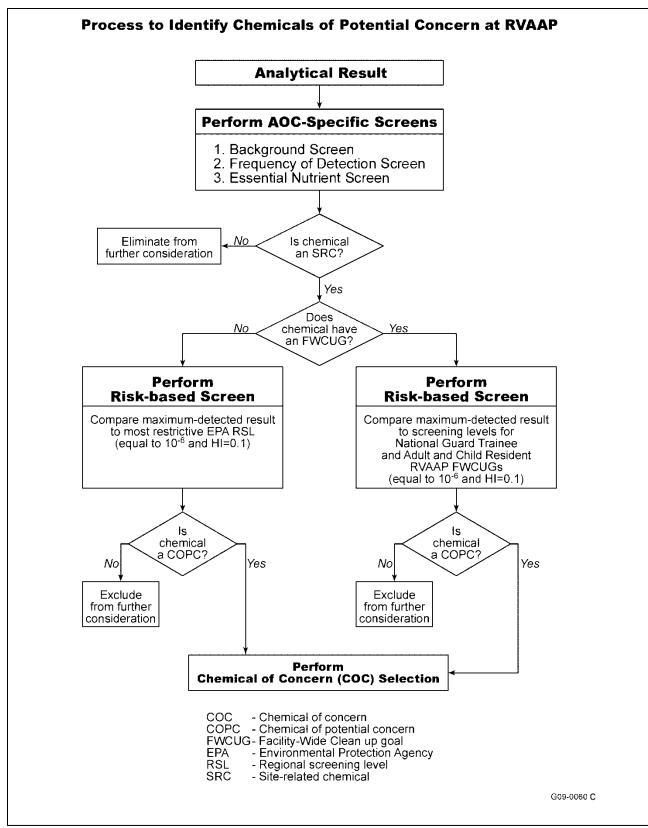


Figure 4-5. Process to Identify RVAAP COPCs in the HHRA (USACE 2010a)

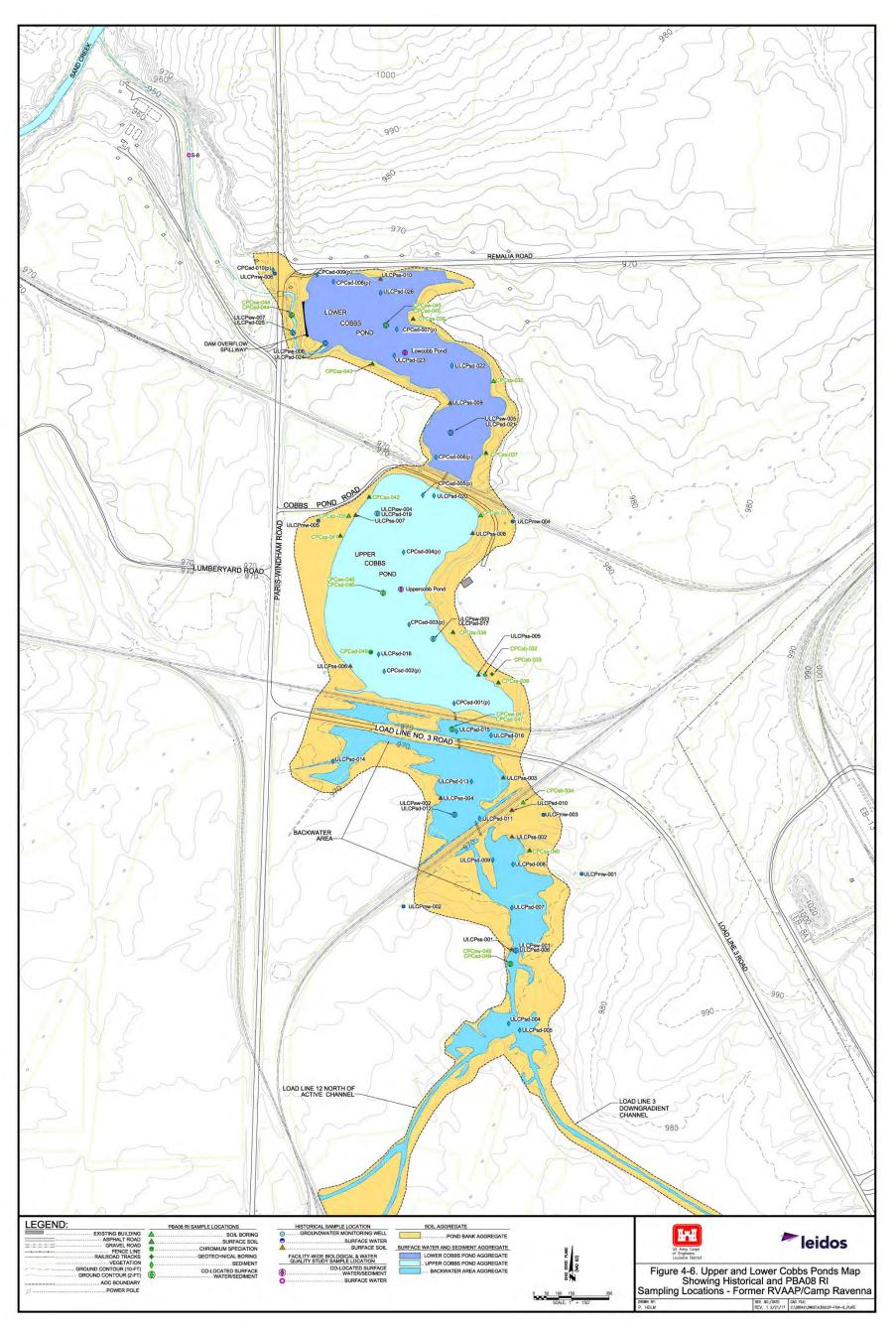


Figure 4–6. All Upper and Lower Cobbs Ponds RI Sample Locations

Upper and Lower Cobbs Ponds

5.0 NATURE AND EXTENT OF CONTAMINATION 1

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This section evaluates the nature and extent of contamination at the Upper and Lower Cobbs Ponds AOC and presents results of the analytical data obtained during the 1996 Phase I RI, 2001 Phase II 5 RI, 2003 FWBWQS, and 2010 PBA08 RI. To support the evaluation of nature and extent of 6 contamination, SRC concentrations were compared to SLs corresponding to the lowest FWCUG for the Resident Receptor (Adult and Child) and National Guard Trainee at a target HQ of 0.1 or TR of 8 1E-06, as presented in the FWCUG Report. If a chemical did not have a FWCUG, the SL was the 9 lower of the USEPA Residential RSL for HQ of 0.1 or TR of 1E-06. The following figures illustrate 10 the concentration and distribution of SRCs that exceed SLs:

- 12 Figure 5-1 – Data Aggregates and Sample Locations at Upper and Lower Cobbs Ponds. •
- 13 • Figure 5-2 – Detected Concentrations of Explosives and Propellants in Soil (Discrete Soil 14 Borings).
- 15 • Figure 5-3 – Exceedances of FWCUG (HQ of 0.1, TR of 10-6) for Aluminum, Arsenic, 16 Cobalt, and Cyanide in Soil (Discrete Soil Borings).
- 17 • Figure 5-4 – PAH Exceedances of FWCUG (HQ of 0.1, TR of 10-6) in Soil (Discrete Soil 18 Borings).
- 19 • Figure 5-5 – Detected Concentrations of VOCs, Pesticides, and PCBs in Soil (Discrete Soil 20 Borings).
- 21 • Figure 5-6 – Detected Concentrations of Explosives and Propellants in Surface Water and 22 Sediment.
- 23 • Figure 5-7 – Exceedances of FWCUG (HQ of 0.1, TR of 10-6) for Aluminum, Arsenic, 24 Hexavalent Chromium, Cobalt, Cvanide, Manganese, and Selenium in Surface Water and 25 Sediment.
- 26 • Figure 5-8 – PAH Exceedances of FWCUG (HQ of 0.1, TR of 10-6) in Surface Water and 27 Sediment.
- Figure 5-9 Detected Concentrations of VOCs, Pesticides, and PCBs in in Surface Water and 28 29 Sediment.
- 30

31 As discussed in Section 4.0, data from all eligible samples were combined and screened to identify 32 SRCs representing current conditions at the Upper and Lower Cobbs Ponds AOC. All validated 33 Upper and Lower Cobbs Ponds AOC data from the RIs (1996 Phase I RI, 2001 Phase II RI, and 2010 34 PBA08 RI) and the 2003 FWBWQS are included in Appendix D. Complete analytical data packages 35 from the PBA08 RI are also included in Appendix D.

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5.1 DATA AGGREGATES AND EVALUATION

Data were grouped (aggregated) in three ways for evaluation of contaminant nature and extent and the 39 40 HHRA and ERA. The initial basic aggregation of data was by environmental medium: surface soil, subsurface soil, sediment, and surface water. For each medium-specific aggregate, an evaluation was 41 42 conducted to determine if further aggregation was warranted with respect to AOC characteristics, spatial aggregates). Data for soil were further aggregated based on depth and sample type for consistency with guidance established in the FWHHRAM and FWCUG Report. Data aggregates for evaluating the nature and extent of contamination at the Upper and Lower Cobbs Ponds AOC are as follows:

- 5
- 6 Surface Soil (0-1 ft bgs): Using the above data aggregation criteria, surface soil within the 7 geographic area of Upper Cobbs Pond, Lower Cobbs Pond, and the Backwater Area were 8 evaluated as an AOC-wide spatial aggregate, designated as the Pond Bank aggregate. The 9 Pond Bank is defined as the perimeter of the natural shorelines throughout the entirety of the 10 AOC. It extends from the Backwater Area to the south at the confluence of the channels 11 exiting Load Lines 3 and 12 and north to the intersection of Remalia Road and Paris-12 Windham Road. The areas encompassed within the Pond Bank are dry the majority of the 13 time, although they may be intermittently wet in association with heavy rainfall events when 14 they may receive runoff from the surrounding areas or when water levels within the AOC are 15 elevated.
- Subsurface Soil (less than 1 ft bgs): This medium was classified as an AOC-wide spatial
 aggregate on the same basis as surface soil (i.e., Pond Bank).
- Sediment: Sediment samples were divided into three spatial aggregates: Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. These three aggregates are perennially inundated water bodies. The Backwater Area is in the southern portion of the complex, extending from the confluence of the drainage channels exiting Load Line 3 and Load Line 12 north to Track 3. Upper Cobbs Pond is the portion of the water body north of Track 33 to Cobbs Pond Road. Lower Cobbs Pond is the portion north of the Cobbs Pond Road, bounded by Remalia Road on the north and Paris-Windham Road to the west.
- Surface Water: This medium is subdivided into the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond spatial aggregates on the same basis as sediment. The direction of flow within the AOC is from the south (Backwater Area) towards the dam overflow spillway at the north (Lower Cobbs Pond).
- 29

Surface soil, surface water, and sediment samples were collected during prior investigations such as
the 1996 Phase I RI (USACE 1998), 2003 FWBWQS (USACE 2005a), 2001 Phase II RI (MKM
2005), and 2010 PBA08 RI. All media, with the exception of sediment during the 2003 FWBWQS,
were sampled using discrete methods.

34

35 The media sampled during each of these events are summarized below.

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- 1996 Phase I RI
- o Tł
 - The only medium sampled in the 1996 Phase I RI was sediment.
- Since more recent sample data were available, the Phase I samples were utilized for nature and extent purposes only.
- 40 41

| 1 | • 2001 Phase II RI | | | | | | | |
|----------|---|--|--|--|--|--|--|--|
| 2 | • Media sampled were surface soil, sediment, and surface water. | | | | | | | |
| 3 | • 2003 FWBWQS | | | | | | | |
| 4 | • Media sampled were sediment and surface water. | | | | | | | |
| 5 | • Each pond bottom sediment sample was an incremental sampling methodology | | | | | | | |
| 6 | (ISM) sample. | | | | | | | |
| 7 | • Pond water samples were discrete grab samples. | | | | | | | |
| 8 | • 2010 PBA08 RI | | | | | | | |
| 9 | • Media sampled were surface soil, subsurface soil, sediment (0–0.5 ft bgs), subsurface | | | | | | | |
| 10 | sediment (0.5–2 ft bgs) and surface water. | | | | | | | |
| 11 | • The PBA08 RI samples collected within the 0.5–2 ft bgs interval were screened as | | | | | | | |
| 12 | subsurface sediment. | | | | | | | |
| 13 | | | | | | | | |
| 14 | All available sample data were evaluated to determine suitability for use in the various key RI data | | | | | | | |
| 15 | screens and evaluations (nature and extent, fate and transport, risk assessment). Evaluation of data | | | | | | | |
| 16 | suitability for use involved two primary considerations: (1) representativeness with respect to current | | | | | | | |
| 17 | AOC conditions; and (2) sample collection methods (e.g., discrete vs. ISM). | | | | | | | |
| 18 | | | | | | | | |
| 19 | Samples from the 1996 Phase I RI and 2001 Phase II RI data sets were evaluated to determine if | | | | | | | |
| 20 | conditions had changed substantively between earlier characterization efforts and the 2010 PBA08 RI | | | | | | | |
| 21 | activities. No disturbance activities have occurred at the AOC between 1996 Phase I RI and 2010 | | | | | | | |
| 22 | PBA08 RI activities. For soil, all data collected during these RI data sets were utilized in screening | | | | | | | |
| 23 | for SRCs and COPCs and were carried forward into the risk assessment. The full Phase II RI and | | | | | | | |
| 24 | PBA08 RI data sets were incorporated into the sediment and surface water SRC screening process. | | | | | | | |
| 25 | Since more recent sample data were available, the Phase I samples were utilized for nature and extent | | | | | | | |
| 26 | purposes only. | | | | | | | |
| 27 | | | | | | | | |
| 28 | The surface sediment screening data set includes both the PBA08 RI samples from the $0-0.5$ ft bgs | | | | | | | |
| 29 | interval and historical sediment samples. ISM sediment samples were screened separately from | | | | | | | |
| 30 | discrete samples. | | | | | | | |
| 31 | 5.4 SUDEACE SOLL SAMDLE DESULTS FOR CHROMIUM SPECIATION | | | | | | | |
| 32 | 5.2 SURFACE SOIL SAMPLE RESULTS FOR CHROMIUM SPECIATION | | | | | | | |
| 33 34 | During the PBA08 RI, a sediment sample was collected from one discrete sample location (CPCsd- | | | | | | | |
| 34 35 | 049) and analyzed for hexavalent chromium and total chromium. The result is presented in Table 5-1. | | | | | | | |
| 36 | Since historical samples at the AOC also incorporated hexavalent chromium analysis, this speciation | | | | | | | |
| 37 | sample was also utilized in the SRC screening data set. This sampling was conducted to determine the | | | | | | | |
| 38 | contribution of hexavalent chromium to total chromium in soil at the AOC for use in the HHRA | | | | | | | |
| 39 | (Section 7.2). | | | | | | | |
| 40 | | | | | | | | |
| 41 | 5.3 CONTAMINANT NATURE AND EXTENT IN SOIL | | | | | | | |
| 42 | | | | | | | | |
| 43 | The Pond Bank SRC screening data set included 10 discrete surface soil samples collected in 2001 | | | | | | | |

under the Phase II RI and 13 discrete surface soil samples collected during the PBA08 RI. Eleven of

1 the PBA08 RI samples were analyzed for TAL metals, explosives, and SVOCs; two samples were

2 analyzed for RVAAP full-suite analytes. Table 4-22 presents the results of the SRC screening for

3 surface soil samples for the Pond Bank.

4

5 Subsurface soil samples were not collected during historical investigations at the AOC. The Pond 6 Bank SRC screening data set for subsurface soil was comprised of six discrete subsurface soil 7 samples collected during the PBA08 RI. Four of the samples were analyzed for TAL metals, 8 explosives, and SVOCs; two samples collected from one boring location were analyzed for RVAAP 9 full-suite analytes. Table 4-23 presents the results of the SRC screening for subsurface soil samples 10 for the Pond Bank.

11

12 Tables 4-13 and 4-15 summarize the analytical results for all detected analytes in the Phase II RI 13 surface samples and the PBA08 RI surface and subsurface soil samples at the Pond Bank. Complete 14 copies of all the laboratory analytical data packages are presented in Appendix D.

15

16 5.3.1 Explosives and Propellants

17

18 Surface and Subsurface Soil

19

20 The propellant nitrocellulose was identified as an SRC in surface and subsurface soil at the Pond 21 Bank. Nitrocellulose was detected in all three surface soil samples where analyzed: PBA08 RI 22 locations CPCsb-035 (1.5J mg/kg), CPCss-039 (1.2J mg/kg), and Phase II RI sample ULCPss-007 23 (1.5 mg/kg). Nitrocellulose was detected in the 1–4 ft bgs interval at the PBA08 RI location CPCsb-24 035 but was not detected in the sample collected from the 4-7 ft bgs interval. As shown on Figure 5-25 2, these soil sample locations were located on the perimeter of the Upper Cobbs Pond. All detected 26 concentrations of nitrocellulose were below the RSL for HQ of 0.1 or TR 1E-06 and were not 27 considered COPCs. No explosives were detected in surface and subsurface soil at the Pond Bank.

- 28 29
 - 5.3.2 Inorganic Chemicals
- 30
- 31 Surface Soil32

Eighteen inorganic chemicals (16 metals, cyanide, and nitrate/nitrite) were identified as SRCs in
 surface soil at the Pond Bank.

35

36 Of these eighteen inorganic chemicals, four (aluminum, arsenic, cobalt, and cyanide) were above the 37 background concentration and exceeded their respective SLs in order to be identified as COPCs as 38 described below.

39

Aluminum was detected above the background concentration of 17,700 mg/kg and its respective SL of 3,496 mg/kg at two surface soil sample locations (ULCPss-004 and ULCPss-005). The maximum concentration of 19,200 mg/kg was observed at Phase II RI sample location ULCPss-005.

Arsenic was detected above the background concentration of 15.4 mg/kg and its respective
 SL of 0.425 mg/kg at four surface soil sample locations, ranging from 16.3 mg/kg (CPCsb-005) to 28.4 mg/kg (ULCPsd-010). The maximum concentration was detected at the Phase II
 RI sample location in the drainage ditch alongside Track FA that crosses the Backwater Area.
 The PBA08 RI sample collected at this same location also exceeds the background
 concentration, but at a lower concentration (CPCsb-034).

- Cobalt was detected above the background concentration of 10.4 mg/kg and its respective SL of 7.03 mg/kg at eight surface soil sample locations, ranging from 10.6 mg/kg (ULCPss-002) to 16.4 mg/kg (CPCsb-032). The Phase II RI sample collected at this same location (Figure 5-3) also exceeds the background concentration and SL, but at a lower concentration (ULCPss-1105).
 - Cyanide was analyzed in 11 of the Phase II RI samples and was detected in 4 samples above its respective SL of 0.27 mg/kg, ranging from 0.31 mg/kg (ULCPss-004) to 1.2 mg/kg (ULCPsd-010). The maximum concentration was detected in the drainage ditch alongside Track FA that crosses the Backwater Area.

Figure 5-3 presents the locations with concentrations that were above the background concentration
and exceeded their respective SLs for these chemicals. Concentrations of chromium exceeded the SL
for hexavalent chromium (1.64 mg/kg) but did not exceed the SL for trivalent chromium (8,147
mg/kg).

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22 Subsurface Soil

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24 Four inorganic chemicals (cadmium, lead, selenium, and silver) were identified as SRCs in 25 subsurface soil samples at the Pond Bank. Background concentrations are not established for 26 cadmium and silver. Cadmium was detected in all six subsurface soil samples, with concentrations 27 ranging from 0.032J (CPCsb-032) to 0.18J mg/kg (CPCsb-035). Silver was detected in two 28 subsurface soil samples, with concentrations of 0.029J mg/kg (CPCsb-032) and 0.013J mg/kg 29 (CPCsb-035) from the 4-7 ft bgs interval at both locations. Lead was detected in all six subsurface 30 soil samples but was detected above the background concentration of 19.1 mg/kg at one location 31 (CPCsb-035 from the 1-4 ft bgs interval) with a concentration of 23.3 mg/kg. Selenium also was 32 detected in all subsurface soil samples; however, one concentration was detected above the background concentration of 1.5 mg/kg in one location (CPCsb-032 from the 1-4 ft bgs interval) with 33 34 a concentration of 1.6 mg/kg. At the two soil borings where multiple intervals were analyzed, the 35 higher concentrations were generally observed in the shallowest interval. No concentrations detected 36 in the subsurface soil at the Pond Bank exceeded their respective SLs and were not considered 37 COPCs.

38

39 5.3.3 Semi-volatile Organic Compounds40

- 41 Surface Soil
- 42

SVOCs do not have background concentrations for comparison to chemical results; consequently a
 large number of SVOCs were identified as SRCs in the surface soil. Sixteen SVOCs were identified

1 as SRCs in surface soil at the Pond Bank. Twelve of the SVOC SRCs were PAHs. As shown on 2 Figure 5-4, benzo(a)pyrene was detected at two sample locations (CPCsb-035 and CPCss-042) with 3 concentrations that exceeded its respective SL of 0.022 mg/kg, and benzo(a)pyrene was identified as 4 a COPC. PBA08 RI sample locations CPCsb-035 and CPCss-042 are located on the western bank of 5 the Upper Cobbs Pond and immediately south of Cobbs Pond Road (Figure 5-4). The benzo(a)pyrene concentrations were detected below the Resident Receptor (Adult and Child) FWCUG at a TR of 6 7 1E-05, HQ of 1. All other SVOC detections in surface soil were observed throughout the Pond Bank 8 and the majority of detections occurred at low, estimated concentrations below laboratory reporting 9 limits.

10

11 Subsurface Soil

12

Thirteen SVOCs, 12 of which were PAHs, were identified as SRCs in the subsurface soil at the Pond Bank. Twelve of the 13 SVOCs were detected from the 1-4 ft bgs interval at PBA08 RI sample location CPCsb-035, and the majority of detections were observed at low, estimated concentrations below laboratory reporting limits. However, benzo(a)pyrene was detected at CPCsb-035 from the 1-4 ft bgs interval at a concentration of 0.066 mg/kg that exceeded its respective SL of 0.022 mg/kg. The concentration was below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1 and was not detected in the 4-7 ft bgs interval at CPCsb-035 (Figure 5-4).

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- 21 22

5.3.4 Volatile Organic Compounds, Pesticides, and Polychlorinated Biphenyls

23 Surface and Subsurface Soil

24

The pesticide beta-hexachlorocyclohexane (BHC) was identified as an SRC in surface soil at the Pond Bank, as shown in Figure 5-5. Beta-BHC was detected in only one surface soil sample, PBA08 RI location CPCsb-035, at a concentration of 0.0035J mg/kg. The concentration was detected below its respective SL and was not detected in the two subsurface intervals collected at CPCsb-035. VOCs and PCBs were not detected in surface soil at the Pond Bank. No VOCs, PCBs, or pesticides were detected in subsurface soil at the Pond Bank.

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5.3.5 Geotechnical Subsurface Soil Samples

One soil boring was completed at the Upper and Lower Cobbs Ponds AOC for the purposes of obtaining geotechnical parameters to perform vadose zone soil leaching and groundwater transport modeling. Two samples were collected from this soil boring, from the 4-5.4 ft bgs and 8-9.7 ft bgs intervals. Soil boring CPCsb-033 was advanced to a depth of 13 ft bgs, with groundwater encountered at 10 ft bgs. Bedrock was not encountered at this boring location. Table 5-2 summarizes the results of the geotechnical characteristics of soil at the Upper and Lower Cobbs Ponds. Laboratory analytical data package results are presented in Appendix D.

1 5.4 CONTAMINANT NATURE AND EXTENT IN SEDIMENT AND SURFACE WATER

2

3 Nine discrete sediment samples (0-0.5 ft bgs and 0.5-2 ft bgs) were collected at the at the Upper and 4 Lower Cobbs Ponds AOC during the PBA08 RI; four were located within the Backwater Area (two 5 surface and two subsurface), two were located within the Upper Cobbs Pond (one surface and one subsurface), and three were located within the Lower Cobbs Pond (two surface and one subsurface). 6 7 Eight sediment samples were analyzed for RVAAP full-suite analytes. Tables 4-24 through 4-31 8 present the results of the SRC screening for sediment at the Backwater Area, Upper Cobbs Pond, and 9 Lower Cobbs Pond aggregates, respectively. The 2003 FWBWQS sediment samples were evaluated 10 separately from the other studies because they were ISM samples and the other studies took discrete 11 samples. The results of the detected analytes for the PBA08 RI sediment samples are summarized in 12 Table 4-18 for the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond aggregates, 13 respectively. Complete copies of the laboratory analytical packages are presented in Appendix D.

14

15 Under the PBA08 RI sampling activities, samples were collected at the 0-0.5 and 0.5-2 ft bgs intervals at a subset of locations in order to provide characterization of vertical extent. The PBA08 RI 16 17 samples collected within the 0.5-2 ft bgs interval were screened as subsurface sediment. The surface 18 sediment screening data set includes both the PBA08 RI samples from the 0-0.5 ft bgs interval and 19 historical sediment samples. In each of the SRC screening tables, the results for surface sediment and 20 subsurface sediment are presented. Historical sediment samples from the 1996 Phase I RI are retained 21 in the discussion of nature and extent, but were omitted from the SRC screening process since more 22 recent data from the 2001 Phase II RI and the 2010 PBA08 RI activities are available.

23

24 Surface water is present perennially within the AOC, which receives drainage from the Load Line 3 25 and Load Line 12 channels and conveys the flow north through the Backwater Area, Upper Cobbs 26 Pond, and Lower Cobbs Pond. Five surface water locations were sampled during the PBA08 RI; two 27 at the Backwater Area, one at Upper Cobbs Pond, and two at Lower Cobbs Pond. The 2010 PBA08 28 RI, the 2001 Phase II RI, and 2003 FWBWQS surface water samples were utilized in the SRC 29 screening data set. Tables 4-32 through 4-34 summarize the results of the surface water SRC 30 screening for the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond, respectively. The 31 results of the detected analytes for the PBA08 RI surface water samples are summarized in Table 4-32 17 for the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. Water quality parameters measured at the time of surface water sample collection are presented in Table 5-3. Complete copies 33 34 of laboratory analytical packages are presented in Appendix D.

35

36 5.4.1 Explosives and Propellants

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38 Backwater Area

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40 Three explosives (TNT; 2,6-DNT; and tetryl) and one propellant (nitrocellulose) were identified as 41 SRCs in surface sediment samples at the Backwater Area. The maximum detections of TNT and 2,6-42 DNT were observed at Phase II RI sample location ULCPsd-005 at the mouth of the channel draining 43 from Load Line 3 into the Backwater Area (Figure 5-6). Nitrocellulose and tetryl were detected at

44 their maximum concentrations at PBA08 RI location CPCsd-047 in surface sediment but not in the

1 subsurface interval. The detected concentrations were below their respective SLs and were not

2 considered to be COPCs. No explosives or propellants were detected in the subsurface sediment

- 3 samples at the Backwater Area.
- 4

5 The explosive 4-amino-2,6-DNT was identified as an SRC in surface water at the Backwater Area. 4-6 amino-2,6-DNT was detected in both PBA08 RI sample locations for this aggregate at low, estimated 7 concentrations below laboratory reporting limits. No propellants were detected in the Backwater Area 8 surface water.

9

10 Upper Cobbs Pond

11

Five explosives (1,3-dinitrobenzene; TNT; 4-amino-DNT; HMX; and tetryl) and one propellant (nitrocellulose) were identified as SRCs in surface sediment at Upper Cobbs Pond. All of the SRCs were detected at PBA08 RI sample location CPCsd-046, but were not detected in the subsurface sediment sample collected at the same location. Sample location CPCsd-046 was collected from the center of the Upper Cobbs Pond (Figure 5-6). The detected concentrations were below their respective SLs and were not considered to be COPCs. No explosives or propellants were detected in the subsurface sediment at the Upper Cobbs Pond.

19

The explosive 4-amino-2,6-DNT was identified as an SRC in surface water at Upper Cobbs Pond. 4-Amino-2,6-DNT was detected in the PBA08 RI sample CPCsw-046 at a low, estimated concentration below the laboratory reporting limit (0.00007J mg/L) and in the June 2003 sample from the FWBWQS Upper Cobbs Pond station (0.00024J mg/L). No propellants were detected in the Upper Cobbs Pond surface water.

25

26 Lower Cobbs Pond

27

Two explosives (HMX and tetryl) and one propellant (nitrocellulose) were identified as SRCs in surface sediment at Lower Cobbs Pond. HMX was detected only at PBA08 RI location CPCsd-044, and nitrocellulose and tetryl were only detected at PBA08 RI location CPCsd-045. The detected concentrations were below their respective SLs and were not considered to be COPCs.

32

HMX and propellant nitrocellulose also were identified as SRCs in subsurface sediment at Lower Cobbs Pond. Nitrocellulose was detected at a higher concentration in the corresponding surface sediment sample at location CPCsd-045, at 7.8J mg/kg versus 1.9J mg/kg. HMX was not detected in the corresponding surface sediment sample.

37

The explosive 4-amino-2,6-DNT was identified as an SRC in surface water at Lower Cobbs Pond. 4-Amino-2,6-DNT was detected at a concentration of 0.00029J mg/L in the August 2003 FWBWQS

40 sample at Lower Cobbs Pond. 4-Amino-2,6-DNT was not detected in the subsequent PBA08 RI

41 samples taken at CPCsw-044 and CPCsw-045 in 2010. No propellants were detected in surface water

- 42 at the Lower Cobbs Pond.
- 43

| 1 2 | 5.4.2 | Inorganic Chemicals |
|--|---|---|
| 2 | Rockw | ater Area |
| 4 | Dackw | alti Aita |
| 5 6 7 | | inorganic chemicals (14 metals, cyanide, and nitrate/nitrite) were identified as SRCs in sediment samples in the Backwater Area. |
| 8 9 10 | backgro | sixteen inorganic chemicals, three (aluminum, cobalt, and cyanide) were detected above the bund concentration and/or exceeded their respective SLs in order to be identified as COPCs as ed below. |
| 11 12 13 14 15 16 17 18 19 20 21 22 | • | Aluminum was detected above the background concentration of 13,900 mg/kg and its respective SL of 3,496 mg/kg at four surface sediment sample locations. The maximum concentration of 18,600 mg/kg was observed at Phase II RI sample location ULCPsd-013. Cobalt was detected above the background concentration of 9.1 mg/kg and its respective SL of 2.3 mg/kg at eleven surface sediment sample locations, ranging from 10.6 mg/kg (CPCsd-048) to 22.8 mg/kg (CPCsd-047). Cobalt was detected below the background concentration at the Phase II RI sample collected at a nearby location (ULCPsd-015) (See Figure 5-7). Cyanide was analyzed in 12 of the surface sediment samples and was detected in 3 samples; however, only one sample exceeded its respective SL of 0.27 mg/kg. The concentration of 0.55 mg/kg was detected in in the Phase II RI sample UCLPsd-009. |
| 23 24 25 26 27 | and exc | 5-7 presents the locations with concentrations that were above the background concentration ceeded their respective SLs for these chemicals. Concentrations of chromium exceeded the SL cavalent chromium (1.64 mg/kg) but did not exceed the SL for trivalent chromium (8,147 |
| 28 29 30 31 32 33 34 | cadmius were go the five above t | subsurface sediment in the Backwater Area, five inorganic chemicals (antimony, beryllium, m, cobalt, and silver) were identified as SRCs. Overall concentrations of the inorganic SRCs enerally lower in the subsurface interval than their corresponding surface sample. Only one of inorganic chemicals identified as SOCs were determined to be a COPC. Cobalt was detected he background concentration (9.1 mg/kg) and its respective SL (2.3 mg/kg) in one subsurface nt location (CPCsd-048) in the interval from 0.5-2 ft bgs. |
| 35 36 37 | • | n inorganic chemicals (16 metals, sulfate, and sulfide) were identified as SRCs in surface n the Backwater Area. |
| 38 39 40 41 | the bac | eighteen inorganic chemicals, four (arsenic, cobalt, manganese, selenium) were detected above kground concentration and/or exceeded their respective SLs in order to be identified as COPCs ribed below. |
| 42 43 44 | • | Arsenic was detected above the background concentration of 0.0032 mg/L and its respective SL of 0.0011 mg/L at two surface water sample locations. The maximum concentration of 0.0104 mg/L was observed at Phase II RI sample location ULCPsw-001. |

Cobalt exceeded its respective SL of 0.0006 mg/L at two surface water sample locations. The 1 • 2 maximum concentration of 0.0107 mg/L was observed at Phase II RI sample location 3 ULCPsw-001. 4 • Manganese was detected above the background concentration of 0.391 mg/L and its 5 respective SL of 0.633 mg/L at two surface water sample locations. The maximum 6 concentration of 15.8 was observed at Phase II RI sample location ULCPsw-002. 7 Selenium exceeded its respective SL of 0.01 mg/L at one surface water sample location • 8 (ULCPsw-001) with a concentration of 0.0104 mg/L. 9 10 As shown on Figure 5-7, concentrations of manganese at ULCPsw-001 and ULCPsw-002 and cobalt 11 at ULCPsw-001 exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ 12 of 1. 13 14 **Upper Cobbs Pond** 15 Fifteen inorganic chemicals were identified as SRCs in the surface sediment in the Upper Cobbs 16 17 Pond. Of the fifteen inorganic chemicals, three [aluminum, chromium (hexavalent), and cobalt] were 18 detected above the background concentration and/or exceeded their respective SLs in order to be 19 identified as COPCs as described below. 20 21 Aluminum was detected above the background concentration of 13.900 mg/kg and its • 22 respective SL of 3,496 mg/kg at four surface sediment sample locations. The maximum 23 concentration of 21,400 mg/kg was observed at Phase II RI sample location ULCPsd-017. 24 • Hexavalent chromium exceeded its respective SL of 1.64 mg/kg at one surface sediment 25 location (ULCPsd-018) with a concentration of 10.6 mg/kg. 26 • Cobalt was detected above the background concentration of 9.1 mg/kg and its respective SL 27 of 2.3 mg/kg at four surface sediment sample locations, ranging from 10.9 mg/kg (CPCsd-28 020) to 19.3 mg/kg (ULCPsd-017). 29 30 Figure 5-7 presents the locations with concentrations that were above the background concentration 31 and exceeded their respective SLs for these chemicals. Concentrations of total chromium exceeded 32 the SL for hexavalent chromium (1.64 mg/kg) but did not exceed the SL for trivalent chromium 33 (8,147 mg/kg). 34 35 In subsurface sediment in the Upper Cobbs Pond, five inorganic chemicals (antimony, beryllium, 36 cadmium, cobalt, and silver) were identified as SRCs. These also were identified as SRCs in surface 37 sediment. Overall concentrations were lower in the subsurface interval than their corresponding 38 surface sample, as fourteen inorganic chemicals were detected above background concentrations in 39 the 0-0.5 ft bgs interval at location CPCsd-046. The concentration of cobalt in the 0.5-2 ft bgs 40 interval (10.1 mg/kg) was detected above the background concentration of 9.1 mg/kg and exceeded 41 its respective SL of 2.3 mg/kg. Cobalt was the only subsurface sediment SRC to be identified as a 42 COPC. 43

1 Seven metals, nitrate/nitrite, sulfate, and sulfide were identified as SRCs in surface water in the Upper 2 Cobbs Pond. All detections in PBA08 RI sample CPCsw-046 were at low, estimated concentrations 3 below laboratory reporting limits. Although manganese was detected above the background 4 concentration of 0.391 mg/L in historical sample ULCPsw-003 (0.528 mg/L), manganese was not 5 detected above the background concentration in the samples from the 2003 FWBWQS or PBA08 RI. No inorganic chemicals exceeded their respective SLs and were considered to be COPCs. 6 7 Nitrate/nitrite, sulfate, and sulfide were detected but do not have respective SLs. 8 9 **Lower Cobbs Pond** 10 11 Seventeen inorganic chemicals were identified as SRCs in the surface sediment in the Lower Cobbs

Pond. Of the seventeen inorganic chemicals, five [aluminum, arsenic, chromium (hexavalent), cobalt, and cyanide] were detected above the background concentration and/or exceeded their respective SLs in order to be considered COPCs as described below.

- 15
- Aluminum was detected above the background concentration of 13,900 mg/kg and its
 respective SL of 3,496 mg/kg at four surface sediment sample locations. The maximum
 concentration of 16,400 mg/kg was observed at Phase II RI sample location ULCPsd-026.
- Arsenic was detected above the background concentration of 19.5 mg/kg and its respective
 SL of 0.425 mg/kg at two surface sediment sample locations. The maximum concentration of
 34.3 mg/kg was observed at Phase II RI sample location ULCPsd-026.
 - Hexavalent chromium exceeded its respective SL of 2.3 mg/kg at two surface sediment location (ULCPsd-021 and ULCPsd-022) with concentrations of 5.7 and 5 mg/kg.
- Cobalt was detected above the background concentration of 9.1 mg/kg and its respective SL
 of 2.3 mg/kg at eight surface sediment sample locations, ranging from 9.3 mg/kg (ULCPsd 024 and ULCOsd-026) to 21.3 mg/kg (ULCPsd-025).
 - Cyanide was analyzed in 6 surface sediment samples and was detected in 1 sample (ULCPsd-024) with a concentration of 0.4 mg/kg that exceeded its respective SL of 0.27 mg/kg.
- 28 29

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Figure 5-7 presents the locations with concentrations that were above the background concentration
and exceeded their respective SLs for these chemicals. Concentrations of total chromium exceeded
the SL for hexavalent chromium (1.64 mg/kg) but did not exceed the SL for trivalent chromium
(8,147 mg/kg).

34

35 Seven inorganic chemicals were identified as SRCs in subsurface sediment (Table 4-31). These also 36 were identified as SRCs in surface sediment at Lower Cobbs Pond. Overall concentrations were lower 37 in the subsurface interval than their corresponding surface sample, as 15 inorganic chemicals were 38 detected above the background concentrations in the 0-0.5 ft bgs interval at location CPCsd-045, but 39 only 8 were above background from the 0.5-2 ft bgs interval. Concentrations of cobalt in the 0.5-2 ft 40 bgs interval were detected above the background concentration of 9.1 mg/kg and exceeded its 41 respective SL of 2.3 mg/kg. Cobalt was the only subsurface sediment SRC that was considered to be a 42 COPC.

Eleven metals, nitrate/nitrite, sulfate, and sulfide were identified as SRCs in surface water in the Lower Cobbs Pond. Of these inorganic chemicals, only manganese was considered a COPC. Manganese was detected above the background concentration of 0.391 mg/L at Phase II RI sample location ULCPsw-007 (0.654 mg/L) but was not detected above the background concentration in the two samples from the 2003 FWBWQS or the two PBA08 RI samples collected for this aggregate. Nitrate/nitrite, sulfate, and sulfide were detected but do not have respective SLs.

7 8

5.4.3 Semi-volatile Organic Compounds

9

10 Backwater Area

11

Fourteen SVOCs were identified as SRCs in surface sediment samples in the Backwater Area. Of the 12 13 fourteen SVOCs, 5 PAHs [benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, 14 dibenzo(a,h)anthracene, and indeno(1,2,3-cd)pyrene] exceeded their respective SLs. Four of the five 15 PAHs were detected at their maximum concentrations in the Phase II RI sample ULCPsd-015. PBA08 16 RI sample CPCsd-047 was collected adjacent to ULCPsd-015, and although all 14 SRCs were 17 detected at CPCsd-047, only 3 of these SVOCs were observed at their maximum concentrations, 18 indicating concentrations have declined since the 2001 Phase II RI sampling event. Concentrations of 19 benzo(a)pyrene at ULCPsd-006, ULCPsd-015, and CPCsd-047 exceeded the Resident Receptor 20 (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. The distribution of SVOCs in sediment in the 21 Backwater Area is shown in Figure 5-8.

22

Eleven SVOC SRCs were identified in subsurface sediment in the Backwater Area. These also were identified as SRCs for surface sediment. Only benzo(a)pyrene was detected above its respective SL (0.022 mg/kg) at PBA08 RI sample location CPCsd-047 in the 0.5-2 ft bgs interval and was considered to be a COPC in the subsurface sediment in the Backwater Area. The concentration was below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1.

28

29 No SVOCs were detected in surface water in the Backwater Area.

30

31 Upper Cobbs Pond

32

33 Ten SVOCs were identified as SRCs in the surface and subsurface sediment in the Upper Cobbs Pond 34 at the PBA08 RI sample location CPCsd-046. Two additional SVOCs (acenaphthylene and 35 anthracene) were identified in the subsurface sediment. The ten SVOCs occurred at higher 36 concentrations in surface sediment than in the corresponding subsurface sediment sample. 37 Benzo(a)pyrene was detected above its respective SL in both the surface and subsurface sediment at 38 CPCsd-046; however, the concentrations were below the Resident Receptor (Adult and Child) 39 FWCUG at a TR of 1E-05, HQ of 1. SVOCs were not detected in the Phase II RI sediment sample 40 ULCPsd-020. PAHs were not detected in the ISM sample from the 2003 FWBWQS; however, three 41 phthalates [bis(2-ethylhexyl)phthalate, butyl benzyl phthalate, and di-n-butyl phthalate] were detected 42 at low concentrations.

1 SVOCs bis(2-ethylhexyl)phthalate and di-n-butyl phthalate were identified as SRCs in surface water 2 in the Upper Cobbs Pond. The detected concentrations were below their respective SLs; therefore, 3 they were not considered to be COPCs. Bis(2-ethylhexyl)phthalate was detected at low, estimated 4 concentrations below the laboratory reporting limit at PBA08 RI sample location CPCsw-046 and at 5 the 2003 FWBWQS sample from August 2003. Di-n-butyl phthalate was detected at low, estimated 6 concentrations below the laboratory reporting limit in the 2003 FWBWQS samples from June and 7 August 2003.

8

9 Lower Cobbs Pond

10

11 Eleven SVOCs were identified as SRCs in the surface sediment in the Lower Cobbs Pond and 12 eighteen SVOCs were identified as SRCs in the subsurface sediment (Table 4-31). The 11 SVOC 13 SRCs that were identified for surface sediment also were observed in subsurface sediment and 14 occurred at their highest concentrations in the PBA08 RI subsurface sediment sample CPCsd-045 15 (Table 4-31). Concentrations of benzo(a)pyrene exceeded its respective SL in the surface and 16 subsurface sediment at CPCsd-045; however, only the subsurface sediment concentration exceeded 17 the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. There were no SVOCs 18 detected above their SLs in the ISM sediment sample taken in 2003.

19

Four SVOCs [benzenemethanol, bis(2-ethylhexyl)phthalate, butyl benzyl phthalate, and di-n-butyl phthalate] were identified as SRCs in surface water in the Lower Cobbs Pond. Benzenemethanol and butyl benzyl phthalate were detected at low, estimated concentrations at PBA08 RI location CPCsw-045 at the center of the aggregate. No SVOCs were detected at CPCsw-044, located at the dam overflow spillway from Lower Cobbs Pond. Bis(2-ethylhexyl)phthalate and di-n-butyl phthalate were detected in both samples taken in the 2003 FWBWQS samples from June and August 2003. There were no SVOCs detected above their respective SLs in surface water at Lower Cobbs Pond.

27

28 5.4.4 Volatile Organic Compounds, Pesticides, and Polychlorinated Biphenyls

29

30 Backwater Area

31

32 Three VOCs (acetone, methylene chloride, and 2-butanone) were identified as SRCs in the surface 33 sediment samples in the Backwater Area. As shown on Figure 5-9, 2-butanone also was identified as 34 the only VOC SRC in subsurface sediment. The maximum concentration of 2-butanone and acetone 35 was observed at the PBA08 RI sample location CPCsd-047. PCB-1254 was identified as an SRC in 36 surface sediment, and it was detected in the Phase II RI sample locations ULCPsd-006 and ULCPsd-37 011. No PCBs were detected in any of the PBA08 RI surface or subsurface sediment samples. No 38 pesticides were detected in the surface or subsurface sediment samples collected in the Backwater 39 Area. No VOCs, pesticides, or PCBs exceeded their respective SLs in sediment at the Backwater 40 Area.

41

42 Three VOCs (acetone, chloromethane, and toluene) were identified as SRCs in surface water in the 43 Backwater Area. All three of the VOCs were detected in the Phase II RI sample ULCPsw-002. No 44 VOCs were detected in the two PBA08 RI surface water samples at this aggregate. No VOCs, 1 detected exceeded their respective SLs in surface water at the Backwater Area. No PCBs or pesticides

2 were detected in surface water at the Backwater Area.

3

4 Upper Cobbs Pond

5

Two VOCs (acetone and 2-butanone) were identified as SRCs in the surface sediment in the Upper 6 7 Cobbs Pond. The maximum concentrations were detected at the PBA08 RI sample location CPCsd-8 046. These VOCs also were detected at the 2001 Phase II RI sample location ULCPsd-020 at lower 9 concentrations. PCB-1254 also was identified as an SRC in the surface sediment in the Upper Cobbs 10 Pond and it was detected in ULCPsd-020. However, PCB-1254 was not detected in CPCsd-046. 11 Pesticides were not detected in surface sediment. One pesticide (delta-BHC) and one VOC (2-12 butanone) were identified as SRCs in subsurface sediment at Upper Cobbs Pond. PCBs were not 13 detected in the subsurface sediment. No VOCs, pesticides, or PCBs exceeded their respective SLs in 14 sediment at Upper Cobbs Pond (although delta-BHC has no SL).

15

16 No VOCs, PCBs, or pesticides were detected in surface water in the Upper Cobbs Pond.

17

18 Lower Cobbs Pond

19

20 Three VOCs (2-butanone, acetone, and carbon disulfide) were detected in the surface sediment in the 21 Lower Cobbs Pond. All three of the VOCs were detected at the PBA08 RI sample location CPCsd-22 045 at their maximum concentrations. Acetone was detected at 0.33J mg/L adjacent to CPCsd-045 23 (0.17 mg/L) in the 1996 sample CPCsd-007(p) which was not used in the screening process because 24 more recent data was available. VOCs were also detected in the Phase II RI sample location ULCPsd-25 021. The VOC 2-butanone was identified as an SRC in the subsurface sediment the Lower Cobbs 26 Pond. 2-Butanone was detected in the corresponding surface sediment sample at location CPCsd-045 27 at a higher concentration than observed in the subsurface interval. No PCBs or pesticides were 28 detected in either surface or subsurface sediment at Lower Cobbs Pond.

29

30 No VOCs, PCBs, or pesticides were detected in surface water in the Lower Cobbs Pond.

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32 33

5.5 SUMMARY OF CONTAMINANT NATURE AND EXTENT

From 1965–1967, Hercules Alcor, Inc. leased Building FF-19 in Load Line 12 for producing aluminum chloride. On November 15, 1966, a fish kill occurred at Lower Cobbs Pond as a result of improper handling of aluminum chloride during manufacturing operations. The bulk of the aluminum chloride was collected and disposed at Ramsdell Quarry Landfill (RVAAP-01). The pond, receiving the contaminating waste from drainage ditches, was settled, drained, and the contaminants were removed to Ramsdell Quarry.

40

Data from the 2001 Phase II RI, the 2003 FWBWQS, and the 2010 PBA08 RI, collected after contaminant removal response from the 1966 fish kill, were used to identify remaining SRCs at the Upper and Lower Cobbs Ponds AOC. This data set effectively characterizes the nature and extent of contamination at the AOC. To support the evaluation of the nature and extent of contamination, SRC concentrations were compared to SLs corresponding to the lowest FWCUG for the Resident Receptor
 (Adult and Child) and the National Guard Trainee at a target HQ of 0.1 or TR of 1E-06, as presented
 in the FWCUG Report. If there was no FWCUG for a chemical, the USEPA RSL was used as the SL.
 Based on the information provided earlier in this section and the summary below, it can be concluded
 that the vertical and horizontal extent of contamination is defined, and no further sampling is needed
 to evaluate the Upper and Lower Cobbs Ponds.

- 7
- 8 5.5.1 Soil
- 9
- 10 5.5.1.1 <u>Surface Soil</u>
- 11

The predominant SRCs for surface soil at the Pond Bank at the AOC were inorganic chemicals and 12 13 SVOCs; the majority of which were PAHs. Eighteen inorganic chemicals (16 metals, cyanide, and 14 nitrate/nitrite) were identified as SRCs in surface soil. Elevated inorganic chemical concentrations 15 above background concentrations are observed in surface soil throughout the Pond Bank, and only 1 16 of 24 surface soil samples collected exhibited no inorganic chemicals above background 17 concentrations (ULCPss-009). However, the majority of elevated inorganic chemical concentrations 18 occurred at levels marginally in excess of background concentrations, as all but two concentrations 19 occurred at magnitudes less than twice their respective background concentrations. As such, no 20 apparent spatial trend in the distribution of inorganic chemicals throughout the surface soil in the 21 Pond Bank is evident.

22

Sixteen SVOCs, predominantly PAHs, were identified as SRCs in surface soil at the Pond Bank. The highest frequency and magnitudes of detection were observed at CPCsb-035, located on the western bank of the Upper Cobbs Pond, immediately south of Cobbs Pond Road. The pesticide beta-BHC and propellant nitrocellulose were also identified as SRCs in surface soil at the Pond Bank. PAH concentrations are an order of magnitude higher in sediment than those observed in the adjacent Pond Bank surface soil.

29

One propellant (nitrocellulose) and one pesticide (BHC) were identified as SRCs in the surface soil at
 the Pond Bank. Neither exceeded their respective SLs. No explosives, VOCs, or PCBs were detected
 or identified as SRCs in Pond Bank surface soil surrounding Upper Cobbs Pond and Lower Cobbs
 Pond.

34

35 5.5.1.2 <u>Subsurface Soil</u>

36

Four inorganic chemicals (cadmium, lead, selenium, and silver) were identified as SRCs in subsurface soil samples at the Pond Bank. Lead and selenium were detected above the background concentrations at only one sample location. Selenium was detected at 1.6 mg/kg versus the background concentration of 1.5 mg/kg, and lead was detected at 23.3 mg/kg versus the background concentration of 19.1 mg/kg. As such, no apparent spatial trend in the distribution of inorganic chemicals throughout the subsurface soil in the Pond Bank is evident.

1 Thirteen SVOCs, 12 of which were PAHs, were detected and identified as SRCs in subsurface soil at 2 the Pond Bank. Twelve of the 13 SVOCs were detected from the 1-4 ft bgs interval at CPCsb-035; the 3 majority of which were observed at low, estimated concentrations below laboratory reporting limits. 4 Sample location CPCsb-035 is on the western bank of Upper Cobbs Pond and near Cobbs Pond Road. 5 Concentrations of PAHs in the subsurface soil samples were generally comparable to those observed in the corresponding surface soil sample (i.e., within the same order of magnitude). One propellant 6 7 (nitrocellulose) was identified as an SRC in subsurface soil at the Pond Bank. No explosives, VOCs, 8 PCBs, or pesticides were detected or identified as SRCs in Pond Bank subsurface soil surrounding 9 Upper Cobbs Pond and Lower Cobbs Pond.

10

11 5.5.2 Sediment

12

14

13 5.5.2.1 Backwater Area

15 The Backwater Area was evaluated using fourteen surface sediment and two subsurface sediment 16 samples. Three explosives (TNT; 2,6-DNT; and tetryl) and one propellant (nitrocellulose) were 17 detected in the surface sediment. All of the detected concentrations were below the Resident Receptor 18 (Adult and Child) FWCUG and RSL at a TR of 1E-06, HO of 0.1. No explosives or propellants were 19 detected in the subsurface sediment in the Backwater Area. In surface sediment, concentrations of 20 aluminum, cobalt, and cyanide exceeded a TR of 1E-06, HQ of 0.1, but not a TR of 1E-05, HQ of 1. 21 Only concentrations of cobalt exceeded a TR of 1E-06, HQ of 0.1 in the subsurface sediment in one 22 subsurface soil location, but not at a TR of 1E-05, HQ of 1. Five PAHs (benzo(a)anthracene, 23 benzo(b)fluoranthene, dibenzo(a,h)anthracene, and benzo(a)pyrene, indeno(1,2,3-cd)pyrene) 24 exceeded their respective SLs in the surface sediment. Four of the five PAHs were detected at their 25 maximum concentrations in the Phase II RI sample ULCPsd-015. The detected concentrations in the 26 adjacent PBA08 RI sample CPCsd-047 indicates that concentrations have declined since the 2001 27 Phase II RI sampling event. Concentrations of benzo(a)pyrene at ULCPsd-006, ULCPsd-015, and 28 CPCsd-047 exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. 29 Only benzo(a)pyrene was detected above its respective SL (0.022 mg/kg) at PBA08 RI sample 30 location CPCsd-047 in the 0.5-2 ft bgs interval. The concentration was below the Resident Receptor 31 (Adult and Child) FWCUG at a TR of 1E-05, HO of 1. Three VOCs (acetone, methylene chloride, 32 and 2-butanone) and one VOC (2-butanone) were detected below their respective SLs in the surface 33 and subsurface sediment, respectively. PCB-1254 also was detected in the Phase II RI surface 34 sediment samples ULCPsd-006 and ULCPsd-011; however, it was not detected in the PBA08 RI 35 samples. No pesticides were detected in surface sediment. Pesticides and PCBs were not detected in 36 the Backwater Area subsurface sediment.

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38 5.5.2.2 Upper Cobbs Pond

39

Five surface discrete sediment, one surface ISM, and one subsurface sediment samples were used to
evaluate the Upper Cobbs Pond. Five explosives (1,3-dinitrobenzene; TNT; 4-amino-2,6-DNT;
HMX; and tetryl) and one propellant (nitrocellulose) were detected in the surface sediment at PBA08
RI sample location CPCsd-046. All the detected concentrations were below Resident Receptor (Adult
and Child) FWCUG and RSL at a TR of 1E-06, HQ of 0.1. No explosives or propellants were

1 detected in the subsurface sediment. In surface sediment, concentrations of aluminum, hexavalent 2 chromium, and cobalt exceeded a TR of 1E-06, HQ of 0.1, but not a TR of 1E-05, HQ of 1. Only 3 concentrations of cobalt exceeded a TR of 1E-06, HQ of 0.1 in the subsurface sediment, but not at a 4 TR of 1E-05, HQ of 1. Ten SVOCs were detected in the surface and subsurface sediment at the 5 PBA08 RI sample location CPCsd-046. Two additional SVOCs (acenaphthylene and anthracene) were detected in the subsurface sediment. The ten SVOCs occurred at higher concentrations in 6 7 surface sediment than in the corresponding subsurface sediment sample. Benzo(a)pyrene was 8 detected above its respective SL in both the surface and subsurface sediment at CPCsd-046; however, 9 the concentrations were below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, 10 HQ of 1. Two VOCs (acetone and 2-butanone) were detected in surface sediment below their 11 respective SLs, both of which were detected at their maximum concentrations at PBA08 RI location CPCsd-046. These VOCs also were detected at the 2001 Phase II RI sample location ULCPsd-020 at 12 13 lower concentrations than CPCsd-046. PCB-1254 also was detected in the surface sediment at Upper 14 Cobbs Pond in ULCPsd-020 below its respective SL. PCB-1254 was not detected in the PBA08 RI sample. Pesticides were not detected in surface sediment at Upper Cobbs Pond. The VOC 2-butanone 15 16 was detected in the subsurface sediment below the TR of 1E-05, HQ of 1. The pesticide delta-BHC 17 does not have an SL. PCBs were not detected in subsurface sediment.

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5.5.2.3 Lower Cobbs Pond

21 The Lower Cobbs Pond was evaluated using eight discrete surface sediment, one surface ISM, and 22 one subsurface sediment samples. Two explosives (HMX and tetryl) and one propellant 23 (nitrocellulose) were detected in the surface sediment, and the explosive HMX and propellant 24 nitrocellulose were detected in the subsurface sediment at Lower Cobbs Pond. The concentrations 25 were below the Resident Receptor (Adult and Child) FWCUG and RSL at a TR of 1E-06, HQ of 0.1. 26 In surface sediment, concentrations of aluminum, arsenic, hexavalent chromium, cobalt, and cyanide 27 exceeded a TR of 1E-06, HO of 0.1, but only arsenic exceeded a TR of 1E-05, HO of 1. Only 28 concentrations of cobalt exceeded a TR of 1E-06, HQ of 0.1 in the subsurface sediment, but not at a 29 TR of 1E-05, HQ of 1. Eleven SVOCs were detected in the surface sediment in the Lower Cobbs 30 Pond and eighteen SVOCs were detected in the subsurface sediment. The 11 SVOC SRCs that were 31 identified for surface sediment also were observed in subsurface sediment and occurred at their 32 highest concentrations in the PBA08 RI subsurface sediment sample CPCsd-045. Concentrations of 33 benzo(a)pyrene exceeded its respective SL in the surface subsurface sediment at CPCsd-045; 34 however, only the subsurface sediment concentration exceeded the Resident Receptor (Adult and 35 Child) FWCUG at a TR of 1E-05, HQ of 1. Three VOCs (2-butanone, acetone, and carbon disulfide) 36 were detected in surface sediment at Lower Cobbs Pond. All three of these VOCs were detected at 37 PBA08 RI location CPCsd-045, with 2-butanone and carbon disulfide observed at their maximum 38 concentrations in this sample location. 2-Butanone also was detected in the subsurface sediment. All 39 the VOCs detected in surface and subsurface soil were below their respective SLs. No PCBs or 40 pesticides were detected in either the surface or subsurface sediment in the Lower Cobbs Pond.

3 5.5.3.1 **Backwater** Area

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2

5 Four surface water samples were used to evaluate the Backwater Area. One explosive (4-amino-2,6-DNT) was detected in surface water at low, estimated concentrations below laboratory reporting 6 7 limits. No propellants were detected in the Backwater Area surface water. Concentrations of four 8 inorganic chemicals (arsenic, cobalt, manganese, and selenium) exceeded the exceeded the RSL at a 9 TR of 1E-06, HQ of 0.1. Concentrations of manganese at Phase II RI sample locations ULCPsw-001 10 and ULCPsw-002 and cobalt at ULCPsw-001 exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HO of 1. Three VOCs (acetone, chloromethane, and toluene) were 11 detected in surface water for the Backwater Area, all of which were detected in ULCPsw-002. No 12 13 propellants, SVOCs, PCBs, or pesticides were detected in surface water at the Backwater Area. 14 Elevated inorganic chemical concentrations and 4-amino-2.6-DNT are likely attributable to 15 continuing upstream sources as these constituents are observed in the PBA08 RI locations immediately downstream of the confluence of the drainage channels from Load Line 3 and Load 16 17 Line 12.

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5.5.3.2 **Upper Cobbs Pond**

21 Surface water in the Upper Cobbs Pond was evaluated using five surface water samples. One 22 explosive (4-amino-2,6-DNT) was detected at a low, estimated concentration in two surface water 23 samples. The concentrations were below the Resident Receptor (Adult and Child) FWCUG and RSL 24 at a TR of 1E-06, HQ of 0.1. No surface water concentrations for inorganic chemicals in the Upper 25 Cobbs Pond exceeded the SL at a TR of 1E-06, HQ of 0.1. The SVOCs bis(2-ethylhexyl)phthalate and di-n-butyl phthalate were detected at a low, estimated concentration below the laboratory 26 27 reporting limit at PBA08 RI sample location CPCsw-046 and the August 2003 FWBWOS location 28 respectively. No VOCs, PCBs, or pesticides were detected in surface water at Upper Cobbs Pond.

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30 5.5.3.3 Lower Cobbs Pond

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32 Lower Cobbs Pond surface water was evaluated using seven surface water samples. The explosive 4-33 amino-2,6-DNT was identified as an SRC in surface water at Lower Cobbs Pond. 4-Amino-2,6-DNT 34 was detected at a concentration of 0.00029J mg/L in the August 2003 FWBWOS sample at Lower 35 Cobbs Pond. 4-Amino-2.6-DNT was not detected in the subsequent PBA08 RI samples taken at 36 CPCsw-044 and CPCsw-045 in 2010. No propellants were detected. Only manganese was detected at 37 a concentration which exceeded the Resident Receptor (Adult and Child) FWCUG and RSL at a TR 38 of 1E-06, HQ of 0.1 at one sample location. Two SVOCs (benzenemethanol and butyl benzyl 39 phthalate) were detected at low, estimated concentrations at PBA08 RI location CPCsw-045, located 40 at the center of the aggregate. No VOCs, PCBs, or pesticides were detected in surface water in the 41 Lower Cobbs Pond.

| Sample Location | Hexavalent Chromium Concentration (mg/kg) | Total Chromium Concentration ^a (mg/kg) | Percent Hexavalent Chromium (%) | | | | | | |
|--------------------|---|---|---------------------------------------|--|--|--|--|--|--|
| | Sediment – Upper Cobbs Pond | | | | | | | | |
| CPCsd-049 <3.1 U | | 34.3 | NA | | | | | | |

^aBackground concentration for total chromium = 18.1 mg/kg. No background concentration is available for hexavalent chromium.

mg/kg = M illigrams per kilogram. NA = Not applicable; hexavalent chromium not detected in sample.

U = Not detected.

2

Table 5–2. Summary of Geotechnical Parameters

| Sample ID: | | |
|------------------------|---|--|
| Parameters | CPCsb-033-5117-SO | CPCsb-033-5118-SO |
| Depth | 4–5.4 ft bgs | 8–9.7 ft bgs |
| Porosity | 37.3 % | 39.8 % |
| Density | 1.75 g/cm^3 | 1.67 g/cm^3 |
| Moisture content | 19.2 % | 21.4 % |
| Total organic carbon | 2,700 mg/kg | 5,400 mg/kg |
| Size fraction analysis | 1.2 % gravel, 10.3 % sand, 51.3 % silt, | 0.7 % gravel, 0.6 % sand, 63.9 % silt, |
| Size fraction analysis | 37.2% clay | 34.8 % clay |
| Permeability (K) | 2.1E-07 cm/sec | 7.5E-07 cm/sec |

bgs = Below ground surface.

ID = Identification.

3

Table 5-3. Surface Water Field Measurements

| Location | Temperature (°C) | рН (S.U.) | Conductivity (mS/cm) | Dissolved Oxygen (mg/L) | Turbidity (NTU) |
|-----------|---------------------|--------------|-------------------------|----------------------------|--------------------|
| CPCsw-044 | 9.13 | 7.10 | 0.128 | 9.32 | 6.92 |
| CPCsw-045 | 10.92 | 7.87 | 0.116 | 12.86 | 5.48 |
| CPCsw-046 | 10.19 | 7.00 | 0.118 | 9.96 | 6.53 |
| CPCsw-047 | 12.94 | 9.61 | 0.140 | 12.25 | 6.17 |
| CPCsw-048 | 15.79 | 8.23 | 0.147 | 10.95 | 4.98 |

°C = degrees Celsius

mg/L = milligrams per Liter mS/cm =milli-Siemens per centimeter NTU = nephelometric turbidity units

S.U. = standard unit

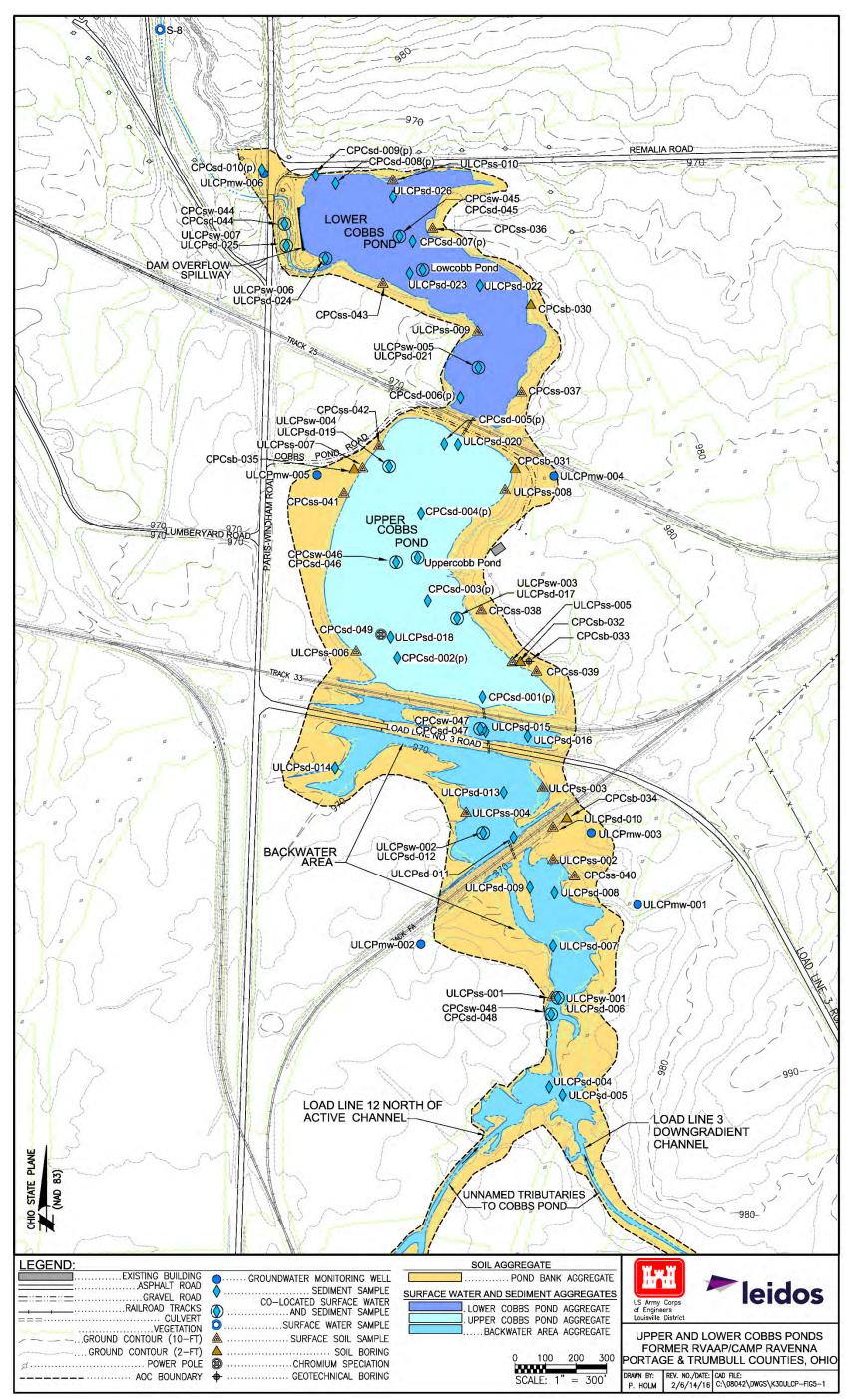


Figure 5-1. Data Aggregates and Sample Locations at Upper and Lower Cobbs Ponds

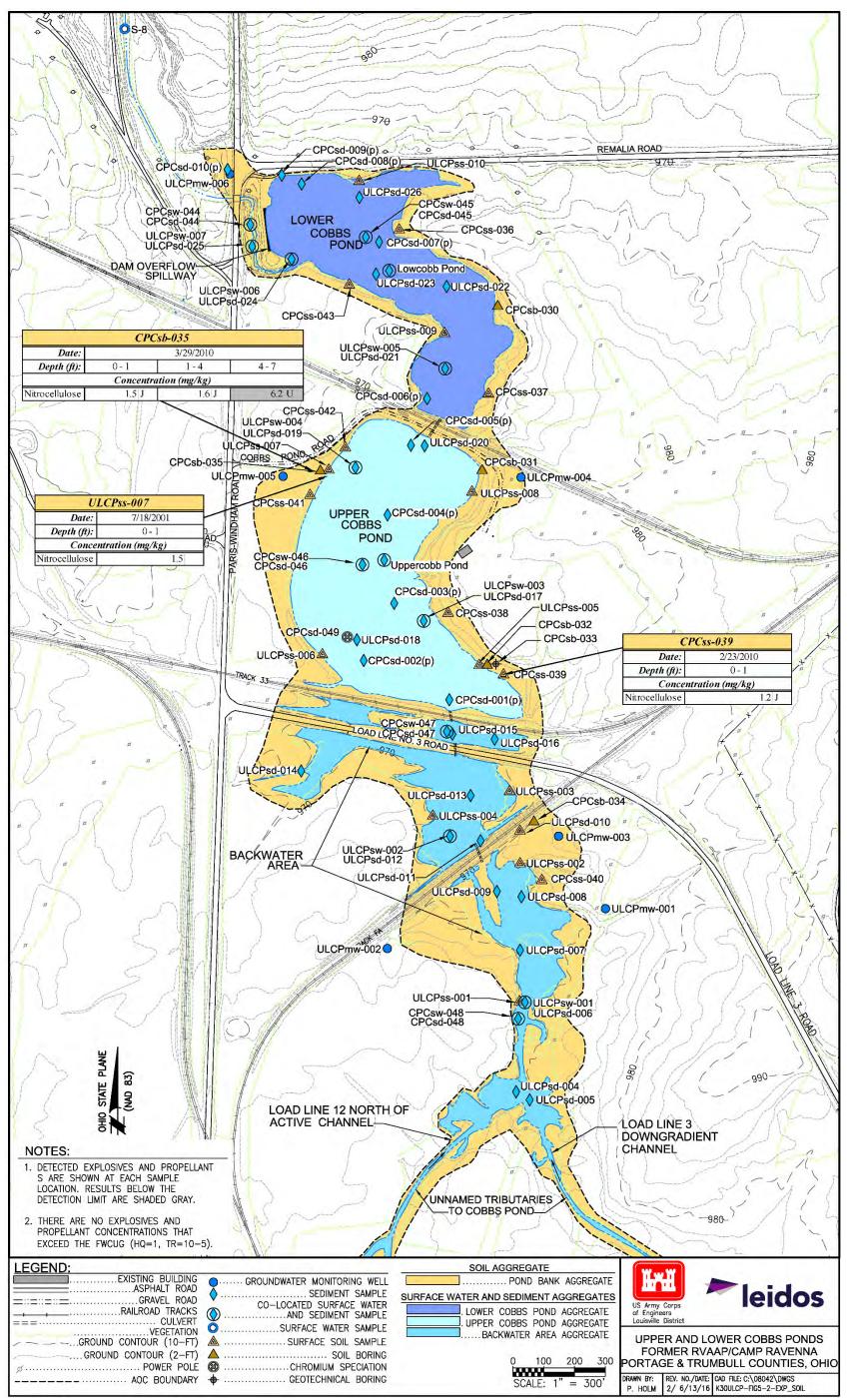
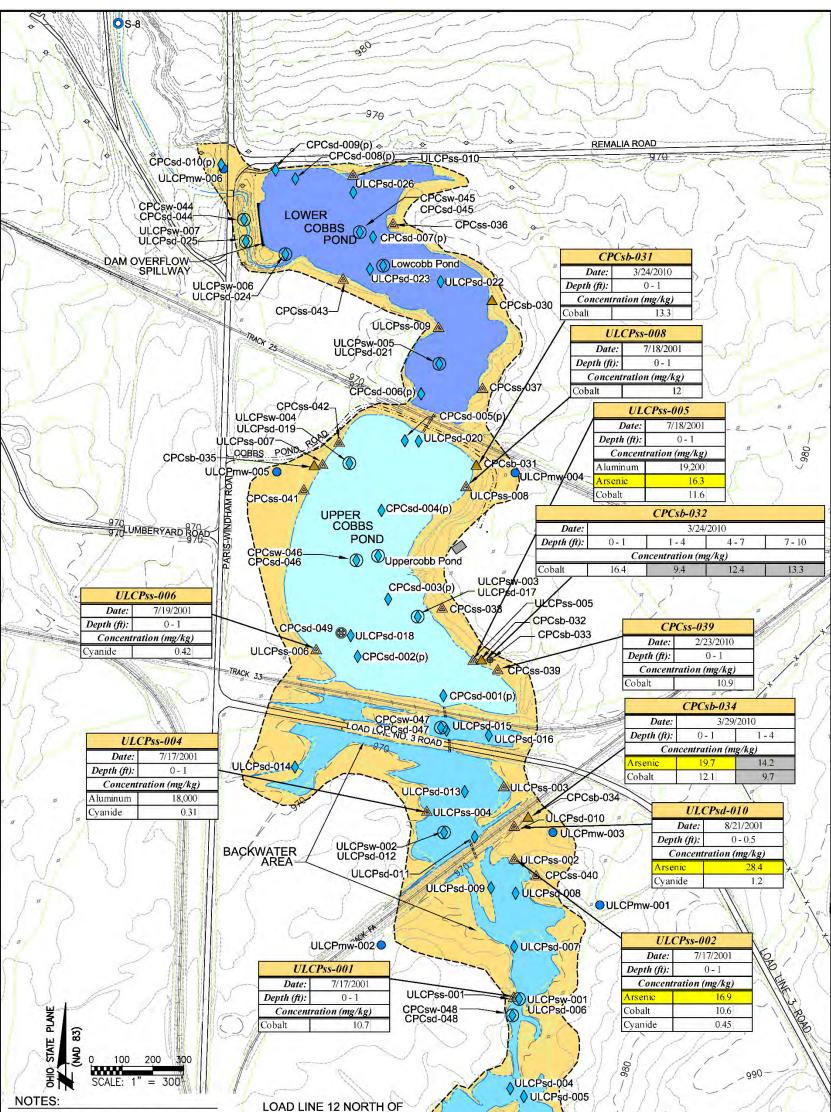


Figure 5-2. Detected Concentrations of Explosives and Propellants in Soil (Discrete Soil Borings)



| CONCENTRATIONS OF ALUMINUM, ARSENIC, COBALT, AND CYANIDE ARE SHOWN ONLY WHEN THEY EXCEEDED BACKGROUND AND THE FWCUG (HQ=0.1, TR=10-6). CONCENTRATIONS BELOW BACKGROUND OR THE FWCUG (HQ=0.1, TR=10-6) ARE SHADED GRAY. TOTAL CHROMIUM CONCENTRATIONS WERE SCREENED AGAINST THE FWCUG FOR TRIVALENT CHROMIUM AND DID NOT EXCEED THE FWCUG. | UNNAMED TRIBUTARIES TO COBBS POND | LOAD LINE 3 DOWNGRADIENT CHANNEL -980- |
|---|--------------------------------------|---|
| LEGEND: | LOWER COBBS POND ACCRECATE | US Army Corps of Engineers Louisville District |
| VEGETATION ✓ SURFACE WATER SAMPLE GROUND CONTOUR (10-FT) ▲ SURFACE SOIL GROUND CONTOUR (2-FT) ▲ SOIL BORING Ø POWER POLE Ø OC BOUNDARY ← GEOTECHNICAL BORING | CONCENTRATION EXCEEDS | UPPER AND LOWER COBBS PONDS FORMER RVAAP/CAMP RAVENNA PORTAGE & TRUMBULL COUNTIES, OHIO DRAWN BY: P. HOLM 2/6/15/16 K30ULCP-FIC5-3-MET_SOIL |

Figure 5–3. Exceedances of FWCUG (HQ of 0.1, TR of 10-6) for Aluminum, Arsenic, Cobalt, and Cyanide in Soil (Discrete Soil Borings)

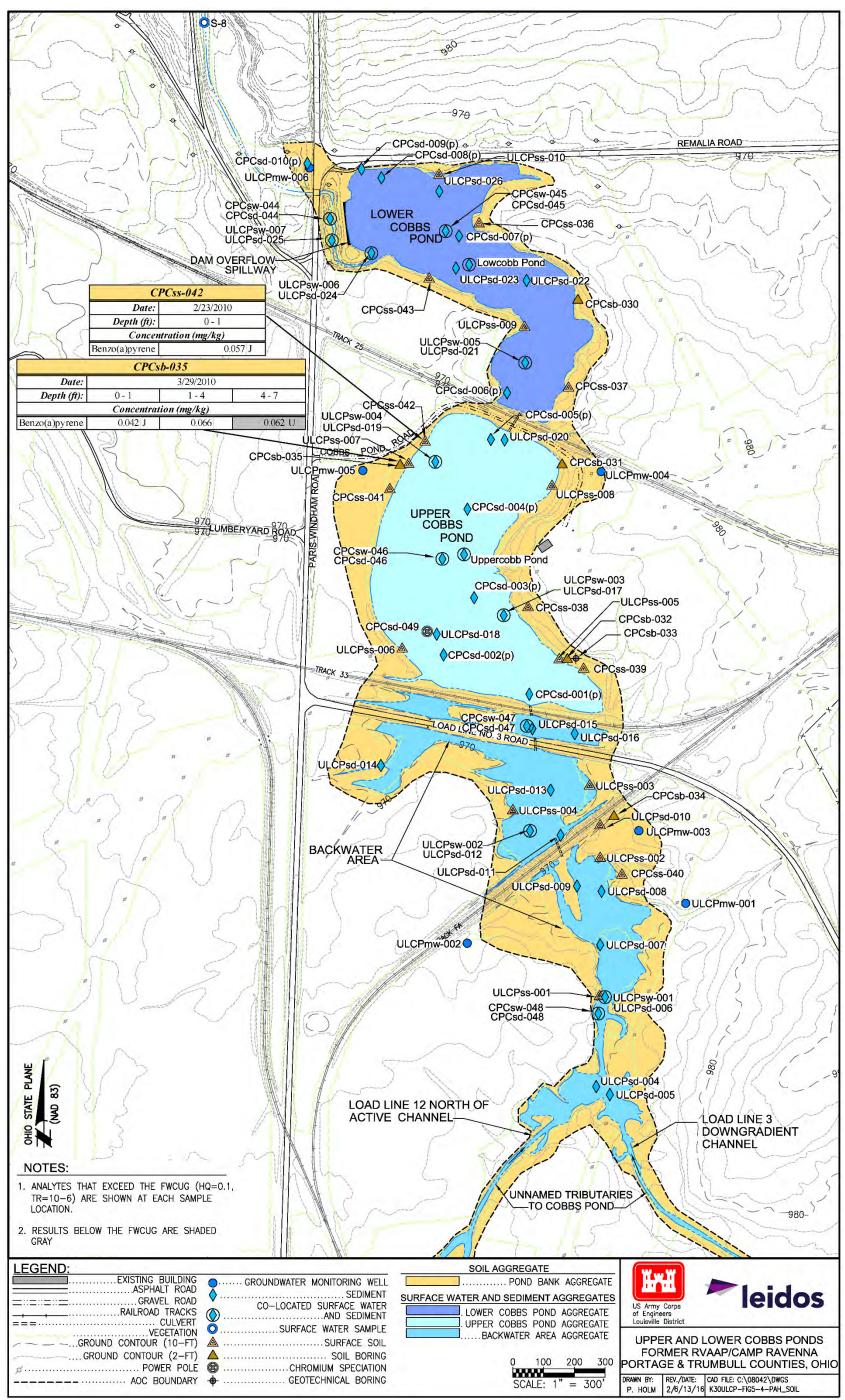


Figure 5-4. PAH Exceedances of FWCUG (HQ of 0.1, TR of 10-6) in Soil (Discrete Soil Borings)

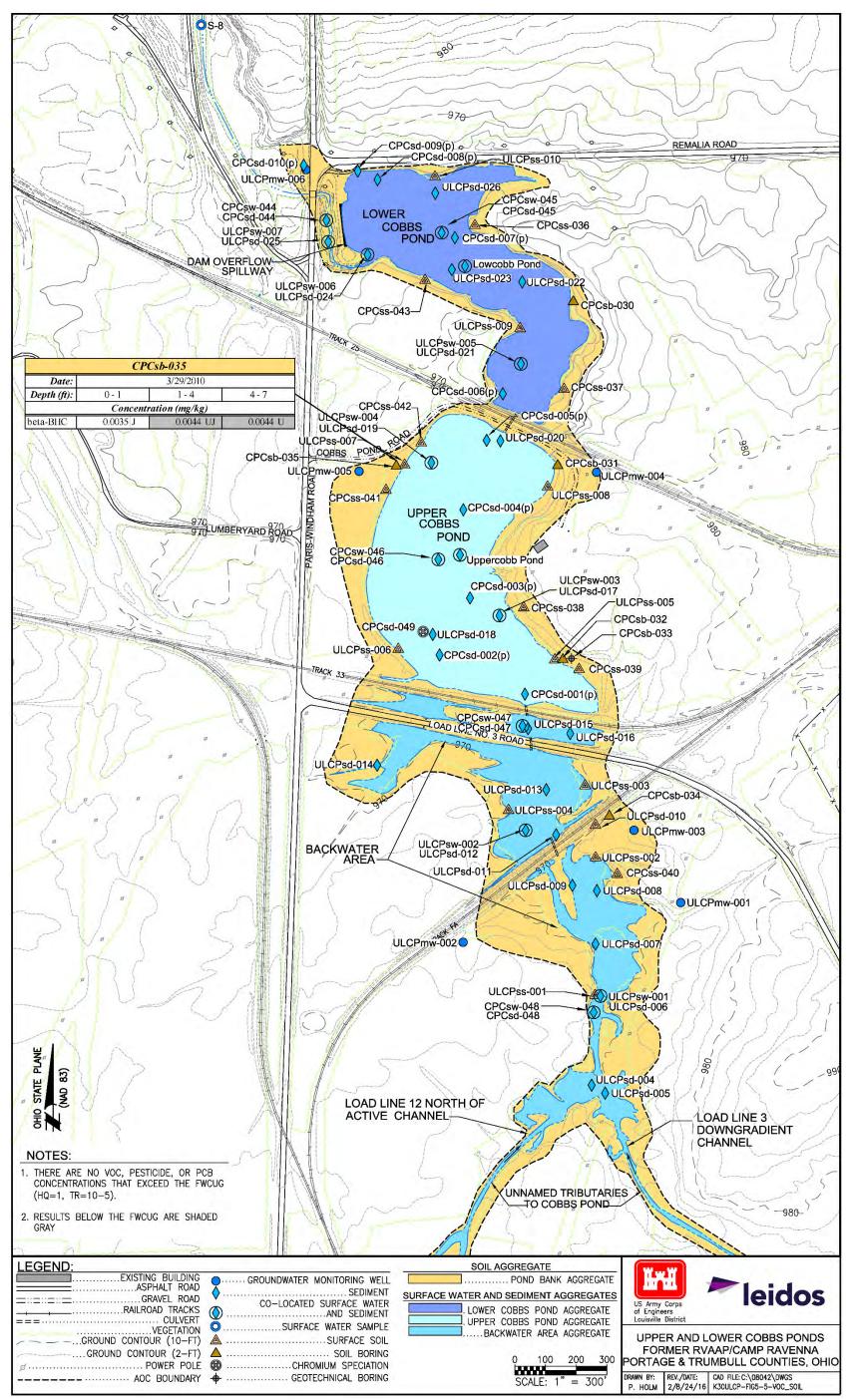


Figure 5-5. Detected Concentrations of VOCs, Pesticides, and PCBs in Soil (Discrete Soil Borings)

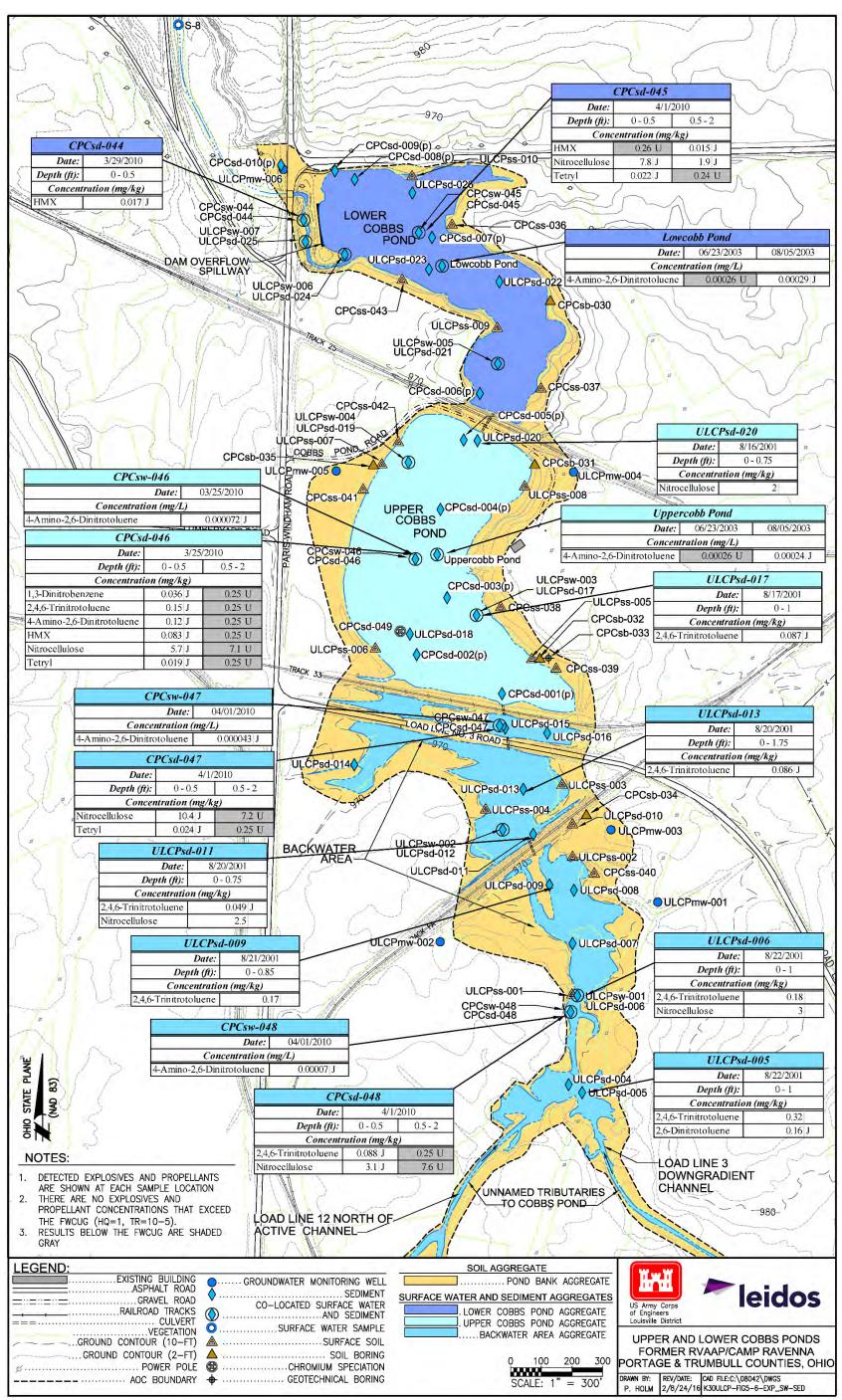


Figure 5-6. Detected Concentrations of Explosives and Propellants in Surface Water and Sediment

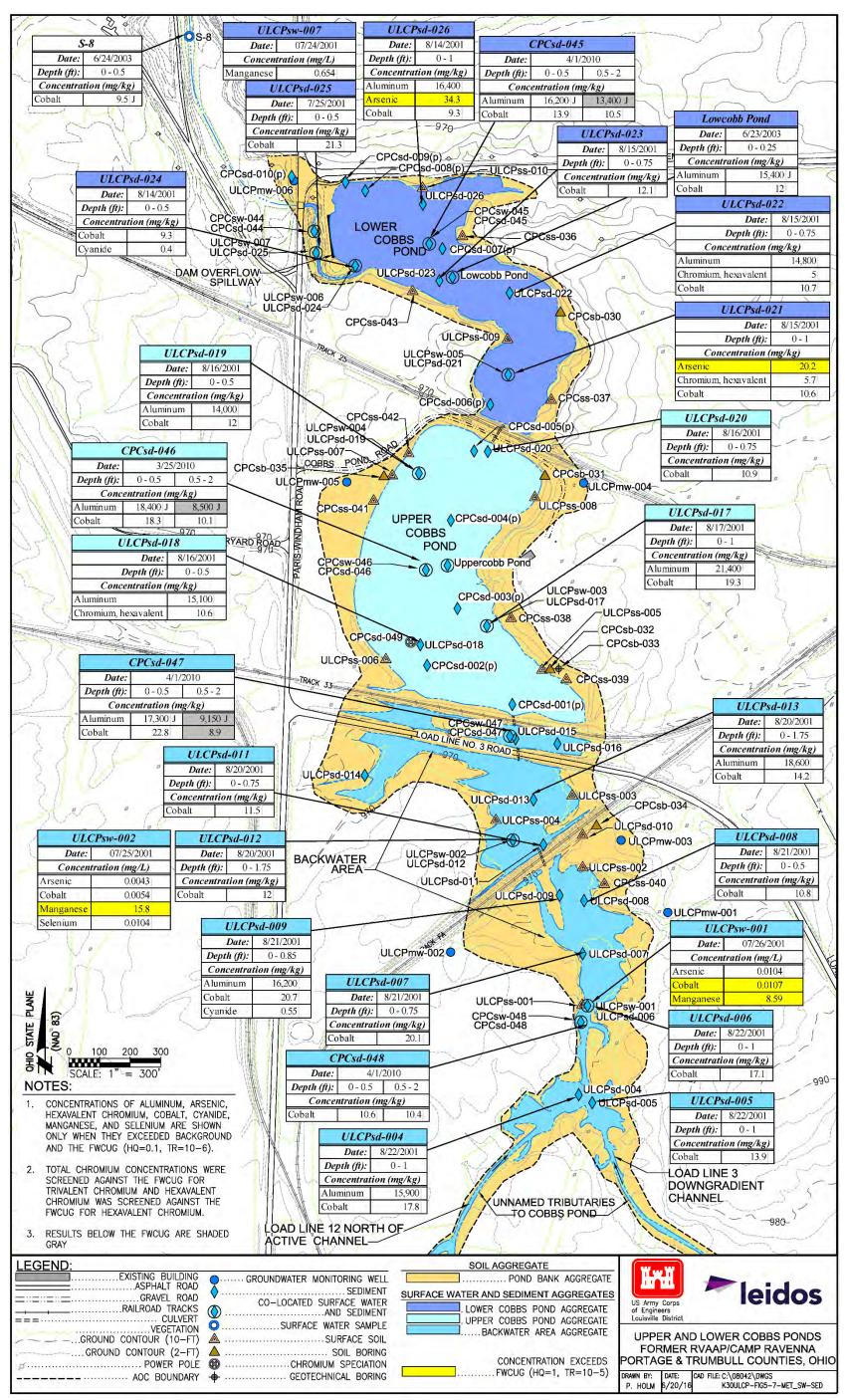


Figure 5–7. Exceedances of FWCUG (HQ of 0.1, TR of 10-6) for Aluminum, Arsenic, Hexavalent Chromium, Cobalt, Cyanide, Manganese, and Selenium in Surface Water and Sediment

1

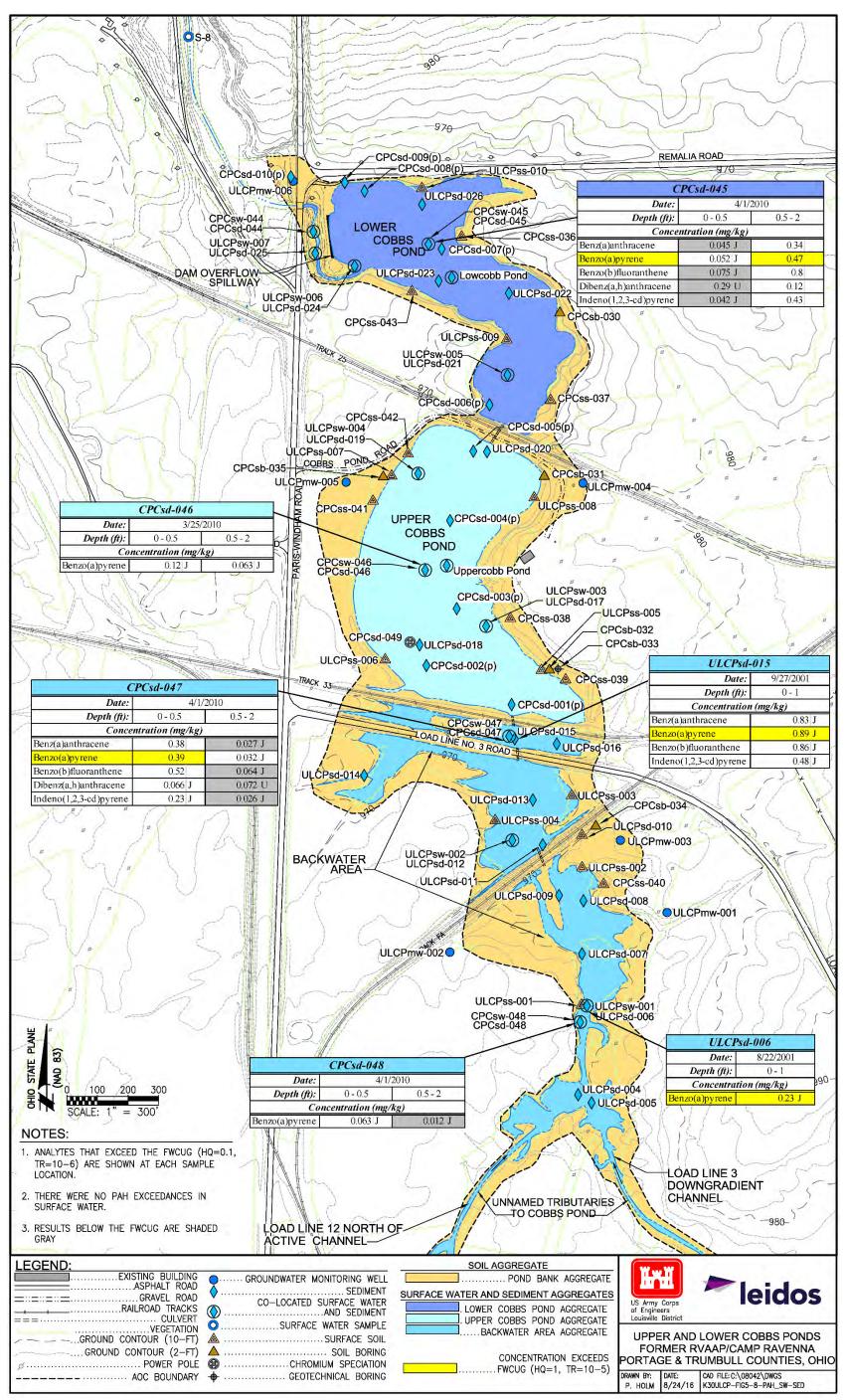


Figure 5-8. PAH Exceedances of FWCUG (HQ of 0.1, TR of 10-6) in Surface Water and Sediment

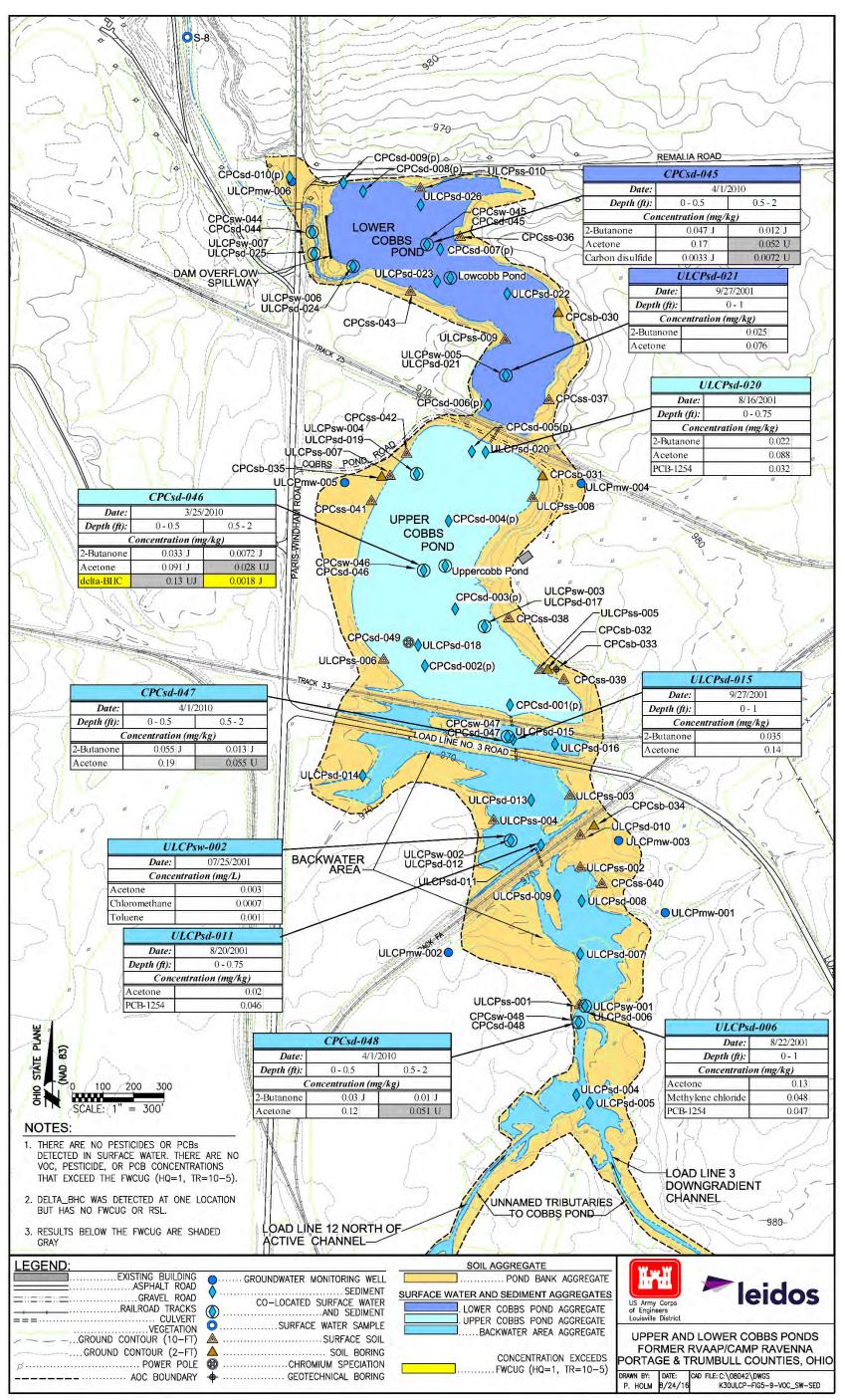


Figure 5-9. Detected Concentrations of VOCs, Pesticides, and PCBs in in Surface Water and Sediment

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6.0 **CONTAMINANT FATE AND TRANSPORT** 1

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Contaminant fate and transport modeling assesses the potential for SRCs to leach from surface soil, subsurface soil, and sediment sources at the Upper and Lower Cobbs Ponds AOC and impact groundwater beneath the sources and downgradient receptor locations. Modeling results were included in the decision-making process to determine whether performing remedial actions may be necessary to protect groundwater resources. Surface water exposure pathways are evaluated in the HHRA and ERA presented in Sections 7.2 and 7.3, respectively. A summary of the principles of contaminant fate and transport are presented in this section along with the results of the modeling.

9 10

11 Section 6.1 describes physical and chemical properties of SRCs found in soil and sediment at the 12 AOC. Section 6.2 presents a conceptual model for contaminant fate and transport that considers AOC 13 topography, hydrogeology, contaminant sources, and release mechanisms. Section 6.3 presents a soil 14 screening analysis, and Section 6.4 presents a sediment screening analysis to identify SRCs with the 15 potential to migrate from soil and sediment to groundwater as initial CMCOPCs. Section 6.5 describes fate and transport modeling of CMCOPCs and presents CMCOCs. Section 6.6 provides a 16 17 list of the remaining CMCOCs and a qualitative assessment of the results and considerations of the 18 limitations and assumptions. Section 6.7 presents the summary and conclusions of this fate and 19 transport analysis.

20

21 22

PHYSICAL AND CHEMICAL PROPERTIES OF SITE-RELATED CONTAMINANTS 6.1

23 The 1978 Installation Assessment identified the major contaminants of the former RVAAP to be 24 TNT, composition B, sulfates, nitrates, lead styphnate, and lead azide. The assessment identified the 25 pink wastewater and washout of residue, dusts, and spills at Load Lines 3 and 12, which were 26 discharged through surface drainage channels towards Upper Cobbs Pond (USATHAMA 1978), as 27 contaminants. Potential site-specific contaminants include TNT, RDX, HMX, nitrate, nitrocellulose, and heavy metals (i.e., lead, chromium, mercury, and arsenic) from the pink water. Other 28 29 contaminants identified include washout from explosives melt-out and demilitarization activities at 30 Load Lines 3 and 12 and ammonium nitrate and aluminum chloride from Load Line 12.

31

32 The evaluation of contaminant fate and transport not only includes chemicals identified as potential 33 contaminants from previous use but also includes chemicals that were evaluated as part of the overall 34 RI. The comprehensive list of surface and subsurface soil SRCs (18 inorganic chemicals and 19 35 organic chemicals) and surface and subsurface sediment SRCs (16 inorganic chemicals and 23 36 organic chemicals in the Backwater Area, 18 inorganic and 25 organic chemicals in Upper Cobbs 37 Pond, and 20 inorganic and 24 organic chemicals in Lower Cobbs Pond) were detailed in Section 4.0 38 and are summarized below:

39

40 41

Inorganic SRCs in surface and subsurface soil: aluminum, arsenic, barium, cadmium, • chromium, hexavalent chromium, cobalt, copper, cyanide, lead, mercury, nickel, selenium, 42 silver, thallium, vanadium, zinc, and nitrate/nitrite.

- Inorganic SRCs in sediment: aluminum, antimony, arsenic, barium, beryllium, cadmium,
 chromium, hexavalent chromium, cobalt, copper, cyanide, lead, mercury, nickel,
 nitrate/nitrite, selenium, silver, vanadium, ammonia, and total phosphorous.
- Organic SRCs in surface and subsurface soil: 2-methylnaphthalene, acenaphthylene, 4 • benz(a)anthracene, 5 benzenemethanol, benzo(a)pyrene, benzo(b)fluoranthene, bis(2-ethylhexyl)phthalate, 6 benzo(ghi)perylene, benzo(k)fluoranthene, benzoic acid, 7 chrysene, di-n-butyl phthalate, dibenz(a,h)anthracene, fluoranthene, indeno(1,2,3-cd)pyrene, 8 phenanthrene, pyrene, nitrocellulose, and beta-BHC.
- 9 Organic SRCs in sediment: 2-methylnaphthalene; acenaphthene; acenaphthylene; anthracene; • 10 benz(a)anthracene; benzo(a)pyrene; benzo(b)fluoranthene; benzo(ghi)pervlene; 11 benzo(k)fluoranthene; bis(2-ethylhexyl)phthalate; butyl benzyl phthalate; chrysene; 12 dibenz(a,h)anthracene; di-n-butyl phthalate; fluoranthene; fluorene; indeno(1,2,3-cd)pyrene; naphthalene; phenanthrene; pyrene; 2-butanone; acetone; carbon disulfide; methylene 13 14 chloride; 1,3-dinitrobenzene; TNT; 2,6-DNT; 4-amino-2,6-DNT; HMX; nitrocellulose; tetryl; 15 delta-BHC; and PCB-1254.
- 16

17 Chemicals released into the environment are susceptible to several degradation pathways, including 18 hydrolysis, oxidation, reduction, isomerization, photolysis, photo-oxidation, biotransformation, and 19 biodegradation. Transformed products resulting from these processes may behave differently than 20 their parent chemical in the environment.

21

22 The migration of chemicals is governed by their physical and chemical properties and the surface and 23 subsurface media through which chemicals are transferred. In general, chemicals and structures with 24 similar physical and chemical characteristics will show similar patterns of transformation, transport, 25 or attenuation in the environment. Solubility, vapor pressure data, chemical partitioning coefficients, 26 degradation rates, and Henry's Law Constant (HLC) provide information that can be used to evaluate 27 contaminant mobility in the environment. Partitioning coefficients are used to assess relative affinities 28 of chemicals for solution or solid phase adsorption. However, the synergistic effects of multiple 29 migrating chemicals and complexity of soil/water interactions, including pH and oxidation-reduction 30 potential, grain size, and clay mineral variability, are typically unknown.

31

The physical properties of the chemicals defined as SRCs in surface and subsurface soil and sediment are summarized in Appendix E, Tables E-1 and E-2. These properties are used to assess the anticipated behavior of each chemical under environmental conditions. The physical properties of the chemicals defined as SRCs detected in soil and sediment are summarized in Sections 6.1.1 through 6.1.5.

37

38

6.1.1 Chemical Factors Affecting Fate and Transport

39

The water solubility of a chemical is a measure of the saturated concentration of the chemical in water at a given temperature and pressure. The tendency for a chemical to be transported by groundwater is directly related to its solubility and inversely related to its tendencies to adsorb to soil and volatilize from water (OGE 1988). Chemicals with high water solubilities tend to desorb from soil, are less likely to volatilize from water, and are susceptible to biodegradation. The water solubility of a 1 chemical varies with temperature, pH, and the presence of other dissolved chemicals (including

- 2 organic carbon and humic acids).
- 3

4 The octanol-water partition coefficient (Kow) can be used to estimate the tendency for a chemical to 5 partition between environmental phases of different polarity. The Kow is a laboratory-determined ratio 6 of the concentration of a chemical in the n-octanol phase of a two-phase system to the concentration 7 in the water phase. Chemicals with log Kow values less than one are highly hydrophilic, while

- 8 chemicals with log K_{ow} values greater than four will partition to soil particles (Lyman et al. 1990).
- 9

10 The water/organic carbon partition coefficient (K_{oc}) is a measure of the tendency of an organic 11 chemical to partition between water and organic carbon in soil. K_{og} is defined as the ratio of the 12 absorbed chemical per unit weight of organic carbon to the aqueous solute concentration. This 13 coefficient can be used to estimate the degree to which an organic chemical will adsorb to soil and thus not migrate with groundwater. The higher the Koc value, the greater is the tendency of the 14 15 chemical to partition into soil (OGE 1988). The soil/water partitioning coefficient (K_d) is calculated 16 by multiplying the K_{oc} value by the fraction of organic carbon in the soil.

17

18 Vapor pressure is a measure of the pressure at which a chemical and its vapor are in equilibrium. The 19 value can be used to determine the extent to which a chemical would travel in air, as well as the rate of 20 volatilization from soil and solution (OGE 1988). In general, chemicals with vapor pressures lower than 21 10^{-7} mm mercury will not be present in the atmosphere or air spaces in soil in significant amounts, while 22 chemicals with vapor pressures higher than 10⁻² mm mercury will exist primarily in the air 23 (Dragun 1988).

24

25 The HLC value for a chemical is a measure of the ratio of the chemical's vapor pressure to its 26 aqueous solubility. The HLC value can be used to make general predictions about a chemical's 27 tendency to volatilize from water. Chemicals with HLC values less than 10⁻⁷ atm-m³/mol will 28 generally volatilize slowly, while chemicals with a HLC greater than 10⁻³ atm-m³/mol will volatilize 29 rapidly (Lyman et al. 1990).

30

32

31 6.1.2 **Biodegradation**

33 Organic chemicals with differing chemical structures will biodegrade at different rates. Primary 34 biodegradation consists of any biologically induced structural change in an organic chemical. 35 Complete biodegradation is the biologically mediated degradation of an organic chemical into carbon 36 dioxide, water, oxygen, and other metabolic inorganic products (Dragun 1988). The first order 37 biodegradation rate of an organic chemical is proportional to the concentration:

38

Where:

$$-dC/dt = kC$$
 (Equation 6-1)

| 41 | C = concentration |
|----|---|
| 42 | t = time |
| 43 | $k = biodegradation \ rate \ constant = ln \ 2 \ / \ t_{1/2}$ |
| 44 | $t_{1/2}$ = biodegradation half-life |

1 The biodegradation half-life is the time necessary for half of the chemical to degrade. The 2 biodegradation rate of an organic chemical generally depends on the presence and population size of 3 soil microorganisms that are capable of degrading the chemical.

4

6.1.3 **Inorganic Chemicals**

5 6

7 Inorganic chemicals detected in soil and sediment samples are associated with the aqueous phase and 8 leachable metal ions on soil particles. The transport of this material from unsaturated soil to the 9 underlying water table is controlled by the physical processes of precipitation percolation, chemical 10 interaction with the soil, and downward transport of metal ions by continued percolation. The chemistry of inorganic chemical interactions with percolating precipitation and varying soil 11 12 conditions is complex and includes numerous chemical transformations that may result in altered 13 oxidation states, including ion exchange, adsorption, precipitation, or complexation. The chemical reactions, which are affected by environmental conditions (i.e., pH, oxidation-reduction conditions, 14 15 type and amount of organic matter, clay content, and the presence of hydrous oxides), may act to 16 enhance or reduce the mobility and toxicity of metal ions. In general, these reactions are reversible 17 and add to the variability commonly observed in distributions of inorganic chemicals in soil.

18

19 The chemical form of an inorganic chemical determines its solubility and mobility in the environment; 20 however, chemical speciation is complex and difficult to delineate in routine laboratory analysis. 21 Inorganic chemicals in soil are commonly found in several forms, including dissolved concentrations in 22 soil pore water, metal ions occupying exchange sites on inorganic soil constituents (adsorbed to 23 inorganic soil constituents), metal ions associated with insoluble organic matter, precipitated inorganic 24 chemicals as pure or mixed solids, and metal ions present in the structure of primary or secondary 25 minerals.

26

27 The dissolved (aqueous) fraction and its equilibrium sorbed fraction are important when considering 28 the migration potential of inorganic chemicals through soil. Of the inorganic chemicals that are likely 29 to form, chlorides, nitrates, and nitrites are commonly the most soluble. Sulfate, carbonate, and 30 hydroxides generally have low to moderate solubility. Soluble chemicals are transported in aqueous 31 forms subject to attenuation, whereas less soluble chemicals remain as a precipitate and limit the 32 overall dissolution of metal ions. The solubility of the metal ions is also regulated by ambient 33 chemical conditions, including pH and oxidation/reduction.

34

35 The attenuation of metal ions in the environment can be estimated numerically using the retardation 36 factor (R), dispersion in higher flow systems (high conductivity environments), and diffusion in low 37 conductivity environments. R defines the extent to which the velocity of the contaminant is slowed, 38 which is largely derived from the K_d. R is calculated using the following equation:

39 40

$$R = 1 + (K_{d} \rho_{b})/\theta_{w}$$
 (Equation 6-2)

41 Where:

42 ρ_b = the soil bulk dry density (g/cm³)

 $\theta_{\rm w}$ = soil moisture content (dimensionless) 43

1 Metal ion concentrations in the environment do not attenuate by natural or biological degradation 2 because of low volatility and solubility of the ions. Inorganic chemicals may be biotransformed or 3 bioconcentrated through microbial activity.

4

6.1.4 Organic Chemicals

5 6

7 Organic chemicals, such as SVOCs or VOCs, may be transformed or degraded in the environment by 8 processes including hydrolysis, oxidation/reduction, photolysis, volatilization, biodegradation, or 9 biotransformation. The half-life of organic chemicals in transport media can vary from minutes to 10 years, depending on environmental conditions and chemical structures. Some types of organic 11 chemicals are very stable, and degradation rates can be very slow. Organic degradation may either 12 enhance (by producing more toxic byproducts) or reduce (reducing concentrations) the toxicity of a 13 chemical in the environment.

14 15

16

6.1.5 Explosives-Related Chemicals

17 Nitrocellulose was detected in surface and subsurface soil at the Upper and Lower Cobbs Ponds 18 AOC, and several explosive chemicals were detected in sediment, including HMX, nitrocellulose; 19 tetryl; 1,3-DNB; TNT; 2,6-DNT; and 4-amino-2,6-DNT. Microbiological and photochemical 20 transformation may affect the fate and transport of explosive compounds in the environment. For 21 example, based on the results of culture studies involving the removal of TNT by activated sludge 22 microorganisms, it has been concluded that TNT undergoes biotransformation but not biodegradation 23 (USABRDL 1989). The biotransformation of TNT occurs with the reduction of the nitro groups by 24 microbial reduction, typically under anaerobic conditions. Beneficial bacteria in these reactions 25 include pseudomonas, escherichia, bacillus, citrobacter, enterobacter, klebseilla, veillonella, and 26 clostridium (USACHPPM 2000). It has been found that anaerobic metabolism occurs in two stages 27 (Funk et al. 1993). The first stage is the reductive stage in which TNT is reduced to its amino 28 derivatives. In the second stage, degradation to non-aromatic products begins after the reduction of 29 the third nitro group. The biotransformation rate of TNT can be rapid (ERDC 2007) and may be 30 increased with the presence of carbon (USACHPPM 2000). Fungi and photolysis can also 31 biotransform TNT. The predominant transformation products due to photolysis of TNT are 1,3,5-32 TNB: 4,6-dinitroanthranil; 2,4,6-trinitrobezadehyde; and 2,4,6-trinitrobenzonitrite (USACHPPM 33 2000). The biotransformation pathway for TNT is shown in Appendix E, Figure E-1 (Kaplan and 34 Kaplan 1982). The nitro groups of TNT are reduced to form 4-amino-2,6-DNT and can undergo 35 further reduction of the nitro groups to form 2,4,6-triaminotoluene (Cockerham and Shane 1994).

36

The biotransformation of 2,6-DNT has been systematically studied in laboratory cell cultures. The biotransformation pathway is shown in Appendix E, Figure E-2 (Singh et al. 2015). The reduction products include the amino and azoxy derivatives as observed with TNT biotransformation. As with TNT and DNT, the principal mode of microbial transformation of the nitroaromatic chemical 1,3-DNB is reduction of nitro groups to form amino groups (Walsh 1990). DNB is a photolytic product of TNT, but will undergo further photolysis. DNB will resist hydrolysis. The breakdown of DNB is primarily through microbial degradation (USACHPPM 2001a).

1 Limited information exists regarding biotransformation or biodegradation of HMX, nitrocellulose, 2 and tetryl. Biotransformation of HMX, primarily by anaerobic degradation (ERDC 2007), occurs at a 3 slow rate in the environment (USACHPPM 2001b). HMX is primarily broken down by photolysis 4 and has a photolytic rate constant of 0.15 days (USEPA 1988a). Breakdown products of HMX 5 include nitrate, nitrite, and formaldehyde (USACHPPM 2001b). Appendix E, Figure E-3 shows the biotransformation pathway for HMX (ATSDR 1997). Nitrocellulose is an aliphatic nitrate ester that 6 7 will gelatinize with nitroglycerin when the two are mixed. Nitrocellulose occurs as a fibrous solid that 8 can act as a sorbent that will dissolve in water under highly basic conditions with high temperatures. 9 Nitrocellulose can undergo denitrification as a degradation pathway. Degradation of nitrocellulose to 10 non-reactive nitrocellulose has been observed under methanogenic and fungus-mediated reducing conditions (USACE 2006). Tetryl may undergo biotransformation, hydrolysis, and photodegradation 11 12 to produce several different transformation products, such as N-methylpicramide and 4-amino-N-13 methyl-N,2,6-trinitroaniline (USACE 1994).

- 14
- 15 16

6.2 CONCEPTUAL MODEL FOR FATE AND TRANSPORT

17 The conceptual site model (CSM), which defines the framework for fate and transport modeling, 18 describes the conditions at the Upper and Lower Cobbs Ponds AOC, including the contaminant 19 sources, surficial and subsurface hydrogeologic conditions, contaminant migration and pathways, and 20 contaminant release mechanisms.

21

22 AOC conditions described in Sections 2.0 through 5.0 include contaminant source information, the 23 surrounding geologic and hydrologic conditions, and the magnitude of SRCs and their current spatial 24 distribution. Information from Section 3.0 and the nature and extent evaluation in Section 5.0 were 25 used to develop the CSM for fate and transport modeling by identifying SRCs and migration 26 pathways. The CSM is based on information and data collected for historical investigations, this RI 27 Report, and informed assumptions about the AOC. Assumptions contained in the CSM are reiterated 28 throughout this section. The better the information is and the more accurate the assumptions are, the 29 more accurately the CSM describes the AOC, and therefore, the more reliable the fate and transport 30 modeling predictions can be. A summary of the salient elements of the CSM that apply to fate and 31 transport modeling are summarized in the following sections.

32

33 6.2.1 Contaminant Sources

34

No primary contaminant sources are located on the AOC. Secondary sources (contaminated soil) were identified in previous investigations and are further evaluated in this report. Another potential secondary source of contamination at the AOC is contaminated sediment in the ponds, which is in contact with the surface water and has the potential to leach contaminants to the groundwater.

- 6.2.2 Hydrogeology
- A description of regional and AOC-specific geology and hydrology are provided in Sections 3.3.3
 and 3.4.2, respectively, and are summarized below.
- 5

1

2

6 The topography at the Upper and Lower Cobbs Ponds AOC ranges from approximately 960 7 ft amsl to approximately 980 ft amsl and is relatively flat with the exception of the bank areas 8 around the ponds. Surface water drainage associated with heavy rainfall events follows the 9 topography and drains into the ponds located in the center of the AOC (Figure 3-1). Upper Cobbs Pond is located in the center of the AOC and Lower Cobbs Pond is located in the 10 11 northern portion of the AOC (downstream of Upper Cobbs Pond). Surface water exits Lower 12 Cobbs Pond via a dam overflow spillway, which leaves the AOC and flows into an unnamed 13 tributary to Sand Creek (MKM 2005). 14 • Soil along the banks at the AOC consists of clay to sand-rich silt tills with interbedded sands 15 scattered throughout, as observed in subsurface borings installed during the PBA08 RI 16 (Appendix A). Soil across the AOC is low to moderately permeable and poorly drained. 17 Six groundwater monitoring wells were installed in the unconsolidated zone at the AOC to an • 18 average depth of 21.5 ft bgs. Potentiometric data indicate the groundwater flow direction is to 19 the northwest with an average hydraulic gradient of 0.011 ft/ft (Figure 3-1). 20 • The water table in the unconsolidated bank soil occurs at approximately 5 ft bgs (ranging 21 from 1 ft bgs to greater than 13 ft bgs) or around 965 ft amsl as observed in subsurface 22 borings installed during the PBA08 RI (Appendix A). 23 Contaminant leaching pathways from bank soil and sediment at the bottom of the ponds to 24 the water table are typically through preferential flow paths created from the interbedded 25 sands scattered throughout the smaller-grained unconsolidated soil. 26 27 6.2.3 **Contaminant Release Mechanisms and Migration Pathways** 28 29 Based on the information presented above, the following contaminant release mechanisms and 30 migration pathways have been identified at the AOC: 31 32 Contaminant leaching from bank soil to the water table (vertical migration) and lateral • 33 transport to a downgradient receptor (the ponds); 34 Contaminated sediment transported to potential downstream receptors; • 35 Contaminated surface water migrating to potential downstream receptors; • Contaminated sediment from the ponds as a secondary source of leaching to the water table 36 • (vertical migration) and lateral transport to downgradient receptors; and 37 38 Contaminated sediment at the bottom of surface water bodies, assuming equilibrium with 39 groundwater, and mixing with surface water based on a calculated, sample-specific dilution 40 attenuation factor (DAF). 41 42 The first of these pathways, which considers a primary groundwater transport pathway, is treated 43 explicitly in this fate and transport section. Sediment and surface water exposure pathways are

1 pathways listed above, which consider a secondary groundwater transport pathway and a primary

2 groundwater transport pathway, respectively, are evaluated using the sediment screening analysis

- 3 presented in Section 6.4.
- 4

5 One of the principal migration pathways at the AOC is percolation through the unsaturated soil to the 6 water table (i.e., vertical leaching of contaminants from soil into groundwater). However, because of 7 the very heterogeneous nature of the unconsolidated glacial material, groundwater flow patterns 8 within the unconsolidated soil are difficult to predict. Precipitation that does not leave the AOC as 9 surface runoff percolates into the subsurface. Some of the percolating water leaves this environment 10 via evapotranspiration after little or no vertical migration.

11

12 The remainder of the water percolates into the water table. As discussed in Section 6.2.4, the rate of 13 percolation is controlled by soil cover, ground slope, saturated conductivity of the soil, and 14 meteorological conditions. Figure 6-1 illustrates the contaminant migration conceptual model.

15

16 Once the contaminant leachate percolates through the soil and reaches the water table, it migrates 17 with the local groundwater and discharges into the ponds. Groundwater flow likely occurs along 18 preferential pathways (e.g., sand seams, channel deposits, or other stratigraphic discontinuities) 19 having higher permeabilities. For inorganic chemicals, lateral migration through groundwater will be 20 very limited due to their high retardation by the bedrock material (USACE 2003b).

21

Additional factors that affect the leaching rate include a chemical's solubility, sorption capacity (expressed by K_d), and the amount of percolation. Insoluble chemicals will precipitate out of the solution in the subsurface or remain in insoluble forms with little leaching.

25

Another factor that affects whether a chemical will reach the water table through percolation of precipitation is the chemical's rate of decay. Most organic compounds decay at characteristic rates proportional to the chemical's half-life. For a given percolation rate, those chemicals with long halflives have a greater potential for contaminating groundwater than those with shorter half-lives. For this analysis, the rate of decay/half-life was not considered.

31

Contaminant releases through gaseous emissions and airborne particulates are not significant at the Upper and Lower Cobbs Ponds AOC. The AOC is vegetated, located in a humid temperate climate, and soil moisture is typically high, which prevents dust borne contaminant migration. Therefore, there is likely little to no gaseous emission, and contaminant levels in the air pathway are minor to nonexistent.

37

39

38 6.2.4 Water Budget

The potential for contaminant transport begins with precipitation. Percolation is the driving mechanism for soil contaminants leaching to groundwater. The actual amount of rainwater available for flow and percolation to groundwater is highly variable and depends upon soil type and climatic conditions. A water balance calculation can be used as a tool to quantitatively account for all components of the hydrologic cycle. The quantified elements of the water balance are used for inputs

| 1 2 | to the soil leaching and groundwater transport models discussed later. The components of a simple steady-state water balance model include precipitation, evapotranspiration, surface runoff, and |
|--------|---|
| 3 | groundwater recharge or percolation. |
| 4 | |
| 5 | These terms are defined as follows: |
| 6 | |
| 7 | $\mathbf{P} = \mathbf{ET} + \mathbf{Sr} + \mathbf{q} \qquad (\text{Equation 6-3})$ |
| 8 | or |
| 9 | Rainwater available for flow = $Sr + q = P - ET$ (Equation 6-4) |
| 10 | Where: |
| 11 | $\mathbf{P} = \mathbf{precipitation}$ |
| 12 | Sr = surface runoff |
| 13 | ET = evapotranspiration |
| 14 | q = groundwater recharge or percolation |
| 15 | |

16 It is expected that loss of runoff also occurs in the form of evaporation. The remaining water, after 17 runoff and evaporation, is available for percolation, which includes loss to the atmosphere by 18 evapotranspiration. The water balance estimations were developed using the Hydrologic Evaluation 19 of Landfill Performance (HELP) model (USEPA 1994a). See Appendix E, Table E-3 for parameters 20 used in the HELP model to develop the water budget estimates used in the evaluation. Calculations 21 using precipitation and temperature data for a 100-year period were generated synthetically using 22 coefficients for Cleveland, Ohio (e.g., the nearest weather station to Camp Ravenna with HELP 23 model coefficients).

24

25 The annual average water balance estimates indicate an evapotranspiration of 28% (10.3 inches) of 26 total precipitation (37 inches). The remaining 72% (26.7 inches) of rainwater is available for surface 27 water runoff and percolation to groundwater. Of the 72% (26.7 inches) of water available for runoff 28 or percolation, groundwater recharge (percolation) accounts for 13% (3.6 inches), and surface runoff 29 (along downgradient topography to nearest surface water bodies) accounts for the remaining 87% 30 (23.1 inches).

31

- 32 6.3 SOIL SCREENING ANALYSIS
- 34 Soil screening analyses are screening evaluations performed to identify SRCs with the potential to 35 leach to groundwater as initial CMCOPCs. The five steps for the soil leachability analysis are 36 illustrated in Figure 6-2.
- 37
- 38 6.3.1 Soil Screening Analysis
- 39
- 40 The first step of the soil screening analysis is developing SRCs, as presented in Section 4.0. A summary
- 41 of SRCs identified for soil and sediment is presented in Section 6.1.
- 42
- 43 The second step of the soil screening process (Figure 6-2) involves comparing the maximum 44 concentrations of the SRCs with MCL-based generic soil screening levels (GSSLs). GSSLs were

1 developed for Superfund sites for contaminant migration to groundwater (USEPA 1996b, 2015). The 2 GSSL is defined as the concentration of a chemical in soil that represents a level of contamination 3 below which there is no concern for impacts to groundwater under CERCLA, provided conditions 4 associated with USEPA risk-based soil screening levels (SSLs) are met. Generally, if chemical 5 concentrations in soil fall below the GSSL, and there are no groundwater receptors of concern or anticipated exposures, then no further study or action is warranted for that chemical. If the GSSL for a 6 7 chemical is not available, the USEPA risk-based SSL for groundwater migration, dated June 2015 8 (USEPA 2015), will be obtained from the USEPA RSL website and used. If neither the GSSL nor the 9 USEPA risk-based SSL for a chemical are available, then no further evaluation of the chemical is 10 performed and it is eliminated from the list of initial CMCOPCs. However, some chemicals have been assigned surrogates by risk assessors if the chemical without an SSL is similar to another 11 12 chemical with an SSL. Surrogates used for this analysis include acenaphthene for acenaphthylene and 13 pyrene for benzo(ghi)perylene and phenanthrene.

14

15 The initial CMCOPC screen, as presented in Appendix E, Table E-4, eliminated 9 inorganic 16 chemicals and 15 organic chemicals from further consideration. Nine inorganic and four organic 17 SRCs exceeded their GSSLs and were carried forward to the next screening step.

18

19 The third step of the soil screening process (Figure 6-2) involves comparing the maximum chemical 20 concentrations with the site-specific soil screening levels (SSSLs). The SSSL is defined as the GSSL 21 (or the USEPA risk-based SSL for groundwater protection if a GSSL is not available) multiplied by 22 the AOC-specific DAF. Direct partitioning is used to derive GSSLs, assuming groundwater is in 23 contact with the chemicals in soil and the groundwater concentration is equal to the leachate 24 concentration. However, as leachate moves through soil, chemical concentrations are attenuated by 25 adsorption and degradation. When the leachate reaches the water table, dilution by groundwater 26 further reduces leachate concentrations. This concentration reduction can be expressed by a DAF. 27 DAFs can vary based on AOC-specific characteristics (e.g., hydrogeologic properties, contaminated 28 source area, and depth to contamination). As described in the Soil Screening Guidance: Technical 29 Background Document (USEPA 1996b), chemical dilution in groundwater is estimated at each AOC 30 from an AOC-specific DAF. The DAF, which is defined as the ratio of soil leachate concentration to 31 receptor point concentration, is minimally equal to one. Dilution in groundwater is derived from a 32 simple mixing zone equation (Equation 6-5) and relies upon estimating the mixing zone depth 33 (Equation 6-6).

34

35

$$DAF = 1 + \frac{(K \times i \times d)}{(q \times L)}$$

(Equation 6-5)

| 36 | Where: |
|----|--|
| 37 | DAF = dilution attenuation factor |
| 38 | K = aquifer hydraulic conductivity (m/yr) |
| 39 | i = horizontal hydraulic gradient (m/m) |
| 40 | q = percolation rate (m/yr) |
| 41 | L = source length parallel to groundwater flow (m) |
| 42 | d = mixing zone depth (m) (which is defined below) |
| 43 | |

$$d = \sqrt{0.0112 \times L^2} + d_a \times \left[1 - \exp\left(\frac{-L \times I}{K \times i \times d_a}\right) \right]$$
(Equation 6-6)

2 Where:

 $d_a = aquifer thickness (m)$

 $d \leq d_{a}$

4

3

6 As stated above, if the aquifer thickness is less than the calculated mixing zone depth, then the aquifer 7 thickness is used for "d" in the DAF calculation. The DAF calculation for the AOC is presented in 8 Appendix E, Table E-5. However, the DAF calculated for the soil screening at this AOC was 9 estimated to be 1.00. Therefore, initial CMCOPCs following the GSSL screening remained initial 10 CMCOPCs throughout the SSSL analysis step. The SRCs identified as initial CMCOPCs are 11 presented in Appendix E, Table E-6. It should be noted that the purpose of this screen is not to 12 identify the chemicals that may pose risk at downgradient locations, but to target those chemicals that 13 pose the greatest problem if they migrate from the AOC.

14

15 The fourth step of the soil screening process (Figure 6-2) involves eliminating initial CMCOPCs 16 identified in the SSSL evaluation that require more than 1,000 years to leach through the unsaturated 17 zone before reaching the water table from further consideration. A period of 1,000 years was 18 conservatively selected to evaluate eventual migration of the contaminant front to the water table 19 despite uncertainties in vadose zone hydraulic parameters and groundwater recharge over time. 20 Additionally, USACE suggests a screening value of 1,000 years be used due to the high uncertainty 21 associated with predicting conditions beyond that time frame (USACE 2003b). Therefore, the initial 22 CMCOPCs at the selected sources were screened against a travel time of greater than 1,000 years. 23 The travel time in this screen is the time required for an initial CMCOPC to migrate vertically from 24 the base of the soil interval detected above the background concentration to the water table. This 25 distance is known as the leaching zone, which is evaluated in Appendix E, Table E-7, which may 26 vary across the AOC based on the varying depths of soil sample concentrations above the facility-27 wide background concentrations and the elevation of the water table. The estimated travel time for 28 each initial CMCOPC to reach the water table is determined using the following equations:

29

30

$$T = \frac{Lz \times R}{V_p}$$
(Equation 6-7)

31 Where:

32 T = leachate travel time (year)

33 Lz =thickness of attenuation zone (ft)

34 R = retardation factor (dimensionless) (Equation 6-2)

35 $V_p = porewater velocity (ft/year)$

| 2 | |
|----------------------|--|
| 3 | $V_{\rm r} = \frac{q}{2} \tag{Equation 6-8}$ |
| 5 | $V_p = \frac{q}{\theta_v} $ (Equation 6-8) |
| 4 | Where: |
| 5 | q = percolation rate (ft/year) |
| 6 | $\theta_{\rm w}$ = fraction of total porosity that is filled by water |
| 7 | |
| 8 | If the travel time for a chemical from a source area exceeded 1,000 years, then the chemical was |
| 9 | eliminated from the list of initial CMCOPCs. Seven inorganic and seven organic SRCs were |
| 10 | eliminated from further consideration based on their travel times exceeding 1,000 years. Initial |
| 11 | CMCOPCs with travel times less than 1,000 years were retained for further evaluation (Appendix E, |
| 12 | Table E-7). The constituents selected for further evaluation with Seasonal Soil Compartment |
| 13 | (SESOIL) model are listed in Table 6-1. |
| 14 | |
| 15 | In the fifth step (Figure 6-2), the initial CMCOPCs (presented in Table 6-1) are further evaluated |
| 16 | using fate and transport models provided in Section 6.5. |
| 17 | |
| 18 | 6.3.2 Limitations and Assumptions of Soil Screening Analysis |
| 19 20 | |
| 20 | It is important to recognize that acceptable soil concentrations for individual chemicals are highly |
| 21 | AOC-specific. The GSSLs used in this screening are based on a number of default assumptions |
| 22 23 | chosen to be protective of human health for most AOC conditions (USEPA 1996b). These GSSLs are expected to be more conservative than SSSLs based on AOC conditions. The conservative |
| 23 24 | assumptions included in this analysis are: (1) no adsorption in the unsaturated zone or in the aquifer, |
| 2 4 25 | (2) no biological or chemical degradation in the soil or aquifer, and (3) contamination is uniformly |
| 23 26 | distributed throughout the source. However, the GSSL does not incorporate the existence of |
| 20 27 | contamination already present within the aquifer. |
| 28 | containing the order of the second within the aquiter. |
| 29 | 6.4 SEDIMENT SCREENING ANALYSIS |
| 30 | |
| 31 | Sediment SRCs were developed in Section 4.0 and are presented in Section 6.1. The purpose of this |
| 32 | screening analysis is to identify CMCOPCs based on contaminant migration from sediment to |
| 33 | groundwater. The steps for the sediment screening analysis are illustrated in Figure 6-3. |
| 34 | |
| 35 | Three different aggregates exist for the sediment screening analysis at the Upper and Lower Cobbs |
| 36 | Pond AOC: Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. Each aggregate was |
| 37 | screened separately to identify potential CMCOPCs. Sediment SRCs were screened by developing |
| 38 | leachate concentrations assuming equilibrium between sediment and groundwater. The predicted |
| 39 | leachate concentrations were diluted based on a sample-specific DAF calculated by dividing the |
| 40 | calculated leachate concentrations by the co-located surface water concentrations. The DAF was |
| 41 | calculated for each chemical that was detected in sediment and surface water at the same sample |
| 42 | location. The calculated DAF was then used to calculate the maximum groundwater concentration, |
| 43 | considering dilution for sediment SRCs and assuming that sediment is in direct contact with |
| | Upper and Lower Cobbs Ponds Remedial Investigation Report Page 6-12 |

1

and

1 groundwater. The lowest DAF calculated for the sample area was used for sediment SRCs that did not

2 have a sample-specific DAF. The DAFs calculated for each chemical are shown in Tables 6-2, 6-3,

- 3 and 6-4.
- 4

5 This sediment screening analysis assumed that the sediment concentration and the recharging groundwater concentration were in equilibrium and a sample-specific DAF for each chemical was 6 7 used for dilution in the aquifer. Based on this screening analysis, there were four initial sediment 8 CMCOPCs [silver; 2,6-DNT; benz(a)anthracene; and methylene chloride] in the Backwater Area 9 (Table 6-2 and Appendix E, Table E-8), two initial sediment CMCOPCs (hexavalent chromium and 10 cobalt) in Upper Cobbs Pond (Table 6-3 and Appendix E, Table E-9), and four initial sediment CMCOPCs [chromium, hexavalent chromium, benz(a)anthracene, and naphthalene] in Lower Cobbs 11 Pond (Table 6-4 and Appendix E, Table E-10). 12

13

These initial sediment CMCOPCs were further evaluated by comparing their maximum detected 14 15 surface water concentrations against their respective MCL/RSLs. This analysis assumes that 16 groundwater beneath sediment and surface water are in contact and in equilibrium. An initial 17 sediment CMCOPC was eliminated if the detected concentration in surface water was below its 18 MCL/RSL (Table 6-5) because there is no evidence that the initial sediment CMCOPC is contributing 19 to groundwater/surface water that is in equilibrium. Based on this evaluation, all the initial sediment 20 CMCOPCs from the Backwater Area were eliminated from further evaluation, and only hexavalent 21 chromium from Upper Cobbs Pond and Lower Cobbs Pond was identified as a final sediment 22 CMCOPC (Table 6-5) and was further evaluated in the WOE discussion to identify sediment 23 CMCOCs.

- 24
- 25
- 26

6.5 FATE AND TRANSPORT MODELING

27 Contaminant fate and transport modeling represents the fifth step in the fate and transport screening 28 and evaluation process (Figure 6-2). SESOIL modeling was performed for chemicals identified as 29 initial CMCOPCs from the soil screening analysis presented in Section 6.3 and summarized in 30 Table 6-1. SESOIL modeling was performed to predict chemical concentrations in the leachate 31 immediately beneath the selected source areas and just above the water table. If the predicted 32 maximum leachate concentration of an initial CMCOPC was higher than the facility-wide background 33 concentration and the lowest risk-based screening value (Resident Receptor Adult FWCUG, MCL, or 34 RSL), it was identified as a final CMCOPC and was further evaluated using the Analytical Transient 35 1-, 2-, 3- Dimensional (AT123D) model to predict future maximum concentrations in groundwater 36 beneath the source, as well as at downgradient receptor locations, if applicable. The downgradient 37 receptor location (if required) is the closest surface water body feature downgradient from the source 38 areas that is connected to the groundwater, and for this AOC the nearest surface water body feature is 39 the Cobbs Pond itself. If the AT123D predicted maximum concentration in groundwater beneath the 40 source was higher than its facility-wide background concentration and the lowest risk-based screening 41 value (i.e., Resident Receptor Adult FWCUG, MCL, or RSL), the chemical was retained as a 42 CMCOC for WOE discussion including a comparison to groundwater monitoring results for the AOC 43 (if available). 44

1 6.5.1 Modeling Approach

2

3 Contaminant transport includes the movement of water and dissolved material from the source areas to groundwater. This occurs as rainwater infiltrates the surface and percolates through the area of 4 5 contamination, its surrounding soil, and into the saturated zone. The downward movement of water, 6 driven by gravitational potential, capillary pressure, and other components of total fluid potential, 7 mobilizes the contaminants and carries them through the soil into the mixing zone with the water 8 table. Lateral transport within the unconsolidated zone is controlled by the groundwater gradient. 9 Vertical transport (evaluated with SESOIL) through the overburden to the water table and horizontal 10 transport (evaluated with AT123D) through the unconsolidated zone to downgradient receptor 11 locations are illustrated in Figure 6-1.

12

The output of the contaminant fate and transport modeling is presented as the expected maximum concentration of modeled contaminants at the selected receptor locations. For SESOIL, the receptor location is the groundwater table beneath the source area. For this analysis, four discrete sample locations were considered as sources of contamination based on the results of the soil screening analysis. A separate SESOIL analysis was performed for each initial CMCOPC listed in Table 6-1 and is presented in Figure 6-4.

19

20 The predicted maximum leachate concentration just above the water table, observed in the SESOIL 21 results, was compared against its applicable RVAAP facility-wide background concentration, as well 22 as the Resident Receptor Adult FWCUG, MCL, and RSL. If the predicted maximum leachate 23 concentration of an initial CMCOPC was higher than the facility-wide background concentration, and 24 the lowest risk-based screening value (Resident Receptor Adult FWCUG, MCL, or RSL), the initial 25 CMCOPC was identified as a final CMCOPC and was further evaluated using AT123D to predict 26 future maximum concentrations in groundwater beneath the source, as well as at downgradient 27 receptor locations, if applicable.

28

If a predicted maximum leachate concentration was lower than the screening criteria, the chemical wasno longer considered a CMCOPC.

31

32 For chemicals identified as final CMCOPCs, maximum concentrations predicted by AT123D in 33 groundwater directly below the source areas and at the downgradient receptor locations were 34 compared to the applicable RVAAP facility-wide background concentrations, as well as the Resident 35 Receptor Adult FWCUG, MCL, and RSL. If the predicted maximum concentration of a final 36 CMCOPC was higher than its facility-wide background concentration, and the lowest risk-based 37 screening value (i.e., Resident Receptor Adult FWCUG, MCL, or RSL), the chemical was retained as a 38 CMCOC for WOE evaluation. If the predicted maximum concentration of a final CMCOPC in 39 groundwater directly below the source areas and at the downgradient receptor location was lower than 40 the screening criteria, the chemical was eliminated as a CMCOC and was not considered in the WOE 41 evaluation.

42

43 CMCOCs identified by modeling results were evaluated with respect to WOE for retaining or 44 eliminating CMCOCs from further consideration as a basis for potential soil remedial actions. Lines 45 of evidence include validating modeling results using available AOC-specific groundwater

monitoring data. Modeled timelines for potential leaching and lateral transport were evaluated with 1 2 respect to estimated times for contaminant releases during RVAAP operations to determine if peak 3 leaching concentrations would likely have occurred in the past. Some CMCOCs present at or below 4 RVAAP soil background concentrations may have predicted leachate or groundwater concentrations exceeding risk-based criteria due to conservative model assumptions; therefore, these were also 5 identified and considered in the evaluation. Additionally, identified CMCOCs were compared to 6 7 COCs identified in the HHRA to determine if they had an associated risk related to direct exposure to 8 soil or if CMCOCs and COCs were co-located and may be addressed simultaneously under a 9 potential remedial action.

10

12

11 6.5.2 Model Applications

The SESOIL model (GSC 1998) used for leachate modeling, when applicable, estimates pollutant concentrations in the soil profile following introduction via direct application and/or interaction with transport media. AT123D (DOE 1992) is an analytical groundwater pollutant fate and transport model. It computes the spatial-temporal concentration distribution of waste in the aquifer system and predicts the transient spread of a contaminant plume through a groundwater aquifer. The application of both of these models is discussed in the following subsections.

19

20 6.5.2.1 SESOIL Modeling

21

22 SESOIL defines the soil column as compartments extending from the ground surface through the 23 unsaturated zone and to the upper level of the saturated soil zone or top of bedrock. Processes 24 simulated in SESOIL are categorized in three cycles: hydrologic, sedimentation, and pollutant. Each 25 cycle is a separate submodule in the SESOIL code. The hydrologic cycle includes rainfall, surface 26 runoff, percolation, soil-water content, evapotranspiration, and groundwater recharge. The sediment 27 washload cycle includes erosion and sediment transport. The pollutant cycle includes convective 28 transport, volatilization, adsorption/desorption, and degradation/decay. A chemical in SESOIL can 29 partition in up to four phases: liquid, adsorbed, air, and pure. Data requirements for SESOIL are not 30 extensive and utilize a minimum of AOC-specific soil and chemical parameters and monthly or 31 seasonal meteorological values as input.

32

33 SESOIL output includes pollutant concentrations at various soil depths and pollutant loss from the 34 unsaturated soil zone in terms of surface runoff, percolation to groundwater, volatilization, and 35 degradation. The mathematical representations in SESOIL generally consider the rate at which the 36 modeled processes occur, the interaction of different processes with each other, and the initial 37 conditions of the waste area and surrounding subsurface matrix material.

38

The input data for SESOIL can be grouped into four types: climatic, chemical, soil, and application. There are 61 separate parameters contained in these 4 data groups. Wherever possible, AOC-specific parameter values were used for modeling. However, certain parameters were not available for the source areas and were estimated based on pertinent scientific literature, geochemical investigations, and checks for consistency between model results and historical data. Conservative estimates were used when a range of values existed or parameter values were not available.

1 6.5.2.2 <u>Climate Data</u>

2

3 The climatic data file of SESOIL consists of an array of mean monthly temperature, mean monthly 4 cloud cover fraction, average monthly relative humidity, average monthly reflectivity of the earth's 5 surface (i.e., shortwave albedo), average daily evapotranspiration, monthly precipitation, mean 6 number of storm events per month, mean duration of rainfall, and mean length of rainy season. The 7 climatic data are presented in Appendix E, Table E-11. The data set was taken from the Youngstown 8 National Weather Service Office weather station at the Youngstown-Warren Regional Airport in 9 Vienna, Ohio, as it was determined to be most appropriate in corresponding to the latitude and 10 longitude at Camp Ravenna.

11

12 Climate data from the Youngstown weather station did not have all of the necessary climatic 13 parameters for the HELP model simulation. Accordingly, the water balance evaluation presented in 14 Section 6.2.4 was based on the nearest available weather station data with all necessary coefficients 15 stored within the HELP model (Cleveland, Ohio). Inputs for the SESOIL (Youngstown station) and 16 HELP models (Cleveland station) produced virtually the same recharge rate (9.40 cm/yr for 17 Cleveland and 9.42 cm/yr for Youngstown) for each location. Therefore, the use of the two different 18 weather station data sets did not impact modeling results.

19

20 6.5.2.3 Chemical Data

21

22 The pollutant fate cycle of SESOIL focuses on various chemical transport and transformation 23 processes that may occur in the soil zone. These processes include volatilization/diffusion, 24 adsorption/desorption, cation exchange, biodegradation and hydrolysis, and metal complexation. The 25 chemical-specific parameters used for SESOIL are presented in Appendix E, Table E-12. The distribution coefficients (K_ds) for inorganic chemicals and the K_{oc} values for organic chemicals were 26 27 obtained from the chemical-specific parameter table associated with the USEPA Risk Based Generic 28 Screening Tables (USEPA 2015). The K_ds for organic chemicals were estimated from organic, carbon-29 based K_{oc} using the relationship $K_d = (f_{oc})(K_{oc})$, where $f_{oc} = mass$ fraction of the organic carbon soil 30 content obtained from AOC-specific measurements. In general, biodegradation rates are not applicable 31 for inorganic CMCOPCs and biodegradation was not considered for the organic chemicals in this 32 evaluation. 33

34 6.5.2.4 Soil Data

36 The soil data file of SESOIL contains input parameters describing the physical characteristics of the 37 subsurface soil and is presented in Table 6-6. These parameters include soil bulk density, intrinsic 38 permeability, soil disconnectedness index, soil porosity, organic carbon content, and cation exchange 39 capacity. AOC-specific data were used from geotechnical samples collected at the AOC during the 40 PBA08 RI. There is, however, no measurement method for the soil disconnectedness index or a 41 measured value of the Freundlich exponent. The soil disconnectedness index is a parameter that 42 relates the soil permeability to the moisture content, and the Freundlich exponent relates to the 43 concentration of a solute on the surface of an adsorbent to the concentration of the solute in the liquid 44 with which it is in contact. Thus, SESOIL default values were used for these two parameters.

45

1 An average intrinsic permeability for the vadose zone, representing the unconsolidated zone above 2 the water table, was calibrated using the percolation rate of 9.42 cm/yr (3.6 inches/year) as the 3 calibration target. The model was calibrated against the percolation rate by varying the intrinsic 4 permeability and keeping all other AOC-specific geotechnical parameters fixed. The final hydrogeologic parameter values used in this modeling are shown in Table 6-6. The soil porosity was 5 set to the AOC-specific value. The intrinsic permeability, calibrated in SESOIL to the percolation rate 6 7 (determined from a water balance estimated in HELP), was found to match the AOC-specific 8 measurements from geotechnical samples.

9

10 The soil disconnectedness index replaces the moisture retention curves (or characteristic curves) used 11 by other unsaturated zone leaching models. SESOIL's User Guide defines this parameter to be the exponent relating the "wetting" and "drying" time-dependent permeability of soil to its saturated 12 permeability (Hetrick and Scott 1993). This "one variable" approach of using the soil 13 14 disconnectedness index in SESOIL simplifies the data estimation process and reduces computational 15 time. In addition, this parameter was calibrated for four different soil types ranging from sandy loam 16 to clay (Hetrick et al. 1986), and calibrated values fell within the default range specified in the 17 SESOIL's User Guide.

- 18
- 19 6.5.2.5 <u>Source Terms</u>
- 20

Analytical data from surface and subsurface soil collected at the AOC were used as source terms for SESOIL. All the samples collected at different depth intervals were compiled to provide a detailed loading option for SESOIL. The maximum soil concentrations for each initial CMCOPC, listed in Table 6-1, were used as source term concentrations for SESOIL.

25

27

26 6.5.2.6 Application Data

Four different layering schemes were developed for sample locations within the AOC due to varying thicknesses of the loading and leaching zones that are based on varying soil sample and groundwater depths throughout the AOC. A model with a 3-ft vadose zone was used to model arsenic, barium, hexavalent chromium, and thallium. A model with a 10-ft vadose zone was used to model nickel and selenium. A model with a 13-ft vadose zone was used to model beta-BHC. Details of the model layers utilized in this modeling are presented in Appendix E, Table E-13.

34

35 Each model was arranged in four layers. The top layer (layer 1) of the 3-ft-thick model contained two 36 sublayers, each 0.5 ft thick, for a total thickness of 1 ft for layer 1. The top 0.5 ft of layer 1 constituted the 37 loading zone for arsenic and hexavalent chromium and corresponded to the maximum contamination 38 depth for arsenic and hexavalent chromium at their respective sample locations. All of Layer 1 (the top 1 39 ft) constituted the loading zone for barium and thallium. The second and third layers of the 3-ft-thick 40 model were 0.75 ft thick each with three 0.25-ft-thick sublayers for each model layer. The fourth layer 41 was 0.5 ft thick and did not contain any sublayers. Layer 4 was included just above the water table to 42 read output results at the water table/vadose zone interface. Layers 2 through 4 served as the leaching 43 zone. 44

1 The vadose zones varied for nickel, selenium, and beta-BHC based on the depth to water observed in 2 soil borings CPCsb-032 and CPCsb-035, respectively. Based on the depth to water in CPCsb-032, the 3 vadose zone was 10 ft thick for nickel and selenium. Nickel and selenium contamination were 4 detected throughout the entire soil interval in CPCsb-032. Therefore, nickel and selenium were loaded in all model layers in SESOIL with a 0.25-ft-thick fourth layer just above the water table to read 5 output results at the water table/vadose zone interface. CPCsb-035 extended to a depth of 13 ft bgs, 6 7 but groundwater was not encountered in the soil boring. Therefore, the depth to groundwater was set 8 to the minimum thickness available at 13 ft bgs. Beta-BHC was detected at a maximum depth of 1 ft 9 bgs in CPCsb-035, which resulted in a 12-ft-thick leaching zone for beta-BHC.

10 11

6.5.3 SESOIL Modeling Results

12

13 SESOIL modeling was performed for initial CMCOPCs (i.e., arsenic, barium, hexavalent chromium, 14 nickel, selenium, thallium, and beta-BHC) that have the potential to reach the water table within 15 1,000 years based on the soil screening analysis results (Table 6-1). Table 6-7 presents the predicted 16 peak leachate concentrations beneath the source areas relative to the discrete sample locations 17 corresponding to the time of peak leachate concentrations. The Resident Receptor Adult FWCUGs, RVAAP facility-wide background concentrations, and MCL/RSL values for the initial CMCOPCs, if 18 19 available, are also shown in this table for comparison purposes. The maximum concentrations of 20 arsenic, barium, nickel, selenium, thallium, and beta-BHC in the leachate below the source and just 21 above the water table exceeded their respective screening criteria. Therefore, these chemicals were 22 selected as the final CMCOPCs for AT123D modeling. Hexavalent chromium was eliminated based 23 on the SESOIL modeling results. Appendix E, Figures E-4 through E-10 show the leachate mass flux 24 versus time plots generated by SESOIL to be used as input to AT123D modeling.

25

Arsenic, barium, nickel, selenium, thallium, and beta-BHC were identified as the final soil CMCOPCs based on SESOIL results for each sample location within the AOC where the leachate concentration exceeded its screening criteria. This leachate concentration does not reflect the groundwater concentration beneath the source. When the leachate reaches the water table, dilution by groundwater further reduces leachate concentrations.

31 32

33

6.5.4 AT123D Modeling in the Saturated Zone

34 The fate and transport processes accounted for in the AT123D model include advection, dispersion, 35 adsorption/retardation, and decay. This model can be used as a tool for estimating the dissolved 36 concentration of a chemical in three dimensions in the groundwater resulting from a mass release over 37 a source area (point, line, area, or volume source). The model can handle instantaneous, as well as 38 continuous, source loadings of CMCOPC concentrations. AT123D is frequently used by the scientific 39 and technical community to perform quick and conservative estimates of groundwater plume 40 movements in space and time. SESOIL and AT123D are linked in a software package (RISKPRO) so 41 that mass loading to the groundwater predicted by SESOIL can be transferred directly to AT123D. 42 Therefore, AT123D was chosen to predict the maximum concentration of contaminants in 43 groundwater after mixing with the leachate and the future concentrations for the contaminants in 44 groundwater at the receptor locations.

1 The hydrogeologic parameter values used in this modeling are shown in Table 6-6. Most of the 2 parameters presented in this table are AOC-specific values, unless otherwise indicated. The chemical-3 specific parameters used for AT123D are presented in Appendix E, Table E-14. A discussion of 4 model assumptions and limitations is presented in Section 6.5.6.

5

Appendix E, Figures E-11 through E-16 show the predicted concentration versus time curves based
on AT123D modeling for final CMCOPCs for soil.

8

9 6.5.5 AT123D Modeling Results

10

11 Results of AT123D modeling for final soil CMCOPCs are shown in Table 6-8. The results show 12 predicted groundwater concentrations for final CMCOPCs beneath the source area and at the selected 13 downgradient receptor location (i.e., surface water at the Backwater Area and Upper Cobbs Pond). 14 Observed groundwater concentrations from AOC monitoring wells are included in Table 6-8; 15 however, it should be noted that these wells may not exist at the sample location with the maximum 16 concentration and should not be considered in direct correlation. The observed groundwater 17 concentrations were added for comparison, not for screening criteria. The distances to the 18 downgradient receptors were based on the distance along the groundwater flow direction to the edge 19 of the pond/surface water.

20

The maximum predicted concentrations of arsenic, nickel, selenium, and thallium were predicted to exceed the screening criteria in groundwater beneath the source area and were modeled to the downgradient receptors (i.e., surface water at the Backwater Area and Upper Cobbs Pond). The maximum predicted concentrations of barium and beta-BHC in the groundwater were not predicted to exceed the screening criteria beneath the source area and were eliminated as CMCOCs.

26

Lateral transport modeling showed the maximum predicted concentrations of all remaining final soil CMCOPCs (i.e., nickel, selenium, and thallium) except arsenic did not exceed the screening criteria at their downgradient receptor location (i.e., surface water at the Backwater Area and Upper Cobbs Pond). However, because arsenic, nickel, selenium, and thallium exceeded screening criteria in groundwater beneath their respective source areas, they were retained for further evaluation. Figure 6-5 presents CMCOCs identified for WOE evaluation based on AT123D modeling.

33

34 6.5.6 Limitations/Assumptions

35

36 In general, a conservative modeling approach was used, which may overestimate the contaminant 37 concentration in the leachate for migration from observed soil concentrations. Important assumptions 38 used in this analysis include:

- 39
- The contaminant fate and transport evaluation included not only chemicals identified as being previously used during historical operations but also chemicals identified during the RVAAP
 SRC screening process.
- Some SRCs were identified due to the lack of background concentration data available or having limited or slight exceedances of established background concentrations.

| 1 | • Chemical and biological degradation rates for organic CMCOPCs were not considered in the |
|----|---|
| 2 | SESOIL and AT123D models. |
| 3 | • The use of K_d and R to describe the reaction term of the transport equation assumes that an |
| 4 | equilibrium relationship exists between the solid-phase and solution-phase concentrations and |
| 5 | that the relationship is linear and reversible. |
| 6 | • Since AOC-specific data are not available, the K_d and K_{oc} values used in this analysis for all |
| 7 | CMCOPCs represent literature or calculated values and may not represent conditions at the |
| 8 | AOC. |
| 9 | • The K _d for inorganic chemicals used here assumed a pH of 6.8 [i.e., the middle value in the |
| 10 | USEPA's evaluation presented in the soil screening guidance document (USEPA 1996b)]. |
| 11 | The K _d for inorganic chemicals varies with pH (generally decreasing with decreasing pH, |
| 12 | although there are few exceptions); therefore, if AOC-specific pH measurements are greater |
| 13 | or less than 6.8, the K _d and calculated screening parameters (such as R) will deviate from |
| 14 | those presented here. |
| 15 | • Flow and transport in the vadose zone is one-dimensional (i.e., only in the vertical direction). |
| 16 | • This modeling used the current soil concentrations that were collected approximately |
| 17 | 65 years after historical operations were terminated at the AOC. Therefore, it does not |
| 18 | account for constituents that have already leached to groundwater. |
| 19 | • Flow and transport are not affected by density variations. |
| 20 | • A realistic distribution of soil contamination was not considered. The maximum |
| 21 | concentration value was used as the source term concentrations for SESOIL model layers; |
| 22 | this is a highly conservative assumption that is expected to produce higher leachate |
| 23 | concentrations for CMCOPCs than the average condition. The horizontal distribution of soil |
| 24 | contamination was assumed based on concentration levels from nearby sample locations as |
| 25 | opposed to taking into account the entire area. |
| 26 | • The water balance represents an overall average rainwater recharge and assumes an even |
| 27 | distribution of infiltration in the modeled area. An average water balance assumes some areas |
| 28 | will have higher or lower recharge based on the heterogeneity of the soil and varying |
| 29 | topography. |
| 30 | • Soil samples were taken along the banks of the AOC which resulted in a maximum horizontal |
| 31 | transport distance for the AT123D model of 20 ft. Hydraulic parameters presented in |
| 32 | Appendix E, Table E-5 for the aquifer, such as the horizontal hydraulic gradient and aquifer |
| 33 | thickness, are assumed applicable within 20 ft of the ponds for evaluating lateral contaminant |
| 34 | migration with AT123D. |
| 35 | |
| 36 | The inherent uncertainties associated with using these assumptions must be recognized. K_d values are |
| 37 | highly sensitive to changes in the major chemistry of the solution phase. Therefore, it is important |
| 38 | that the values be measured or estimated under conditions that will closely represent those of the |
| 39 | contaminant plume. Deviations from actual AOC-specific parameter values from assumed literature |
| 40 | values may significantly affect contaminant fate predictions. It is also important to note that the |
| 41 | contaminant plume will change over time and will be affected by multiple solutes present at the AOC. |

- 41 contaminant plume will change over time and will be affected by multiple solutes plume
 42 The effects of heterogeneity and anisotropy are not addressed in these simulations.
- 43

1 The discrepancy between the contaminant concentrations measured in the field and the values 2 predicted by the model could be investigated by performing sensitivity analyses on the following model input parameters that have the most influence on the model predictions: (1) biodegradation rate 3 4 constants for organic chemicals; (2) saturated hydraulic conductivity; (3) soil porosity; (4) foc for organic chemicals; (5) K_d for inorganic chemicals; and (6) longitudinal, transverse, and vertical 5 6 dispersivity values. Generally, higher biodegradation rates will produce lower concentrations, and 7 lower rates will produce higher concentrations for organic chemicals without impacting the results of 8 the inorganic chemicals. Higher hydraulic conductivity and dispersivity causes higher advection and 9 dispersion, thereby producing lower peaks near the source area, but increasing the migration distance. 10 The reverse will be true with lower hydraulic conductivity and dispersivity values. Higher f_{oc} values have a similar effect on organic chemicals as higher K_d has on inorganic chemicals; they decrease the 11 12 mobility of the chemicals as well as produce lower concentrations in groundwater.

- 13
- 14 15

6.6 EVALUATION TO IDENTIFY CMCOCS

This evaluation of contaminant fate and transport uses a soil screening analysis to identify SRCs that have potential to leach to groundwater, performs SESOIL modeling to conservatively estimate final CMCOPC leachate concentrations before the SRCs enter the groundwater system beneath the sources with highest level of contamination, and uses AT123D modeling to present a conservative maximum concentration in groundwater of final CMCOPCs beneath the sources and at downgradient receptor locations.

22

This analysis also includes a sediment screening analysis that was performed for sediment samples at the AOC. Chemical-specific DAFs were calculated using co-located surface water and sediment concentrations for identified sediment SRCs. The estimated groundwater concentrations using these DAFs along with observed maximum surface water concentrations were used in the sediment screening analysis to identify the final sediment CMCOPCs for further evaluation.

28

The limitations and assumptions of the overall process are presented in Section 6.5.6. The text below provides a list of the remaining CMCOCs after AT123D modeling, final sediment CMCOPCs, and a qualitative assessment of the results and considerations of the limitations and assumptions.

32

6.6.1 Evaluation of Remaining Soil CMCOCs

33 34

35 Arsenic. Of the 30 surface and subsurface soil samples collected, only 1 sample (ULCPsd-010-0001-36 SD) exceeded the subsurface soil background concentration of 19.8 mg/kg at a concentration of 28.4 37 mg/kg. Arsenic was not considered a COC in the HHRA. Arsenic was detected in the AOC 38 groundwater samples collected from 2008–2011 at concentrations above its MCL (Table 6-8). Using 39 the maximum soil concentration, arsenic modeling results indicate it would take more than 150 years 40 for a breakthrough in groundwater beneath the source at a concentration above its MCL (0.01 mg/L), 41 and approximately 400 years to migrate to the downgradient receptor location at a concentration 42 above its MCL (Table 6-8). However, it should be noted that the background concentration of arsenic 43 in unconsolidated groundwater also exceeds the MCL. Therefore, if arsenic is detected in

1 groundwater above its MCL, it should be considered background related, and not due to 2 contamination from the site.

3

4 Nickel. Although the maximum concentration of nickel (31.7 mg/kg) was detected above its surface 5 soil background concentration (21.1 mg/kg), it is well below the subsurface soil background concentration of 60.7 mg/kg. Nickel was not considered a COPC in the HHRA. Although nickel was 6 7 detected in the AOC groundwater samples collected from 2008-2011, the maximum detection 8 (0.0038 mg/L) was below its tap water RSL of 0.39 mg/L (Table 6-8). Using the maximum detected 9 soil concentration, nickel modeling results indicate it would take more than 350 years for a 10 breakthrough in groundwater beneath the source at a concentration above its RSL, and nickel is not predicted to migrate to the downgradient receptor location at concentrations exceeding its RSL within 11 12 1,000 years (Table 6-8).

13

14 Selenium. Of the 30 surface and subsurface soil samples collected, only 1 sample (CPCsb-032-5114-15 SO) at a concentration of 1.6 mg/kg exceeded the surface or subsurface soil background concentrations (1.4 and 1.5 mg/kg, respectively). Selenium was not considered a COPC in the HHRA. 16 17 This MDC in soil is well below the RSL of 39 mg/kg. Although selenium was detected in the AOC groundwater samples collected from 2008–2011, the maximum detection (0.0046 mg/L) was below 18 19 its MCL of 0.05 mg/L (Table 6-8). Using the maximum soil concentration, selenium modeling results 20 indicate it would take about 15 years for a breakthrough in groundwater beneath the source at a 21 concentration above its MCL (0.05 mg/L), and selenium is not predicted to migrate to the 22 downgradient receptor location at concentrations exceeding its MCL within 1,000 years (Table 6-8). 23 Based on the AOC period of operations, selenium should have already been detected in groundwater 24 with concentrations exceeding its MCL. Therefore, this evaluation concludes that the model-predicted 25 concentrations are very conservative and selenium would be expected to continue to be below its 26 MCL based on its more realistic attenuation rate.

27

28 Thallium. None of the 30 surface and subsurface soil samples collected for thallium exceeded the subsurface soil background concentration of 0.91 mg/kg. Thallium was not considered a COPC in the 29 30 HHRA. Thallium was not detected in the AOC groundwater samples collected from 2008-2011 31 (Table 6-8). Using the maximum soil concentration, thallium modeling results indicate it would take 32 more than 600 years for a breakthrough in groundwater beneath the source at a concentration above 33 its MCL (0.002 mg/L), and thallium is not predicted to migrate to the downgradient receptor location 34 at concentrations exceeding its MCL within 1,000 years (Table 6-8). Because the maximum soil 35 concentration of thallium does not exceed its subsurface soil background, this evaluation concludes 36 that the model-predicted groundwater concentrations are very conservative and thallium would be 37 expected to be below its MCL based on its more realistic attenuation rate.

38 39

40

6.6.2 Evaluation of Remaining Sediment CMCOPCs

41 Hexavalent Chromium. The maximum sediment concentrations for hexavalent chromium (10.6 42 mg/kg in Upper Cobbs Pond and 5.7 mg/kg in Lower Cobbs Pond) were below the Resident Receptor 43 Adult FWCUG at a TR of 1E-06, HQ of 0.1, and hexavalent chromium was not identified as a 44 sediment COC for the Resident Receptor in the HHRA. Hexavalent chromium was not analyzed in 1 the AOC groundwater samples collected from 2008–2011; however, total chromium was detected in

- 2 groundwater at a maximum detection of 0.009 mg/L which was well below its MCL of 0.1 mg/L.
- 3

4 It should be noted here that the reason hexavalent chromium in groundwater is considered to be of 5 concern is because the evaluation assumes the DAF calculated using chromium data in surface water 6 and sediment can be applicable to hexavalent chromium. This assumption was made because 7 hexavalent chromium was not analyzed in surface water.

8

9 This qualitative assessment concludes that the soil and sediment contaminants identified for WOE 10 evaluation, due to predicted groundwater concentrations beneath the source and at the downgradient 11 receptor location, are not adversely impacting groundwater quality based on current data and are not 12 predicted to have future impacts. Potential additional investigation under the Facility-wide 13 Groundwater AOC may be warranted, but based on the fate and transport evaluation, CMCOCs were 14 not identified for the Backwater Area and Upper and Lower Cobbs Ponds, and no further action is 15 required of soil and sediment to be protective of groundwater.

16

17 6.7 SUMMARY AND CONCLUSIONS

18

Inorganic and organic SRCs exist in surface soil, subsurface soil, and sediment at the Upper and Lower Cobbs Ponds AOC. These SRCs include chemicals that were identified as potential contaminants from previous site usage and chemicals that were identified from the SRC screening process using available data. All SRCs were further evaluated to determine if residual concentrations in soil and sediment may potentially impact groundwater quality and warrant evaluation in an FS.

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All SRCs identified in the surface soil, subsurface soil, and sediment at Upper and Lower Cobbs Ponds were evaluated through the stepwise fate and transport evaluation. Evaluation of modeling results identified the following CMCOCs for soil and final CMCOPC for sediment:

- Arsenic, nickel, selenium, and thallium in soil were predicted to exceed the screening criteria in groundwater beneath the source area; arsenic was predicted to exceed the screening criteria in groundwater at the downgradient receptor location.
- Hexavalent chromium in Upper Cobbs Pond and Lower Cobbs Pond sediment was predicted to exceed the screening criteria in groundwater beneath the source areas.
- 33 34

A qualitative assessment of the sample results and considerations of the limitations and assumptions of the models were performed to identify if any CMCOCs are present in soil or sediment at the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond that may potentially impact groundwater at the AOC beneath the source or at the downgradient receptor location. This qualitative assessment concluded that there were no CMCOCs present in soil and sediment that may impact the groundwater beneath the source or at the downgradient receptor location. No further action is required of soil and sediment to be protective of groundwater.

| SRC | Maximum Concentration (mg/kg) | Discrete Sample Location | Sample Depth (ft bgs) | Leachate Modeling Required ^a ? (Yes/No) | | | | | | | | | |
|---------------------|-------------------------------------|--------------------------------|-----------------------------|---|--|--|--|--|--|--|--|--|--|
| Inorganic Chemicals | | | | | | | | | | | | | |
| Arsenic | 28.4 | ULCPsd-010 | 0-0.5 | Yes | | | | | | | | | |
| Barium | 90.4 | ULCPss-001 | 0-1 | Yes | | | | | | | | | |
| Hexavalent chromium | 1.00 | ULCPsd-010 | 0-0.5 | Yes | | | | | | | | | |
| Nickel | 31.7 | CPCsb-032 | 7–10 | Yes | | | | | | | | | |
| Selenium | 1.60 | CPCsb-032 | 0–4 | Yes | | | | | | | | | |
| Thallium | 0.250 | ULCPss-001 | 0-1 | Yes | | | | | | | | | |
| | Pesticio | le/PCB | • | | | | | | | | | | |
| beta-BHC | 0.0035 | CPCsb-035 | 0-1 | Yes | | | | | | | | | |

bgs = Below ground surface.

BHC = Hexachlorocy clohexane.

CMCOPC = Contaminant migration chemical of potential concern.

ft = Feet.

mg/kg = M illigrams per kilogram. PCB = Polychlorinated biphenyl.

SESOIL = Seasonal Soil Compartment Model.

SRC = Site-related contaminant.

| Analyte | CAS Number | Background Criteria (mg/kg)ª | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | Koc (L/kg) | Reference | K _d (L/kg) | Reference | Maximum Calculated Leachate Concentration (C _L) (mg/L) ^b | DAF | Maximum Calculated Groundwater Concentration (C1/DAF) (mg/L) | MCL or RSL (mg/L) | MCL or RSL? | Maximum Calculated Groundwater Concentration > MCL or RSL? | Initial Sediment CMCOPC? |
|---|------------|------------------------------------|---|--------------------|---|-----------|-----------------------|-----------|---|-------|--|----------------------|-------------------|--|--------------------------------|
| | | | г — г | | ``````````````````````````````````````` | ganic (| Chemicals | | | | 1 | 1 | | | |
| Aluminum | 7429-90-5 | 1.39E+04 | 1.86E+04 | ULCPsd-013-0001-SD | NA | - | 1.50E+03 | f | 1.24E+01 | 32 | 3.84E-01 | 2.00E+01 | RSL | No | No |
| Antimony | 7440-36-0 | 0.00E+00 | 2.10E+00 | CPCsd-047-5025-SD | NA | - | 4.50E+01 | f | 4.67E-02 | 49 | 9.60E-04 | 6.00E-03 | MCL | No | No |
| Barium | 7440-39-3 | 1.23E+02 | 1.51E+02 | CPCsd-047-5025-SD | NA | - | 4.10E+01 | f | 3.68E+00 | 171 | 2.16E-02 | 2.00E+00 | MCL | No | No |
| Beryllium | 7440-41-7 | 3.80E-01 | 1.10E+00 | ULCPsd-004-0001-SD | NA | - | 7.90E+02 | f | 1.39E-03 | 21 | 6.60E-05 | 4.00E-03 | MCL | No | No |
| Cadmium | 7440-43-9 | 0.00E=00 | 2.30E+00 | CPCsd-047-5025-SD | NA | - | 7.50E+01 | f | 3.07E-02 | 511 | 6.00E-05 | 5.00E-03 | MCL | No | No |
| Chromium | 7440-47-3 | 1.81E+01 | 5.65E+01 | ULCPsd-009-0001-SD | NA | - | 1.90E+01 | f | 2.97E+00 | 2,063 | 1.44E-03 | 1.00E-01 | MCL | No | No |
| Cobalt | 7440-48-4 | 9.10E+00 | 2.28E+01 | CPCsd-047-5025-SD | NA | - | 4.50E+01 | f | 5.07E-01 | 1,299 | 3.90E-04 | 6.00E-03 | RSL | No | No |
| Copper | 7440-50-8 | 2.76E+01 | 6.23E+01 | CPCsd-047-5025-SD | NA | - | 3.50E+01 | f | 1.78E+00 | 848 | 2.10E-03 | 1.30E+00 | MCL | No | No |
| Cyanide | 57-12-5 | 0.00E+00 | 5.50E-01 | ULCPsd-009-0001-SD | NA | - | 9.90E+00 | f | 5.56E-02 | 21 | 2.63E-03 | 2.00E-01 | MCL | No | No |
| Lead | 7439-92-1 | 2.74E+01 | 5.79E+01 | ULCPsd-009-0001-SD | NA | - | 9.00E+02 | f | 6.43E-02 | 96 | 6.67E-04 | 1.50E-02 | MCL | No | No |
| Mercury | 7439-97-6 | 6.00E-02 | 1.10E-01 | ULCPsd-005-0001-SD | NA | - | 5.20E+01 | f | 2.12E-03 | 21 | 1.00E-04 | 2.00E-03 | MCL | No | No |
| Nickel | 7440-02-0 | 1.77E+01 | 3.55E+01 | CPCsd-047-5025-SD | NA | - | 6.50E+01 | f | 5.46E-01 | 287 | 1.90E-03 | 3.90E-01 | RSL | No | No |
| Selenium | 7782-49-2 | 1.70E+00 | 2.70E+00 | CPCsd-047-5025-SD | NA | - | 5.00E+00 | f | 5.40E-01 | 2,348 | 2.30E-04 | 5.00E-02 | MCL | No | No |
| Silver | 7440-22-4 | 0.00E=00 | 2.30E+01 | ULCPsd-007-0001-SD | NA | - | 8.30E+00 | f | 2.77E+00 | 21 | 1.31E-01 | 9.40E-02 | RSL | Yes | Yes |
| Vanadium | 7440-62-2 | 2.61E+01 | 3.00E+01 | ULCPsd-013-0001-SD | NA | - | 1.00E+03 | f | 3.00E-02 | 51 | 5.86E-04 | 8.60E-02 | RSL | No | No |
| | | | | | | Anio | ns | | | | | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 0.00E+00 | 2.20E+00 | ULCPsd-016-0001-SD | None | - | None | - | NA | NA | NA | 1.00E+01 | MCL | No | No |
| | | | I I | | • | Explos | sives | <u> </u> | | | | | | | |
| 2,4,6-Trinitrotoluene | 118-96-7 | None | 3.20E-01 | ULCPsd-005-0001-SD | 2.81E+03 | f | 1.15E+01 | g | 2.78E-02 | 21 | 1.32E-03 | 2.50E-03 | RSL | No | No |
| 2,6-Dinitrotoluene | 606-20-2 | None | 1.60E-01 | ULCPsd-005-0001-SD | 5.87E+02 | f | 2.41E+00 | g | 6.64E-02 | 21 | 3.15E-03 | 4.80E-05 | RSL | Yes | Yes |
| Nitrocellulose | 9004-70-0 | None | 1.04E+01 | CPCsd-047-5025-SD | 1.00E+01 | f | 4.10E-02 | g | 2.54E+02 | 21 | 1.20E+01 | 6.00E+04 | RSL | No | No |
| Tetryl | 479-45-8 | None | 2.40E-02 | CPCsd-047-5025-SD | 4.61E+03 | f | 1.89E+01 | g | 1.27E-03 | 21 | 6.02E-05 | 3.90E-02 | RSL | No | No |
| | | | | | Semi-volati | ile Orga | inic Compound | s | | | | | | | |
| Anthracene | 120-12-7 | None | 2.20E-01 | ULCPsd-015-0002-SD | 1.64E+04 | f | 6.71E+01 | g | 3.28E-03 | 21 | 1.55E-04 | 1.80E+00 | RSL | No | No |
| Benz(a)anthracene | 56-55-3 | None | 8.30E-01 | ULCPsd-015-0002-SD | 1.77E+05 | f | 7.26E+02 | g | 1.14E-03 | 21 | 5.42E-05 | 1.20E-05 | RSL | Yes | Yes |
| Benzo(a)pyrene | 50-32-8 | None | 8.90E-01 | ULCPsd-015-0002-SD | 5.87E+05 | f | 2.41E+03 | g | 3.70E-04 | 21 | 1.75E-05 | 2.00E-04 | MCL | No | No |
| Benzo(b)fluoranthene | 205-99-2 | None | 8.60E-01 | ULCPsd-015-0002-SD | 5.99E+05 | f | 2.46E+03 | g | 3.50E-04 | 21 | 1.66E-05 | 3.40E-05 | RSL | No | No |
| Benzo(ghi)perylene ^d | 191-24-2 | None | 4.90E-01 | ULCPsd-015-0002-SD | 1.07E+07 | h | 4.39E+04 | g | 1.12E-05 | 21 | 5.29E-07 | 1.20E-01 | RSL | No | No |
| Benzo(k)fluoranthene | 207-08-9 | None | 8.00E-01 | ULCPsd-015-0002-SD | 5.87E+05 | f | 2.41E+03 | g | 3.32E-04 | 21 | 1.58E-05 | 3.40E-04 | RSL | No | No |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | None | 1.60E-01 | CPCsd-047-5025-SD | 1.20E+05 | f | 4.90E+02 | g | 3.26E-04 | 21 | 1.55E-05 | 6.00E-03 | MCL | No | No |
| Chrysene | 218-01-9 | None | 9.30E-01 | ULCPsd-015-0002-SD | 1.81E+05 | f | 7.42E+02 | g | 1.25E-03 | 21 | 5.94E-05 | 3.40E-03 | RSL | No | No |
| Di-n-butyl phthalate | 84-74-2 | None | 3.40E-02 | CPCsd-048-5786-SD | 1.16E+03 | f | 4.76E+00 | g | 7.15E-03 | 21 | 3.39E-04 | 9.00E-01 | RSL | No | No |
| Dibenz(a,h)anthracene | 53-70-3 | None | 6.60E-02 | CPCsd-047-5025-SD | 1.91E+06 | f | 7.84E+03 | g | 8.42E-06 | 21 | 3.99E-07 | 3.40E-06 | RSL | No | No |
| Fluoranthene | 206-44-0 | None | 1.80E+00 | ULCPsd-015-0002-SD | 5.55E+04 | f | 2.28E+02 | g | 7.91E-03 | 21 | 3.75E-04 | 8.00E-01 | RSL | No | No |
| Fluorene | 86-73-7 | None | 5.30E-02 | CPCsd-047-5025-SD | 9.16E+03 | f | 3.76E+01 | g | 1.41E-03 | 21 | 6.69E-05 | 2.90E-01 | RSL | No | No |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | None | 4.80E-01 | ULCPsd-015-0002-SD | 1.95E+06 | f | 8.00E+03 | g | 6.00E-05 | 21 | 2.85E-06 | 3.40E-05 | RSL | No | No |
| Phenanthrene ^e | 85-01-8 | None | 9.10E-01 | ULCPsd-015-0002-SD | 1.82E+04 | h | 7.46E+01 | g | 1.22E-02 | 21 | 5.78E-04 | 1.20E-01 | RSL | No | No |
| Pyrene | 129-00-0 | None | 1.90E+00 | ULCPsd-015-0002-SD | 5.43E+04 | f | 2.23E+02 | <u>σ</u> | 8.53E-03 | 21 | 4.04E-04 | 1.20E-01 | RSL | No | No |

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Table 6–2. Sediment Screening Results for the Backwater Area Aggregate (continued)

| Analyte | CAS Number | Background Criteria (mg/kg)ª | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | K _{oc} (L/kg) | Reference | K _d (L/kg) | Reference | Maximum Calculated Leachate Concentration (C _L) (mg/L) ^b | DAF ^c | Maximum Calculated Groundwater Concentration (C1/DAF) (mg/L) | MCL or RSL (mg/L) | MCL or RSL? | Maximum Calculated Groundwater Concentration > MCL or RSL? | Initial Sediment CMCOPC? |
|--------------------|------------------------|------------------------------------|---|--------------------|------------------------|-----------|-----------------------|-----------|---|------------------|--|----------------------|-------------------|--|--------------------------------|
| | Organics-Pesticide/PCB | | | | | | | | | | | | | | |
| PCB-1254 | 11097-69-1 | None | 4.70E-02 | ULCPsd-006-0001-SD | 1.31E+05 | f | 5.35E+02 | g | 8.78E-05 | 21 | 4.16E-06 | 7.80E-06 | RSL | No | No |
| | · | | | | Org | anics- | Volatile | | | | | | | | |
| 2-Butanone | 78-93-3 | None | 5.50E-02 | CPCsd-047-5025-SD | 4.51E+00 | f | 1.85E-02 | g | 2.97E+00 | 21 | 1.41E-01 | 5.60E+00 | RSL | No | No |
| Acetone | 67-64-1 | None | 1.90E-01 | CPCsd-047-5025-SD | 2.36E+00 | f | 9.69E-03 | g | 1.96E+01 | 21 | 9.29E-01 | 1.40E+01 | RSL | No | No |
| Methylene chloride | 75-09-2 | None | 4.80E-02 | ULCPsd-006-0001-SD | 2.17E+01 | f | 8.91E-02 | g | 5.39E-01 | 21 | 2.55E-02 | 5.00E-03 | MCL | Yes | Yes |

Sediment samples were taken from 0-1.8 ft below ground surface at discrete sample locations in the Backwater Area south of the railroad trackbed that is just north of Load Line 3 Road.

^aBackground criteria are the criteria for sediment from final facility-wide background values for the Ravenna Army Ammunition Plant, published in the Final Phase II Remedial Investigation Report for Winklepeck Burning Grounds at Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 2001b). ^bM aximum calculated leachate concentration = maximum sediment concentration divided by the distribution coefficient (K_d).

'An aggregate-specific DAF was calculated based on the sediment and co-located surface water concentrations. The lowest calculated DAF (21 for beryllium) was used for analytes that did not have an aggregate-specific DAF. ^dPyrene RSL was used as a surrogate for benzo(*ghi*)perylene.

^ePyrene RSL was used as a surrogate for phenanthrene.

^fU.S. Environmental Protection Agency (USEPA) RSL generic tables June 2015; found at: < <u>http://www2.epa.gov/risk/risk-based-screening-table-generic-tables></u>.

^gK_d value for organic chemicals calculated by multiplying K_{oc} by f_{oc} of 0.0041 (the average of the Performance-Based Acquisition 2008 RI geotechnical samples CPCSB-033-5117-SO and CPCSB-033-5118-SO). ^hUSEPA 1994b. Risk Reduction Engineering Laboratory Treatability Database, Version 5.0, Office of Research and Development, Cincinnati, Ohio.

CAS = Chemical Abstract Service.

CMCOPC = Contaminant migration chemical of potential concern.

DAF = Dilution attenuation factor.

ID = Identification.

 $K_d = Distribution coefficient.$

K_{oc} = Organic carbon distribution coefficient.

L/kg = Liters per kilogram.

MCL = Maximum contaminant level.

mg/kg = M illigrams per kilogram. mg/L = M illigrams per liter.

NA = Not applicable.

PCB = Polychlorinated biphenyl.

RSL = Regional screening level (USEPA 2015).

Bold = Identified initial sediment CMCOPC.

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|---|---------------------|---------------------------------|---|--------------------------------------|--|-----------------------|-----------|---|----------|---|----------------------|-------------------|--|--------------------------------|
| Analyte | CAS Number | Background Criteria (mg/kg)ª | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | Supervision of the second seco | K _d (L/kg) | Reference | Maximum Calculated Leachate Concentration (C _L) (mg/L) ^b | DAF° | Maximum Calculated Groundwater Concentration (C _L /DAF) (mg/L) | MCL or RSL (mg/L) | MCL or RSL? | Maximum Calculated Groundwater Concentration > MCL or RSL? | Initial Sediment CMCOPC? |
| | · | | · · | | Inorganic Ch | emicals | | | | · | | | · | |
| Aluminum | 7429-90-5 | 3.39E+04 | 2.14E+04 | ULCPsd-017-0001-SD | NA - | 1.50E+03 | f | 1.43E+01 | 37 | 3.86E-01 | 2.00E+01 | RSL | No | No |
| Antimony | 7440-36-0 | 0.00E+00 | 1.90E+00 | CPCsd-046-5024-SD | NA - | 4.50E+01 | f | 4.22E-02 | 45 | 9.40E-04 | 6.00E-03 | MCL | No | No |
| Barium | 7440-39-3 | 1.23E+02 | 1.57E+02 | CPCsd-046-5024-SD | NA - | 4.10E+01 | f | 3.83E+00 | 228 | 1.68E-02 | 2.00E+00 | MCL | No | No |
| Beryllium | 7440-41-7 | 3.80E-01 | 1.30E+00 | ULCPsd-017-0001-SD | NA - | 7.90E+02 | f | 1.65E-03 | 31 | 5.31E-05 | 4.00E-03 | MCL | No | No |
| Cadmium | 7440-43-9 | 0.00+00 | 2.30E+00 | CPCsd-046-5024-SD | NA - | 7.50E+01 | f | 3.07E-02 | 31 | 9.89E-04 | 5.00E-03 | MCL | No | No |
| Chromium | 7440-47-3 | 1.81E+01 | 9.16E+01 | CPCsd-046-5024-SD | NA - | 1.90E+01 | f | 4.82E+00 | 8,171 | 5.90E-04 | 1.00E-01 | MCL | No | No |
| Chromium, hexavalent | 18540-29-9 | 0.00+00 | 1.06E+01 | ULCPsd-018-0001-SD | NA - | 1.90E+01 | f | 5.58E-01 | 8,171 | 6.83E-05 | 3.50E-05 | RSL | Yes | Yes |
| Cobalt | 7440-48-4 | 9.10E+00 | 1.93E+01 | ULCPsd-017-0001-SD | NA - | 4.50E+01 | f | 4.29E-01 | 31 | 1.38E-02 | 6.00E-03 | RSL | Yes | Yes |
| Copper | 7440-50-8 | 2.76E+01 | 1.06E+02 | ULCPsd-017-0001-SD | NA - | 3.50E+01 | f | 3.03E+00 | 1,564 | 1.94E-03 | 1.30E+00 | MCL | No | No |
| Lead | 7439-92-1 | 2.74E+01 | 5.24E+01 | CPCsd-046-5024-SD | NA - | 9.00E+02 | f | 5.82E-02 | 149 | 3.90E-04 | 1.50E-02 | MCL | No | No |
| Mercury | 7439-97-6 | 6.00E-02 | 1.50E-01 | CPCsd-046-5024-SD | NA - | 5.20E+01 | f | 2.88E-03 | 31 | 9.30E-05 | 2.00E-03 | MCL | No | No |
| Nickel | 7440-02-0 | 1.77E+01 | 4.11E+01 | CPCsd-046-5024-SD | NA - | 6.50E+01 | f | 6.32E-01 | 333 | 1.90E-03 | 3.90E-01 | RSL | No | No |
| Selenium | 7782-49-2 | 1.70E+00 | 2.90E+00 | CPCsd-046-5024-SD | NA - | 5.00E+00 | f | 5.80E-01 | 2,417 | 2.40E-04 | 5.00E-02 | MCL | No | No |
| Silver | 7440-22-4 | 0.00E+00 | 1.10E+01 | ULCPsd-017-0001-SD | NA - | 8.30E+00 | f | 1.33E+00 | 31 | 4.28E-02 | 9.40E-02 | RSL | No | No |
| Vanadium | 7440-62-2 | 2.61E+01 | 3.34E+01 | ULCPsd-018-0001-SD | NA - | 1.00E+03 | f | 3.34E-02 | 31 | 1.08E-03 | 8.60E-02 | RSL | No | No |
| | | | | | Anions | 5 | | | | | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 0.00+00 | 5.30E+00 | FSW-SD-030-0000 | None - | None | - | NA | NA | NA | 1.00E+01 | MCL | No | No |
| | 1 | | | | Miscellan | eous | | | • | 1 | | | 1 | , , |
| Ammonia | 7664-41-7 | 0.00+00 | 6.90E+01 | FSW-SD-030-0000 | None - | None | - | NA | NA | NA | None | NA | No | No |
| Total Phosphorus as P | 7723-14-0 | 0.00+00 | 2.70E+02 | FSW-SD-030-0000 | None - | None | - | NA | NA | NA | None | NA | No | No |
| | 1 | | · · · · · · | | Explosi | | | | 1 | 1 | T | | 1 | |
| 1,3-Dinitrobenzene | 99-65-0 | None | 3.60E-02 | CPCsd-046-5024-SD | 3.52E+02 f | 1.44E+00 | g | 2.50E-02 | 31 | 8.06E-04 | 2.00E-03 | RSL | No | No |
| 2,4,6-Trinitrotoluene | 118-96-7 | None | 1.50E-01 | CPCsd-046-5024-SD | 2.81E+03 f | 1.15E+01 | g | 1.30E-02 | 31 | 4.20E-04 | 2.50E-03 | RSL | No | No |
| 4-Amino-2,6-Dinitrotoluene | 19406-51-0 | None | 1.20E-01 | CPCsd-046-5024-SD | 2.83E+02 f | 1.16E+00 | g | 1.03E-01 | 1,480 | 6.99E-05 | 3.90E-02 | RSL | No | No |
| HMX | 2691-41-0 | None | 8.30E-02 | CPCsd-046-5024-SD | 5.32E+02 f | 2.18E+00 | g | 3.81E-02 | 31 | 1.23E-03 | 1.00E+00 | RSL | No | No |
| Nitrocellulose | 9004-70-0 | None | 5.70E+00 | CPCsd-046-5024-SD | 1.00E+01 f | 4.10E-02 | g | 1.39E+02 | 31 | 4.48E+00 | 6.00E+04 | RSL | No | No |
| Tetryl | 479-45-8 | None | 1.90E-02 | CPCsd-046-5024-SD | 4.61E+03 f | 1.89E+01 | g | 1.01E-03 | 31 | 3.25E-05 | 3.90E-02 | RSL | No | No |
| | 200.04.0 | N T | 0.007.00 | | Semi-volatile Organ | | | | | 0.107.05 | | Dat | | |
| Acenaphthylene | 208-96-8 | None | 2.00E-02 | CPCsd-046-5784-SD | 7.40E+03 f | 3.03E+01 | g | 6.59E-04 | 31 | 2.13E-05 | 5.30E-01 | RSL | No | No |
| Anthracene | 120-12-7 | None | 1.00E-02 | CPCsd-046-5784-SD | 1.64E+04 f | 6.72E+01 | g | 1.49E-04 | 31 | 4.80E-06 | 1.80E+00 | RSL | No | No |
| Benz(a)anthracene | 56-55-3 | None | 7.80E-02 | CPCsd-046-5024-SD | 1.77E+05 f | 7.26E+02 | g | 1.07E-04 | 31 | 3.47E-06 | 1.20E-05 | RSL | No | No |
| Benzo(a)pyrene | 50-32-8 | None | 1.20E-01 | CPCsd-046-5024-SD | 5.87E+05 f | 2.41E+03 | g | 4.99E-05 | 31 | 1.61E-06 | 2.00E-04 | MCL | No | No |
| Benzo(b)fluoranthene | 205-99-2 | None | 1.90E-01 | CPCsd-046-5024-SD | 5.99E+05 f | 2.46E+03 | g | 7.74E-05 | 31 | 2.50E-06 | 3.40E-05 | RSL | No | No |
| Benzo(ghi)perylene ^d | 191-24-2 | None | 1.10E-01 | CPCsd-046-5024-SD | 1.07E+07 h | 4.39E+04 | g | 2.51E-06 | 31 | 8.09E-08 | 1.20E-01 | RSL | No | No |
| Benzo(k)fluoranthene | 207-08-9 | None | 6.70E-02 | CPCsd-046-5024-SD | 5.87E+05 f | 2.41E+03 | g | 2.78E-05 | 31 | 8.98E-07 | 3.40E-04 | RSL | No | No |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | None | 8.40E-01 | FSW-SD-030-0000 | 1.20E+05 f | 4.92E+02 | g | 1.71E-03 | 31 | 5.51E-05 | 6.00E-03 | RSL | No | No |
| Butyl benzyl phthalate Chrysene | 85-68-7 218-01-9 | None None | 1.60E-01 1.10E-01 | FSW-SD-030-0000 CPCsd-046-5024-SD | 7.16E+03 f 1.81E+05 f | 2.94E+01 7.42E+02 | g | 5.45E-03 1.48E-04 | 31 31 | 1.76E-04 4.78E-06 | 1.60E-02 3.40E-03 | RSL RSL | No No | No No |
| Di-n-butyl phthalate | 84-74-2 | None | 2.70E+00 | FSW-SD-030-0000 | 1.16E+03 f | 4.76E+02 | g g | 5.68E-01 | 31 | 1.83E-02 | 9.00E-01 | RSL | No | No |
| Fluoranthene | 206-44-0 | None | 2.20E-01 | CPCsd-046-5024-SD | 5.55E+04 f | 2.28E+02 | g | 9.67E-04 | 31 | 3.12E-05 | 8.00E-01 | RSL | No | No |

Table 6-3. Sediment Screening Results for Upper Cobbs Pond Aggregate (continued)

| Analyte | CAS Number | Background Criteria (mg/kg)ª | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | K _{oc} (L/kg) | Reference | K _d (L/kg) | Reference | Maximum Calculated Leachate Concentration (C _L) (mg/L) ^b | DAF ^c | Maximum Calculated Groundwater Concentration (C1/DAF) (mg/L) | MCL or RSL (mg/L) | MCL or RSL? | Maximum Calculated Groundwater Concentration > MCL or RSL? | Initial Sediment CMCOPC? |
|---------------------------|------------|---------------------------------|---|--------------------|------------------------|-----------|-----------------------|-----------|---|------------------|--|----------------------|-------------------|--|--------------------------------|
| Indeno(1,2,3-cd)pyrene | 193-39-5 | None | 9.00E-02 | CPCsd-046-5024-SD | 1.95E+06 | f | 8.00E+03 | g | 1.13E-05 | 31 | 3.63E-07 | 3.40E-05 | RSL | No | No |
| Phenanthrene ^e | 85-01-8 | None | 7.00E-02 | CPCsd-046-5024-SD | 1.82E+04 | h | 7.46E+01 | g | 9.38E-04 | 31 | 3.03E-05 | 1.20E-01 | RSL | No | No |
| Pyrene | 129-00-0 | None | 1.80E-01 | CPCsd-046-5024-SD | 5.43E+04 | f | 2.23E+02 | g | 8.09E-04 | 31 | 2.61E-05 | 1.20E-01 | RSL | No | No |
| | | | | | Pest | cide/P | PCB | | | | | | | · | |
| delta-BHC | 319-86-8 | None | 1.80E-03 | CPCsd-046-5784-SD | None | - | None | - | NA | NA | NA | None | NA | No | No |
| PCB-1254 | 11097-69-1 | None | 3.20E-02 | ULCPsd-020-0001-SD | 1.31E+05 | f | 5.35E+02 | g | 5.98E-05 | 31 | 1.93E-06 | 7.80E-06 | RSL | No | No |
| | | | | | Volatile Org | anic C | Compounds | | | • | • | • | • | | |
| 2-Butanone | 78-93-3 | None | 3.30E-02 | CPCsd-046-5024-SD | 4.51E+00 | f | 1.85E-02 | g | 1.78E+00 | 31 | 5.76E-02 | 5.60E+00 | RSL | No | No |
| Acetone | 67-64-1 | None | 9.10E-02 | CPCsd-046-5024-SD | 2.36E+00 | f | 9.69E-03 | g | 9.39E+00 | 31 | 3.03E-01 | 1.40E+01 | RSL | No | No |

Sediment samples were taken from 0–2 ft below ground surface at discrete sample locations and one incremental sampling method sample in the Upper Cobbs Pond Area.

^aBackground criteria for sediment from final facility-wide background values for the Ravenna Army Ammunition Plant, published in the Final Phase II Remedial Investigation Report for Winklepeck Burning Grounds at Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 2001b). ^bM aximum calculated leachate concentration = maximum sediment concentration divided by the distribution coefficient (K_d) .

"An aggregate-specific DAF was calculated based on the sediment and co-located surface water concentrations. The lowest calculated DAF (31 for vanadium) was used for analytes that did not have an aggregate-specific DAF. ^dPyrene RSL was used as a surrogate for benzo(*ghi*)perylene.

^ePyrene RSL was used as a surrogate for phenanthrene.

^fU.S. Environmental Protection Agency (USEPA) regional screening level generic tables June 2015; found at: < <u>http://www2.epa.gov/risk/risk-based-screening-table-generic-tables</u>>. ^gK_d value for organics calculated by multiplying K_{oc} by f_{oc} of 0.0041 (the average of the Performance-Based Acquisition 2008 RI geotechnical samples CPCSB-033-5117-SO and CPCSB-033-5118-SO). ^hUSEPA 1994b. Risk Reduction Engineering Laboratory Treatability Database, Version 5.0, Office of Research and Development, Cincinnati, Ohio

BHC = Hexachlorocy clohexane.

CAS = Chemical Abstract Service.

CMCOPC = Contaminant migration chemical of potential concern.

DAF = Dilution attenuation factor.

ID = Identification.

 $K_d = Distribution coefficient.$

 $K_{oc} = Organic carbon distribution coefficient.$

L/kg = Liters per kilogram.

MCL = Maximum contaminant level.

mg/kg = Milligrams per kilogram.

mg/L = M illigrams per liter.

NA = Not applicable.

PCB = Polychlorinated biphenyl.

RSL = Regional screening level (USEPA 2015).

Bold = Identified initial sediment CMCOPC.

| Analyte | CAS Number | Background Criteria (mg/kg)ª | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | K _{oc} (L/kg) | Reference | K _d (L/kg) | Reference | Maximum Calculated Leachate Concentration (CL) (mg/L) ^b | DAF ^c | Maximum Calculated Groundwater Concentration (C1/DAF) (mg/L) | MCL or RSL (mg/L) | MCL or RSL? | Maximum Calculated Groundwater Concentration > MCL or RSL? | Initial Sediment CMCOPC? |
|---|---------------|---------------------------------|---|--------------------|------------------------|-----------|-----------------------|-----------|--|------------------|--|----------------------|-------------------|---|--------------------------------|
| | 1 | | | * | Inorganic | Chemi | cals | | | | ,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,, | | | | |
| Aluminum | 7429-90-5 | 1.39E+04 | 1.64E+04 | ULCPsd-026-0001-SD | NA | - | 1.50E+03 | f | 1.09E+01 | 21 | 5.23E-01 | 3.65E+01 | RSL | No | No |
| Antimony | 7440-36-0 | 0.00+00 | 1.40E+00 | CPCsd-045-5023-SD | NA | - | 4.50E+01 | f | 3.11E-02 | 35 | 8.80E-04 | 6.00E-03 | MCL | No | No |
| Arsenic | 7440-38-2 | 1.95E+01 | 3.43E+01 | ULCPsd-026-0001-SD | NA | _ | 2.90E+01 | f | 1.18E+00 | 389 | 3.04E-03 | 1.00E-02 | MCL | No | No |
| Barium | 7440-39-3 | 1.23E+02 | 1.49E+02 | CPCsd-045-5023-SD | NA | - | 4.10E+01 | f | 3.63E+00 | 148 | 2.46E-02 | 2.00E+00 | MCL | No | No |
| Beryllium | 7440-41-7 | 3.80E-01 | 9.30E-01 | CPCsd-045-5023-SD | NA | - | 7.90E+02 | f | 1.18E-03 | 39 | 3.00E-05 | 4.00E-03 | MCL | No | No |
| Cadmium | 7440-43-9 | 0.00+00 | 1.50E+00 | CPCsd-045-5023-SD | NA | - | 7.50E+01 | f | 2.00E-02 | 500 | 4.00E-05 | 5.00E-03 | MCL | No | No |
| Chromium | 7440-47-3 | 1.81E+01 | 1.50E+02 | ULCPsd-023-0001-SD | NA | - | 1.90E+01 | f | 7.89E+00 | 16 | 4.78E-01 | 1.00E-01 | MCL | Yes | Yes |
| Chromium, hexavalent | 18540-29-9 | 0.00+00 | 5.70E+00 | ULCPsd-021-0001-SD | NA | - | 1.90E+01 | f | 3.00E-01 | 16 | 1.82E-02 | 3.50E-05 | RSL | Yes | Yes |
| Cobalt | 7440-48-4 | 9.10+00 | 2.13E+01 | ULCPsd-025-0001-SD | NA | - | 4.50E+01 | f | 4.73E-01 | 2,059 | 2.30E-04 | 6.00E-03 | RSL | No | No |
| Copper | 7440-50-8 | 2.76E+01 | 1.49E+02 | ULCPsd-023-0001-SD | NA | - | 3.50E+01 | f | 4.26E+00 | 229 | 1.86E-02 | 1.30E+00 | MCL | No | No |
| Cyanide | 57-12-5 | 0.00 | 4.00E-01 | ULCPsd-024-0001-SD | NA | - | 9.90E+00 | f | 4.04E-02 | 16 | 2.45E-03 | 2.00E-01 | MCL | No | No |
| Lead | 7439-92-1 | 27.40 | 6.04E+01 | ULCPsd-024-0001-SD | NA | - | 9.00E+02 | f | 6.71E-02 | 43 | 1.57E-03 | 1.50E-02 | MCL | No | No |
| Mercury | 7439-97-6 | 0.06 | 1.00E-01 | FSW-SD-031-0000 | NA | - | 5.20E+01 | f | 1.92E-03 | 16 | 1.17E-04 | 2.00E-03 | MCL | No | No |
| Nickel | 7440-02-0 | 17.70 | 3.02E+01 | CPCsd-045-5023-SD | NA | - | 6.50E+01 | f | 4.65E-01 | 147 | 3.16E-03 | 3.90E-01 | RSL | No | No |
| Selenium | 7782-49-2 | 1.70 | 2.20E+00 | CPCsd-045-5023-SD | NA | - | 5.00E+00 | f | 4.40E-01 | 1,375 | 3.20E-04 | 5.00E-02 | MCL | No | No |
| Silver | 7440-22-4 | 0.00 | 2.40E+00 | ULCPsd-021-0001-SD | NA | - | 8.30E+00 | f | 2.89E-01 | 16 | 1.75E-02 | 9.40E-02 | RSL | No | No |
| Vanadium | 7440-62-2 | 26.10 | 3.22E+01 | ULCPsd-026-0001-SD | NA | - | 1.00E+03 | f | 3.22E-02 | 24 | 1.33E-03 | 8.60E-02 | RSL | No | No |
| | • | | | | An | ions | | • | | | | | | | |
| Nitrate/Nitrite (NO ₃ /NO ₂ -N) | 14797-55-8 | 0.00 | 8.50E+00 | FSW-SD-031-0000 | None | - | None | - | NA | NA | NA | 1.00E+01 | MCL | No | No |
| · · · · · · · · · · · · · · · · · · · | • | | · | | Miscel | laneous | 5 | • | | | | | | | |
| Ammonia | 7664-41-7 | 0.00 | 6.10E+01 | FSW-SD-031-0000 | None | - | None | - | NA | NA | NA | None | NA | No | No |
| Total Phosphorus as P | 7723-14-0 | 0.00 | 3.60E+02 | FSW-SD-031-0000 | None | - | None | - | NA | NA | NA | None | NA | No | No |
| | | | | | Expl | losives | | | | | | | | | |
| HMX | 2691-41-0 | None | 1.70E-02 | CPCsd-044-5022-SD | 5.32E+02 | f | 2.18E+00 | g | 7.80E-03 | 16 | 4.73E-04 | 1.00E+00 | RSL | No | No |
| Nitrocellulose | 9004-70-0 | None | 7.80E+00 | CPCsd-045-5023-SD | 1.00E+01 | f | 4.10E-02 | g | 1.90E+02 | 16 | 1.15E+01 | 6.00E+04 | RSL | No | No |
| Tetryl | 479-45-8 | None | 2.20E-02 | CPCsd-045-5023-SD | 4.61E+03 | f | 1.89E+01 | g | 1.17E-03 | 16 | 7.06E-05 | 3.90E-02 | RSL | No | No |
| | | | | S | emi-volatile Or | ganic C | ompounds | | | | | | | | |
| 2-Methylnaphthalene | 91-57-6 | None | 2.50E-02 | CPCsd-045-5783-SD | 2.48E+03 | f | 1.02E+01 | g | 2.46E-03 | 16 | 1.49E-04 | 3.60E-02 | RSL | No | No |
| Acenaphthene | 83-32-9 | None | 9.90E-03 | CPCsd-045-5783-SD | 5.03E+03 | f | 2.06E+01 | g | 4.80E-04 | 16 | 2.91E-05 | 5.30E-01 | RSL | No | No |
| Acenaphthylene | 208-96-8 | None | 9.10E-02 | CPCsd-045-5783-SD | 7.40E+03 | f | 3.03E+01 | g | 3.00E-03 | 16 | 1.82E-04 | 5.30E-01 | RSL | No | No |
| Anthracene | 120-12-7 | None | 7.00E-02 | CPCsd-045-5783-SD | 1.64E+04 | f | 6.72E+01 | g | 1.04E-03 | 16 | 6.31E-05 | 1.80E+00 | RSL | No | No |
| Benz(a)anthracene | 56-55-3 | None | 3.40E-01 | CPCsd-045-5783-SD | 1.77E+05 | f | 7.26E+02 | g | 4.69E-04 | 16 | 2.84E-05 | 1.20E-05 | RSL | Yes | Yes |
| Benzo(a)pyrene | 50-32-8 | None | 4.70E-01 | CPCsd-045-5783-SD | 5.87E+05 | f | 2.41E+03 | g | 1.95E-04 | 16 | 1.18E-05 | 2.00E-04 | MCL | No | No |
| Benzo(b)fluoranthene | 205-99-2 | None | 8.00E-01 | CPCsd-045-5783-SD | 5.99E+05 | f | 2.46E+03 | g | 3.26E-04 | 16 | 1.97E-05 | 3.40E-05 | RSL | No | No |
| Benzo(ghi)perylene ^d | 191-24-2 | None | 4.70E-01 | CPCsd-045-5783-SD | 1.07E+07 | h | 4.39E+04 | g | 1.07E-05 | 16 | 6.49E-07 | 1.20E-01 | RSL | No | No |
| Benzo(k)fluoranthene | 207-08-9 | None | 2.10E-01 | CPCsd-045-5783-SD | 5.87E+05 | f | 2.41E+03 | g | 8.73E-05 | 16 | 5.29E-06 | 3.40E-04 | RSL | No | No |
| Chrysene | 218-01-9 | None | 4.00E-01 | CPCsd-045-5783-SD | 1.81E+05 | f | 7.42E+02 | g | 5.39E-04 | 16 | 3.27E-05 | 3.40E-03 | RSL | No | No |
| Di-n-butyl phthalate | 84-74-2 | None | 1.04E+00 | FSW-SD-031-0000 | 1.16E+03 | f | 4.74E+00 | g | 2.18E-01 | 16 | 1.32E-02 | 9.00E-01 | RSL | No | No |

Table 6-4. Sediment Screening Results for Lower Cobbs Pond Aggregate (continued)

| Analyte | CAS Number | Background Criteria (mg/kg)ª | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | K _{oc} (L/kg) | Reference | Kd (L/kg) | Reference | Maximum Calculated Leachate Concentration (C _L) (mg/L) ^b | DAF ^c | Maximum Calculated Groundwater Concentration (C1/DAF) (mg/L) | MCL or RSL (mg/L) | MCL or RSL? | Maximum Calculated Groundwater Concentration > MCL or RSL? | Initial Sediment CMCOPC? |
|---------------------------|------------|---------------------------------|---|--------------------|------------------------|-----------|-----------|-----------|---|------------------|--|----------------------|-------------------|---|--------------------------------|
| Dibenz(a,h)anthracene | 53-70-3 | None | 1.20E-01 | CPCsd-045-5783-SD | 1.91E+06 | f | 7.83E+03 | g | 1.53E-05 | 16 | 9.29E-07 | 3.40E-06 | RSL | No | No |
| Fluoranthene | 206-44-0 | None | 4.80E-01 | CPCsd-045-5783-SD | 5.55E+04 | f | 2.28E+02 | g | 2.11E-03 | 16 | 1.28E-04 | 8.00E-01 | RSL | No | No |
| Fluorene | 86-73-7 | None | 2.00E-02 | CPCsd-045-5783-SD | 9.16E+03 | f | 3.76E+01 | g | 5.33E-04 | 16 | 3.23E-05 | 2.90E-01 | RSL | No | No |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | None | 4.30E-01 | CPCsd-045-5783-SD | 1.95E+06 | f | 8.00E+03 | g | 5.38E-05 | 16 | 3.26E-06 | 3.40E-05 | RSL | No | No |
| Naphthalene | 91-20-3 | None | 3.80E-02 | CPCsd-045-5783-SD | 1.54E+03 | f | 6.31E+00 | g | 6.02E-03 | 16 | 3.65E-04 | 1.70E-04 | RSL | Yes | Yes |
| Phenanthrene ^e | 85-01-8 | None | 1.50E-01 | CPCsd-045-5783-SD | 1.82E+04 | h | 7.46E+01 | g | 2.01E-03 | 16 | 1.22E-04 | 1.20E-01 | RSL | No | No |
| Pyrene | 129-00-0 | None | 4.10E-01 | CPCsd-045-5783-SD | 5.43E+04 | f | 2.23E+02 | g | 1.84E-03 | 16 | 1.12E-04 | 1.20E-01 | RSL | No | No |
| | • • | | · · · · · · | | Volatile Orga | nic Com | pounds | | | | | | | | |
| 2-Butanone | 78-93-3 | None | 4.70E-02 | CPCsd-045-5023-SD | 4.51E+00 | f | 1.85E-02 | g | 2.54E+00 | 16 | 1.54E-01 | 5.60E+00 | RSL | No | No |
| Acetone | 67-64-1 | None | 1.70E-01 | CPCsd-045-5023-SD | 2.36E+00 | f | 9.69E-03 | g | 1.75E+01 | 16 | 1.06E+00 | 1.40E+01 | RSL | No | No |
| Carbon Disulfide | 75-15-0 | None | 3.30E-03 | CPCsd-045-5023-SD | 2.17E+01 | f | 8.91E-02 | g | 3.70E-02 | 16 | 2.24E-03 | 8.10E-01 | RSL | No | No |

samples were taken from 0–2 ft below ground surface at discrete sample locations and one incremental sampling method sample in Lower Cobbs Pond Area.

^aBackground criteria for sediment from final facility-wide background values for the Ravenna Army Ammunition Plant, published in the Final Phase II Remedial Investigation Report for Winklepeck Burning Grounds at Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 2001b). ^bMaximum calculated leachate concentration = maximum sediment concentration divided by the distribution coefficient (K_d).

*An aggregate-specific DAF was calculated based on the sediment and co-located surface water concentrations. The lowest calculated DAF (16 for thallium) was used for analytes that did not have an aggregate-specific DAF. ^dPyrene RSL was used as a surrogate for benzo(*ghi*)perylene.

^ePyrene RSL was used as a surrogate for phenanthrene.

^fU.S. Environmental Protection Agency (USEPA) regional screening level generic tables June 2015; found at: < <u>http://www2.epa.gov/risk/risk-based-screening-table-generic-tables</u>>.

^gK_d value for organic chemicals calculated by multiplying K_{oc} by f_{oc} of 0.0041 (the average of the Performance-Based Acquisition 2008 RI geotechnical samples CPCSB-033-5117-SO and CPCSB-033-5118-SO). ^hUSEPA 1994b. Risk Reduction Engineering Laboratory Treatability Database, Version 5.0, Office of Research and Development, Cincinnati, Ohio

CAS = Chemical Abstract Service.

CMCOPC = Contaminant migration chemical of potential concern.

DAF = Dilution Attenuation Factor.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

ID = Identification.

 $K_d = Distribution coefficient.$

 $K_{oc} = Organic carbon distribution coefficient.$

L/kg = Liters per kilogram.

- MCL = Maximum Contaminant Level.
- mg/kg = M illigrams per kilogram.
- mg/L = M illigrams per liter.
- NA = Not applicable.

RSL = Regional screening level (USEPA 2015).

Bold = Identified initial sediment CMCOPC.

| Initial Sediment CMCOPCs | CAS Number | Maximum Detected Surface Water Concentration (mg/L) | Surface Water Sample ID at Maximum Detect | MCL or RSL (mg/L) | MCL or RSL? | Maximum Detected Surface Water Concentration > MCL or RSL? | Final Sediment CMCOPC?ª |
|-----------------------------|---------------|---|---|----------------------|-------------------|--|-------------------------------|
| | |] | Backwater Area Aggregate | | | | |
| Silver | 7440-22-4 | 1.70E-03 | ULCPsw-002-0001-SW | 9.40E-02 | RSL | No | No |
| 2,6-Dinitrotoluene | 606-20-2 | ND | NA | 4.80E-05 | RSL | No | No |
| Benz(a)anthracene | 56-55-3 | ND | NA | 1.20E-05 | RSL | No | No |
| Methylene chloride | 75-09-2 | ND | NA | 5.00E-03 | MCL | No | No |
| | | Uj | oper Cobbs Pond Aggrega | te | | | |
| Chromium, hexavalent | 18540-29-9 | Not Analyzed | NA | 3.50E-05 | RSL | NA | Yes |
| Cobalt | 7440-48-4 | ND | NA | 6.00E-03 | RSL | No | No |
| | | Le | ower Cobbs Pond Aggrega | te | | | |
| Chromium | 7440-47-3 | 1.10E-03 | FSW-SW-031-0000 | 1.00E-01 | MCL | No | No |
| Chromium, hexavalent | 18540-29-9 | Not Analyzed | NA | 3.50E-05 | RSL | NA | Yes |
| Benz(a)anthracene | 56-55-3 | ND | NA | 1.20E-05 | RSL | No | No |
| Naphthalene | 91-20-3 | ND | NA | 1.70E-04 | RSL | No | No |

^aAnalyte is a final sediment CMCOPC if the maximum detected surface water concentration exceeds the MCL or RSL, or if there is no surface water data available. This analysis assumes that groundwater and surface water are in contact and in equilibrium.

CAS = Chemical Abstract Service.

CMCOPC = Contaminant migration chemical of potential concern.

ID = Identification.

MCL = M aximum contaminant level.

mg/L = M illigrams per liter.

NA = Not applicable.

ND = Not detected.

RSL = Regional screening level (USEPA 2015).

Bold = Identified final sediment CMCOPC.

Table 6-6. Unit-Specific Parameters Used in SESOIL and AT123D Modeling

| Parameters | Symbol | Units | Value | Source for Value |
|--|------------------|-----------------|-------------------|--|
| | | | SESOIL | |
| Percolation Rate (Recharge Rate) | q | m/yr | 9.42E-02 | 0.1 * SESOIL Precipitation for Youngstown, Ohio |
| Horizontal Area of Aggregate | Ap | cm ² | Sample specific | Sample specific |
| Intrinsic Permeability - clayey sand | р | cm ² | 1.05E-10 | Calibrated from SESOIL model |
| Disconnectedness Index | с | unitless | 11 | Calibrated from SESOIL model |
| Freundlich Equation Exponent | n | unitless | 1 | SESOIL default |
| Fraction Organic Carbon | foc | unitless | 4.10E-03 | |
| Bulk Density | ρь | kg/L | 1.71 | |
| Moisture Content | W | wt % | 20.3 | The average of the PBA08 RI Geotechnical Samples CPCSB-033-5117- |
| Water-filled Soil Porosity | Tw | unitless | 0.347 | SO and CPCSB-033-5118-SO |
| Air-filled Soil Porosity | Та | unitless | 0.039 | |
| Porosity – total | n _T | unitless | 0.386 | |
| Vadose Zone Thickness | Vz | m | 1.52 | Average depth to water table from potentiometric surface map (Figure 6-4) |
| Leaching Zone Thickness | Th | m | 0 to 3.66 | Based on distance from deepest detection of a CMCOPC above criteria to the water table from potentiometric surface maps (Figure 6-4) |
| | | | AT123D | |
| Aquifer Thickness | h | m | 6 | Facility-wide assumption for the unconsolidated aquifer presented the Load Line 1 investigation (USACE 2003b) |
| Hydraulic Conductivity in Saturated Zone | Ks | cm/s | 5.64E-06 | Average of slug test results from MKM (2005) |
| Hydraulic Gradient | i | unitless | 1.10E-02 | Calculated from Cobbs Pond potentiometric surface map (Figure 6-4) |
| Effective porosity | ne | unitless | 0.2 | Assumed for sandstone (USEPA 1985) |
| Dispersivity, longitudinal | $\alpha_{\rm L}$ | m | 30 | Assumed |
| Dispersivity, transverse | ατ | m | 3 | 0.1 α _L |
| Dispersivity, vertical | $\alpha_{\rm V}$ | m | 0.3 | 0.01 α _L |
| Retardation factor | Rd | unitless | chemical-specific | Presented in Table E-7 in Appendix E |

MKM (MKM Engineers) 2005. Phase II Remedial Investigation Report for Upper and Lower Cobbs Ponds at the Ravenna Army Ammunition Plant, Ravenna, Ohio. September 2005. USACE (U.S.Army Corps of Engineers) 2003b. Phase II Remedial Investigation Report for the Load Line 1 at the Ravenna Army Ammunition Plant, Ravenna, Ohio. June 2003. USEPA (U.S. Environmental Protection Agency) 1985. Water Quality Assessment: A Screening Procedure for Toxic and Conventional Pollutants in Surface and Ground Water, Revised

1985 Parts 1 and 2, USEPA/600/6-85/002. Office of Research and Development, Environmental Research Laboratory, Athens, Georgia. September 1985. AT123D = Analytical Transient 1-,2-,3-Dimensional modeling.

PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation.

SESOIL = Seasonal Soil Compartment model.

wt % = Weight percent.

| A I I 23D = Analytical I | I ransient 1-,2-,3-Dimensional modeling. | |
|---|--|--|
| $am^2 = \mathbf{S}$ and a a set in at a | | |

- $cm^2 = Square centimeters.$
- cm/s = Centimeters per second.
- kg/L = Killograms per liter.
- 2

Table 6-7. Summary of SESOIL Modeling Results

| | | | | | | | | Resident | | |
|------------|---------------|------------|----------------------|-------------|--------------------|--------------------|----------|--------------------|--------------------|----------------------|
| | | | Maximum | | | Time | | Receptor | Facility-wide | |
| | Maximum Soil | Discrete | Depth of | Depth to | Predicted CL, max | Required to | | Adult | Background | Final |
| Initial | Concentration | Sample | Contamination | Groundwater | Beneath the | Reach CL, max | MCL/RSL | FWCUG ^a | Uncons olidate d | CMCOPC? ^b |
| CMCOPC | (mg/kg) | Location | (ft bgs) | (ft bgs) | Source (mg/L) | (years) | (mg/L) | (mg/L) | Groundwater (mg/L) | (Yes/No) |
| | | | | 1 | Inorganic Chemical | 's | | | | |
| Arsenic | 28.4 | ULCPsd-010 | 0.5 | 3 | 5.80E-01 | 239 | 1.00E-02 | 5.60E-05 | 1.17E-02 | Yes |
| Barium | 90.4 | ULCPss-001 | 1 | 3 | 2.61E+00 | 299 | 2.00E+00 | 7.06E-01 | 8.21E-02 | Yes |
| Hexavalent | | | | | | | | | | |
| Chromium | 1 | ULCPsd-010 | 0.5 | 3 | 3.11E-02 | 157 | 1.10E-01 | None | 0.00E+00 | No |
| Nickel | 31.7 | CPCsb-032 | 10 | 10 | 7.72E-01 | 1000 | 3.90E-01 | 7.29E-02 | 0.00E+00 | Yes |
| Selenium | 1.6 | CPCsb-032 | 10 | 10 | 4.30E-01 | 102 | 5.00E-02 | None | 0.00E+00 | Yes |
| Thallium | 0.25 | ULCPss-001 | 1 | 3 | 4.18E-03 | 516 | 2.00E-03 | 2.91E-04 | 0.00E+00 | Yes |
| | | | | | Pesticide/PCB | | | | | |
| beta-BHC | 0.0035 | CPCsb-035 | 1 | 13 | 6.73E-05 | 433 | 2.50E-05 | 4.70E-05 | None | Yes |

^aThe FWCUG is based on a target risk of 10⁻⁶ and/or a hazard quotient of 0.1.

^bThe final CMCOPC was identified comparing predicted maximum leachate concentration to MCL/RSL, FWCUGs, and facility-wide background concentrations. A constituent is a final CMCOPC if its predicted leachate concentration exceeds its MCL/RSL within 1,000 years.

bgs = Below ground surface.

BHC = Hexachlorocy clohexane.

 $C_{L, max} = M$ aximum leachate concentration.

CMCOPC = Contaminant migration chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

ft = Feet.

MCL = Maximum contaminant level.

mg/kg = Milligram per kilogram.

mg/L = M illigram per liter.

PCB = Polychlorinated biphenyl.

RSL = Regional screening level.

SESOIL = Seasonal Soil Compartment model.

Bold = CMCOPCs exceeding MCL/RSL, FWCUGs, and facility-wide background concentrations.

| Final CMCOPC | Maximum Leachate Concentration, C _{L, max} ^a (mg/L) | Predicted Max Groundwater Concentration ^b (C _{gw} , _{MAX}) Beneath Source (mg/L) | Predicted Max Groundwater Concentration ^b (C _R , _{MAX}) Downgradient Receptor (mg/L) | Distance to Downgradient Receptor (ft) | Observed Maximum Groundwater Concentration ^c (mg/L) | MCL/RSL (mg/L) | Resident Adult FWCUG ^d (mg/L) | Facility-wide Background Unconsolidated Groundwater (mg/L) | CMCOC for Further WOE Evaluation? ^e (Yes/No) |
|-----------------|--|---|---|--|--|-------------------|--|--|---|
| | | | | Inorganic Che | emicals | | | | |
| Arsenic | 5.80E-01 | 3.14E-01 | 2.91E-02 | 2.00E+01 | 6.51E-02 | 1.00E-02 | 5.60E-05 | 1.17E-02 | Yes |
| Barium | 2.61E+00 | 1.33E+00 | 6.94E-01 | 1.00E+01 | 1.55E-01 | 2.00E+00 | 7.06E-01 | 8.21E-02 | No |
| Nickel | 7.72E-01 | 7.26E-01 | 1.37E-02 | 2.00E+01 | 3.80E-03 | 3.90E-01 | 7.29E-02 | 0.00E+00 | Yes |
| Selenium | 4.30E-01 | 4.15E-01 | 4.76E-02 | 2.00E+01 | 4.60E-03 | 5.00E-02 | None | 0.00E+00 | Yes |
| Thallium | 4.18E-03 | 2.15E-03 | 1.12E-03 | 1.00E+01 | ND | 2.00E-03 | 2.91E-04 | 0.00E+00 | Yes |
| | Pesticide/PCB | | | | | | | | |
| beta-BHC | 6.73E-05 | 9.26E-07 | 8.79E-07 | 1.00E+01 | 1.60E-05 | 2.50E-05 | 4.70E-05 | None | No |

^aRepresents Seasonal Soil Compartment (SESOIL) modeling predicted maximum leachate concentration just above the water table.

^bThe predicted concentration was estimated using the results from SESOIL and applying AT123D model.

^cObserved groundwater concentrations were reported in the Facility-wide Groundwater Monitoring Program Annual Report for 2009 (EQM 2010b) and sampling in January 2011.

^dThe FWCUG is based on a target risk of 10⁻⁶ and/or a hazard quotient of 0.1.

^eThe CMCOC was identified comparing predicted concentration in groundwater beneath the source to MCL/RSL, Resident Receptor Adult FWCUG, and facility-wide background concentration. A constituent is a CMCOC if its predicted concentration in groundwater exceeds all its screening criteria within 1,000 years.

AT123D = Analytical Transient 1-,2-,3-Dimensional model.

BHC = Hexachlorocyclohexane.

 $C_{gw} = Groundwater \ concentration.$

 C_L = Leachate concentration.

 C_R = Groundwater concentration at the downgradient receptor location.

CMCOC = Contaminant migration chemical of concern.

CMCOPC = Contaminant migration chemical of potential concern.

ft = Feet.

FWCUG = Facility-wide clean-up goal.

MCL = M aximum contaminant level.

mg/L = M illigrams per liter.

ND = Not detected.

PCB = Polychlorinated biphenyl.

RSL = Regional screening level.

WOE = Weight of evidence.

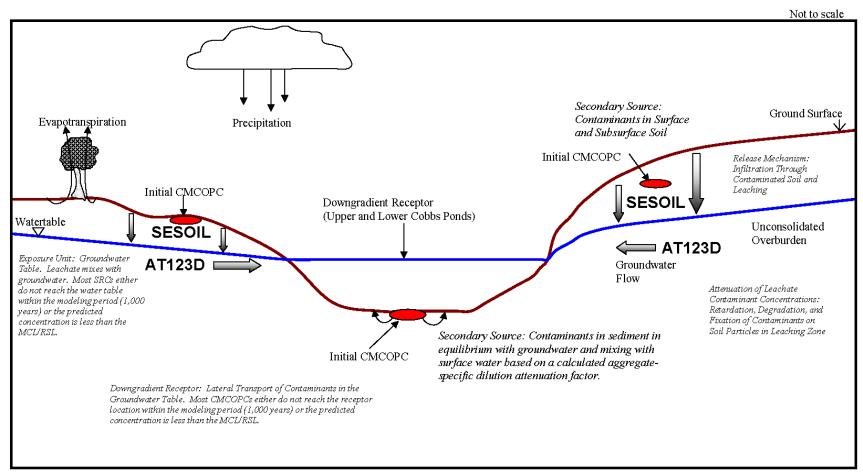


Figure 6-1. Contaminant Migration Conceptual Model

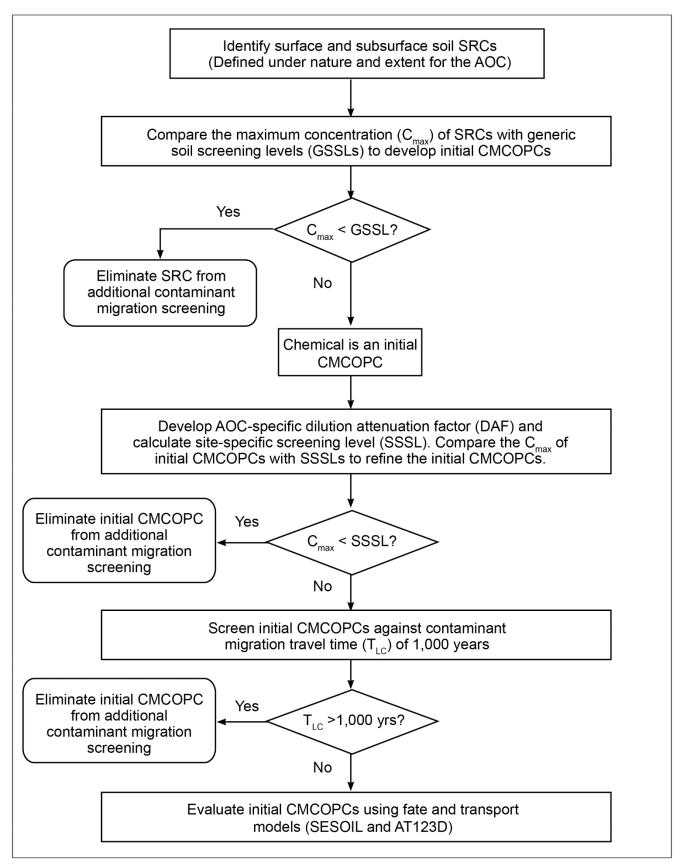


Figure 6-2. AOC Fate and Transport Modeling Approach - Soil

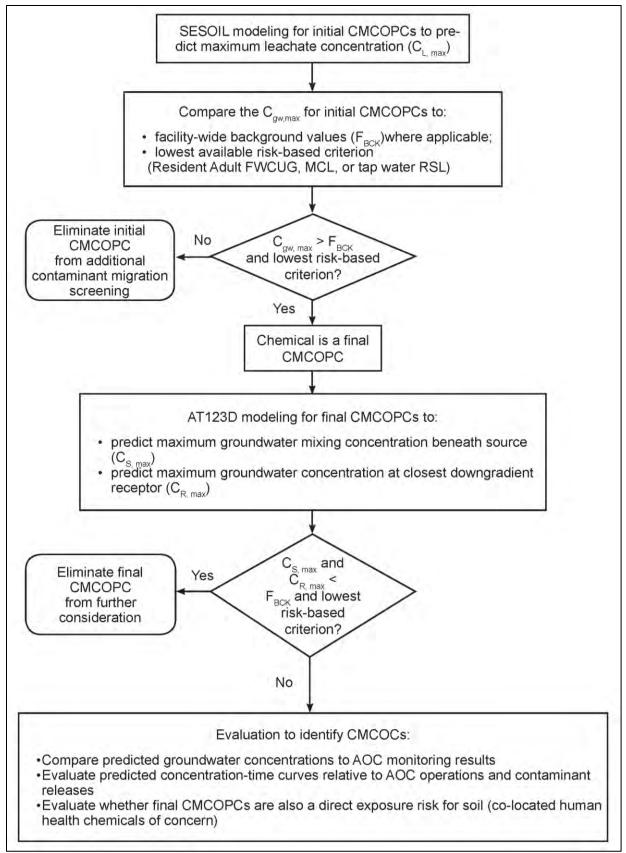


Figure 6-2. AOC Fate and Transport Modeling Approach - Soil (continued)

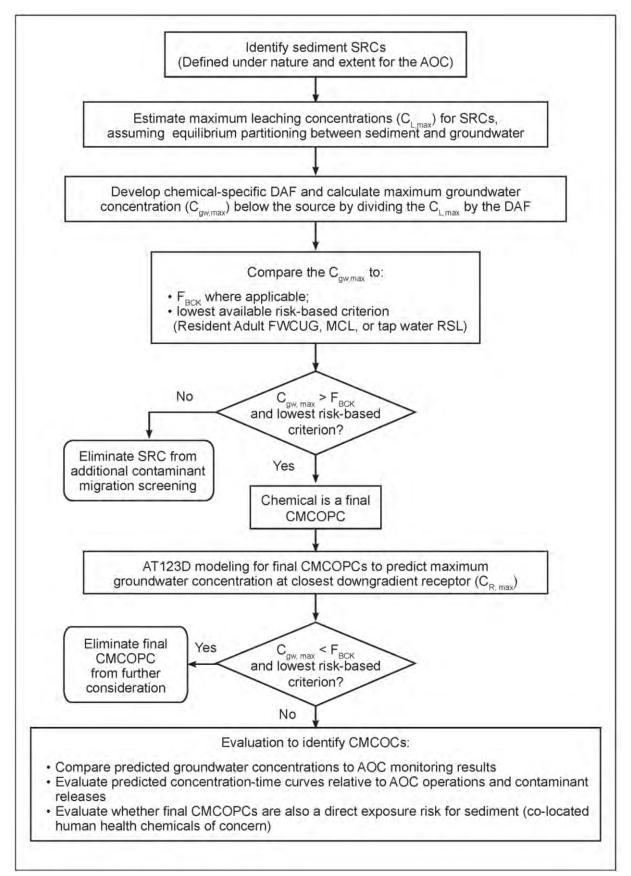


Figure 6-3. AOC Fate and Transport Modeling Approach - Sediment

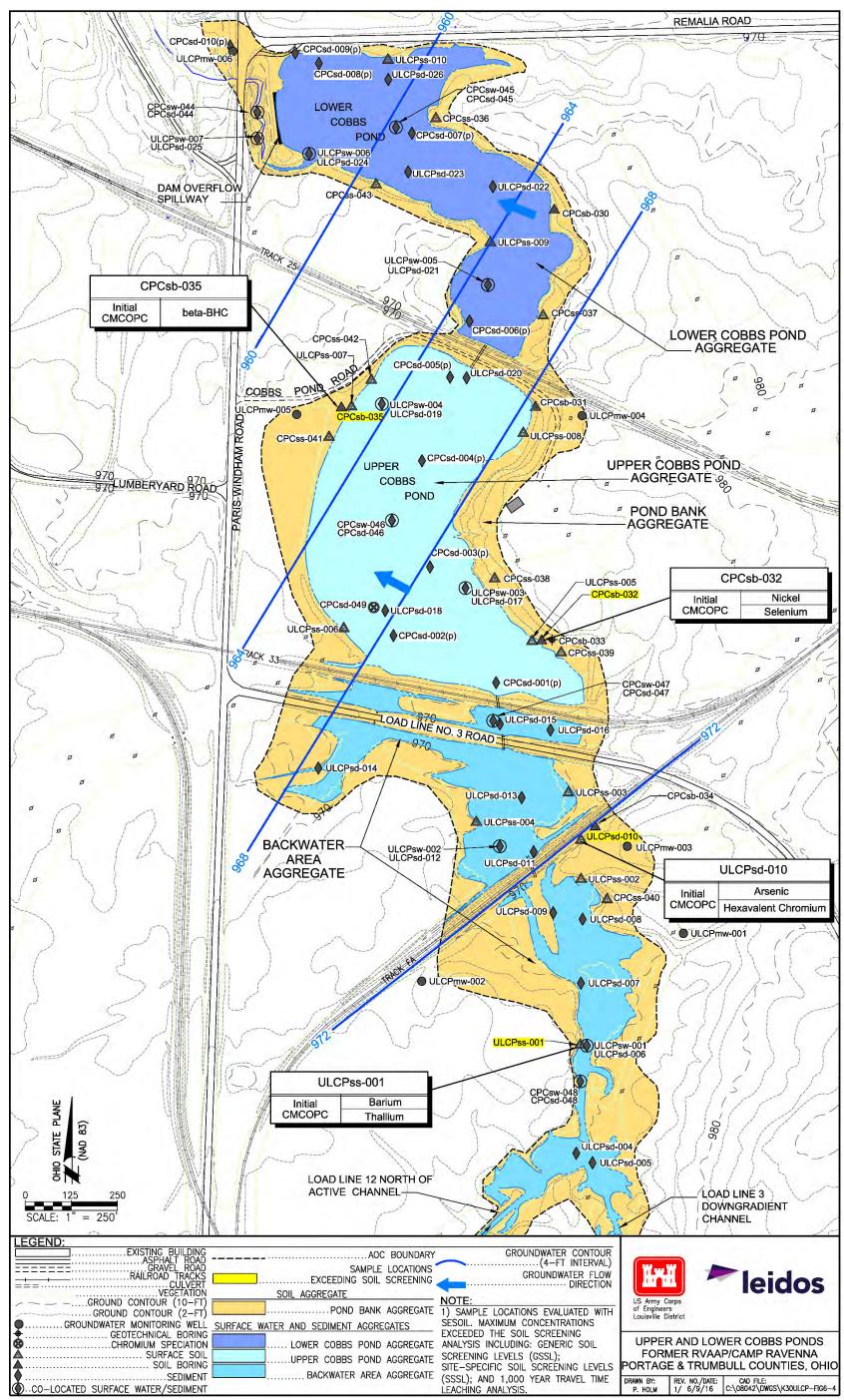


Figure 6-4. Initial CMCOPCs Identified in Soil Screening Analysis for SESOIL Evaluation

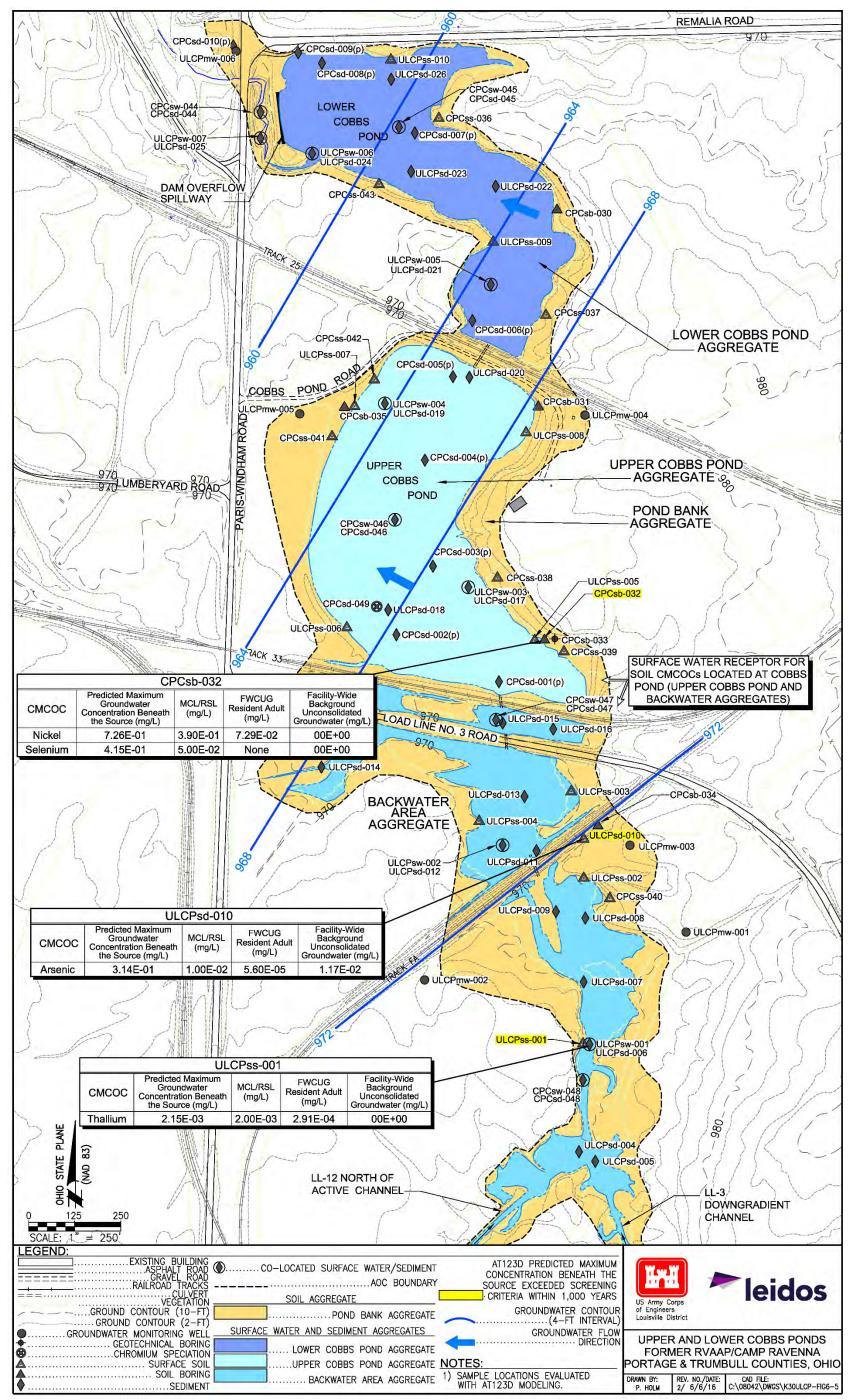


Figure 6-5. CMCOCs Identified for Further Weight-of-Evidence Evaluation Based on AT123D Modeling

1 7.0 RISKASSESSMENT

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5

Camp Ravenna is a controlled-access facility. The Upper and Lower Cobbs Ponds AOC is located in the east-central portion of the facility. The AOC is approximately 39 acres and is comprised of the Backwater Area (5.2 acres), Upper Cobbs Pond (9.4 acres), Lower Cobbs Pond (6.4 acres), and surrounding pond banks (18 acres).

6 7

8 Upper Cobbs Pond ranges from 3–8 ft in depth, and Lower Cobbs Pond ranges from 2–7 ft in depth 9 (MKM 2005). The Backwater Area is located to the south (upstream) of Upper Cobbs Pond and was 10 created through beaver dam construction activity. Flow from the Backwater Area into Upper Cobbs 11 Pond and, subsequently, into Lower Cobbs Pond is through culverts beneath the railroad trackbeds 12 and roads that cross the AOC. Lower Cobbs Pond terminates in a dam overflow spillway adjacent to 13 the intersection of Remalia and Paris-Windham Roads.

14

Upper Cobbs Pond and Lower Cobbs Pond are used for recreational purposes. In addition, both ponds are used by OHARNG for fishing. The AOC may also be accessed by hunters (during annual controlled hunts) and trappers. Security patrols and maintenance activities also occur at the AOC. Future OHARNG Land Use for the AOC (which includes the ponds and their associated sediment and banks) includes recreational use, dam maintenance, wildlife management activities, and training purposes. Surrounding adjacent areas may be used for recreational activities as well as maneuver and multipurpose military training operations.

22

Three Land Uses for the RVAAP restoration program are specified in the Technical Memorandum
 (ARNG 2014) for consideration in the RI along with the following Representative Receptors:

25 26

1. Unrestricted (Residential) Land Use – Resident Receptor (Adult and Child).

- 2. Military Training Land Use National Guard Trainee.
- 28 3. Commercial/Industrial Land Use Industrial Receptor (USEPA's Composite Worker).
- 29

27

 $\mathbf{S}_{\mathbf{A}} = \mathbf{S}_{\mathbf{A}} =$

30 Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp 31 Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then 32 the AOC is also considered to have met the requirements of the other Land Uses (i.e., 33 Commercial/Industrial and Military Training), and those other Land Uses do not require evaluation.

34

35 7.1 DATA EVALUATION FOR HUMAN HEALTH AND ECOLOGICAL RISK 36 ASSESSMENTS

37

38 The purpose of this data evaluation is to develop a set of chemical data suitable for use in the HHRA 39 and ERA. The Upper and Lower Cobbs Ponds AOC data were evaluated to establish data aggregates 40 and identify a list of SRCs. 1 2

7.1.1 Data Aggregates

This section describes the data aggregates for the media for which human and ecological receptors are potentially exposed, followed by a summary of SRCs in Section 7.1.2. Section 4.0 provides a summary of available data.

6 7

7.1.1.1 <u>Soil Data</u>

Soil at the Pond Bank was aggregated by the following depth intervals:

9 10 11

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15

16

8

- Surface soil with an exposure depth of 0–1 ft bgs was evaluated for the Resident Receptor (Adult and Child) and for potential risk to ecological receptors because this layer is the most active biological zone (USACE 2003a). Table 7-1 presents the risk assessment data set for surface soil (0–1 ft bgs). For this risk assessment, discrete samples collected in July–August 2001 during the Phase II RI (MKM 2005) and February–March 2010 during the PBA08 RI were used to characterize surface soil.
- Subsurface soil with an exposure depth of 1–13 ft bgs was evaluated for the Resident Receptor (Adult and Child). Discrete samples collected in March 2010 during the PBA08 RI with a starting depth within these intervals were used to characterize subsurface soil. Table 7-20 presents the risk assessment data sets for subsurface soil.
- 21 22

23

7.1.1.2 <u>Sediment and Surface Water Data</u>

The AOC is naturally divided into three sediment and surface water EUs (Figure 2-2): Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. Two channels from Load Lines 3 and 12 drain northward through the Backwater Area and into Upper Cobbs Pond. Overflow from Upper Cobbs Pond discharges through a spill way into Lower Cobbs Pond and to a receiving unnamed tributary to Sand Creek. The channels tend to hold water for extended periods of time due to the low permeability of soil in the AOC.

30

31 Discrete sediment samples collected in August-September 2001 during the Phase II RI (MKM 2005) 32 and in March-April 2010 during the PBA08 RI and ISM samples collected in June 2003 for the 33 FWBWOS (USACE 2005a) were used to characterize risks. Sediment samples included in the risk 34 assessments are limited to sediment samples collected with a starting depth of 0 ft bgs (e.g., 0-0.5 or 35 0-1.75 ft bgs). Sediment samples collected from 0.5-2 ft bgs were not used in the risk assessment 36 because exposure of human and ecological receptors is generally limited to surface sediment. 37 Sediment data from samples collected from 0.5-2 ft bgs and older (1996) data not included in the risk 38 assessment were used to characterize nature and extent of contamination (Section 5.0) and are only 39 considered for evaluating trends over time and depth. Samples included in the risk assessment data 40 sets for sediment are listed in Table 7-3. It is inappropriate to combine ISM and discrete sample data due to different levels of variability in these two data types; therefore, the discrete and ISM sediment 41 42 data were evaluated separately.

Surface water samples collected in July 2001 during the Phase II RI (MKM 2005), in March–April
 2010 during the PBA08 RI, and in June 2003 for the FWBWQS (USACE 2005a) were used to
 characterize risks from exposure to chemicals in these EUs. Samples included in the risk assessment
 data sets for surface water are listed in Table 7-4.

- 6 7.1.2 Identification of SRCs
- 7

5

8 Section 4.4 presents the statistical methods and screening criteria used to identify SRCs. The purpose 9 of identifying SRCs is to determine the presence or absence of contamination that is above naturally 10 occurring levels.

11

The 1978 Installation Assessment identified the pink wastewater and washout of residue, dusts, and spills at Load Lines 3 and 12, which were discharged through surface drainage channels towards Upper Cobbs Pond (USATHAMA 1978), as contaminants. Potential site-specific contaminants include TNT, RDX, HMX, nitrate, nitrocellulose, and heavy metals (i.e., lead, chromium, mercury, and arsenic) from the pink water. Other contaminants identified include washout from explosives melt-out and demilitarization activities at Load Lines 3 and 12 and ammonium nitrate and aluminum chloride from Load Line 12.

19

The SRC screen was not limited to only contaminants that may have been products of previous site use. Rather, the SRC screen followed the three steps outlined in the FWCUG Report, as summarized below, using all chemical data available:

23

Background screening. MDCs of naturally occurring inorganic chemicals were compared to
 the facility-wide background concentrations for RVAAP, which are summarized in the
 FWCUG Report. Inorganic constituents detected above facility-wide background
 concentrations or having no background concentrations were retained as SRCs. All detected
 organic chemicals were retained as SRCs.

- 29 Screening of essential human nutrients. Chemicals considered essential nutrients (e.g., • 30 calcium, chloride, iodine, iron, magnesium, potassium, phosphorous, and sodium) are an 31 integral part of the human food supply and are often added to foods as supplements. USEPA 32 recommends these chemicals not be evaluated provided they are: (1) present at low 33 concentrations (i.e., only slightly above naturally occurring levels) and (2) toxic at only very 34 high doses (i.e., much higher than those that could be associated with contact at the AOC) 35 (USEPA 1989). Essential nutrients detected near or below their RDA/RDI-based SLs were eliminated as SRCs. 36
- Frequency of detection screening. In accordance with the FWCUG Report and as revised in
 the FWCUG Position Paper (USACE 2012a), analytes detected in less than 5% of the
 samples are screened out from further consideration with the exception of explosives and
 propellants. A frequency of detection screen was included in the SRC screening for data sets
 having at least 20 discrete samples; however, no SRCs were screened out on this basis.
- 42

1 Details of the SRC screening for each exposure medium are provided in Appendix G, Tables G-1 2 through G-4. The SRCs identified for the Upper and Lower Cobbs Ponds AOC are summarized in 3 Table 7-5.

4

7.2 HUMAN HEALTH RISK ASSESSMENT

5 6

7 This HHRA identifies COCs that may pose potential health risks to humans resulting from exposure to contamination at the Upper and Lower Cobbs Ponds AOC. This HHRA was conducted as part of 8 9 the PBA08 RI and is based on the methods from the following guidance documents:

10 11

12

13

- FWHHRAM (USACE 2005b), •
- FWCUG Report (USACE 2010a), •
- FWCUG Position Paper (USACE 2012a), and •
- Technical Memorandum (ARNG 2014). •
- 14 15

16 To accomplish the goal of streamlined decision making, the FWCUG Report was developed to 17 support risk assessments of the remaining AOCs within the former RVAAP. The FWCUG Report 18 contains calculated FWCUGs and guidance for applying FWCUGs to accelerate the risk assessment 19 process. This approach takes advantage of the many risk assessment inputs and decisions that have 20 previously been accepted by stakeholders applying the CERCLA process at the former RVAAP.

21

22 The agreed-upon risk assessment methods have been documented in the FWHHRAM (USACE 23 2005b) and follow standard USEPA-approved risk assessment guidance. Other approaches, such as 24 calculating the sum-of-ratios (SOR), were developed in the FWCUG Report (USACE 2010a) and 25 FWCUG Position Paper (USACE 2012a). The Technical Memorandum (ARNG 2014) identifies 26 future Land Uses and prescribes the applicable receptors for these Land Uses to be evaluated in an RI. 27

28

Using the FWCUGs and information from the RI sampling, the approach to the HHRA is as follows:

29 30

- 1. Specify Land Use(s) and Representative Receptor(s).
- 2. Identify Media of Concern.
- 32 3. Synthesize and analyze data to Identify SRCs – Follow the requirements specified in 33 the FWHHRAM (USACE 2005b) and the FWCUG Position Paper (USACE 2012a), 34 perform data analysis and mapping to identify SRCs, establish EUs, and calculate 35 exposure point concentrations (EPCs) for each COPC. The results of the mapping and 36 data analysis for the Upper and Lower Cobbs Ponds AOC to identify SRCs are presented 37 in Sections 4.0 and 5.0 and are summarized in Section 7.1.
- 38 4. Identify COPCs – To identify COPCs, the MDC of all SRCs are screened against the 39 most stringent chemical-specific FWCUG of all RVAAP receptors at a target cancer risk 40 level of 1E-06 and non-carcinogenic target HQ of 0.1 for the Resident Receptor (Adult 41 and Child) and National Guard Trainee. If no FWCUGs exist for an SRC, the USEPA residential RSLs (from RSL table dated June 2015) are used for this screen. 42

| 1 | 5. Compare to Appropriate FWCUGs and Identify COCs - Comparing COPC exposure |
|----------|---|
| 2 | concentrations to FWCUGs and determining COCs follows guidance presented in the |
| 3 | FWCUG Position Paper (USACE 2012a) and Technical Memorandum (ARNG 2014). |
| 4 | The COC determination process is as follows: |
| 5 | • Report all carcinogenic- and non-carcinogenic-based FWCUGs corresponding to a |
| 6 | TR of 1E-05 and target HQ of 1 using the most stringent of the Resident Receptor |
| 7 | (Adult and Child) FWCUGs to evaluate Unrestricted (Residential) Land Use for |
| 8 | each COPC. If no FWCUG is available for a COPC, the residential RSL, adjusted |
| 9 | to represent a TR of 1E-05 or target HQ of 1, is used. |
| 10 | • Report critical effect and target organ for each non-carcinogenic-based FWCUG. |
| 11 | • Compare the selected FWCUG to the EPC, including an SOR. |
| 12 | • For non-carcinogens, compare the EPC to the target HQ FWCUG. Sum the |
| 13 | ratios of EPC/FWCUG for COPCs that affect similar target organs or do not |
| 14 | have an identified target organ. |
| 15 | • For carcinogens, compare the EPC to the TR FWCUG. Sum the ratios of |
| 16 | EPC/FWCUG for all carcinogens. |
| 17 | • Identify the COPC as a COC for a given receptor if: |
| 18 | • The EPC exceeds the most stringent of the Resident Receptor (Adult and |
| 19 | Child) FWCUGs for either the 1E-05 target cancer risk or the target HQ of 1; |
| 20 | or |
| 21 | • The SOR for all carcinogens or non-carcinogens that may affect the same |
| 22 | organ is greater than one. Chemicals contributing at least 10% to an SOR |
| 23 | greater than one are also considered COCs. In accordance with the FWCUG |
| 24 25 | Position Paper (USACE 2012a), chemicals contributing greater than 5% but |
| 23 26 | less than 10% to the SOR must be further evaluated before being eliminated as COCs. |
| 20 27 | as cocs. |
| 27 | The process for calculating FWCUGs rearranges the cancer risk or non-cancer hazard equations to |
| 20 29 | obtain a concentration that will produce a specific risk or hazard level (USEPA 1991, USACE |
| 30 | 2010a). For example, the FWCUG for arsenic at the cancer risk level of 1E-05 for the Resident |
| 31 | Receptor Adult is the concentration of arsenic that produces a risk of 1E-05 when using the exposure |
| 32 | parameters specific to the Resident Receptor Adult. |
| 33 | |
| 34 | For carcinogens, risk is expressed as the probability that an individual will develop cancer over a |
| 35 | lifetime as a result of exposure to the carcinogen. Cancer risk from exposure to contamination is |
| 36 | expressed as the increased chance of cancer above the normal background rate. In the United States, |
| 37 | the background chance of contracting cancer is a little more than 3 in 10 for women and a little less |
| 38 | than 5 in 10 for men, or 3E-01 to 5E-01 (American Cancer Society 2015). The calculated incremental |
| 39 | lifetime cancer risk (ILCR) is compared to the range specified in the NCP of 10 ⁻⁶ to 10 ⁻⁴ , or 1 in a |
| 40 | million to 1 in 10,000 exposed persons developing cancer (USEPA 1990). Cancer risks above 10 ⁻⁴ are |
| 41 | considered unacceptable. The range between 10^{-6} and 10^{-4} is of concern, and any decisions to address |
| 42 | risk further in this range, either through additional study or engineered control measures, should |
| 43 | account for uncertainty in the risk estimates. The Ohio EPA Division of Environmental Response and |

1 Revitalization (DERR) program has adopted a human health cumulative ILCR goal within this range

2 of 1E-05 to be used as the level of acceptable excess cancer risk and for developing remediation goals

- for the site. The DERR notes that the defined risk goal should be applied as a goal, recognizing the
 need to retain flexibility during the evaluation and selection of remedial alternatives.
- 5

6 In addition to developing cancer from exposure to chemicals, an individual may experience other 7 adverse effects. The term "adverse effects" is used here to describe a wide variety of systemic effects 8 ranging from minor irritations, such as eye irritation and headaches, to more substantial effects, such 9 as kidney or liver disease and neurological damage. The risk associated with non-carcinogenic 10 chemicals is evaluated by comparing an estimated exposure (i.e., intake or dose) from AOC media to an acceptable exposure expressed as a reference dose (RfD). The RfD is the threshold level below 11 12 which no adverse effects are expected to occur in a population, including sensitive subpopulations. 13 The ratio of intake over the RfD is the HQ (USEPA 1989).

14

15 The SOR is used to account for potential additive effects from exposure to multiple chemicals that 16 can cause the same effect (e.g., cancer) or affect the same target organ. Cancer risk is assumed to be 17 additive for all carcinogens. Non-cancer risk is assumed to be additive for chemicals with similar sites 18 of toxicological action (i.e., target organ such as liver or critical effect such as adversely affecting the 19 ability to reproduce). This approach compares the EPC of each COPC to the FWCUG to determine a 20 ratio. The sum of these individual ratios is then compared to one (using one significant figure). The 21 SOR method is based on the principle that a ratio greater than one represents unacceptable cumulative 22 exposure (i.e., above FWCUGs if adjusted for exposure to multiple COPCs), and a ratio less than or 23 equal to one represents acceptable cumulative exposure (i.e., below FWCUGs if adjusted for 24 exposure to multiple COPCs). The FWCUGs for some chemical/receptor combinations are less than 25 the background concentration. In these instances, the chemical concentrations are compared to 26 background concentrations to identify COCs. Since the background concentration is not risk-based, 27 these chemicals are not included in the SOR calculations.

28

29 COCs identified by comparing EPCs to FWCUGs are further evaluated in an uncertainty analysis to30 identify COCs requiring evaluation in an FS.

- 31
- 32 33

34

- Uncertainty Assessment Assess sources of uncertainty, as well as the potential bias they impart to the risk assessment (i.e., whether conservatism is increased or decreased) and approaches for minimizing their impact on the conclusions of the RI.
- 35
 2. Identify COCs for Potential Remediation Make a final determination of COCs
 36 requiring evaluation in an FS and potential remediation.
- 37

38 These steps are executed in the following subsections.

1 2

7.2.1 Land Use and Representative Receptors

Three Land Uses for the RVAAP restoration program are specified in the Technical Memorandum
(ARNG 2014) for consideration in the RI along with the following Representative Receptors:

- 5 6
- 1. Unrestricted (Residential) Land Use Resident Receptor (Adult and Child).
- 2. Military Training Land Use National Guard Trainee.
- 3. Commercial/Industrial Land Use Industrial Receptor (USEPA's Composite Worker).
- 8 9

14

16

7

Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp
Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, the
AOC is also considered to have met the requirements of the other Land Uses (i.e.,
Commercial/Industrial and Military Training), and those other Land Uses do not require evaluation.

15 7.2.2 Identify Media of Concern

Media of concern at the Upper and Lower Cobbs Ponds AOC are surface soil, subsurface soil, surface water, and sediment. Groundwater is present at this AOC but will be evaluated (including risk assessment) in a separate document, as described in Section 1.2.

- 20
- 21 22

7.2.3 Data Synthesis and Analysis to Identify SRCs

The results of the mapping and data analysis for the Upper and Lower Cobbs Ponds AOC to identify SRCs are presented in Sections 4.0 and 5.0 and are summarized in Section 7.1.

26 **7.2.4** Identify COPCs

27

25

Details of the COPC screening for each exposure medium are provided in Appendix G, Tables G-1
through G-4. The COPCs identified for the media of concern at the Upper and Lower Cobbs Ponds
AOC are presented in Table 7-6 and are summarized below.

31

No RfD or cancer potency factors are available for acenaphthylene, benzo(*ghi*)perylene, and phenanthrene; therefore, the RSL for pyrene was used for these PAHs (NDEP 2006). Hexavalent chromium was detected in 1 of 2 soil samples and 3 of 23 sediment samples analyzed for hexavalent and total chromium at the AOC. Since hexavalent chromium was detected as part of the conservative screening approach for identifying COPCs, the FWCUG for hexavalent chromium (the more toxic of the two chromium species evaluated) was used at this stage.

38

39 7.2.4.1 COPCs in Surface Soil

40

Of the 44 chemicals detected in surface soil (0-1 ft bgs) samples collected from the Pond Bank, 36
(18 inorganic chemicals, 16 SVOCs, 1 pesticide, and 1 explosive) were identified as SRCs. Risk-

based screening identified five inorganic chemicals (aluminum, arsenic, chromium, cobalt, and
cyanide) and one SVOC [benzo(a)pyrene] as COPCs in surface soil (0–1 ft bgs).

3 4

5

7.2.4.2 <u>COPCs in Subsurface Soil</u>

6 Of the 37 chemicals detected in the subsurface soil (1–13 ft bgs) samples collected from the Pond 7 Bank, 18 (4 inorganic chemicals, 13 VOCs, and 1 explosive) were identified as SRCs. Risk-based 8 screening identified only benzo(a)pyrene as a COPC.

9 10

7.2.4.3 <u>COPCs in Sediment</u>

11

12 Backwater Area

13

Of the 47 chemicals detected in discrete sediment samples collected at the Backwater Area, 38 (16 inorganic chemicals, 14 SVOCs, 3 VOCs, 4 explosives, and 1 PCB) were identified as SRCs. Riskbased screening identified three inorganic chemicals (aluminum, chromium, and cobalt) and five SVOCs [benz(a)anthracene, benzo(a)pyrene, benzo(a)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] as COPCs in the Backwater Area.

19

20 Upper Cobbs Pond

21

Of the 43 chemicals detected in discrete sediment samples collected at Upper Cobbs Pond, 34 (15 inorganic chemicals, 10 SVOCs, 2 VOCs, 6 explosives, and 1 PCB) were identified as SRCs. Riskbased screening identified four inorganic chemicals (aluminum, chromium, hexavalent chromium, and cobalt) and 1 SVOC [benzo(a)pyrene] as COPCs in Upper Cobbs Pond. Of the 27 chemicals detected in the sediment ISM sample collected at Upper Cobbs Pond, 10 (7 inorganic chemicals and 3 SVOCs) were identified as SRCs. Risk-based screening identified two inorganic chemicals (chromium and ammonia) as COPCs in this sample.

29

30 Lower Cobbs Pond

31

Of the 42 chemicals detected in discrete sediment samples collected at Lower Cobbs Pond, 34 (17 inorganic chemicals, 11 SVOCs, 3 VOCs, and 3 explosives) were identified as SRCs. Risk-based screening identified six inorganic chemicals (aluminum, arsenic, chromium, hexavalent chromium, cobalt, and cyanide) and one SVOC [benzo(a)pyrene] as COPCs in Lower Cobbs Pond. Of the 24 chemicals detected in the sediment ISM sample collected at Lower Cobbs Pond, 14 (13 inorganic chemicals and 1 SVOC) were identified as SRCs. Risk-based screening identified four inorganic chemicals (aluminum, ammonia, chromium, and cobalt) as COPCs in this sample.

39

Ammonia was initially identified as a COPC in two sediment ISM samples because no FWCUGs or
 RSLs are available. Ammonia is reported for evaluation of water quality for protection of aquatic

42 receptors and has very low toxicity to humans. Therefore, ammonia is not evaluated further in this

43 HHRA.

```
1
      7.2.4.4
               COPCs in Surface Water
 2
 3
      Backwater Area
 4
 5
      Of the 28 chemicals detected in surface water samples collected at the Backwater Area, 22 (18
      inorganic chemicals, 3 VOCs, and 1 explosive) were identified as SRCs. Risk-based screening
 6
 7
      identified five inorganic chemicals (arsenic, cobalt, manganese, sulfate, and sulfite) as COPCs in the
 8
      Backwater Area.
 9
10
      Upper Cobbs Pond
11
12
      Of the 21 chemicals detected in surface water samples collected at Upper Cobbs Pond, 11 (9
13
      inorganic chemicals, 1 SVOC, and 1 explosive) were identified as SRCs. Risk-based screening
14
      identified sulfate and sulfite as COPCs in Upper Cobbs Pond.
15
16
      Lower Cobbs Pond
17
18
      Of the 24 chemicals detected in surface water samples collected at Lower Cobbs Pond, 14 (12
19
      inorganic chemicals and 2 SVOCs) were identified as SRCs. Risk-based screening identified three
20
      inorganic chemicals (manganese, sulfate, and sulfite) as COPCs in Lower Cobbs Pond.
21
22
      Sulfate and sulfite were initially identified as COPCs in surface water because no FWCUGs or RSLs
23
      are available. These anions are reported for evaluation of water quality for protection of aquatic
24
      receptors and have very low toxicity to humans. There is a secondary maximum contaminant level
25
      (SMCL) of 250 mg/L established for sulfate in drinking water. SMCLs are established only as
      guidelines for aesthetic considerations, such as taste, color, and odor. Chemicals are not considered to
26
27
      present a risk to human health at the SMCL, and the concentration of sulfate and sulfite detected in all
28
      three EUs are well below the sulfate SMCL. Therefore, sulfate and sulfite are not evaluated further in
29
      this HHRA.
30
31
      7.2.5
             Compare to Appropriate FWCUGs
32
33
      7.2.5.1
               Selection of Appropriate FWCUGs
34
      As specified in the Technical Memorandum (ARNG 2014), EPCs for each AOC should initially be
35
```

36 evaluated to determine if no further action is necessary at an AOC to attain Unrestricted (Residential) 37 Land Use. Unrestricted (Residential) Land Use is evaluated using FWCUGs for the Resident 38 Receptor (Adult and Child). The Resident Receptor (Adult and Child) FWCUGs provided in Tables 39 7-7 through 7-9 are the lower of the Resident Receptor (Adult and Child) values for each COPC and 40 endpoint (non-cancer and cancer) corresponding to a TR of 1E-05 and target HQ of 1. The critical 41 effect or target organ associated with the toxicity values used to calculate the non-cancer FWCUGs 42 are also provided.

1 Chromium Speciation

- FWCUGs and RSLs are available for hexavalent and trivalent chromium. Existing data at other AOCs, such as Building 1200 and Anchor Test Area (USACE 2012b, USACE 2012c), indicate chromium exists predominantly in the trivalent state rather than the more toxic hexavalent state.
- 6

2

7 Total chromium concentrations in soil at this AOC are at or near background concentrations. The 8 MDC of chromium in soil at the Upper and Lower Cobbs Ponds AOC is 24.8 mg/kg compared to the 9 RVAAP background concentrations of 17.4 mg/kg in surface soil and 27.2 mg/kg in subsurface soil. 10 Historical and current earth moving activities at Camp Ravenna have reduced the distinction between 11 surface and subsurface soil. One surface soil sample was analyzed for both hexavalent and total chromium. The reported concentration of hexavalent chromium (1 mg/kg) was 10.4% of the total 12 13 chromium concentration of 9.6 mg/kg [i.e., less than the 14% present in the study used to calculate the unit risk factor (URF) for hexavalent chromium]. This hexavalent chromium concentration is also 14 15 less than the residential RSL (3 mg/kg based on a TR of 1E-05) for hexavalent chromium, indicating 16 hexavalent chromium is not present above the residential RSL for hexavalent chromium. Therefore, 17 total chromium results for surface soil were evaluated using the FWCUGs and RSLs for trivalent 18 chromium.

19

20 Chromium concentrations appear elevated in sediment at the Upper and Lower Cobbs Ponds AOC 21 with reported concentrations in 12 of 30 samples exceeding the Ohio EPA SRV of 29 mg/kg. Most of 22 the discrete sediment samples (23 of 28) were analyzed for both hexavalent and total chromium. The 23 results indicate most of the chromium present is not in the hexavalent form (i.e., hexavalent 24 chromium was detected in only 3 of 23 samples). However, in the samples where hexavalent 25 chromium was detected it made up 5%, 15%, and 53% of the total chromium. The detected 26 hexavalent chromium concentrations also exceed the hexavalent chromium-specific residential RSL 27 of 3 mg/kg. Because adequate hexavalent chromium data are available to characterize the sediment 28 EUs, total chromium results are evaluated using the trivalent chromium FWCUGs and RSLs because 29 the hexavalent portion of the total chromium result is evaluated separately.

30

No surface water samples were analyzed for hexavalent chromium. Chromium was not identified as a
 COPC in surface water using the screening values for hexavalent chromium.

33

34 7.2.5.2 Exposure Point Concentrations for Comparison to FWCUGs

35

36 Soil EPCs

37

EPCs were calculated for each depth interval using analytical results from the discrete samples presented in Tables 7-1 and 7-2. As indicated in Section 7.1.1.1, the depth intervals for the Resident Receptor are 0–1 ft bgs for surface soil and 1–13 ft bgs for subsurface soil. Per the FWHHRAM, the EPC is either the 95% upper confidence limit (UCL) of the mean or the MDC, whichever value is lowest. If the 95% UCL could not be determined, the EPC is the MDC. 1

Sediment and Surface Water EPCs

- 2
- 3 EPCs were calculated for each EU using analytical results from the discrete samples presented in
- 4 Tables 7-3 and 7-4. As indicated in Section 7.1.1.2, the sediment risk assessment for the Resident 5 Receptor includes sediment samples collected with a starting depth of 0 ft bgs (e.g., 0-0.5 or 0-1.75 ft
- bgs). Per the FWHHRAM, the EPC is either the 95% UCL of the mean or the MDC, whichever value
- 7 is lowest. If the 95% UCL could not be determined, the EPC is the MDC. In addition to the EPC
- 8 calculated from the discrete data, the reported concentration in each sediment ISM sample was9 evaluated separately.
- 10
- 11 12

7.2.5.3 Identification of COCs for Unrestricted (Residential) Land Use

- 13 The Upper and Lower Cobbs Ponds AOC COCs for Unrestricted (Residential) Land Use, as 14 represented by the Resident Receptor (Adult and Child), are presented below.
- 16 COCs for Surface Soil (0–1 ft bgs)
- 17

15

18 COC screening for surface soil (0-1 ft bgs) for the Resident Receptor (Adult and Child) is detailed in
19 Appendix G, Table G-5. No COCs were identified in surface soil (0-1 ft bgs) for the Resident
20 Receptor (Adult and Child), as explained below.

- 21
- 22 23

COPCs with EPCs Lower than the Resident Receptor (Adult and Child) FWCUG

- All aluminum, chromium, cobalt, cyanide, and benzo(a)pyrene EPCs are lower than the Resident Receptor (Adult and Child) FWCUG.
- 26

28

27 COPCs with EPCs Exceeding the Resident Receptor (Adult and Child) FWCUG

The EPC of arsenic (13.5 mg/kg) exceeds the FWCUG of 4.25 mg/kg but is below the surface soil facility-wide background concentration of 15.4 mg/kg. Because the FWCUG is less than the background concentration, the background concentration is used as the CUG for this inorganic chemical. Thus, arsenic is representative of background and is not a COC in surface soil.

33 34

4 SOR Analysis

35

36 No COCs were identified based on the SOR analysis as summarized below.

37

Five COPCs (aluminum, arsenic, chromium, cobalt, and cyanide) identified in surface soil have FWCUGs for non-cancer endpoints. The EPCs for aluminum, arsenic, chromium, and cobalt are less than the facility-wide background concentrations for surface soil; therefore, these metals are not included in the SOR. Only one COPC (cyanide) with a non-cancer endpoint is present above background concentrations; therefore, no SOR was calculated.

Three surface soil COPCs [arsenic, cobalt, and benzo(a)pyrene] have FWCUGs for the cancer endpoint. The FWCUG for arsenic is less than the background concentration for this inorganic chemical; therefore, the background concentration is used as the CUG, and arsenic is not included in the SOR. The EPC for cobalt is less than the facility-wide background concentration for surface soil; therefore, this metal is not included in the SOR. Only one COPC [benzo(a)pyrene] with a cancer endpoint is present above background concentrations; therefore, no SOR was calculated.

8

COCs for Subsurface Soil (1–13 ft bgs)

9 10

11 COC screening for subsurface soil (1-13 ft bgs) for the Resident Receptor (Adult and Child) is 12 detailed in Appendix G, Table G-6. No COCs were identified in subsurface soil (1–13 ft bgs) for the 13 Resident Receptor (Adult and Child) because the EPC of the only COPC [benzo(a)pyrene] is lower 14 than the Resident Receptor (Adult and Child) FWCUG. Because only one COPC is present in 15 subsurface soil, no SOR was calculated.

16

17 COCs for Sediment

18

19 COC screening for sediment for the Resident Receptor (Adult and Child) is detailed in Appendix G, 20 Tables G-7 through G-9. Arsenic, benzo(a)pyrene, benz(a)anthracene, benzo(b)fluoranthene, and 21 dibenz(a,h)anthracene were identified as COCs for the Resident Receptor (Adult and Child), as 22 explained below.

24 COPCs with EPCs Lower than the Resident Receptor (Adult and Child) FWCUG

25

23

All aluminum, total chromium, hexavalent chromium, cobalt, cyanide, benz(a)anthracene,
benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene EPCs are lower than the
Resident Receptor (Adult and Child) FWCUG.

29 30

COPCs with EPCs Exceeding the Resident Receptor (Adult and Child) FWCUG

31

The EPCs of benzo(a)pyrene at the Backwater Area and arsenic at Lower Cobbs Pond exceed theFWCUG.

- 34
- 35 36 37

38

- The EPC for benzo(a)pyrene (0.89 mg/kg) at the Backwater Area exceeds the FWCUG of 0.221 mg/kg. The EPC is the MDC at ULCPsd-015 collected in 2001. Detected benzo(a)pyrene concentrations in the other Backwater Area samples ranged from 0.063–0.39 mg/kg. Benzo(a)pyrene was identified as a COC at the Backwater Area. PAHs are not COPCs at Upper Cobbs Pond or Lower Cobbs Pond.
- Arsenic was identified as a COC at Lower Cobbs Pond because the EPC (26.9 mg/kg)
 exceeds both the FWCUG for the Resident Receptor Adult of 4.25 mg/kg and facility-wide
 background concentration of 19.5 mg/kg. The MDC of arsenic (34.3 mg/kg) at ULCPsd-026,
 collected in 2001, is the only sample concentration greater than the facility-wide background

2 ranged from 5.1–20.2 mg/kg. 3 4 SOR Analysis 5 6 Three additional COCs, all PAHs, were identified based on the SOR analysis, as summarized below. 7 8 Four COPCs [aluminum, chromium (as trivalent chromium), cobalt, and cyanide] identified 9 in sediment at the Backwater Area have FWCUGs for non-cancer endpoints. The total SOR 10 for these COPCs is less than or equal to one; therefore, no COCs were identified. 11 • Four COPCs [aluminum, chromium (as trivalent chromium), hexavalent chromium, and 12 cobalt] identified in sediment at Upper Cobbs Pond have FWCUGs for non-cancer endpoints. 13 The total SOR for these COPCs is less than or equal to one; therefore, no COCs were 14 identified. 15 Six COPCs [aluminum, arsenic, chromium (as trivalent chromium), hexavalent chromium, • cobalt, and cyanide] identified in sediment at Lower Cobbs Pond have FWCUGs for non-16 17 cancer endpoints. The total SOR for these COPCs is less than or equal to one; therefore, no 18 COCs were identified. 19 Six **COPCs** [cobalt, benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, • 20 dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] identified in sediment at the Backwater 21 Area have FWCUGs for the cancer endpoint. The SOR of five is primarily due to 22 benzo(a)pyrene having an EPC above the FWCUG. Benzo(a)pyrene was previously 23 identified as a COC at this EU. Benz(a)anthracene, benzo(b)fluoranthene, and 24 dibenz(a,h)anthracene each contribute at least 5% to the SOR; therefore, the SOR identifies 25 three additional PAHs as COCs at the Backwater Area. 26 Three COPCs [hexavalent chromium, cobalt, and benzo(a)pyrene] identified in sediment at • 27 Upper Cobbs Pond have FWCUGs for the cancer endpoint. The calculated SOR is less than 28 or equal to one; therefore, no additional COCs were identified at Upper Cobbs Pond. 29 Four COPCs [arsenic, hexavalent chromium, cobalt, and benzo(a)pyrene] identified in • 30 sediment at Lower Cobbs Pond have FWCUGs for the cancer endpoint. The FWCUG for 31 arsenic is less than the background concentration for this inorganic chemical; therefore, the 32 background concentration is used as the CUG, and arsenic is not included in the SOR. The 33 calculated SOR for the remaining chemicals is less than or equal to one; therefore, no 34 additional COCs were identified at Lower Cobbs Pond. 35 36 **COCs for Surface Water** 37 38 COC screening for surface water for the Resident Receptor (Adult and Child) is detailed in Appendix 39 G, Tables G-10 and G-11. Cobalt and manganese were identified as COCs as explained below. 40 41 **COPCs** with EPCs Lower than the Resident Receptor (Adult and Child) FWCUG 42 43 All arsenic and selenium EPCs are lower than the Resident Receptor (Adult and Child) FWCUG.

concentration. Reported arsenic concentrations in the other Lower Cobbs Pond samples

1 2 COPCs with EPCs Exceeding the Resident Receptor (Adult and Child) FWCUG

The EPCs of cobalt and manganese at the Backwater Area exceed the Resident Receptor (Adult and
Child) FWCUG.

5

6 No FWCUGs are available for cobalt in surface water. The EPC for cobalt (0.01 mg/L) 7 exceeds the USEPA tap water RSL of 0.006 mg/L. Reported concentrations of cobalt range 8 from 0.00039-0.0107 mg/L with the MDC reported in sample ULCPsw-001 collected in 9 2001. Cobalt is identified as a COC at the Backwater Area. Cobalt was not detected at Upper 10 Cobbs Pond. Cobalt was detected in one sample at Lower Cobbs Pond at a concentration of 11 0.00015 mg/L. Thus, cobalt is not a COC at these two EUs. 12 The EPC for manganese (15.1 mg/L) in the Backwater Area exceeds the Resident FWCUG of 13 6.3 mg/L. The highest concentration of manganese (15.8 mg/L) was detected in the 14 Backwater Area at sample location ULCPsw-002 in 2001. The field duplicate for this sample 15 had a reported concentration of 7.36 mg/L. Manganese was identified as a COC at the 16 Backwater Area. The MDCs of manganese in surface water in Upper Cobbs Pond and Lower 17 Cobbs Pond are 0.528 mg/L and 0.654 mg/L, respectively. Thus, manganese is not a COC in 18 surface water at these two EUs. 19 20 SOR Analysis 21 22 No additional COCs were identified based on the SOR analysis, as explained below. 23 24 Four COPCs (arsenic, cobalt, manganese, and selenium) identified in surface water at the • 25 Backwater Area have FWCUGs for non-cancer endpoints. The total SOR for these COPCs is 26 four due to cobalt and manganese each having a ratio of two. Therefore, no additional COCs 27 were identified. 28 • Only one COPC (arsenic) identified in surface water at the Backwater Area has a FWCUG 29 for the cancer endpoint; therefore, no SOR was calculated. 30 No COPCs were identified in surface water at Upper Cobbs Pond and only one COPC • 31 (manganese) was identified in surface water at Lower Cobbs Pond; therefore, no SOR was 32 conducted for these EUs. 33 34 7.2.6 **Uncertainty Assessment** 35 36 The sources of uncertainty, as well as the potential bias they impart to the risk assessment (i.e., 37 whether conservatism is increased or decreased) and approaches for minimizing their impact on the 38 conclusions of the RI, are briefly discussed below. 39 40 7.2.6.1 **Uncertainty in Estimating Potential Exposure** 41 42 Sources of uncertainty in estimating potential human exposure include sampling and analysis

43 limitations, comparison to background concentrations to identify SRCs, and estimation of EPCs.

Sampling Limitations. Uncertainties arise from limits on the media sampled, the total number and specific locations that can be sampled, and the parameters chosen for analysis to characterize the AOC. A total of 75 surface soil (0-1 ft bgs) samples, 32 sediment samples, and 16 surface water samples were available for the HHRA. Samples were collected from areas biased toward areas anticipated to have the highest level of potential contamination to delineate potential sources. The results of surface soil sampling were used to efficiently guide selection of locations for subsurface soil sampling with a bias toward the areas of highest potential contamination.

8

9 Analytical Limitations. Uncertainty is associated with the chemical concentrations detected and 10 reported by the analytical laboratory. The quality of the analytical data used in the risk assessment 11 was maximized and uncertainty was minimized by implementing QA/QC procedures that specify 12 how samples are selected and handled; however, sampling errors, laboratory analysis errors, and data 13 analysis errors can occur. Beyond the potential for errors, there is normal variability in analytical 14 results.

15

16 Some current analytical methods are limited in their ability to achieve detection limits at or below 17 risk-based SLs. Under these circumstances, it is uncertain whether the true concentration is above or 18 below the SLs that are protective of human health. When analytes have a mixture of detected and 19 non-detected concentrations, EPC calculations may be affected by these detection limits. Risks may 20 be overestimated as a result of some sample concentrations being reported as non-detected at the 21 method detection limit (MDL), when the actual concentration may be much smaller than the MDL. 22 Risks may also be underestimated if some analytes that were not detected in any sample were 23 removed from the COPC list. If the concentrations of these analytes are below the MDL but are above 24 the SL, the risk from these analytes would not be included in the risk assessment results.

25

Identifying SRCs. Part of determining SRCs is to identify chemicals detected above the established RVAAP facility-wide background concentrations. This screen does not account for the potential sources of chemicals, and background values are only available for inorganic chemicals.

29

Uncertainty associated with screening against background results from statistical limitations and natural variation in background concentrations. Because of this variation, inorganic chemical concentrations below the background concentration are likely representative of background conditions. Inorganic chemical concentrations above the background concentration may be above background concentrations or may reflect natural variation. This is especially true for measured concentrations close to the background concentration.

36

At the Upper and Lower Cobbs Ponds AOC, the MDCs of all inorganic chemicals in all soil intervals are less than 2.4x times their background concentrations. Nine inorganic chemicals detected in sediment had MDCs that were one to two times their background concentrations. The consequences of carrying most inorganic chemicals forward as SRCs, even if they actually represent background concentrations, is negligible because they are not toxic at near background concentrations. By contrast, naturally occurring (background) arsenic in soil and sediment exceeds some risk-based CUGs. Therefore, the consequence of identifying arsenic as an SRC if it is, in fact, representative of

1 background can have a significant impact on the conclusions of the risk assessment. The MDCs of 2 arsenic at the AOC are 28.4 mg/kg in surface soil, 17.9 mg/kg in subsurface soil, and 34.3 mg/kg in 3 sediment. The facility-wide background concentrations for arsenic at RVAAP are 15.4 mg/kg for 4 surface soil, 19.8 mg/kg for subsurface soil, and 19.5 mg/kg for sediment. The Ohio SRV for arsenic 5 is 25 mg/kg (Ohio EPA 2008). Vosnakis and Perry (2009) recently published the results of arsenic background studies that included 313 samples of Ohio soil. Naturally-occurring arsenic in these 6 7 samples ranged from 1.6-71.3 mg/kg with 95th percentiles of 21.7 mg/kg in surface soil, 25.5 mg/kg 8 in subsurface soil, and upper tolerance limits of 22.8 and 29.6 mg/kg for surface and subsurface soil, 9 respectively. In other studies, native soil concentrations of arsenic in Ohio have been reported as 10 ranging from 0.5-56 mg/kg (Ohio EPA 1996) and the U.S. Geological Survey's Certificate of 11 Analysis of the Devonian Ohio Shale estimates arsenic concentrations of 68.5 mg/kg are naturally 12 present in bedrock shales (USGS 2004). Based on this information, arsenic appears to be present at 13 the Upper and Lower Cobbs Ponds AOC at naturally occurring concentrations.

14

15 Organic chemicals are not screened against background concentrations even though some organic 16 compounds are present in the environment as a result of natural or human activities not related to the 17 AOC. For example, PAHs are present in the environment as a result of burning fossil fuels and as a component of road dust, vehicle exhaust, tire wear particles, pavement, and slag used as railroad 18 19 ballast and fill. Samples collected near roadways or parking areas may represent normal "urban" 20 sources of PAHs. These issues represent significant sources of uncertainty at sites where low levels of 21 PAHs are found over large areas of the AOC. At the Upper and Lower Cobbs Ponds AOC, PAHs 22 were detected in sediment across the ponds; one or more PAHs were detected in 7 of 10 discrete 23 sediment samples analyzed for SVOCs. PAH concentrations were less than the Resident Receptor 24 (Adult and Child) FWCUGs in all but three sediment samples where concentrations of 25 benzo(a)pyrene ranged from 0.23-0.89, exceeding the FWCUG of 0.221 mg/kg. The highest 26 concentrations were reported near a culvert between Load Line 3 Road and Track 33 in 2001 (0.89 27 mg/kg) and 2010 (0.39 mg/kg).

28

Background concentrations of PAHs in sediment are limited; however, numerous studies have been conducted that examine environmental levels of PAHs in rural and urban surface soil (e.g., ATSDR Bradley et al. 1994, IEPA 2005, MADEP 2002, and Teaf et al. 2008). Reported minimum, maximum, and 95th percentile concentrations of benzo(a)pyrene from numerous studies are shown in Table 7-10. These studies demonstrate the high variability in environmental levels of PAHs within a single study area and among multiple studies.

35

The lack of established RVAAP background concentrations for identifying SRCs for PAHs is a source of uncertainty. Evaluating potential RVAAP process-related sources and other common anthropogenic (non-CERCLA) sources using available PAH environmental data minimizes the impact of this uncertainty on the conclusions of the RI (see Section 7.2.6).

40

41 **Exposure Point Concentrations**. Generally, the 95% UCL on the arithmetic mean was adopted as 42 the EPC for discrete sample results and is considered to represent a conservative estimate of the 43 average concentration. This imparts a small but intentional conservative bias to the risk assessment, provided the sampling captured the most highly contaminated areas. Thus, representative EPCs for the EUs were calculated from discrete data based on the assumption that the samples collected from the EUs were truly random samples. This assumption is not true for the Upper and Lower Cobbs Ponds AOC where sample locations were biased to identify areas of highest contaminant concentrations. Therefore, EPCs generated from these data are likely to represent an upper bound of potential exposure concentrations.

7

8 In addition to calculating EPCs for each EU, individual discrete sample results are evaluated to 9 identify whether potential hotspots are present as a result of specific source areas.

10

11 The EPC of arsenic is less than the FWCUG in surface soil at the Pond Bank, but the MDC exceeds 12 the FWCUG. The MDC of arsenic in surface soil at the Pond Bank was 28.4 mg/kg at sample 13 location ULCPsd-010, collected in 2001, near the railroad trackbeds in the Backwater Area. The 14 reported concentrations in the other 23 surface soil samples range from 4.2–19.7 mg/kg. The RVAAP 15 background concentration for arsenic in surface soil is 15.4 mg/kg and in subsurface soil is 19.8 16 mg/kg. The RVAAP background concentration of arsenic in sediment is 19.5 mg/kg and the Ohio 17 EPA SRV is 25 mg/kg. As noted previously, other studies indicate arsenic may be naturally occurring in Ohio soils at greater than 20 mg/kg. For example, an environmental study of three locations in 18 19 Cuyahoga County performed for Ohio EPA (Weston 2012) showed arsenic ranged from 4.6-25.2 20 mg/kg (22.9 mg/kg excluding statistical outliers) in surface soil (0-2 ft bgs) and 5.3-34.8 mg/kg (22.6 21 mg/kg excluding statistical outliers) in subsurface soil (2-4 ft bgs). Based on this evaluation, arsenic 22 does not represent a hotspot and is not identified as a COC.

23

There are no other instances of COPCs having an MDC greater than a FWCUG with an EPC less than
 the FWCUG. Therefore, no hotspots are identified at the Upper and Lower Cobbs Ponds AOC.

26 27

28

7.2.6.2 <u>Uncertainty in Use of FWCUGs</u>

Sources of uncertainty in the FWCUGs used to identify COCs include selecting appropriate receptors
 and exposure parameters, exposure models, and toxicity values used in calculating FWCUGs.

31

32 Selection of Representative Receptors. Camp Ravenna is a controlled-access facility. The Upper 33 and Lower Cobbs Ponds AOC is located in the east-central portion of the facility and is used for 34 recreational purposes. In addition, both ponds are used by OHARNG for fishing. The AOC may also 35 be accessed by hunters (during annual controlled hunts) and trappers. Security patrols and 36 maintenance activities also occur at the AOC. Future OHARNG Land Use for the AOC (which 37 includes ponds and associated sediment and banks) includes recreational use, dam maintenance, 38 wildlife management activities, and training purposes. Surrounding adjacent areas may be used for 39 recreational activities as well as maneuver and multipurpose military training operations.

40

41 While residential Land Use is unlikely, an evaluation using Resident Receptor (Adult and Child) 42 FWCUGs is included to provide an Unrestricted (Residential) Land Use evaluation. As stated in

43 Paragraph 6.d of the Technical Memorandum, if an AOC fails to meet the Unrestricted (Residential)

1 Land Use, then all three Land Uses [i.e., Unrestricted (Residential) Land Use, Military Training Land

2 Use, and Commercial/Industrial Land Use] will be evaluated.

3

4 Exposure Parameters and Exposure Models. For each primary exposure pathway included in the 5 FWCUGs, assumptions are made concerning the exposure parameters (e.g., amount of contaminated media a receptor can be exposed to and intake rates for different routes of exposure) and the routes of 6 7 exposure. Most exposure parameters have been selected so that errors occur on the side of human 8 health protection. When several of these upper-bound values are combined in estimating exposure for 9 any one pathway, the resulting risks can be in excess of the 99th percentile and outside of the range 10 that may be reasonably expected. Therefore, consistently selecting upper-bound parameters generally 11 leads to overestimation of the potential risk.

12

13 Sediment FWCUGs are based on the same exposure assumptions as the soil FWCUGs and assume 14 Resident Receptors are exposed daily to sediment, as they would be exposed to soil in a residential 15 vard (i.e., exposure occurs daily from age 0-30 years old and children ingest significant amounts of 16 material while playing in the yard). Exposure to sediment in a water body and under more than a foot 17 of water will be less frequent and for a shorter duration. Also, very young children (e.g., less than two 18 years) are not expected to be exposed to sediment located under water and sediment ingestion rates 19 are expected to be less than soil because the presence of overlying water will reduce adherence of 20 sediment and thus ingestion by hand-to-mouth contact. To evaluate the uncertainty associated with 21 the assumption that exposure to sediment is the same as exposure to soil in a residential yard, the 22 online USEPA RSL calculator at https://epa-prgs.ornl.gov/cgi-bin/chemicals/csl search was used to 23 calculate screening levels based on slightly more realistic, but still conservative exposure 24 assumptions. It was assumed that a resident could contact sediment during the warmer months of the 25 year (i.e., June through September). It was further assumed that toddlers (i.e., children under 2 years old) would not be exposed to underwater sediment, young children (2-6 years) and adults would visit 26 27 the ponds 2 days per month (i.e., 8 days per year) and older children (age 6-16 years) would visit the 28 ponds 2 days per week (35 days/year). All other assumptions were unchanged from default soil 29 assumptions (i.e., ingestion rate, exposure time, and exposed skin surface area). The complete input 30 and output files for the calculator are provided in Appendix G, Exhibit G-1. The calculated RSL for 31 exposure to benzo(a)pyrene in sediment during the summer months as described above is 9 mg/kg. 32 The maximum detected benzo(a)pyrene concentration in sediment is 0.89 mg/kg.

33

Toxicity Values. The toxicity of chemicals is under constant study and values change from time to time. The toxicity values used in calculating FWCUGs were the most recent values available at the time of those calculations (September 2008). These values are designed to be conservative and provide an upper-bound estimate of risk.

38

The toxicity and mobility of many inorganic chemicals in the environment depends on the chemical species present. Two important examples are arsenic and chromium. The toxicity values used in developing FWCUGs are for inorganic arsenic; however, these values do not distinguish between arsenite and arsenate. Chromium is generally present in the environment as either the trivalent (Cr+3)

43 or hexavalent (Cr+6) species, with the trivalent form generally being more stable and therefore more

1 common. FWCUGs are available for hexavalent and trivalent chromium. Trivalent chromium has not 2 been shown to be carcinogenic. It is an essential micronutrient but can also be toxic at high doses 3 (i.e., above the RfD used to calculate the FWCUG). FWCUGs for trivalent chromium are based on 4 non-cancerous effects. Hexavalent chromium is much more toxic than trivalent chromium. It is 5 classified as a "known human carcinogen" and may also cause non-cancerous effects. The cancer URF for hexavalent chromium published in USEPA's Integrated Risk Information System (IRIS) is 6 7 based on epidemiological data on lung cancer in workers associated with chromate production. 8 Workers in the chromate industry are exposed to trivalent and hexavalent compounds of chromium. 9 The cancer mortality in the study used to establish the URF was assumed to be due to hexavalent 10 chromium. It was further assumed that hexavalent chromium constituted no less than one seventh of 11 the total chromium in air that the workers were exposed to. As noted in IRIS, the assumption that the 12 ratio of hexavalent to trivalent chromium was 1:6 in this study may lead to a sevenfold 13 underestimation of risk when using this URF to evaluate exposure to hexavalent chromium alone.

14

To avoid underestimating risk, selecting the FWCUG for chromium includes a step that compares the maximum concentration of hexavalent chromium detected in chromium speciation samples to the residential RSL for hexavalent chromium of 3 mg/kg. Concentrations of hexavalent chromium in chromium speciation samples that are less than or equal to 3 mg/kg indicate that hexavalent chromium is not present above hexavalent chromium FWCUGs and supports using trivalent chromium FWCUGs for evaluating total chromium. Using speciation samples to identify the appropriate FWCUG minimizes the associated uncertainty.

22

23 FWCUGs and RSLs below Background Concentrations. One purpose of the HHRA process is to 24 identify COCs and CUGs for evaluating remedial alternatives for remediating residual contamination 25 that has resulted from process operations at the AOC. FWCUGs and RSLs are risk-based values. In 26 some cases, natural or anthropogenic background concentrations, unrelated to process operations, 27 exceed the risk-based FWCUGs and RSLs. For naturally occurring inorganic chemicals this problem 28 is addressed by using the background concentration as the CUG. This introduces uncertainty in the 29 chosen CUG because there is uncertainty in assigning a specific value to background, which can be 30 highly variable.

31

32 No background concentrations are available for organic chemicals, although PAHs are often present 33 in the environment from natural and anthropogenic sources and regulatory standards are often much 34 lower than environmental levels of PAHs in urban and rural surface soil, especially near areas of 35 vehicle traffic (e.g., roads and parking areas). Given their frequent presence in environmental media, 36 and especially in areas influenced by vehicle exhaust and tire particles, it is important to compare 37 risk-based cleanup levels with typical environmental concentrations before utilizing unrealistically 38 low cleanup targets. Numerous studies have been conducted that examine ambient levels of PAHs in 39 rural and urban surface soil (e.g., ATSDR 1995, Bradley et al. 1994, MADEP 2002, and Teaf et al. 40 2008). These studies indicate that given the multitude of non-point mobile sources for PAHs, it is not 41 uncommon for ambient concentrations to exceed health-based regulatory recommendations. Some 42 states have begun to consider ambient anthropogenic levels by establishing minimum screening levels 43 based on environmental studies. For example, the New York State Department of Environmental 1 Conservation has established a minimum soil cleanup objective of 1 mg/kg for benz(a)anthracene, 2 benzo(a)pyrene, and benzo(b)fluoranthene and 0.1 mg/kg for dibenz(a,h)anthracene, based on the 95th 3 percentile concentrations of these PAHs in rural areas near roads (NYSDEC 2006).

4

7.2.6.3 Uncertainty in the Identification of COCs

5 6

7 All of the sources of uncertainty described in the previous sections potentially impact the 8 identification of COCs. The exposure and toxicity values used to calculate FWCUGs and RSLs as 9 well as the approach for identifying SRCs, COPCs, and ultimately COCs based on the FWCUGs and 10 RSLs were designed to ensure overestimation, rather than underestimation, of potential risk. The 11 uncertainty assessment attempts to put the identified COCs in perspective to facilitate informed risk 12 management decisions for the AOC.

13

14 The SOR is used to account for the potential additive effects from exposure to multiple chemicals that 15 can cause the same effect or affect the same target organ. Cancer risk is assumed to be additive for all 16 carcinogens. Non-cancer risk is assumed to be additive for chemicals with similar sites of 17 toxicological action. In the event that any combination of COPCs results in synergistic effects, risk might be underestimated. Conversely, the assumption of additivity would overestimate risk if a 18 19 combination of COPCs acted antagonistically. It is unclear whether the potential for chemical 20 interaction has been inadvertently understated or overstated. It seems unlikely that the potential for 21 chemical interaction contributes significant uncertainty to the conclusions of the risk assessment.

- 22
- 23

7.2.7 Identification of COCs for Potential Remediation

24

25 COCs were identified in Section 7.2.5 as any COPC having an EPC greater than an applicable 26 FWCUG or contributing more than 5-10% to an SOR greater than one. For inorganic chemicals with 27 FWCUGs below background concentrations, the background concentration was used as the point of 28 comparison. The TR for the FWCUGs used to identify COCs is 1E-05 per the Ohio EPA DERR 29 program, which has adopted a human health cumulative ILCR goal of 1E-05 to be used as the level of 30 acceptable excess cancer risk and for developing site remediation goals.

31

32 The results of the COC screening (Section 7.2.5) are combined with the results of the uncertainty 33 assessment (Section 7.2.6) to identify COCs to be carried forward for potential remediation.

34

35 Several PAHs were identified as COCs in sediment. The MDC of benzo(a)pyrene (0.89 mg/kg) at the 36 Backwater Area exceeds the FWCUG (0.221 mg/kg). The MDC is at ULCPsd-015, which was 37 collected in 2001. Sample ULCPsd-047 was collected at approximately the same location in 2010 and showed a benzo(a)pyrene concentration of 0.39 mg/kg. Detected benzo(a)pyrene concentrations in the 38 39 other Backwater Area samples ranged from 0.063–0.23 mg/kg. Three additional PAHs, present below 40 FWCUGs, were identified as COCs based on the SOR analysis.

1 Benzo(a)pyrene, and by association the other PAHs, were not identified as COCs to be carried 2 forward for potential remediation based on the following WOE factors:

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- 4
- 5 6

7

• Samples ULCPsd-015 and ULCPsd-047 are located near a culvert between Load Line 3 Road and Track 33. The source of the PAHs in sediment is most likely runoff from the road and railroad. Samples collected upgradient (in the Backwater area) and downgradient (in Upper Cobbs Pond) show lower concentrations.

- The FWCUGs for exposure to sediment are the same as the FWCUGs for soil (i.e., exposure is assumed to be the same for soil in a residential yard and sediment in a pond). Using more realistic (but still conservative) exposure assumptions (e.g., exposure only occurs during the warmer months of the year) results in a screening level of 9 mg/kg. The MDC at the backwater area (0.89 mg/kg) is below this still conservative value.
- 13

The EPC of arsenic in sediment at Lower Cobbs Pond (26.9 mg/kg) exceeds the FWCUG for the Resident Receptor Adult of 4.25 mg/kg, the facility-wide background concentration of 19.5 mg/kg, and the Ohio EPA SRV of 25 mg/kg. The MDC of arsenic (34.3 mg/kg) at ULCPsd-026, collected in 2001, is the only sample concentration greater than the facility-wide background concentration and Ohio EPA SRV. Reported arsenic concentrations in the other Lower Cobbs Pond samples ranged from 5.1–20.2 mg/kg. Arsenic was not identified as a COC for potential remediation because the reported concentrations appear to represent naturally occurring levels in sediment.

21

No FWCUG is available for cobalt in surface water; the screening value used for the Resident (Adult and Child) is the tap water RSL of 0.006 mg/L. The EPC of cobalt (0.01 mg/L) is less than two times the tap water RSL. The Backwater Area is a shallow pond created through beaver dam construction activity upstream of Upper Cobbs Pond. This limited shallow surface water is not a potential source of residential drinking water; therefore, this low exceedance of the tap water RSL does not warrant potential remediation in this EU.

28

29 The EPC of manganese in surface water at the Backwater Area EU (15.1 mg/L) exceeds the FWCUG 30 for the Resident Receptor (Adult and Child) (6.326 mg/L). The EPC is strongly influenced by the 31 MDC (15.8 mg/L) reported in sample ULCPsw-002 collected in 2001. The field duplicate of this 32 sample reported a concentration of 7.36 mg/L. The manganese concentrations in all samples collected 33 in 2010 are below the facility-wide background concentration of 0.391 mg/L. Because the field 34 duplicate of the MDC concentration is much closer to the Resident Receptor FWCUG and recent 35 sample data indicate manganese is present at naturally occurring background concentrations, 36 manganese does not warrant potential remediation in this EU.

37

38 7.2.8 Summary of HHRA

39

This HHRA documents COCs that may pose potential health risk to human receptors resulting from exposure to contamination at the Upper and Lower Cobbs Ponds AOC. This HHRA was conducted as part of the RI and was based on the streamlined approach described in the FWCUG Report (USACE 2010a), FWCUG Position Paper (USACE 2012a), and Technical Memorandum (ARNG 2014). The 1 components of the risk assessment (i.e., receptors, exposure media, EPCs, and results) are 2 summarized below.

3

Receptors. Camp Ravenna is a controlled-access facility. The Upper and Lower Cobbs Ponds AOC is located in the east-central portion of the facility. The Upper Cobbs Pond and Lower Cobbs Pond are used for recreational purposes. In addition, both ponds are used by OHARNG for fishing. Security patrols and maintenance activities also occur at the AOC. Future OHARNG Land Use for the AOC (which includes ponds and associated sediment and banks) includes recreational use, dam maintenance, wildlife management activities, and training purposes. Surrounding adjacent areas may be used for recreational activities as well as military training operations.

11

Three Land Uses for the RVAAP restoration program are specified in the Technical Memorandum (ARNG 2014) for consideration in the RI along with their Representative Receptors. Unrestricted (Residential) Land Use [Resident Receptor (Adult and Child)] is considered protective for all three Land Uses at Camp Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then the AOC is also considered to have met the requirements of the other Land Uses (i.e., Commercial/Industrial and Military Training).

18

Exposure Media. Media of concern at the Upper and Lower Cobbs Ponds AOC are surface soil,
 subsurface soil, surface water, and sediment.

21

Estimation of EPCs. The EPCs were calculated from the results of all of the discrete samples collected from a given depth interval at each EU. The EPC was either the 95% UCL of the mean or the MDC, whichever value is lowest.

25

The EPCs for soil, sediment, and surface water were calculated from the results of all of the discrete samples collected at each EU. The EPC was either the 95% UCL of the mean or the MDC, whichever value is lowest. In addition to the discrete data, the results of sediment ISM data were evaluated separately with each ISM sample representing a decision unit.

30

Results of Human Health Risk Assessment. No COCs were identified for potential remediation for
 soil, sediment, or surface water at the Upper and Lower Cobbs Ponds AOC.

33

34 **7.3 ECOLOGICAL RISK ASSESSMENT**

35

36 7.3.1 Introduction

37

The ERA presented in this Phase III RI Report follows a unified approach of methods integrating Army, Ohio EPA, and USEPA guidance. This ERA approach is consistent with the general approach by these agencies and primarily follows the Level I Scoping ERA, Level II Screening ERA, and Level III Baseline ERA outlined in the *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA 2008), with specific application of components from the FWERWP, *Risk Assessment Handbook*

43 Volume II: Environmental Evaluation (USACE 2010b), and Ecological Risk Assessment Guidance

| 1 | for Superfund: Process for Designing and Conducting Ecological Risk Assessments (USEPA |
|----------|--|
| 2 | 1997). The process implemented in this RI Report combines these guidance documents to meet |
| 3 | requirements of the Ohio EPA and Army, while following previously accepted methods established |
| 4 | for RVAAP. This unified approach resulted from coordination between USACE and Ohio EPA |
| 5 | during the summer of 2011. |
| 6 | |
| 7 | 7.3.1.1 <u>Scope and Objective</u> |
| 8 | |
| 9 | The Upper and Lower Cobbs Ponds AOC contains habitats that support ecological receptors. These |
| 10 | terrestrial and aquatic habitats have known chemical contamination (MKM 2005). Habitat types and |
| 11 | an assessment of the ecological resources found at the AOC are presented in subsequent subsections. |
| 12 | |
| 13 | Additionally, the conclusions of a historical ERA (a SERA) performed as part of the Phase II |
| 14 | Remedial Investigation Report for Upper and Lower Cobbs Ponds (MKM 2005)] state the following: |
| 15 | |
| 16 | "No additional action is recommended based on ecological risk because, based on the |
| 17 | Step 3a refinement step, it was determined that site-related risks were not great |
| 18 | enough to warrant proceeding further into the ecological risk assessment process." |
| 19 | |
| 20 | The results of the historical ERA and PBA08 RI are combined in this report to evaluate the |
| 21 | conclusions of the historical ERA and determine whether a qualitative ERA (Level I) is sufficient, |
| 22 | based on the quality of the habitat and the presence of contamination, or a more rigorous ERA (Level |
| 23 | II or III) will be conducted. |
| 24 | |
| 25 | 7.3.2 Level I: Scoping Level Ecological Risk Assessment |
| 26 | |
| 27 | The ERA method for Level I follows guidance documents listed in Section 7.3.1. Level I is intended |
| 28 | to evaluate if the AOC had past releases, has the potential for current contamination, and has |
| 29 | important ecological resources on or near the AOC. |
| 30 | |
| 31 | The following two questions should be answered when the Level I ERA is complete: |
| 32 | 1 Are comment on most valescent events at the AOC? Comment on most valescent on |
| 33 | 1. Are current or past releases suspected at the AOC? Current or past releases are |
| 34 25 | determined by evidence that chemical contaminants or COPECs are present. |
| 35 | 2. Are important ecological resources present at or in the locality of the AOC? Important |
| 36 27 | ecological resources are defined in the <i>Guidance for Conducting Ecological Risk Assessments</i> |
| 37 | (Ohio EPA 2008) and Technical Document for Ecological Risk Assessment: Process for |
| 38 | Developing Management Goals (BTAG 2005). |
| 39 40 | If an AOC has contaminants but lacks important ecological resources, the ERA process can stop at |
| 40 41 | Level I. Contamination and important ecological resources must both be present to proceed to a Level |
| 41 42 | II Screening Level ERA. |
| 42 43 | |
| ъJ | |

1 7.3.2.1 **AOC Description and Land Use**

2

3 The Upper and Lower Cobbs Ponds AOC is approximately 39 acres. The AOC includes Upper Cobbs

4 Pond (9.4 acres), Lower Cobbs Pond (6.4 acres), the Backwater Area (5.2 acres), and Pond Bank (18 5 acres, including a portion of Wetland 1). Water flows from the Backwater Area to Upper Cobbs Pond and then to Lower Cobbs Pond. The surrounding soil is large enough to support cover and food for 6 7 riparian birds and mammals that typically inhabit areas of about 1 acre (USEPA 1993). The combined 8 area of 39 acres would represent a small portion of the home range of other wildlife such as duck, 9 wild turkey, and deer. Another way of understanding the habitat is the length of the pond perimeter 10 along which riparian animals feed. The perimeter lengths of Upper Cobbs Pond (3,360 ft), Lower Cobbs Pond (3,630 ft), and Backwater Area (8,870 ft) total 15,860 ft of linear habitat for herons and 11 other aquatic birds and mammals. The 39 acres of habitat at the Upper and Lower Cobbs Ponds AOC 12 13 represents 0.18% of the 21,683 acres at Camp Ravenna.

14

15 The Upper and Lower Cobbs Ponds AOC is currently utilized for recreational purposes, as it contains abundant fish and wildlife. OHARNG uses the ponds for fishing. A picnic area, pavilion, and 16 17 playground are located approximately 100 ft from the edge of the eastern bank of Upper Cobbs Pond.

- 18
- 19 20

Evidence of Historical Chemical Contamination 7.3.2.2

21 The 1978 Installation Assessment identified the major contaminants of the former RVAAP to be 22 TNT, composition B, sulfates, nitrates, lead styphnate, and lead azide. The assessment identified the 23 pink wastewater and washout of residue, dusts, and spills at Load Lines 3 and 12, which were 24 discharged through surface drainage channels towards Upper Cobbs Pond (USATHAMA 1978), as 25 contaminants. Potential site-specific contaminants include TNT, RDX, HMX, nitrate, nitrocellulose, and heavy metals (i.e., lead, chromium, mercury, and arsenic) from the pink water. Other 26 27 contaminants identified include washout from explosives melt-out and demilitarization activities at 28 Load Lines 3 and 12 and ammonium nitrate and aluminum chloride from Load Line 12.

29

30 The goal of the historical ERA (MKM 2005) was to identify COPECs in surface soil, sediment, and 31 surface water for the Upper and Lower Cobbs Ponds AOC. The historical ERA followed instructions 32 presented in the Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2003) and consisted of the first two of six steps listed in Figure III of the FWERWP (USACE 2003a). These two 33 34 steps identified the evaluation procedures, which were used to determine AOC-related COPECs. 35 First, the MDC of each chemical was compared to its respective facility-wide background 36 concentration. Chemicals were not considered COPECs if the MDC was below the background 37 concentration. For all chemicals detected above background concentrations, the MDC was compared 38 to an ESV. The hierarchy of screening values was based on the guidance included in the FWERWP 39 and Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2003). In addition to the ESV 40 comparison, it was determined if the chemical was a persistent, bioaccumulative, and toxic (PBT) 41 compound. For soil, in addition to the PBT determination, chemicals that were bioaccumulative 42 (whether or not they were persistent and toxic as well) were also identified. For sediment, chemicals 43 were also compared with the Ohio SRV. Chemicals were retained as COPECs if they exceeded

| 1 | background concentrations and the ESV, if the chemical exceeded background concentrations and |
|----------|---|
| 2 | had no toxicity information, or if the chemical was a bioaccumulative or PBT compound. |
| 3 | |
| 4 5 | Groundwater was not included in the historical ERA. As explained in Section 3.2.2 of the FWERWP, groundwater is not considered an exposure medium to ecological receptors because these receptors |
| 6 | are unlikely to contact groundwater greater than 5 ft bgs. As discussed in Section 3.4.2, groundwater |
| 7 | varies from 5.75–39.75 ft bgs throughout the AOC. |
| 8 | |
| 9 | The historical ERA tables for soil, sediment, and surface water are included in Appendix H, Tables |
| 10 | H-1 to H-10. The screening tables (Appendix H, Tables H-1 to H-3) contain the following: |
| 11 | |
| 12 | • Frequency of detection, |
| 13 | • Average concentration, |
| 14 | • MDC, |
| 15 | • Selected UCL (soil and sediment only), |
| 16 | • Background concentration for each medium (surface soil, sediment, and surface water), |
| 17 | • Comparison of MDC to background concentrations (SRC determination), |
| 18 | • SRVs (sediment only), |
| 19 | • Comparison of MDC to SRVs (sediment only), |
| 20 | • Screening levels (surface soil, sediment, and surface water), |
| 21 | • Comparison of MDC to screening values (HQ, MDC/screening level ratio), |
| 22 | • Bioaccumulative chemical (soil only), |
| 23 | PBT compound identification, |
| 24 | • COPEC determination, and |
| 25 | COPEC rationale. |
| 26 | |
| 27 | Historical COPECs for Soil |
| 28 | |
| 29 | The historical ERA conducted as part of the Phase II RI reported 26 chemicals detected in surface soil |
| 30 | (0–1 ft bgs) along the banks of the Upper Cobbs Pond and Lower Cobbs Pond (MKM 2005). A total |
| 31 | of 25 chemicals either exceeded background concentrations or did not have an associated background |
| 32 | concentration for comparison. Four chemicals (calcium, magnesium, potassium, and sodium) were |
| 33 | essential nutrients and were excluded from the COPEC screen. A total of 11 inorganic chemicals |
| 34 25 | (aluminum, arsenic, cadmium, chromium, copper, iron, mercury, nickel, silver, vanadium, and zinc) |
| 35 | were identified as COPECs because detected concentrations were above ESVs (Table 7-11). Mercury |
| 36 37 | is also a PBT compound. Two other chemicals (nitrocellulose and nitrate) were also identified as COPECs because there were no associated background concentrations or ESVs for comparison. In |
| 38 | total, 13 chemicals were identified as COPECs in soil (Table 7-11). Appendix H, Table H-1 presents |
| 39 | the historical ecological screening for soil at Upper and Lower Cobbs Ponds AOC. |

1 Historical COPECs for Sediment

2

3 The historical ERA conducted as part of the Phase II RI reported 45 chemicals in sediment (MKM 4 2005). A total of 33 chemicals either exceeded background concentrations and SRVs or did not have 5 an associated background concentration for comparison. Four chemicals (calcium, magnesium, potassium, and sodium) were essential nutrients and were excluded from the COPEC screen. A total 6 7 of 8 inorganic chemicals (arsenic, cadmium, chromium, copper, cyanide, lead, nickel, and silver), 11 8 organic chemicals [2,6-DNT; anthracene; benz(a)anthracene; benzo(a)pyrene; benzo(ghi)perylene; 9 benzo(k)fluoranthene; chrysene; fluoranthene; indeno(1,2,3-cd)pyrene; phenanthrene; and pyrene], 10 and total PAHs were identified as COPECs because detected concentrations were above ESVs. Two 11 chemicals (mercury and aroclor-1254) are below ESVs, but they are PBT compounds. Five other chemicals (antimony, beryllium, TNT, nitrocellulose, and nitrate) were also identified as COPECs 12

due to a lack of ESV. In total, 27 COPECs were identified in sediment (Table 7-11). Appendix H,
Table H-2 presents the historical ecological screening for sediment at the Upper and Lower Cobbs
Ponds AOC.

16

17 Historical COPECs for Surface Water

18

19 The historical ERA conducted as part of the Phase II RI reported 22 chemicals in surface water 20 (MKM 2005). A total of 20 chemicals either exceeded background concentrations or did not have an 21 associated background concentration for comparison. Four chemicals (calcium, magnesium, 22 potassium, and sodium) were essential nutrients and were excluded from the COPEC screen. Four 23 inorganic chemicals (iron, selenium, silver, and sulfide) were identified as COPECs because their 24 detected concentrations were above ESVs. Four other chemicals (acetone, chloromethane, 25 manganese, and sulfate) were also identified as COPECs due to a lack of ESVs. In total, eight 26 chemicals were identified as COPECs in surface water (Table 7-11). Appendix H, Table H-3 presents 27 the historical ecological screening for surface water at the Upper and Lower Cobbs Ponds AOC.

28

30

29 Summary of Historical ERA

As explained previously, a historical SERA was performed to determine COPECs in surface soil, sediment, and surface water at the AOC. The COPECs are summarized in Table 7-11. Based on the identified historical COPECs, and given additional data for current (2010) samples and updated ESVs, ecological risk in surface soil, sediment, and surface water is re-evaluated in this report.

35

36 7.3.2.3 Ecological Significance

37

38 Sources of data and information about the ecological resources at the AOC include the *Integrated* 39 *Natural Resource Management Plan* (OHARNG 2014) (herein referred to as the INRMP), FWBWQS

40 (USACE 2005a), previous characterization work (e.g., Phase I and Phase II RIs), and visits to the

- 41 Upper and Lower Cobbs Ponds AOC conducted for the PBA08 RI.
- 42

One of the two key questions to answer in the Level I Scoping ERA is whether there are ecologically important and especially ecologically significant resources at the Upper and Lower Cobbs Ponds AOC. Ecological importance is defined as a place or resource that exhibits unique, special, or other attributes that makes it of great value. Ecological significance is defined as an important resource found at an AOC or in its vicinity that is subject to contaminant exposure. The underlying basis for this distinction can be found in *Ecological Significance and Selection of Candidate Assessment Endpoints* (USEPA 1996a), and is stated as follows:

- 8
- 9 10
- 10

"A critical element in the ERA process requires distinguishing important environmental responses to chemical releases from those that are inconsequential to the ecosystem in which the site resides: in other words, determining the ecological significance of past, current, or projected site-related effects."

12 13

14 Important places and resources identified by the Army and Ohio EPA (Appendix H, Table H-11) 15 include wetlands, terrestrial areas used for breeding by large or dense aggregations of animals, habitat 16 known to be used by threatened or endangered species, state land designated for wildlife or game 17 management, locally important ecological places, and state parks. The Army and Ohio EPA recognize 18 17 important places and resources. The Army recognizes an additional 16 important places (BTAG 19 2005), and the Ohio EPA recognizes another 6 important places (Ohio EPA 2008). In total, there are 20 39 important places. Presence or absence of an ecologically important place can be determined by 21 comparing environmental facts and characteristics of the AOC with each of the important places and 22 resources listed in Appendix H, Table H-11.

23

24 Presence of an important ecological resource or place and proximity to contamination at an AOC 25 makes a resource ecologically significant. Thus, any important places and resources listed in 26 Appendix H, Table H-11 are elevated to ecologically significant when present on the AOC and there 27 is exposure to contaminants. For all 39 important places and resources, it is relatively clear the 28 ecological place or resource is present or absent on the AOC; therefore, the decision process is 29 objective. If no important or significant resource is present at an AOC, the evaluation will not proceed 30 to Level II regardless of the presence of contamination. Instead, the Level I Scoping ERA would 31 acknowledge that there are important ecological places but that those resources are not ecologically 32 significant, and no further evaluation is required.

33

34 Management Goals for the AOC

35

Regardless of whether the evaluation is concluded at Level I or continues to Level II, there is another level of environmental protection for the AOC through the natural resource management goals expressed in the INRMP (OHARNG 2014). OHARNG manages the ecological and natural resources at Camp Ravenna to maintain or enhance the current integrity of the natural resources and ecosystems at the facility. Natural resource monitoring and management activities in place at Camp Ravenna may also be applicable to any degradation noted from contamination.

1 Some natural resources management goals of OHARNG (listed in Appendix H, Table H-12) benefit 2 the AOC. For example, Goal 1 states the natural resources need to be managed in a compatible way 3 with the military mission, and Goal 5 requires the Army to sustain usable training lands and native 4 natural resources by implementing a natural resource management plan that incorporates invasive 5 species management and by utilizing native species mixes for revegetation after ground disturbance activities. These management goals help detect degradation (whether from training activities or 6 7 historical contamination). While the applicability of the remaining 10 management goals to the Upper 8 and Lower Cobbs Ponds AOC varies, all of the management goals are intended to monitor, maintain, 9 or enhance the facility's natural resources and ecosystem. While these goals are for managing all 10 types of resources at and near the AOC, they do not affect decisions concerning the presence or absence of important or significant ecological places or resources at the AOC. 11

12

13 Important Places and Resources

14

15 Ecological importance means a place or resource that exhibits a unique, special, or other attribute that 16 makes it of great value. Examples of important places and resources include wetlands, terrestrial areas 17 used for breeding by large or dense aggregations of animals, and habitat of state- or federally-listed species. An important resource becomes significant when found on an AOC and there is contaminant 18 19 exposure. The ponds and wetlands are important/significant ecological resources at the AOC. In 20 addition, the Least Bittern (Ixobrychus exilis), which is a state of Ohio threatened species, uses the 21 AOC and nests or resides at Camp Ravenna; the Least Bittern (Ixobrychus exilis) and its habitat at the 22 AOC are important/significant ecological resources.

23

24 Terrestrial Resources

25

The habitat area for the Upper and Lower Cobbs Ponds AOC contains terrestrial and aquatic resources in approximately equal amounts (Figure 7-1). There are approximately 18 acres of terrestrial resources surrounding the ponds, Backwater Area, and other wetlands identified in Figure 7-1. A small portion of the 18 acres includes Wetland 1. The terrestrial habitat is described below.

30

31 Habitat Descriptions and Species. SAIC scientists conducted a field survey at the AOC in May 2010 32 and confirmed that terrestrial vegetation habitat surrounding the ponds consists of four forest 33 community types, one shrub community type, and three herbaceous community types (Figure 7-1). 34 The dominant forest community at the AOC is the red maple (Acer rubrum) successional forest. This 35 community surrounds nearly all of Lower Cobbs Pond and a large portion of Upper Cobbs Pond. 36 Other forest communities present include a small area of green ash (Fraxinus pennsylvanica), 37 American elm (Ulmus americana), and hackberry (Celtis occidentalis and laevigata) temporarily 38 flooded forest alliance at the extreme northern end of Lower Cobbs Pond. At Upper Cobbs Pond, 39 other forest communities include small areas of mixed, needle-leaved evergreen, cold-deciduous 40 forest dominated by Norway spruce (Picea abies) and red maple and mixed, cold-deciduous, 41 hardwood forest dominated by white ash (Fraxinus americana), wild black cherry (Prunus serotina), 42 red maple, black locust (Robinia pseudoacacia), quaking aspen (Populus tremuloides), and bigtooth 43 aspen (Populus grandidentata).

Upper and Lower Cobbs Ponds

1 Dry, mid-successional, cold-deciduous shrubland habitat (SU1) occurs on both sides of Upper Cobbs

2 Pond and includes various willows (Salix spp.), gray dogwood (Cornus racemosa), autumn olive

3 (Elaeagnus umbellata), blackberry (Rubus allegheniensis), hawthorn (Crataegus spp.), and multiflora

- 4 rose (*Rosa multiflora*).
- 5

6 Herbaceous communities at the AOC include dry, early-successional field; intermittently flooded, 7 early-successional field; and permanently flooded, herbaceous alliance. Dry, early-successional field 8 is an upland habitat largely maintained by routine periodic disturbance (e.g., mowing). Dominant 9 plants include an assortment of grasses, forbs, and seedlings of trees and shrubs. Common species 10 include several species of goldenrod (Solidago spp.), clasping-leaf dogbane (Apocynum cannabinum), 11 varrow (Achillea millefolium), gray dogwood (Cornus racemosa), blackberry (Rubus allegheniensis), 12 autumn olive (Elaeagnus umbellata), and multiflora rose (Rosa multiflora). The intermittently 13 flooded, early-successional field habitat is dominated by nearly pure stands of narrow-leaf cattail 14 (Typha angustifolia) and broad-leaf (Typha latifolia) cattail, with various sedges (Carex spp.), 15 bulrushes (*Scirpus* spp.), and broad-leaf wetland plants as associate species. The permanently flooded 16 herbaceous alliance is dominated by submerged aquatic plants such as spatterdock (Nuphar lutea) and 17 white water lily (Nymphaea odorata).

18

19 Terrestrial habitat at the AOC includes a diverse mixture of forest, shrubland, and herbaceous 20 communities. The habitats provide cover and food for numerous birds and mammals that typically 21 inhabit areas of about 1 acre (USEPA 1993) as well as animals with much larger home ranges, such 22 as wild turkey (*Meleagris gallopavo*) and white-tailed deer (*Odocoileus virginianus*).

23

Based on August 2008 observations (Photographs 7-1 and 7-2), SAIC scientists assessed the terrestrial habitat at the ponds to be healthy and functioning. Functional habitat was determined by noting the absence of large bare spots and dead vegetation or other obvious visual signs of an unhealthy ecosystem. Additional habitat photographs are provided in Appendix H.

28

29 Threatened, Endangered, and Other State- or Federally-Listed Species. The northern long-eared bat 30 (Myotis septentrionalis; federally threatened) exists at Camp Ravenna. There are no other federally 31 listed wildlife species and no critical habitat on Camp Ravenna. The Upper and Lower Cobbs Ponds 32 AOC has not been previously surveyed for rare, threatened, or endangered species; however, there 33 has been one sighting of a state-threatened species at the AOC—the Least Bittern (*Ixobrychus exilis*) 34 (OHARNG 2014).

35

36 **Other Terrestrial Resources.** While there are no other known important terrestrial places and 37 resources, there are other resources at or near the AOC (e.g., vegetation, animals) that interact in their 38 ecosystems and support nutrient cycling and energy flow. For example, wildlife such as wild turkey 39 and white-tailed deer could use the area. The INRMP provides information about species and habitat 40 surveys at Camp Ravenna (e.g., timber and ecological succession) (OHARNG 2014). There are no 41 other reported surveys of habitats and wildlife at the AOC beyond those summarized in the INRMP 42 (OHARNG 2014).

1 Aquatic Resources

2 3

Aquatic habitats make up approximately 50% of the AOC. There are two types of aquatic resources at the Upper and Lower Cobbs Ponds AOC: ponds and wetlands.

4 5

Habitat Descriptions and Species. Ponds are the dominant aquatic habitat at the AOC (Figure 7-2).
Upper Cobbs Pond (9.4 acres; Photograph 7-1), Lower Cobbs Pond (6.4 acres; Photograph 7-2), and
the Backwater Area (5.2 acres) comprise 21 acres. The ponds support a variety of aquatic animals
(e.g., fish, turtles, frogs) and semi-aquatic wildlife, such as waterfowl (e.g., ducks and geese) and
wading birds (e.g., great blue heron).

11

Wetlands. Wetlands are important resources with water-saturated soil or sediment and plant life that can survive saturation. Wetlands are home to many different species and are also chemical sinks that can serve as detoxifiers and natural water purifiers. It is expected the wetlands at the AOC perform these and other related functions.

16

17 There are planning level survey [i.e., based on desktop surveys conducted for OHARNG of wetland data and resources (e.g., National Wetland Inventory maps, aerials)] wetlands and wetland complexes 18 19 at the AOC (Figure 7-2). These wetlands include a small area around the Lower Cobbs Pond outlet 20 (Wetland 1), fringe areas around the shoreline, shallow coves and embayments, and floodplains of 21 larger tributaries associated with the Backwater Area (Wetlands 2 and 3). No jurisdictional wetlands 22 determination has been conducted at this AOC. For any wetland at the AOC potentially affected by 23 remedial activities, a jurisdictional determination by USACE would be required to determine the 24 regulatory status.

25



26 27

Photograph 7-1. Upper Cobbs Pond on August 12, 2008 (looking south from the access road and dam between the two ponds)



Photograph 7-2. Lower Cobbs Pond on August 12, 2008 (looking northeast from the access road and dam between the two ponds)

5 An SAIC Professional Wetland Scientist used the Ohio Rapid Assessment Method (ORAM) (Ohio 6 EPA 2001) in May 2010 to assess the condition of the wetland complex discussed above and to 7 determine its potential ecological importance (Appendix H, Figures H-1 through H-3).

- 9 Using the ORAM, wetlands are classified into three categories:
- 10

15

16

8

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- Category 1 wetlands are described as "limited quality waters." They are considered to be a
 resource that has been degraded, has limited potential for restoration, or is of such low
 functionality that lower standards for avoidance, minimization, and mitigation can be applied.
 Scores range from 1–29.
 - Category 2 includes wetlands of moderate quality and wetlands that are degraded but exhibit reasonable potential for restoration. Scores range from 30–59.
- Category 3 includes wetlands of very high quality and wetlands of concern regionally and/or
 statewide, such as wetlands that provide habitat for species listed as threatened or
 endangered. Scores range from 60–100.
- 20

21 The field sheets detailing the ORAM evaluations at the Upper and Lower Cobbs Ponds AOC are 22 presented in Appendix H, Figures H-1 through H-3. Figure 7-2 shows the locations of the wetlands at 23 the AOC. Wetlands were consolidated into the following three groups for ORAM evaluations (Table 24 7-12). Wetland 1 is a small, forested wetland associated with the drainage channel below the Lower 25 Cobbs Pond dam on the west side of Paris-Windham Road; its total size is 0.34 acres, with 0.17 acres 26 inside the AOC. Wetland 2 is a wetland complex consisting of several small- and medium-sized 27 wetland habitat types in the northern half of the Backwater Area between Track 33 and Track FA; its 28 total size is 4.75 acres, with 3.60 acres inside the AOC. Wetland 3 is a wetland complex consisting of 29 several small- and medium-sized wetland habitat types in the southern half of the Backwater Area,

1 south of Track FA; its total size is 4.62 acres, with 3.90 acres inside the AOC. Wetland 3 includes 2

portions of the drainage ditches from Load Lines 3 and 12.

3

4 Threatened and Endangered and Other State- or Federally-listed Species. The northern long-eared 5 bat (Myotis septentrionalis; federally threatened) exists at Camp Ravenna. There are no other federally listed wildlife species and no critical habitat on Camp Ravenna. The Upper and Lower 6 7 Cobbs Ponds AOC has not been previously surveyed for rare, threatened, or endangered species; 8 however, there has been one sighting of a state-threatened species at the AOC-the Least Bittern 9 (Ixobrychus exilis) (OHARNG 2014).

10

11 Other Aquatic Resources. There are no other known aquatic resources (e.g., vegetation, animals) at 12 or near the AOC (Appendix H, Table H-11). There are no other reported surveys of aquatic habitats 13 and wildlife at the AOC beyond those summarized in the INRMP (OHARNG 2014). There are three 14 nearby biological and water quality stations. The following subsections provide a summary of the 15 biological/water quality stations in the vicinity of the Upper and Lower Cobbs Ponds AOC.

16

17 Biological/Water Quality Sampling Stations. Ohio EPA and USACE investigated 11 ponds and 18 several streams at Camp Ravenna using a network of biological/water quality sampling stations 19 (USACE 2005a). The purpose of the investigation was to document ecological effects of 20 contamination at AOCs on stream or pond biota and conditions. One stream sampling location (S-8) 21 is in the vicinity of the AOC. Sampling Station S-8 (River Mile 0.1) is located on an unnamed 22 tributary downgradient and northwest of Lower Cobbs Pond (Figure 7-1) and provides information 23 about potential contamination from the Upper and Lower Cobbs Ponds AOC and other upgradient 24 AOCs (e.g., Load Lines 3 and 12). The station is located about 350 ft upstream from the confluence 25 with Sand Creek. If the S-8 sampling station has a positive rating (e.g., good, excellent, full 26 attainment, and other positive terms reported in the study), it means that the Upper and Lower Cobbs 27 Ponds AOC and other upstream AOCs are not adversely impacting the quality of Sand Creek. In 28 addition, both Upper Cobbs Pond and Lower Cobbs Pond were sampled under this program.

29

30 According to the FWBWQS (USACE 2005a), each sampling station included a sediment 31 sampling/assessment, surface water sampling/assessment, fish and macroinvertebrate community 32 assessment, and habitat assessment. The sampling reach for stream sampling stations ranged 120-210 33 meters.

34

35 Sediment evaluations were conducted in June 2003 using guidelines established in the Development 36 and Evaluation of Consensus-based Sediment Ouality Guidelines for Freshwater Ecosystems 37 (MacDonald et al. 2000), SRVs for inorganic chemicals (Ohio EPA 2003), and USEPA Region 5 38 ecological screening levels (ESLs) (USEPA 2003b). Sediment samples were analyzed for SVOCs, 39 pesticides, PCBs, TAL metals, explosives, percent solids, cyanide, ammonia, nitrate, and phosphorus. 40 Surface water grab samples collected in June, August, and September 2003 were evaluated by being 41 compared to Ohio WQS criteria, reference conditions, or those conditions cited in the FWBWQS 42 (USACE 2005a). Surface water samples were analyzed for TAL metals, pesticides, PCBs, explosives,

43 SVOCs, and several nutrients.

1 The fish and macroinvertebrate sampling and assessment at stream locations occurred in August and 2 September 2003, while pond sampling occurred between June and August 2003. Fish were sampled 3 using electrofishing methods. Macroinvertebrate communities were assessed using artificial 4 substrates (quantitative sampling) supplemented with a composite natural substrate sample 5 (qualitative sampling). Both the fish and macroinvertebrate community assessments followed the methods contained in the Biological Criteria for the Protection of Aquatic Life: Volume III, 6 7 Standardized Biological Field Sampling and Laboratory Methods for Assessing Fish and 8 Macroinvertebrate Communities (Ohio EPA 1989). Funnel traps designed to collect any amphibians, 9 fish, or macroinvertebrates that swim or crawl into the funnel openings were used to assess 10 macroinvertebrate communities in the ponds (Ohio EPA 2002b). Laboratory analysis of the funnel trap macroinvertebrate samples followed standardized Ohio EPA procedures (Ohio EPA 1989). 11 12

- 13 Community Similarity Index (CSI) values were used to determine impairment of the fish and 14 macroinvertebrate communities contained in the ponds. For fish, CSI values were calculated based on 15 fish relative abundance and relative biomass, and a pond-by-pond comparison was made. Eight ponds 16 at Camp Ravenna were sampled, with three of them being reference ponds. For macroinvertebrates, 17 CSI values were calculated based on the presence of macroinvertebrate taxa in the artificial 18 substrate/qualitative samples (Hester Dendy sampling) and abundance of macroinvertebrates in 19 funnel trap samples. Using the CSI method, ponds with identical composition would have a value of 1 20 and those with no similarity at all would have a value of 0. Comparisons with values of 0.65 or 21 greater were regarded as having strong resemblances, a criterion used previously by Hanson (1955), 22 Beckett (1978), and Ohio EPA (1981).
- 23

24 The physical habitat assessment for stream locations was conducted in June 2003 and used the 25 Qualitative Habitat Evaluation Index (QHEI) developed by Ohio EPA (Rankin 1989, 1995). The 26 types(s) and quality of substrates; amount and quality of instream cover; channel morphology; and 27 extent and quality of riparian vegetation, pool, run, riffle development and quality, and gradient are 28 some of the habitat characteristics used to determine the QHEI score. Pond habitat assessments were 29 conducted between June and August 2003 and used the Lake Erie shoreline Qualitative Habitat 30 Evaluation Index (L-OHEI) developed by the Ohio EPA (Ohio EPA 2002a). The types(s) and quality 31 of substrates, amount and quality of cover types, shoreline morphology, extent and quality of riparian 32 vegetation, and aquatic vegetation are habitat characteristics used to determine the L-QHEI score.

33

Summary of Station S-8 Sampling/Assessment Results. Data from S-8 (USACE 2005a) showed many positive attribute ratings (e.g., good, excellent, full attainment) and no sign of aquatic impairment, and the station was rated at Full Use Attainment Status, which indicates that all indices met Ohio EPA biological criteria.

38

At S-8, all inorganic chemicals tested in sediment were below Ohio SRVs and TEC levels. All tested explosives and PCBs were not detected in sediment samples from S-8. The few detected pesticides and SVOCs were measured at low levels, with all concentrations below TEC or ESL guidelines. Ammonia and total phosphorus levels were measured below screening guidelines. None of the surface water chemical concentrations at S-8 exceeded Ohio WQS aquatic life maximum or average water quality criteria (WQC), and none of the chemicals measured exceeded criteria protective of the warmwater habitat aquatic life use (USACE 2005a). Overall, the sediment quality and water quality at S-8 were rated "excellent." The fish community at S-8 was rated "very good." The index of biotic integrity score was 48, and nine species were reported. The macroinvertebrate community at S-8 was rated "good." The QHEI score was 61.0, indicating a "good" stream habitat capable of supporting warmwater habitat biological communities.

7

8 Summary of Pond Sampling/Assessment Results. Review of the data from the ponds (USACE
9 2005a) indicates the biological communities of the two ponds are not currently affected by activities
10 that occurred at Load Line 12 and other nearby AOCs.

11

12 At Upper Cobbs Pond, all inorganic chemicals tested in sediment were near or below Ohio SRVs. 13 Silver, bis(2-ethylhexyl)phthalate, and di-n-butyl phthalate were measured above ESLs, and zinc was above the TEC. All tested explosives, pesticides, and PCBs were not detected in sediment samples 14 15 collected from Upper Cobbs Pond. Ammonia and total phosphorus levels were measured below 16 screening guidelines. One explosive compound (4-amino-2,6-DNT) was detected in the surface water 17 of Upper Cobbs Pond, but the measurement was below the Ohio WQS aquatic life maximum and average WOC. Lead and pH were the only parameters that were not within the aquatic life outside 18 19 mixing zone average (OMZA) criteria (USACE 2005a). Overall, the sediment quality and water 20 quality in Upper Cobbs Pond was of sufficient quality to not adversely impact the biological 21 community (USACE 2005a).

22

23 Fish abundance similarity index values revealed a strong resemblance (CSI greater than or equal to 24 0.65) of Upper Cobbs Pond to Lower Cobbs Pond, Fuze and Booster Ponds, and two reference ponds 25 (Frank's Pond and Route 80 Trout Pond located along Hinkley Creek). Stronger similarities were observed based on fish biomass. Similarity index values based on abundance of macroinvertebrates in 26 27 funnel trap samples were highly variable, and it is difficult to make generalizations about the 28 similarity of Upper Cobbs Pond to the other ponds due to the dominance of a few macroinvertebrate 29 taxa. Similarity index values based on the presence of macroinvertebrate species at an AOC were not 30 as variable as the abundance data. Upper Cobbs Pond was most similar to Lower Cobbs Pond and the 31 Load Line 4 Pond. The most similar reference pond for macroinvertebrates was Franks Pond (Hester 32 Dendy CSI of 0.57).

33

The physical habitat was also evaluated at Upper Cobbs Pond, and the L-QHEI score was 43, indicating "poor" habitat quality. This was primarily due to excessive amounts of undesirable species (or low amounts of preferred species) and bottom substrates dominated by muck.

37

At Lower Cobbs Pond, three inorganic chemicals (cadmium, copper, and zinc) were above the Ohio
 SRVs and TEC levels. Chromium was measured above the PEC, a level above which harmful effects

40 to aquatic biota could potentially occur. Silver was above the ESL and SRV. Arsenic and nickel were

41 above TEC levels but below the SRV. All other inorganic chemicals, tested explosives, pesticides,

- 42 PCBs, and SVOCs were low (below TEC or ESL) or not detected in sediment samples collected from
- 43 Lower Cobbs Pond. Ammonia and total phosphorus levels were measured below screening

1 guidelines. One explosive compound (4-amino-2,6-DNT) was detected in the surface water of Lower

- 2 Cobbs Pond but the measurement was below the Ohio WQS aquatic life maximum and average
- 3 WQC. Lead and pH were the only parameters that were not within the aquatic life OMZA criteria. At
- 4 Lower Cobbs Pond, only chromium in sediment exceeded the PEC; surface water quality did not
- 5 appear to impact the biological community (USACE 2005a).
- 6

7 Fish abundance similarity index values revealed a strong resemblance of Lower Cobbs Pond to Upper Cobbs Pond, Fuze and Booster Ponds, and two reference ponds (Frank's Pond and Trout Pond). 8 9 Stronger similarities were observed based on fish biomass. Similarity index values based on 10 abundance of macroinvertebrates in funnel trap samples were highly variable, and it is difficult to make generalizations about the similarity of Lower Cobbs Pond to the other ponds due to the 11 12 dominance of a few macroinvertebrate taxa. However, Lower Cobbs Pond had a strong resemblance 13 to Boy Scout Pond, which is a 3.4-acre pond that resulted from an old concrete dam across from 14 Eagle Creek and west of Wadsworth Road. Similarity index values based on the presence of 15 macroinvertebrate species at an AOC were not as variable as the abundance data, and Lower Cobbs 16 Pond was most similar to Upper Cobbs Pond and Load Line 4 Pond. The most similar reference pond 17 for macroinvertebrates was Trout Pond (Hester Dendy CSI of 0.54).

18

19 The physical habitat was also evaluated at Lower Cobbs Pond, and the L-QHEI score was 42.5, 20 indicating "poor" habitat quality. This was primarily due to excessive amounts of undesirable species 21 (or low amounts of preferred species) and bottom substrates dominated by muck.

22

Overall, the data collected indicate Upper Cobbs Pond and Lower Cobbs Pond are not currently affected by activities that occurred at the former RVAAP when it was in operation, as stated in the conclusions for all ponds in the study (USACE 2005a):

26

27 "Upper Cobbs Pond [is] not currently affected by activities that occurred at RVAAP when it 28 [was] in operation. Biological conditions at [Upper Cobbs Pond] were comparable to 29 reference ponds, sediment chemicals were below benchmark levels, and surface water quality 30 was consistent with reference conditions and/or below surface WOC. Biological communities 31 in Lower Cobbs Pond...were not affected by chemical exposures from past activities at 32 RVAAP. Fish communities from [Lower Cobbs Pond] were strongly similar to reference 33 ponds. Poor quality vegetative cover in [Lower Cobbs Pond] influenced macroinvertebrate 34 results. Surface water quality was consistent with reference conditions and/or below WOC. 35 Although sediment quality (lead, zinc, or chromium) suggest[s] the potential for negative 36 impacts, aquatic results confirmed conditions comparable to reference ponds."

37

38 Ecosystem and Landscape Roles and Relationships

39

40 There were four spatial areas evaluated to assess the ecosystem and landscape roles and relationships

41 at the Upper and Lower Cobbs Ponds AOC: the actual AOC, the vicinity of the AOC, the entire

42 Camp Ravenna, and the ecoregion of northeastern Ohio. Information about the first spatial area (the

43 AOC) is provided in the terrestrial and aquatic resources subsections above.

Vicinity of the AOC. The same vegetation types previously described at the AOC occur to the east, south, west, and north of AOC for many hundreds of feet. These forest, shrub, and herb communities are not unique and are in abundance at Camp Ravenna. The quality of the forest and other habitats appears to be typical of forests at other places at Camp Ravenna and the northeastern Ohio ecoregion. Thus, the type and quality of forest, shrub, and herbaceous habitats are not unique.

6

As discussed previously, there are several wetlands inside the AOC (Table 7-12). There are many
other wetlands of various sizes located near the east, north, west, and south boundaries of the habitat
area (e.g., the wetland complex associated with the Load Line 12 ditch southwest of the AOC) (Table
7-13).

11

One state-threatened bird, the Least Bittern (*Ixobrychus exilis*), has been observed at the AOC (OHARNG 2014). The closest recorded state- or federally-listed species outside the habitat boundary is the state-endangered yellow-bellied sapsucker (*S. varius*) (located within 500 ft of the southern end of the AOC and about 1,700 ft of the northern end of the AOC). The butternut tree (*Juglans cinerea*) is a federal species of concern located at the Paris-Windham Dump about 2,200 ft north of the AOC (OHARNG 2014) (Table 7-13).

18

There are three beaver dams in or near the AOC (Figure 7-1). There is a 100-year floodplain along
Sand Creek, near the northwestern corner of the AOC. There are biological and water quality stations
within 900 ft of the AOC (Table 7-13).

22

23 The Entire Camp Ravenna. The AOC is approximately 39 acres, which represents 0.18% of the total 24 area of Camp Ravenna (21,683 acres). There are approximately 2,310 acres of forest type FL1 25 [temporarily flooded forest alliance (e.g., green ash and American elm)] at Camp Ravenna 26 (OHARNG 2014); this represents 10.7% of the habitat at Camp Ravenna. There are approximately 27 3,500 acres of forest type FU4 (red maple and green/red ash) (OHARNG 2014), representing 16.1% 28 of the habitat at Camp Ravenna. There are approximately 2,050 acres of vegetation type HU1 (goldenrod and clasping-leaf dogbane) (OHARNG 2014), representing 9.5% of the habitat at Camp 29 30 Ravenna. There are approximately 169 acres of vegetation type HL4 (cattail and bulrush) (OHARNG 2014), representing 0.78% of the habitat at Camp Ravenna. There are approximately 79 acres of 31 32 vegetation type HL5 (spatterdock and white water lily) (OHARNG 2014), representing 0.36% of the 33 habitat at Camp Ravenna. There are approximately 136 acres of vegetation type MFU2 (Norway 34 spruce, white ash, and red maple) (OHARNG 2014), representing 0.63% of the habitat at Camp 35 Ravenna. There are approximately 2,900 acres of vegetation type SU1 (gray dogwood and northern 36 arrowwood) (OHARNG 2014), representing 13.4% of the habitat at Camp Ravenna. There are 37 approximately 1,970 acres of wetlands (jurisdictional and planning level survey) as defined in the 38 INRMP (OHARNG 2014), representing 9.1% of the habitat at Camp Ravenna. Most of these types of 39 resources are abundant and are not unique to the Upper and Lower Cobbs Ponds AOC at Camp 40 Ravenna. HL5 occupies a small area at the AOC and represents an especially small percentage of the 41 entire Camp Ravenna.

1 *Ecoregion.* In the area surrounding Camp Ravenna, forests occupy a high percentage of the terrain. 2 Ohio's forests cover approximately 8,000,000 acres or 30% of the state (USDA 2009). The 3 Erie/Ontario Drift and Lake Plain ecoregions (USGS 1998) are located in the northeastern part of 4 Ohio, and both contain communities of temporarily flooded forest alliance (e.g., green ash and 5 American elm); red maple successional forest; herbaceous field (e.g., goldenrod and dogbane); semipermanently flooded herbaceous alliance (e.g., cattail and bulrush); permanently flooded herbaceous 6 7 alliance (e.g., spatterdock and white water lily); mixed, needle-leaved evergreen, cold-deciduous 8 forest (Norway spruce and red maple); and dry, mid-successional, cold-deciduous shrubland (e.g., 9 dogwood and arrowwood). The Erie/Ontario Drift and Lake Plain ecoregion exhibit rolling to level 10 terrain formed by lacustrine and low lime drift deposits. Lakes, wetlands, and swampy streams occur 11 where stream networks converge or where the land is flat and clayey (USGS 1998). The U.S. Forest Service has a Forest Inventory Data Online tool that was queried for the forest types in the 12 13 surrounding counties in or near Camp Ravenna (USFS 2011). In 2009, approximately 93,900 acres of forest type FL1; 265,290 acres of forest type FU4; and approximately 367,200 acres of forest type 14 15 MFU2 were found throughout northwestern Ohio in Cuvahoga, Geauga, Mahoning, Portage, Stark, 16 Summit, and Trumbull counties that surround Camp Ravenna (USFS 2011). The herbaceous fields 17 and shrubland were not individually found in this query because they are not classified as a main 18 group of trees in the forest inventory data tool. However, herbaceous fields (HU1 and HL4) and 19 shrubland (SU1) are common across the ecoregion (USDA 2011). Wetlands across the ecoregion 20 make up 207,800 acres (USEPA 1999a). The vegetation and wetland communities at the Upper and 21 Lower Cobbs Ponds AOC are also found in the surrounding counties in the ecoregion of northeastern 22 Ohio.

23

24 In summary, the current vegetation types found in the vicinity of the AOC consist of: (1) temporarily 25 flooded forest alliance (e.g., green ash and American elm); (2) red maple forest alliance; (3) dry, earlysuccessional, herbaceous field (e.g., golden rod and dogbane); (4) semi-permanently flooded 26 27 herbaceous alliance (e.g., cattail and bulrush); (5) permanently flooded herbaceous alliance (e.g., 28 spatterdock and white water lily); (6) mixed, needle-leaved evergreen, cold-deciduous forest (e.g., 29 Norway spruce and red maple); (7) dry, mid-successional, cold-deciduous shrubland (e.g., dogwood 30 and arrowwood); and wetlands. The forest types, herbaceous fields, shrubland, and wetlands are in 31 abundance at Camp Ravenna and the larger surrounding local ecoregion. There is no known unique 32 resource at the AOC that cannot be found in the immediate vicinity of the AOC, Camp Ravenna, and 33 in the large part of the ecoregion of northeastern Ohio. The forest types, herbaceous fields, shrubland, 34 and wetlands are in abundance at Camp Ravenna (with the exception of HL5) and the larger 35 surrounding local ecoregion. There is one known unique resource (HL5) at the AOC that is not 36 common in the immediate vicinity of the AOC, is not frequently found at Camp Ravenna, and is not 37 believed to be common to the ecoregion of northeastern Ohio.

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- 39 40

7.3.2.4 <u>Evaluation of Historical Chemical Contamination and Ecological Significance</u>

There were 13 surface soil, 27 sediment, and 8 surface water COPECs identified in the historical
SERA as part of the Phase II RI (Table 7-11).

1 The Army and Ohio EPA provide a checklist of important ecological places and resources to 2 determine if such ecological resources are present in an AOC and/or nearby. The ponds and wetlands 3 are important/significant ecological resources. In addition, the Least Bittern (Ixobrychus exilis), a 4 state of Ohio threatened species, has been observed at the AOC. OHARNG environmental 5 management goals and objectives, including Goal 1 that requires natural resources management to be compatible with military mission and Goal 5 that requires the Army to sustain usable training lands 6 7 and native natural resources by implementing a natural resource management plan that incorporates 8 invasive species management and by utilizing native species mixes for revegetation after ground 9 disturbance activities, are applicable to the AOC.

10

11 The AOC is approximately 39 acres and includes Upper Cobbs Pond (9.4 acres), Lower Cobbs Pond (6.4 acres), the Backwater Area (5.2 acres), and the Pond Bank (18 acres, including a portion of 12 13 Wetland 1). The terrestrial vegetation habitat at the AOC consists of four forest community types, one 14 shrub community type, and three herbaceous community types. The dominant forest community at 15 the AOC is the red maple (A. rubrum) successional forest. Wetlands are associated with the 16 Backwater Area and Lower Cobbs Pond. The vegetation types and wetlands at the AOC are found 17 nearby, at Camp Ravenna, and in the ecoregion.

18

19 Because there is contamination and important or ecologically significant resources at the AOC, this 20 ERA will continue to a Level II Screening Level ERA.

21

22

7.3.3 Level II: Screening Level Ecological Risk Assessment

23

24 The Level II method follows the guidance documents listed in Section 7.3.1 and identifies evaluation 25 procedures used for problem formulation and ecological effects evaluation to determine AOC-related 26 COPECs. This work includes defining habitats/environmental setting, suspected contaminants, 27

28 29

30 In addition, technical and refinement factors can be used to assess outcomes of the above procedures.

possible pathways, and mechanisms for ecotoxicity and contaminant transport. Level II also includes

31 The factors include using mean exposure concentrations and discussing approved ESVs, wetland

32 quality at the AOC, and other topics that evaluate and refine the COPECs from the Level II Screening

33 ERA. This type of assessment is called Step 3A in the ERA process (USEPA 1997).

establishing screening values.

1 2

7

7.3.3.1 Generic Ecological Conceptual Exposure Model

The conceptual site exposure model (CSEM) identifies the interconnections of contaminant sources and transport mechanisms for contaminant migration through the environment to the receptors. The SEM provides an understanding of the relationships of all sources, release and transport pathways, potential exposure media, and receptors. The CSEM includes:

- Source Media. Based on historical AOC information, operations associated with various load
 lines upstream of the AOC (e.g., former buildings at Load Line 12) are the contaminant
 source to the surrounding soil, sediment, and surface water (MKM 2005).
- **Transport Mechanisms.** Contaminants in soil, sediment, surface water, and/or groundwater 11 • 12 can migrate via erosion and leaching. Migration to sediment and surface water via erosion is 13 controlled by the amount of precipitation, type of ground cover, and topography of the AOC. 14 Although parts of the land are relatively sloped, little erosion is expected to occur at the 15 AOC. The slopes and pond banks are extensively vegetated. This will help increase 16 infiltration and decrease erosion. While some of the precipitation landing on this area is 17 expected to infiltrate the soil, some rainfall will leave the AOC as runoff into the Backwater 18 Area, Upper Cobbs Pond, and Lower Cobbs Pond and eventually into the unnamed tributary 19 of Sand Creek.
 - **Exposure Media.** Exposure media are media where contaminants are available for exposure to ecological receptors. Potential exposure media at the AOC are sediment, surface water, soil, vegetation, animals, and groundwater.
 - **Exposure Pathways.** A main exposure pathway is ingesting contaminated food. Other pathways may include ingesting soil, sediment, and water and dermal contact by receptors with soil, sediment, or water.
- Ecological Receptors. A variety of ecological receptors, such as terrestrial birds and mammals, are present in the area. Receptors associated with various published toxicological endpoints (e.g., reproduction, physiology) are assumed to represent these various plants and animals.
- 30 31

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7.3.3.2 <u>Habitats and Species (Including Generic Receptors)</u>

- 33 Habitats and species were defined in the Level I ERA (Section 7.3.2). For example, the terrestrial 34 vegetation habitat at the ponds consists of four forest community types, one shrub community type, 35 and three herbaceous community types. Habitats, species, and other resources were analyzed, and it 36 was determined that important or significant ecological resources are present at the Upper and Lower 37 Cobbs Ponds AOC and vicinity. The Backwater Area, Upper Cobbs Pond, Lower Cobbs Pond, 38 associated wetlands, and the state-threatened Least Bittern (*Ixobrychus exilis*) are present at the AOC. 39 Contamination is present; therefore, a Level II analysis is needed. Level II assumes ecological 40 receptors are sensitive to various chemicals based on a variety of toxicological data from field-41 observed effects and laboratory tests. The ESV is utilized as a toxicity metric representing multiple 42 generic receptors, including plants, microorganisms, and animals.
- 43

1 7.3.3.3 Procedure to Identify COPECs

2

3 The screening level approach to evaluate sample results from the PBA08 RI followed a similar 4 approach used in the historical SERA. Section 5.0 details chemical concentration data. This ERA uses current soil, sediment, and surface water data collected during the PBA08 RI, as well as previously 5 collected data used in the historical SERA and sediment and surface water data collected during the 6 7 FWBWQS (USACE 2005a). The PBA08 RI included surface soil (0–1 ft bgs) and subsurface soil 8 samples collected from the Pond Banks and discrete sediment and surface water samples from the 9 Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. Some locations were different from the 10 historical sample locations and some were co-located with previous samples.

11

12 As discussed previously, sediment and surface water data were aggregated into three EUs: the 13 Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. The soil surrounding the ponds was 14 evaluated as one aggregate. This ERA uses updated SRVs (Appendix H, Table H-13) and ESVs that

follow the revised *Ecological Risk Assessment Guidance* (Ohio EPA 2008), as provided in Appendix

16 H, Tables H-14 through H-16.

17

18 The hierarchy of ESVs is based on the information found in the Ohio EPA risk assessment guidance 19 (Ohio EPA 2008) and FWERWP (USACE 2003a). The MDC of each chemical is compared to its 20 respective facility-wide background concentration. Sediment concentrations are also compared to the 21 SRV. Chemicals are not considered site-related if the MDC is below the background concentration. 22 For all chemicals detected above background concentrations, the MDC is compared to the chemical-23 specific ESV (and SRVs for sediment if an ESV is not available). In addition to the ESV comparison, 24 it was determined if the chemical is a PBT compound. Chemicals are retained as COPECs if they 25 exceed background concentrations (and SRVs for sediment) and the ESV, if the chemical exceeds background concentrations (and SRVs for sediment) and had no toxicity information, or if the 26 27 chemical is considered a PBT compound. MDC to ESV ratios are used to determine the integrated 28 COPECs that result from the combined current and historical data sets. A ratio greater than one 29 suggests a possible environmental consequence. Any chemicals with ratios greater than one are 30 identified as integrated COPECs.

31

Based on comment resolution with Ohio EPA in July 2014, the selection of integrated COPECs for surface water was modified to include two screens. Along with comparing the MDC to the Ohio EPA outside mixing zone maximum (OMZM) ESV, when available, Ohio EPA requested the MDC and average concentrations also be compared to the Ohio EPA OMZA ESV at each EU when available.

36

37 Maximum Detected Concentrations

38

39 The MDCs were compared to background concentrations and ESVs (Appendix H, Tables H-17 40 through H-28) for each detected chemical. These comparisons are provided in Appendix H, Table H-41 17 for surface soil; Tables H-18 through H-22 for discrete and ISM sediment; and Tables H-23 42 through H-28 for surface water.

1 Ecological Screening Values

2

Although the historical SERA used ESVs from the 2003 version of the *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA 2003), this ERA uses updated ESVs from the 2008 version of this guidance document (Ohio EPA 2008). The hierarchy for soil is ecological soil screening levels (EcoSSLs), PRGs, and ESLs. The hierarchy for sediment is the Sediment Quality Guidelines, followed by ESLs. The hierarchy for surface water is the Ohio EPA WQC, National Recommended Water Quality Criteria, and ESLs. Appendix H provides values and sources for ESVs in Tables H-14 through H-16.

- 10
- 11 12

7.3.3.4 Integrated COPECs for Soil (0-1 ft bgs)

13 Historical and PBA08 RI surface soil (0-1 ft bgs) data were evaluated as a single aggregate to 14 determine integrated COPECs. During the evaluation, 44 chemicals were detected in surface soil. 15 Five chemicals (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were 16 excluded as SRCs. A total of 17 inorganic chemicals and 18 organic chemicals were determined to be 17 SRCs because they exceeded background concentrations or did not have an associated background 18 concentration for comparison. Of the 35 SRCs, only 8 chemicals (aluminum, arsenic, cadmium, 19 cobalt, lead, mercury, vanadium, and zinc) exceeded the ESVs and were identified as integrated 20 COPECs (Table 7-14). In addition, three other chemicals (nitrate/nitrite, nitrocellulose, and benzoic 21 acid) were selected as COPECs because they do not have an ESV. Beta-BHC is below its ESV, but it 22 is included because it is a PBT compound (Ohio EPA 2008). Mercury is also a PBT compound. In 23 total, 12 chemicals were identified as integrated COPECs in soil (Table 7-14). The calculated ratio of 24 MDC to ESV is shown in Table 7-14 for each integrated COPEC. Appendix H, Table H-17 presents 25 the details of the ESV comparisons for surface soil.

26

27 Most of the surface soil COPECs reported in the historical SERA (Table 7-11) are also identified in 28 the current ERA. The earlier work (Phase II RI) identified five COPECs (chromium, copper, iron, nickel, and silver) that are not integrated COPECs. Chromium, copper, nickel, and silver were below 29 30 the new ESVs (Ohio EPA 2008, Appendix H, Table H-14), and iron was considered an essential 31 nutrient in the new data set. Four new integrated COPECs (cobalt, lead, benzoic acid, and beta-BHC) 32 are identified in the PBA08 RI. The addition of cobalt and lead are due to both being detected at 33 concentrations higher than the updated, more conservative ESV. Benzoic acid is a new integrated 34 COPEC because it was detected in a sample collected during the PBA08 RI, and it lacks a screening 35 value. The pesticide beta-BHC was not identified as a PBT compound in the historical investigation, 36 but it is in the PBA08 RI per the Guidance for Conducting Ecological Risk Assessments (Ohio EPA 37 2008); therefore, it is an integrated COPEC. Based on the presence of 12 integrated COPECs, this 38 ERA predicts the potential for ecological risk in surface soil.

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7.3.3.5 Integrated COPECs for Sediment

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42 Historical and PBA08 RI discrete sediment data were evaluated as three EUs: the Backwater Area,

43 Upper Cobbs Pond, and Lower Cobbs Pond. ISM sediment data from the FWBWQS (USACE 2005a)

1 was available for two of the three EUs (Upper Cobbs Pond and Lower Cobbs Pond) and was 2 evaluated separately from the discrete samples for each EU. Sediment SRVs and ESVs are published 3 in Appendix H, Tables H-13 and H-15, respectively. Detected chemical concentrations were 4 compared to the sediment background concentrations, SRVs, and ESVs (Appendix H, Tables H-18 5 through H-22), and the results are summarized in Tables 7-16 through 7-18. Five of the chemicals at each EU (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were 6 7 excluded as SRCs. A summary of the integrated COPECs identified in sediment in the three sediment 8 EUs follows.

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Integrated COPECs for Sediment at the Backwater Area

12 At the Backwater Area, 35 of 47 chemicals detected in discrete sediment were identified as SRCs 13 because they either exceeded background concentrations or were organic chemicals with no 14 background concentrations (Appendix H, Table H-18). Of the 35 SRCs, 21 integrated COPECs 15 [cadmium; chromium; copper; cvanide; silver; 2.6-DNT; anthracene; benz(a)anthracene; 16 benzo(a)pyrene; benzo(ghi)perylene; benzo(k)fluoranthene; chrysene; dibenz(a,h)anthracene; 17 fluoranthene; indeno(1,2,3-cd)pyrene; lead; nickel; phenanthrene; pyrene; 2-butanone; and acetone] 18 were identified in sediment at the Backwater Area because they exceeded ESVs. In addition, seven 19 chemicals (antimony, beryllium, selenium, nitrate/nitrite, TNT, nitrocellulose, and tetryl) are 20 integrated COPECs because they do not have an ESV and they exceed their SRVs when SRVs are 21 available. Mercury is below its ESV, but it is included as a PBT compound per the Guidance for 22 Conducting Ecological Risk Assessments (Ohio EPA 2008). PCB-1254 is also a PBT compound; 23 therefore, it is an integrated COPEC. In total, 30 chemicals were identified as integrated COPECs in 24 sediment at the Backwater Area (Table 7-15). The calculated ratio of MDC to ESV is shown in Table 25 7-15 for each integrated COPEC. Appendix H, Table H-18 presents the details of the ESV 26 comparisons for sediment at the Backwater Area.

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Integrated COPECs for Sediment at Upper Cobbs Pond

30 At Upper Cobbs Pond, 31 of 43 chemicals detected in discrete sediment were identified as SRCs 31 (Appendix H, Table H-19). Of the 31 SRCs, 8 integrated COPECs (cadmium; chromium; copper; 32 lead; nickel; silver; 1,3-dinitrobenzene; and acetone) were identified in sediment at Upper Cobbs 33 Pond because they exceeded ESVs. In addition, nine chemicals (antimony; beryllium; hexavalent 34 chromium; selenium; TNT; 4-amino-2,6-DNT; HMX; nitrocellulose; and tetryl) are integrated 35 COPECs because they do not have an ESV and exceed their SRVs when available. Mercury is below 36 its ESV, but it is included as a PBT compound per the Guidance for Conducting Ecological Risk 37 Assessments (Ohio EPA 2008). PCB-1254 is also a PBT compound; therefore, it is an integrated 38 COPEC. In total, 19 chemicals were identified as integrated COPECs in discrete sediment at Upper 39 Cobbs Pond (Table 7-16). The calculated ratio of MDC to ESV is shown in Table 7-16 for each 40 integrated COPEC. Appendix H, Table H-19 presents the details of the ESV comparisons for discrete 41 sediment at Upper Cobbs Pond.

1 In the single ISM sediment sample at Upper Cobbs Pond, 8 of 27 chemicals detected were identified 2 as SRCs (Appendix H, Table H-20). Of the 8 SRCs, 3 integrated COPECs [silver, bis(2-3 ethylhexyl)phthalate, and di-n-butyl phthalate] were identified in sediment at Upper Cobbs Pond 4 because they exceeded ESVs. In addition, three chemicals (nitrate/nitrite, ammonia, and total phosphorous) are integrated COPECs because they do not have an ESV or background criteria. 5 Mercury is below its ESV, but it is included as a PBT compound per the Guidance for Conducting 6 7 Ecological Risk Assessments (Ohio EPA 2008). Seven chemicals were identified as integrated 8 COPECs in ISM sediment at Upper Cobbs Pond (Table 7-16). The calculated MDC to ESV ratio is 9 shown in Table 7-16 for each integrated COPEC. Appendix H, Table H-20 presents the details of the 10 ESV comparisons for ISM sediment at Upper Cobbs Pond.

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12 Integrated COPECs for Sediment at Lower Cobbs Pond

13

14 At Lower Cobbs Pond, 30 of 42 chemicals detected in discrete sediment were identified as SRCs 15 (Appendix H, Table H-21). Of the 30 SRCs, 9 integrated COPECs (arsenic, cadmium, chromium, 16 copper, cyanide, lead, silver, 2-butanone, and acetone) were identified in sediment at Lower Cobbs 17 Pond because they exceeded ESVs. In addition, seven chemicals (antimony, beryllium, hexavalent chromium, selenium, HMX, nitrocellulose, and tetryl) are integrated COPECs because they do not 18 19 have an ESV or exceed their SRVs when available. Mercury is below its ESV, but it is included as a 20 PBT compound per the Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008). In 21 total, 17 chemicals were identified as integrated COPECs in discrete sediment at Lower Cobbs Pond 22 (Table 7-17). The calculated ratio of MDC to ESV is shown in Table 7-17 for each integrated 23 COPEC. Appendix H, Table H-21 presents the details of the ESV comparisons for sediment at Lower 24 Cobbs Pond.

25

26 In the single ISM sediment sample at Lower Cobbs Pond, 9 of 24 chemicals detected were identified 27 as SRCs (Appendix H, Table H-22). Of the nine SRCs, three integrated COPECs (cadmium, 28 chromium, and silver) were identified in sediment at Lower Cobbs Pond because they exceeded 29 ESVs. In addition, one chemical (beryllium) is an integrated COPEC because it does not have an ESV 30 and it exceeds the SRV, and three chemicals (nitrate/nitrite, ammonia, and total phosphorous) are integrated COPECs because they do not have an ESV or background criteria. Mercury is below its 31 32 ESV, but it is included as a PBT compound per the Guidance for Conducting Ecological Risk 33 Assessments (Ohio EPA 2008). In total, eight chemicals were identified as integrated COPECs in ISM 34 sediment at Upper Cobbs Pond (Table 7-17). The calculated MDC to ESV ratio is shown in Table 7-35 17 for each integrated COPEC. Appendix H, Table H-22 presents the details of the ESV comparisons 36 for ISM sediment at Lower Cobbs Pond.

37

Most of the sediment COPECs reported in the historical ERA (Table 7-11) are also identified in the current ERA. The earlier work (Phase II RI) identified one COPEC (total PAHs) that is not an integrated COPEC, as total PAHs were not analyzed for during the PBA08 RI. Rather, the PAH assessment during the PBA08 RI was for individual PAHs [e.g., benzo(a)pyrene and dibenz(a,h)anthracene] as currently performed in the Camp Ravenna full-suite parameters and in accordance with the PBA08 SAP. A total of 13 new integrated COPECs [hexavalent chromium;

1 selenium; ammonia; total phosphorous; 1,3-dinitrobenzene; 4-amino-2,6-DNT; HMX; tetryl; bis(2-2 ethylhexyl)phthalate; di-n-butyl phthalate; dibenz(a,h)anthracene; 2-butanone; and acetone] are 3 identified in the PBA08 RI. Of these, hexavalent chromium, ammonia, and phosphorous lack ESVs. 4 The addition of selenium is due to a higher concentration detected in the PBA08 RI samples that 5 exceeded the SRV. 1,3-Dinitrobenzene; 4-amino-2,6-DNT; HMX; and tetryl are new integrated COPECs due to detections in samples collected during the PBA08 RI, and bis(2-ethylhexyl)phthalate 6 7 and di-n-butyl phthalate are new integrated COPECs due to detections in samples collected in the 8 FWBWQS (USACE 2005a). Dibenz(a,h)anthracene, 2-butanone, and acetone were detected during 9 the PBA08 RI at concentrations greater than the updated, more conservative ESV.

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7.3.3.6 Integrated COPECs for Surface Water

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Surface water at the AOC was evaluated as three EUs: the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. Surface water ESVs are published in Appendix H, Table H-16. Detected chemical concentrations were compared to surface water background concentrations and ESVs (Appendix H, Tables H-23 through H-28). Five of the chemicals at each EU (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were excluded as SRCs. A summary of the integrated COPECs identified in surface water is presented below.

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- 20 21

Integrated COPECs for Surface Water at the Backwater Area

- 22 At the Backwater Area, 22 of 28 detected chemicals were identified as SRCs because they exceeded 23 background concentrations or were organic chemicals with no background concentrations (Appendix 24 H, Table H-23). Of the 22 SRCs, 6 chemicals were identified as integrated COPECs at the Backwater 25 Area. Four inorganic chemicals (manganese, selenium, silver, and sulfide) exceeded ESVs, while two 26 other chemicals (sulfate and chloromethane) are COPECs because they do not have an ESV. An 27 additional inorganic chemical (barium) did not exceed the OMZM but did exceed the OMZA, and 28 was selected as an integrated COPEC (Table 7-18). The calculated ratio of MDC to ESV is shown in 29 Table 7-18 for each integrated COPEC. Appendix H, Tables H-23 and H-24 present the details of the 30 OMZM and OMZA ESV comparisons for surface water at the Backwater Area. These comparisons 31 used the OMZM and OMZA from the Ohio EPA WOC, when available, as the ESV; when the 32 OMZM or OMZA were not available, the hierarchy of preferred sources was followed, as presented in Section 7.3.3.3. 33
- 34

35 Integrated COPECs for Surface Water at Upper Cobbs Pond

36

At Upper Cobbs Pond, 13 of 24 detected chemicals were identified as SRCs (Appendix H, Table H-25). Of the 13 SRCs, 4 integrated COPECs (manganese, nitrate/nitrite, sulfate, and sulfide) were identified in surface water at Upper Cobbs Pond. Manganese and sulfide are integrated COPECs because they exceeded ESVs, and nitrate/nitrite and sulfate are integrated COPECs because they do not have an ESV. The calculated ratio of MDC to ESV is shown in Table 7-19 for each integrated COPEC. Appendix H, Tables H-25 and H-26 present the details of the OMZM and OMZA ESV comparisons for surface water at the Upper Cobbs Pond. These comparisons used the OMZM and 1 OMZA from the Ohio EPA WQC, when available, as the ESV; when the OMZM or OMZA were not

2 available, the hierarchy of preferred sources was followed, as presented in Section 7.3.3.3.

3

Integrated COPECs for Surface Water at Lower Cobbs Pond

4 5

At Lower Cobbs Pond, 19 of 30 detected chemicals were identified as SRCs (Appendix H, Table H-

6 7 27). Of the 19 SRCs, 4 integrated COPECs (manganese, nitrate/nitrite, sulfate, and sulfide) were 8 identified in surface water at Lower Cobbs Pond. Manganese and sulfide are integrated COPECs 9 because they exceeded ESVs, and nitrate/nitrite and sulfate are integrated COPECs because they do 10 not have an ESV. The calculated ratio of MDC to ESV is shown in Table 7-20 for each integrated 11 COPEC. Appendix H, Tables H-27 and H-28 present the details of the OMZM and OMZA ESV comparisons for surface water at Lower Cobbs Pond. These comparisons used the OMZM and 12 13 OMZA from the Ohio EPA WQC, when available, as the ESV; when the OMZM or OMZA were not 14 available, the hierarchy of preferred sources was followed, as presented in Section 7.3.3.3.

15

16 Of the eight surface water COPECs identified in the historical ERA (Phase II RI), six COPECs 17 (manganese, selenium, silver, sulfate, sulfide, and chloromethane) were identified as integrated 18 COPECs in the PBA08 RI. Iron was considered an essential nutrient in the integrated data set and 19 acetone was detected at a concentration less than the new ESV. One new integrated COPEC 20 (nitrate/nitrite) was identified from the biological water quality study data.

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7.3.3.7 <u>Step 3A: Refinement of Integrated COPECs</u>

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24 Step 3A refines the list of integrated COPECs to determine if: (1) there are final COPECs requiring 25 further evaluation in a Level III assessment or if remediation is required to protect ecological 26 receptors; or (2) integrated COPECs can be eliminated from further consideration. This section 27 applies and evaluates refinement factors to the integrated COPECs for the AOC. This evaluation is an 28 important part of the Level II assessment and is adapted from USEPA Step 3A, as outlined in the 29 Ecological Risk Assessment Guidance for Superfund: Process for Designing and Conducting 30 Ecological Risk Assessments (USEPA 1997) and Risk Assessment Handbook Volume II: 31 Environmental Evaluation (USACE 2010b). The purpose of Step 3A is stated as follows by the Army 32 (BTAG 2005):

33 34

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"The results of Step 3A will be used to determine if threats to ecological receptors are negligible and an appropriate risk management decision may be made to end the ERA process, or potential threats are still indicated and a BERA should be initiated."

36 37

- 38 The evaluation and refinement factors used in Step 3A are as follows:
- Comparison of average (i.e., mean) concentration to ESV,
- 41 Comparison of mean concentration to background concentration,
- 42 Comparison of background concentration to ESV,
- Frequency of chemical occurrence relative to ESV,

- Magnitude of ESV exceedance (ratio of ESV to chemical concentrations),
- Discussion of Ohio EPA approved and preferred ESVs,
- Category of wetland quality inside the AOC,
 - Geographical relationship of on-site wetlands to AOC exceedance area,
 - Information about on-site migration of chemicals to on-site wetlands, and
 - Evaluation of off-site migration of chemicals at biological/water quality stations.
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8 Surface soil at the Upper and Lower Cobbs Ponds AOC consists of pond bank soil. Twelve integrated 9 COPECs were identified in surface soil, as presented in Table 7-14. Eight chemicals (aluminum, 10 arsenic, cadmium, cobalt, lead, mercury, vanadium, and zinc) exceeded ESVs; three other chemicals 11 (nitrate/nitrite, nitrocellulose, and benzoic acid) were selected as integrated COPECs because they do 12 not have an ESV. Beta-BHC is below its ESV, but it is included because it is a PBT compound (Ohio 13 EPA 2008).

14

15 Sediment at the Upper and Lower Cobbs Ponds AOC is divided into three EUs: the Backwater Area, 16 Upper Cobbs Pond, and Lower Cobbs Pond. Upper Cobbs Pond and Lower Cobbs Pond both include 17 a single ISM sediment sample in addition to the discrete samples taken in each of the three EUs. In 18 the Backwater Area, there are 30 integrated COPECs in sediment, as presented in Table 7-15. 19 Twenty-one integrated COPECs [cadmium; chromium; copper; cyanide; silver; 2,6-DNT; anthracene; 20 benzo(ghi)perylene; benzo(k)fluoranthene; benz(a)anthracene; benzo(a)pyrene; chrysene; 21 dibenz(a,h)anthracene; fluoranthene; indeno(1,2,3-cd)pyrene; lead; nickel; phenanthrene; pyrene; 2-22 butanone; and acetone] were identified because they exceeded ESVs. In addition, seven chemicals 23 (antimony, beryllium, selenium, nitrate/nitrite, TNT, nitrocellulose, and tetryl) are integrated 24 COPECs because they do not have an ESV. Mercury and PCB-1254 did not exceed ESVs, but were 25 identified as integrated COPECs because they are PBT compounds.

26

27 In Upper Cobbs Pond, there are 19 integrated COPECs in discrete sediment and 7 integrated COPECs 28 in ISM sediment, as presented in Table 7-16. In discrete sediment, eight chemicals (cadmium; 29 chromium; copper; lead; nickel; silver; 1,3-dinitrobenzene; and acetone) were integrated COPECs 30 because they exceeded ESVs; nine chemicals (antimony; beryllium; hexavalent chromium; selenium; 31 TNT; 4-amino-2,6-DNT; HMX; nitrocellulose; and tetryl) were integrated COPECs because they do 32 not have an ESV; mercury and PCB-1254 were PBT compounds. In ISM sediment, three integrated 33 COPECs [silver, bis(2-ethylhexyl)phthalate, and di-n-butyl phthalate] were identified at Upper Cobbs 34 Pond because they exceeded ESVs; three chemicals (nitrate/nitrite, ammonia, and total phosphorous) 35 are integrated COPECs because they do not have an ESV or background criteria; and mercury is a 36 PBT compound.

37

In Lower Cobbs Pond, there are 17 integrated COPECs in discrete sediment and 8 integrated COPECs in ISM sediment, as presented in Table 7-17. In discrete sediment, nine chemicals (arsenic, cadmium, chromium, copper, cyanide, lead, silver, 2-butanone, and acetone) were integrated COPECs because they exceeded ESVs; seven chemicals (antimony, beryllium, hexavalent chromium, selenium, HMX, nitrocellulose, and tetryl) were integrated COPECs because they do not have an ESV; mercury was a PBT compound. In ISM sediment, three integrated COPECs (cadmium, 1 chromium, and silver) were identified at Lower Cobbs Pond because they exceeded ESVs; four

2 chemicals (beryllium, nitrate/nitrite, ammonia, and total phosphorous) are integrated COPECs

3 because they do not have an ESV; mercury is a PBT compound.

4

5 Surface water at the Upper and Lower Cobbs Ponds AOC is divided into three EUs: the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. In the Backwater Area, there are seven integrated 6 7 COPEC in surface water, as presented in Table 7-18. Five chemicals (barium, manganese, selenium, 8 silver, and sulfide) exceeded ESVs; two other chemicals (sulfate and chloromethane) were selected as 9 integrated COPECs because they do not have an ESV. In both the Upper Cobbs Pond EU and Lower 10 Cobbs Pond EU, there are four integrated COPEC in surface water, as presented in Tables 7-20 and 7-21. Two chemicals (manganese and sulfide) exceeded ESVs; two other chemicals (nitrate/nitrite and 11 12 sulfate) do not have an ESV. 13

14 Chemicals with no ESVs are discussed later in Step 3A and in the uncertainty section (Section 15 7.3.3.9). PBT compounds are discussed later in Step 3A. All integrated COPECs that have an MDC 16 to ESV ratio greater than one are evaluated based on a series of evaluation or refinement factors. The 17 first four evaluation and refinement factors are organized to compare key quantitative information. 18 These factors are:

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21

22

- Comparison of average or mean concentration to ESV,
- Comparison of mean concentration to background concentration,
 - Comparison of background concentration to ESV, and
 - Frequency of chemical occurrence relative to ESV.
- 23 24

Multiple evaluation factors can be used to define whether an integrated COPEC should be retained or eliminated from further consideration. There are two types of comparisons and associated decisions in the first steps of the refinement process (Table 7-21).

28

29 Comparing the background concentration to the ESV is an important consideration in this part of the 30 evaluation. Additionally, frequency of detection is provided. The evaluations are presented by type of 31 decision on a COPEC-by-COPEC basis (Table 7-22). These evaluations are followed by the 32 application of additional evaluation and refinement factors, when necessary.

33

34 Evaluation of Integrated COPECs in Soil

35

There are 12 integrated COPECs identified in soil at the AOC. Eight integrated COPECs (aluminum, arsenic, cadmium, cobalt, lead, mercury, vanadium, and zinc) had MDCs that exceeded ESVs, and mercury is also a PBT compound. Beta-BHC has a ratio below 1 but is an integrated COPEC because it is a PBT compound. Three other chemicals (nitrate/nitrite, nitrocellulose, and benzoic acid) are integrated COPECs because they do not have an ESV; they are evaluated later in the refinement process.

1 Comparison of Mean Concentration to ESV

2

Three integrated COPECs (arsenic, cadmium, and cobalt) in soil are eliminated in this step because the mean concentration is smaller than the ESV (Appendix H, Table H-29). Each eliminated integrated COPEC is discussed relative to the mean concentration being smaller than the ESV and the related evaluation and refinement factors. Table 7-22 shows the relevant data and various comparisons.

8

9 Arsenic. Arsenic was detected in all 24 surface soil samples analyzed for inorganic chemicals at the 10 AOC. Only four samples had detections above the background concentration, and only two samples 11 had arsenic concentrations exceeding the ESV (Table 7-22). Although the MDC for arsenic in surface 12 soil exceeds the background concentration and ESV, the mean concentration is less than both the 13 ESV and background concentrations. Therefore, arsenic is eliminated from further consideration and 14 will not be a final COPEC.

15

16 *Cadmium*. Cadmium was detected in 19 of 24 surface soil samples analyzed for inorganic chemicals 17 at the AOC. Cadmium was not detected during the facility-wide background investigation. Only one 18 sample had a concentration exceeding the ESV (Table 7-22). Although the MDC for cadmium in 19 surface soil exceeds the background concentration and ESV, the mean concentration is less than the 20 ESV. Therefore, cadmium is eliminated from further consideration and will not be a final COPEC.

21

Cobalt. Cobalt was detected in all 24 surface soil samples analyzed for inorganic chemicals at the AOC. Eight samples had detections above the background concentration, but only two of these samples had cobalt concentrations exceeding the ESV (Table 7-22). Although the MDC for cobalt in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV and background concentration. Therefore, cobalt is eliminated from further consideration and will not be a final COPEC.

28

Of the eight integrated soil COPECs with MDCs exceeding the ESV, arsenic, cadmium, and cobalt were eliminated from further consideration based on the mean comparison to ESV refining factor. Five remaining integrated COPECs with MDCs greater than the ESV (aluminum, lead, mercury, vanadium, and zinc) in surface soil also have mean concentrations larger than the ESV.

33

34 Comparison of Mean Concentration Above ESV to Background Concentration

35

Three integrated COPECs in surface soil (aluminum, lead, and vanadium) are eliminated in this step because the mean concentrations are larger than the ESVs but smaller than background concentrations. Each remaining integrated COPEC is presented below and discussed relative to the various evaluation and refinement factors.

40

Aluminum. Aluminum in surface soil has a mean concentration less than the background
concentration (Table 7-22). The background concentration is more than 350 times greater than the
ESV, so the ESV can be considered conservative. Although aluminum was detected in all 24 surface

1 soil samples at concentrations above the ESV, only two samples had a detection that exceeded the 2 background concentration. Having only two samples exceed the background concentration suggests 3 the concentration of aluminum in surface soil is not likely a concern. Additionally, aluminum is not 4 likely a concern because the soil pH is too high to dissociate the chemical. A typical soil pH is 6-7 at one of the nearby load lines at Camp Ravenna (USACE 2004). Regarding aluminum chemistry and 5 ecological risk in soil, the USEPA states, "aluminum is identified as a COPC only for soil with a pH 6 7 less than 5.5" (USEPA 2003a). Therefore, because the mean concentration is less than the 8 background concentration, the ESV is rather conservative, and the soil pH at Camp Ravenna is higher 9 than the USEPA dissociation limit, aluminum is eliminated from further consideration and will not be 10 a final COPEC.

11

12 Lead. Lead in surface soil has a mean concentration less than the background concentration (Table 7-13 22). The background concentration is more than two times greater than the ESV, so the ESV can be 14 considered conservative. Although lead was detected in 23 of the 24 surface soil samples at 15 concentrations above the ESV, only three samples had detections exceeding the background 16 concentration. Having only 3 of 24 samples exceed the background concentration suggests the 17 concentration of lead in surface soil is not likely a concern. Therefore, because the mean 18 concentration is less than the background concentration and the ESV is conservative, lead is 19 eliminated from further consideration and will not be a final COPEC.

20

21 Vanadium. Vanadium in surface soil has a mean concentration less than the background 22 concentration (Table 7-22). The background concentration is more than three times greater than the 23 ESV, so the ESV can be considered conservative. Although vanadium was detected in all 24 discrete 24 samples at concentrations above the ESV, only one sample had a detection exceeding the background 25 concentration. Having only one sample exceed the background concentration suggests the 26 concentration of vanadium in surface soil is not likely a concern. Therefore, because the mean 27 concentration is less than the background concentration and the ESV is conservative, vanadium is 28 eliminated from further consideration and will not be a final COPEC.

29

30 **Continued Evaluations**

31

The remaining two integrated COPECs (mercury and zinc) with ESVs in soil have mean concentrations greater than the ESV and the background concentration. Each remaining integrated COPEC is presented below and discussed relative to the first four and related evaluation and refinement factors.

36

Mercury. Mercury was detected in 21 of 24 discrete samples. Mercury in surface soil has a mean concentration greater than the background concentration and ESV. Mercury was detected above the background concentration in 10 of 21 surface soil samples, and it was detected at concentrations above the ESV in all 21 samples (Table 7-22). The background concentration is 70 times greater than the ESV, suggesting the ESV may be very conservative. Although mercury has a very conservative ESV, it is a PBT compound, and the mean concentration exceeds both the background concentration and ESV. Mercury requires further evaluation as a COPEC. *Zinc.* Zinc was detected in all 24 discrete samples. Zinc in surface soil has a mean concentration greater than the background concentration and ESV. It was detected above the background concentration in 10 of 24 discrete samples, and it was detected at concentrations above the ESV in all 24 samples (Table 7-22). The ESV is lower than the background concentration, which indicates the ESV for zinc may be conservative. Because the mean concentration in surface soil exceeds both the background concentration and the ESV, zinc requires further evaluation as a COPEC.

7

Additional Aspects of Continued Evaluations

8 9

10 The second refinement factor comparing the mean concentration to background concentration 11 evaluates how much higher the mean soil concentration is compared to the background concentration. 12 Two COPECs (mercury and zinc) have mean concentrations higher than their background 13 concentrations. If the degree of difference between the mean concentration and the background 14 concentration is small, the integrated COPEC may not be considered a final COPEC. Table 7-23 15 shows that while the mean concentration exceeds the background concentration, the exceedance is 16 small in both cases. For example, the mean concentration for zinc is 63.9 mg/kg, while the 17 background concentration is 61.8 mg/kg.

18

20

19 Additional Technical and Refinement Factors

21 The next two evaluation and refinement factors include:

- 22 23
- Evaluating the magnitude of ESV exceedance (ratio of ESV to chemical concentrations), and
- Discussion of Ohio EPA approved and preferred ESVs.
- 24 25

Evaluating the Magnitude of ESV Exceedance. Although the mean concentration to ESV ratios for mercury (84) and zinc (1.4) indicate a possibility of risk, the zinc ratio is relatively small. The small ratio for zinc indicates the potential for toxicity is relatively low, and this likely supports elimination of this integrated COPEC. The *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA 2008) states:

31 32

"If only minor exceedances are detected and other evidence can substantiate, a claim may be made that some or all of the site-associated soils have not been impacted and no additional ecological investigation of the soils is warranted."

34 35

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36 Comparison of Ohio EPA Approved and Preferred ESVs. The Guidance for Conducting Ecological 37 Risk Assessments (Ohio EPA 2008) gives specific guidance on selecting media screening values 38 (ESVs) for Level II evaluation. For soil, three possible sources of ESV values are listed in order of 39 preference: (1) USEPA EcoSSLs, (2) Preliminary Remediation Goals for Ecological Endpoints (DOE 40 1997), and (3) Region 5 Ecological Screening Levels (USEPA 2003b). However, it is important to 41 note the preferred source (EcoSSLs) can have up to four values per chemical—one for each receptor 42 type (plants, soil invertebrates, birds, and mammals). Because Ohio EPA does not provide guidance 43 on which value to select of these four, the most conservative (lowest) value was chosen for this ERA.

1 It is possible the chosen ESV is too conservative. Alternative ESVs are presented below for the 2 remaining three integrated COPECs.

3

4 The Ohio EPA approved and preferred mercury ESV used in this ERA is 0.00051 mg/kg (Appendix 5 H, Table H-14). The ESV is lower than other ESVs for mercury (0.1 mg/kg), the ESV for methylmercury (0.0016 mg/kg) (USEPA 2003b) (Appendix H, Table H-14), and the background 6 7 concentration (0.036 mg/kg) (Appendix H, Table H-29). The Preliminary Remediation Goals for 8 Ecological Endpoints labels the form of mercury as an "inorganic chemical" and notes the ESV for 9 mercury is "so low that it may often be within background soil concentrations" (DOE 1997). 10 Although mercury can bioaccumulate in food chains as a PBT compound, the Ohio EPA approved and preferred ESV is very low; therefore, the selection of mercury as a COPEC is very conservative. 11

12

13 The Ohio EPA approved and preferred zinc ESV used in this ERA is 46 mg/kg. This value is taken 14 from the USEPA EcoSSLs (Appendix H, Table H-14). The zinc ESV used in this ERA is the most 15 conservative ESV of the USEPA EcoSSLs. Other EcoSSLs for zinc include 79, 120, and 160 mg/kg 16 (USEPA 2007b). The ESV of 46 mg/kg is about five times greater than the ESV (8.5 mg/kg) (DOE 17 1997), the next source of ESVs approved by Ohio EPA (Appendix H, Table H-14). However, the 18 preferred ESV used for zinc is lower than the background concentration of 61.8 mg/kg (Appendix H, 19 Table H-29), while all other EcoSSLs are above the background concentration. These factors indicate 20 the Ohio EPA approved and preferred ESV for zinc is somewhat conservative, and the selection of 21 zinc as a COPEC is conservative.

22

The above information about alternative ESVs shows there are less conservative ESVs that could be chosen for the Level II work. Table 7-24 shows the ratio of ESV to mean concentration for the preferred ESV and an alternative ESV. This alternative ESV is the ESV with the closest concentration to the preferred ESV that is above the background concentration. For the two remaining integrated COPECs, using the alternative ESV would decrease the ratios for mercury and zinc to less than one. Thus, if the alternative ESVs were used, mercury and zinc would be eliminated from further consideration and would not be final COPECs.

30

Based on the three previous refinement factors, the two integrated soil COPECs remaining in Table 7-24 are dismissed as final COPECs. More specifically, mean concentrations are only slightly higher than background concentrations (i.e., mercury and zinc). Mean concentration-to-ESV ratios of exceedance are near one (i.e., zinc). Using alternate ESVs results in ratios below one for mercury and zinc.

36

37 Evaluation of Integrated COPECs in Sediment

38

Integrated COPECs in each sediment EU were subjected to some of the same evaluation and refinement factors applied to integrated COPECs for surface soil. Some factors, such as those concerning wetlands and off-site migration, apply to the whole AOC and are discussed after the surface water evaluation.

1 Backwater Area

2

At the Backwater Area there are 30 integrated COPECs. Nitrate/nitrite, TNT, nitrocellulose, tetryl, and PCB-1254 (also a PBT compound) were conservatively identified as integrated COPECs because they do not have ESVs, background concentrations, or SRVs; these chemicals are discussed later in Step 3A and in the uncertainty section (Section 7.3.3.9). Mercury and PCB-1254 are PBT compounds and discussed later in Step 3A. Cadmium; chromium; copper; cyanide; lead; nickel; silver; 2,6-DNT; anthracene; benz(a)anthracene; benzo(a)pyrene; benzo(ghi)perylene; benzo(k)fluoranthene; chrysene; dibenz(a,h)anthracene; fluoranthene; indeno(1,2,3-cd)pyrene; phenanthrene; pyrene; 2-butanone; and

acetone had MDCs that exceeded the respective ESV. Although antimony, beryllium, and selenium

did not have ESVs, their MDCs exceeded SRVs; these 24 COPECs are discussed below.

10 11

12

13 Comparison of Mean Concentration to ESV – Backwater Area

14

Seven integrated COPECs [cadmium, chromium, copper, lead, nickel, indeno(1,2,3-cd)pyrene, and 2butanone] in sediment at the Backwater Area are eliminated in this step because the mean concentration is smaller than the ESV (Appendix H, Table H-30). Each eliminated integrated COPEC is discussed relative to the mean concentration being smaller than the ESV and the related evaluation and refinement factors. Table 7-25 shows the relevant data and various comparisons.

20

Cadmium. Cadmium was detected in all 14 samples. Cadmium was not detected during the facilitywide background investigation. Six samples had a concentration exceeding the SRV, and five samples had a concentration exceeding the ESV (Table 7-25). Although the MDC for cadmium in sediment exceeds the ESV, the mean concentration is less than the ESV. Therefore, cadmium is eliminated from further consideration and will not be a final COPEC.

26

Chromium. Chromium was detected in all 14 samples. Nine samples had detections above the background concentration, but only one sample had a chromium concentration exceeding the ESV (Table 7-25). Although the MDC for chromium in sediment exceeds the ESV, the mean concentration is less than the ESV. Therefore, chromium is eliminated from further consideration and will not be a final COPEC.

32

Copper. Copper was detected in all 14 samples. Only four samples had detections above both the background concentration and the ESV (Table 7-25). Although the MDC for copper in sediment exceeds the background concentration, ESV, and SRV, the mean concentration is less than the ESV and SRV. Therefore, copper is eliminated from further consideration and will not be a final COPEC.

37

Lead. Lead was detected in all 14 samples. Seven samples had detections above the background concentration, and only four samples had lead concentrations exceeding the ESV (Table 7-25). However, only one sample exceeded the SRV. Although the MDC for lead in sediment exceeds the background concentration, ESV, and SRV, the mean concentration is less than the ESV and SRV. Therefore, lead is eliminated from further consideration and will not be a final COPEC.

1 Nickel. Nickel was detected in all 14 discrete samples. Eight samples had detections above the 2 background concentration, and six samples had nickel concentrations exceeding the ESV (Table 7-3 25). However, only one sample exceeded the SRV. Although the MDC for nickel in sediment exceeds 4 the background concentration, ESV, and SRV, the mean concentration is less than the ESV and SRV. 5 Therefore, nickel is eliminated from further consideration and will not be a final COPEC.

6

7 Indeno(1,2,3-cd)pyrene. Indeno(1,2,3-cd)pyrene was detected in three of five discrete samples. There 8 was no background concentration for comparison, but two samples had concentrations exceeding the 9 ESV (Table 7-25). Although the MDC for indeno(1,2,3-cd) pyrene in sediment exceeds the ESV, the 10 mean concentration is less than the ESV. Therefore, indeno(1,2,3-cd)pyrene is eliminated from further consideration and will not be a final COPEC. 11

12

13 2-Butanone. 2-Butanone was detected in three of five samples. There was no background 14 concentration for comparison, but only one sample had a concentration exceeding the ESV (Table 7-15 25). Although the MDC for 2-butanone in sediment exceeds the ESV, the mean concentration is less 16 than the ESV. Therefore, 2-butanone is eliminated from further consideration and will not be a final 17 COPEC.

- 18
- 19

Comparison of Mean Concentration to Background Concentration and SRVs - Backwater Area 20

21 Three integrated COPECs in sediment (antimony, beryllium, and selenium) are eliminated in this step 22 because the mean concentrations are smaller than background concentrations or SRVs (Appendix H, 23 Table H-30). Each eliminated integrated COPEC is discussed relative to the mean concentration 24 being smaller than the background concentration or SRV and the related evaluation and refinement 25 factors. Table 7-25 shows the relevant data and various comparisons.

26

27 Antimony. Antimony was detected in 11 of 14 samples. There was no background concentration or 28 ESV for comparison, but only two samples had a concentration exceeding the SRV (Table 7-25). 29 Although the MDC for antimony in sediment exceeds the SRV, the mean concentration is less than 30 the SRV. Therefore, antimony is eliminated from further consideration and will not be a final 31 COPEC.

32

33 Beryllium. Beryllium was detected in all 14 samples. There was no ESV for comparison, but 13 of 14 34 samples had a concentration above the background concentration (Table 7-25). In addition, 8 of 14 35 samples exceeded the SRV. Although the MDC for beryllium in sediment exceeds the background 36 concentration and SRV, the mean concentration is less than the SRV. Therefore, beryllium is 37 eliminated from further consideration and will not be a final COPEC.

38

39 Selenium. Selenium was detected in 8 of 14 samples. There was no ESV for comparison, but only 1 40 of 14 samples had a concentration above background concentration and the SRV (which is equal to 41 the background concentration) (Table 7-25). Although the MDC for selenium in sediment exceeds the 42 background concentration and SRV, the mean concentration is less than both. Therefore, selenium is

43 eliminated from further consideration and will not be a final COPEC.

1 Continued Evaluations - Backwater Area

2

The remaining 14 integrated COPECs with ESVs in sediment have mean concentrations greater than the ESV and the background concentration. Each remaining integrated COPEC is presented below and discussed relative to the first four and related evaluation and refinement factors.

6

7 Cyanide. Cyanide was detected in 3 of 12 samples. There was no background concentration for comparison, but three samples had detections exceeding the ESV (Table 7-25). Because cyanide also 8 9 had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV 10 exceedance was very high; the mean concentration to ESV ratio was 2,240 (Appendix H, Table H-30). The highest concentration of cvanide (0.55 mg/kg at ULCPSD-009-0001-SD) in all three EUs 11 occurred in the Backwater Area. There is also uncertainty related to cyanide due to the fact it was not 12 13 included in the 2010 sampling event. However, the overall frequency of detection for all three sediment EUs is small-there are only 4 detections in 27 samples, which include a few OA/OC 14 15 samples. Cyanide, being a negatively charged ion, is relatively mobile by itself; however, it forms very strong complexes with iron and other inorganic chemicals in the environment that effectively 16 17 immobilize it. These complexes dramatically reduce the availability of cyanide to benthic organisms 18 because the complexes are difficult to decompose (EC 1999). Therefore, it is not likely cyanide in 19 sediment is causing a risk to aquatic receptors in the Backwater Area. Due to the low frequency of 20 detection (4 of 27) throughout the three sediment EUs and the immobility of the COPEC, cyanide is 21 eliminated from further consideration and will not be a final COPEC.

22

Silver. Silver was detected in 13 of 14 samples. Silver was not detected during the facility-wide background investigation. A total of 13 of 14 samples had a concentration exceeding the ESV (Table 7-25). Because silver also had a mean concentration that exceeded the ESV, it was evaluated further. The mean concentration to ESV ratio was 11.4 (Appendix H, Table H-30). Therefore, silver will be evaluated further.

28

29 2,6-DNT. 2,6-DNT in sediment had a mean concentration above its ESV (Table 7-25). It was detected 30 in only 1 of 14 samples. There was no background concentration for comparison, but the sample had 31 a concentration exceeding the ESV (Table 7-25). Because 2.6-DNT also had a mean concentration 32 that exceeded the ESV, it was evaluated further. Despite only one detection, the mean concentration 33 remained higher than the ESV because of high detection limits in the non-detected samples (one-half 34 the detection limit for the non-detectable concentrations was used to calculate the mean). The 35 magnitude of the ESV exceedance was very low; the mean concentration to ESV ratio was 1.21 36 (Appendix H, Table H-30). Due to the low frequency of detection (1 of 14) and the low mean 37 concentration to ESV ratio, 2,6-DNT is eliminated from further consideration and will not be a final 38 COPEC.

39

Anthracene. Anthracene in sediment had a mean concentration above its ESV (Table 7-25). It was
 detected in two of five samples. There was no background concentration for comparison, but both
 samples had concentrations exceeding the ESV (Table 7-25). Because anthracene also had a mean

43 concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance

1 was low; the mean concentration to ESV ratio was 1.67 (Appendix H, Table H-30). Due to the low

2 mean concentration to ESV ratio, anthracene is eliminated from further consideration and will not be

3 a final COPEC.

4

5 **Benz(a)** anthracene. Benz(a) anthracene in sediment had a mean concentration above its ESV (Table 7-25). It was detected in three of five samples. There was no background concentration for 6 7 comparison, and two samples had concentrations exceeding the ESV (Table 7-25). Because 8 benz(a)anthracene also had a mean concentration that exceeded the ESV, it was evaluated further. 9 The magnitude of ESV exceedance was fairly low; the mean concentration to ESV ratio was 2.53 10 (Appendix H, Table H-30). While the mean concentration to ESV ratio was low, it was not considered low enough to dismiss benz(a)anthracene as a final COPEC. Benz(a)anthracene will be 11 12 evaluated further.

13

Benzo(a)pyrene. Benzo(a)pyrene in sediment had a mean concentration above its ESV (Table 7-25).
It was detected in four of five samples. There was no background concentration for comparison, and
three samples had concentrations exceeding the ESV (Table 7-25). Because benzo(a)pyrene also had
a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV
exceedance was very low; the mean concentration to ESV ratio was 2.15 (Appendix H, Table H-30).
While the mean concentration to ESV ratio was low, it was not considered low enough to dismiss
benzo(a)pyrene as a final COPEC. Benzo(a)pyrene will be evaluated further.

21

Benzo(ghi)perylene. Benzo(ghi)perylene in sediment had a mean concentration above its ESV (Table 7-25). It was detected in three of five discrete samples. There was no background concentration for comparison, and two samples had concentrations exceeding the ESV (Table 7-25). Because benzo(ghi)perylene also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was low; the mean concentration to ESV ratio was 1.29 (Appendix H, Table H-30). Due to the low mean concentration to ESV ratio, benzo(ghi)perylene is eliminated from further consideration and will not be a final COPEC.

29

30 Benzo(k)fluoranthene. Benzo(k)fluoranthene in sediment had a mean concentration above its ESV 31 (Table 7-25). It was detected in three of five samples. There was no background concentration for 32 comparison, and two samples had concentrations exceeding the ESV (Table 7-25). Because 33 benzo(k)fluoranthene also had a mean concentration that exceeded the ESV, it was evaluated further. 34 The magnitude of ESV exceedance was very low; the mean concentration to ESV ratio was 1.1 35 (Appendix H, Table H-30). Due to the very low mean concentration to ESV ratio (1.1), 36 benzo(k)fluoranthene is eliminated from further consideration and will not be a final COPEC.

37

Chrysene. Chrysene in sediment had a mean concentration above its ESV (Table 7-25). It was detected in four of five samples. There was no background concentration for comparison, and three samples had concentrations exceeding the ESV (Table 7-25). Because chrysene also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance

42 was low; the mean concentration to ESV ratio was 2.04 (Appendix H, Table H-30). While the mean

1 concentration to ESV ratio was low, it was not considered low enough to dismiss chrysene as a final

- 2 COPEC. Chrysene will be evaluated further.
- 3

4 **Dibenz(a,h)** anthracene. Dibenz(a,h) anthracene in sediment had a mean concentration above its ESV 5 (Table 7-25). It was detected in one of five samples. There was no background concentration for comparison, but only one sample had a concentration exceeding the ESV (Table 7-25). Because 6 7 dibenz(a,h)anthracene also had a mean concentration that exceeded the ESV, it was evaluated further. 8 The magnitude of ESV exceedance was low; the mean concentration to ESV ratio was 2.88 9 (Appendix H, Table H-30). Due to the low frequency of detection (one of five) in the Backwater Area and low mean concentration to ESV ratio, dibenz(a,h)anthracene is eliminated from further 10 consideration and will not be a final COPEC. 11

12

Fluoranthene. Fluoranthene in sediment had a mean concentration above its ESV (Table 7-25). It was detected in three of five samples. There was no background concentration for comparison, and two samples had concentrations exceeding the ESV (Table 7-25). Because fluoranthene also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was very low; the mean concentration to ESV ratio was 1.44 (Appendix H, Table H-30). Due to the low mean concentration to ESV ratio, fluoranthene is eliminated from further consideration and will not be a final COPEC.

20

Phenanthrene. Phenanthrene in sediment had a mean concentration above its ESV (Table 7-25). It was detected in three of five discrete samples. There was no background concentration for comparison, but only two samples had concentrations exceeding the ESV (Table 7-25). Because phenanthrene also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was low; the mean concentration to ESV ratio was 1.4 (Appendix H, Table H-30). Due to the low mean concentration to ESV ratio, phenanthrene is eliminated from further consideration and will not be a final COPEC.

28

Pyrene. Pyrene in sediment had a mean concentration above its ESV (Table 7-25). It was detected in three of five samples. There was no background concentration for comparison, but only two samples had concentrations exceeding the ESV (Table 7-25). Because pyrene also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was low; the mean concentration to ESV ratio was 3.10 (Appendix H, Table H-30). While the mean concentration to ESV ratio was low, it was not considered low enough to dismiss pyrene as a final COPEC. Pyrene will be evaluated further.

36

37 Acetone. Acetone in sediment had a mean concentration above its ESV (Table 7-25). It was detected 38 in all five samples. There was no background concentration for comparison, but all five samples had 39 concentrations exceeding the ESV (Table 7-25). Because acetone also had a mean concentration that 40 exceeded the ESV, it was evaluated further. The mean concentration to ESV ratio was 12.1 41 (Appendix H, Table H-30). Therefore, acetone will be evaluated further.

Upper and Lower Cobbs Ponds

1 Upper Cobbs Pond

2

3 At Upper Cobbs Pond, there are 19 integrated COPECs in discrete sediment and 7 integrated 4 COPECs in ISM sediment. Nitrate/nitrite, ammonia, and total phosphorous were conservatively 5 identified as integrated COPECs for ISM sediment, because they do not have ESVs, background concentrations, or SRVs; these chemicals are discussed later in Step 3A and in the uncertainty section 6 7 (Section 7.3.3.9). Mercury is a PBT compound and is discussed later in Step 3A. Silver, bis(2-8 ethylhexyl)phthalate, and di-n-butyl phthalate had MDCs that exceeded the ESV; however, because 9 there is only one ISM sample at Upper Cobbs Pond, the average is equal to the MDC, and these three 10 COPECs are thus discussed in further evaluations later in Step 3A. For discrete sediment samples, hexavalent chromium; TNT: 4-amino-2,6-DNT; HMX; nitrocellulose; and tetryl were conservatively 11 identified as integrated COPECs because they do not have ESVs, background concentrations, or 12 13 SRVs; these chemicals are discussed later in Step 3A and in the uncertainty section (Section 7.3.3.9). Mercury and PCB-1254 are PBT compounds and discussed later in Step 3A. Cadmium; chromium; 14 15 copper; lead; nickel; silver; 1,3-dinitrobenzene; and acetone had an MDC that exceeded the ESV. Although antimony, beryllium, and selenium did not have ESVs, their MDCs exceeded SRVs; these 16 11 COPECs are discussed below. 17

18

19 Comparison of Mean Concentration to ESV - Upper Cobbs Pond

20

Two integrated COPECs (cadmium and lead) in sediment at Upper Cobbs Pond are eliminated in this step because the mean concentration is smaller than the ESV (Appendix H, Table H-31). Each eliminated integrated COPEC is discussed relative to the mean concentration being smaller than the ESV and the related evaluation and refinement factors. Table 7-26 shows the relevant data and various comparisons.

26

27 Cadmium. Cadmium was detected in three of five samples. Cadmium was not detected during the 28 facility-wide background investigation. Two samples had a concentration exceeding the ESV (Table 29 7-26). Although the MDC for cadmium in sediment exceeds the ESV, the mean concentration is less 30 than the ESV. Therefore, cadmium is eliminated from further consideration and will not be a final 31 COPEC.

32

Lead. Lead was detected in all five samples. Only two samples had detections above the background concentration and the ESV (Table 7-26). Although the MDC for lead in sediment exceeds the background concentration, ESV, and SRV, the mean concentration is less than the ESV and SRV. Therefore, lead is eliminated from further consideration and will not be a final COPEC.

37

38 Comparison of Mean Concentration to SRV - Upper Cobbs Pond

39

Four integrated COPECs (antimony, beryllium, nickel, and selenium) in sediment at Upper Cobbs
Pond are eliminated in this step because the mean concentration is smaller than the SRV (Appendix
H, Table H-31). Each eliminated integrated COPEC is discussed relative to the mean concentration

Upper and Lower Cobbs Ponds

1 being smaller than the SRV and the related evaluation and refinement factors. Table 7-26 shows the

2 relevant data and various comparisons.

3

Antimony. Antimony was detected in all five samples. Antimony was not detected during the facilitywide background investigation. There also was no ESV for comparison, but only one sample had a concentration exceeding the SRV (Table 7-26). Although the MDC for antimony in sediment exceeds the SRV, the mean concentration is less than the SRV. Therefore, antimony is eliminated from further consideration and will not be a final COPEC.

9

10 **Beryllium.** Beryllium was detected in all five samples. There was no ESV for comparison, but all five 11 samples had concentrations above the background concentration (Table 7-26). In addition, two of five 12 samples exceeded the SRV. Although the MDC for beryllium in sediment exceeds the background 13 concentration and SRV, the mean concentration is essentially equal to the SRV. Therefore, beryllium 14 is eliminated from further consideration and will not be a final COPEC.

15

16 Nickel. Nickel was detected in all five samples. Four samples had detections above the background 17 concentration and the ESV, and two samples had nickel concentrations exceeding the SRV (Table 7-18 26). Although the MDC and mean concentration for nickel in sediment exceeds the background 19 concentration and ESV, the mean concentration is less than the SRV. Therefore, nickel is eliminated 10 from further consideration and will not be a final COPEC.

21

Selenium. Selenium was detected in three of five samples. There was no ESV for comparison, but only one of five samples had a concentration above background concentration and the SRV, which is equal to the background concentration (Table 7-26). Although the MDC for selenium in sediment exceeds the background concentration and SRV, the mean concentration is less than both. Therefore, selenium is eliminated from further consideration and will not be a final COPEC.

27

29

28 Continued Evaluations - Upper Cobbs Pond

The remaining five integrated COPECs with ESVs in sediment have mean concentrations greater than the ESV and the background concentration. Each remaining integrated COPEC is presented below and discussed relative to the first four and related evaluation and refinement factors.

33

Chromium. Chromium was detected in all six samples. All six samples had detections above the background concentration, and three samples had a chromium concentration exceeding the ESV (Table 7-26). Because chromium also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was very low; the mean concentration to ESV ratio was 1.17 (Appendix H, Table H-31). Due to the very low mean concentration to ESV ratio (1.17), chromium is eliminated from further consideration and will not be a final COPEC.

40

41 *Copper*. Copper was detected in all five samples. Four samples had detections above the background 42 concentration, and three samples had detections exceeding the ESV and SRV (Table 7-26). Because

43 copper also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude

1 of ESV exceedance was fairly low; the mean concentration to ESV ratio was 1.81 (Appendix H, 2 Table H-31). The mean concentration to background concentration ratio also is low at 2.07. Due to 3 the low mean concentration to ESV ratio and low mean to background concentration ratio, copper is

4 eliminated from further consideration and will not be a final COPEC.

5

Silver. Silver was detected in three of five samples. Silver was not detected during the facility-wide
background investigation. All three samples had concentrations exceeding the ESV (Table 7-26).
Because silver also had a mean concentration that exceeded the ESV, it was evaluated further. The
mean concentration to ESV ratio was 7.6 (Appendix H, Table H-31). Therefore, silver will be
evaluated further.

11

12 1,3-Dinitrobenzene. 1,3-Dinitrobenzene was detected in only one of five samples. There was no 13 background concentration for comparison, but the sample had a concentration exceeding the ESV 14 (Table 7-26). Because 1.3-dinitrobenzene also had a mean concentration that exceeded the ESV, it 15 was evaluated further. Despite only one detection, the mean concentration remained higher than the 16 ESV because of high detection limits in the non-detected samples (one-half the detection limit for the 17 non-detectable concentrations was used to calculate the mean). The magnitude of ESV exceedance 18 was low; the mean concentration to ESV ratio was 1.67 (Appendix H, Table H-31). Due to the low 19 frequency of detection (one of five) in Upper Cobbs Pond, the absence of any detections in the other 20 two sediment EUs, and the low mean concentration to ESV ratio (1.67), 1,3-dinitrobenzene is 21 eliminated from further consideration and will not be a final COPEC.

22

Acetone. Acetone was detected in both samples collected. There was no background concentration for comparison, but both samples had concentrations exceeding the ESV (Table 7-26). Because acetone also had a mean concentration that exceeded the ESV, it was evaluated further. The mean concentration to ESV ratio was 9 (Appendix H, Table H-31). Therefore, acetone will be evaluated further.

28

29 Lower Cobbs Pond

30

31 At Lower Cobbs Pond, there are 17 integrated COPECs in discrete sediment and 8 integrated 32 COPECs in ISM sediment. For ISM sediment, nitrate/nitrite, ammonia, and total phosphorous were 33 conservatively identified as integrated COPECs because they do not have ESVs, background concentrations, or SRVs, and beryllium exceeds the background and SRV but does not have an ESV; 34 35 these chemicals are discussed later in Step 3A and in the uncertainty section (Section 7.3.3.9). 36 Mercury is a PBT compound and is discussed later in Step 3A. Cadmium, chromium, and silver, had 37 an MDC that exceeded the ESV; however, because there is only one ISM sample at Lower Cobbs 38 Pond and the average is equal to the MDC, these three COPECs are discussed in further evaluations 39 later in Step 3A. For discrete sediment samples, hexavalent chromium, HMX, nitrocellulose, and 40 tetryl were conservatively identified as integrated COPECs because they do not have ESVs, SRVs, or 41 background concentrations; these chemicals are discussed later in Step 3A and in the uncertainty 42 section (Section 7.3.3.9). Mercury is a PBT compound and is discussed later in Step 3A. Arsenic, 43 cadmium, chromium, copper, cyanide, lead, silver, 2-butanone, and acetone had an MDC that

exceeded the ESV. Although antimony, beryllium, and selenium did not have ESVs, their MDCs
 exceeded SRVs; these 12 COPECs are discussed below.

3 4

5

Comparison of Mean Concentration to ESV and SRV- Lower Cobbs Pond

6 Seven integrated COPECs (antimony, arsenic, beryllium, cadmium, lead, selenium, and 2-butanone) 7 in sediment at Lower Cobbs Pond are eliminated in this step because the mean concentration is 8 smaller than the ESV or SRV (Appendix H, Table H-32). Each eliminated integrated COPEC is 9 discussed relative to the mean concentration being smaller than the ESV or SRV and the related 10 evaluation and refinement factors. Table 7-27 shows the relevant data and various comparisons.

11

12 Antimony. Antimony was detected in only one of eight samples. Antimony was not detected during 13 the facility-wide background investigation. There also was no ESV for comparison, but the sample 14 had a concentration exceeding the SRV (Table 7-27). Although the MDC for antimony in sediment 15 exceeds the SRV, the mean concentration is less than the SRV. Therefore, antimony is eliminated 16 from further consideration and will not be a final COPEC.

17

18 Arsenic. Arsenic was detected in all eight samples. Five samples had detections above the ESV, but 19 two samples had detections exceeding the background concentration, and only one sample exceeded 20 the SRV (Table 7-27). Although the MDC and mean concentration for arsenic in sediment exceed the 21 ESV, the mean concentration is less than the background concentration and SRV. Therefore, arsenic 22 is eliminated from further consideration and will not be a final COPEC.

23

Beryllium. Beryllium was detected in all eight samples. There was no ESV for comparison, but all eight samples had concentrations above the background concentration (Table 7-27). In addition, two of eight samples exceeded the SRV. Although the MDC for beryllium in sediment exceeds the SRV, the mean concentration is less than the SRV. Therefore, beryllium is eliminated from further consideration and will not be a final COPEC.

29

30 Cadmium. Cadmium was detected in six of eight samples. Cadmium was not detected during the 31 facility-wide background investigation. Two samples had a concentration exceeding the ESV, and 32 three samples had concentrations exceeding the SRV (Table 7-27). Although the MDC for cadmium 33 in sediment exceeds the ESV and SRV, the mean concentration is less than the ESV and SRV. 34 Therefore, cadmium is eliminated from further consideration and will not be a final COPEC.

35

Lead. Lead was detected in all eight samples. Four samples had detections above the background concentration, and three samples had a lead concentration exceeding the ESV (Table 7-27). Although the MDC for lead in sediment exceeds the background concentration, ESV, and SRV, the mean concentration is less than the ESV and SRV. Therefore, lead is eliminated from further consideration and will not be a final COPEC.

41

42 *Selenium.* Selenium was detected in four of eight samples. There was no ESV for comparison, but 43 one of eight samples had a concentration above the background concentration, and the SRV is equal 1 to the background concentration (Table 7-27). Although the MDC for selenium in sediment exceeds

2 the background concentration and SRV, the mean concentration is less than both. Therefore, selenium

3 is eliminated from further consideration and will not be a final COPEC.

4

2-Butanone. 2-Butanone was detected in both discrete samples. There was no background
concentration for comparison, and one sample had a concentration exceeding the ESV (Table 7-27).
Although the MDC for 2-butanone in sediment exceeds the background concentration and ESV, the
mean concentration is less than the ESV. Therefore, 2-butanone is eliminated from further
consideration and will not be a final COPEC.

10

11 Continued Evaluations – Lower Cobbs Pond

12

13 The remaining five integrated COPECs with ESVs in sediment have mean concentrations greater than 14 the ESV and the background concentration. Each remaining integrated COPEC is presented below 15 and discussed relative to the first four and related evaluation and refinement factors.

16

17 *Chromium.* Chromium was detected in all eight samples. Five samples had detections above the 18 background concentration, and three samples had a chromium concentration exceeding the ESV 19 (Table 7-27). Because chromium also had a mean concentration that exceeded the ESV, it was 20 evaluated further. The magnitude of ESV exceedance was very low; the mean concentration to ESV 21 ratio was 1.36 (Appendix H, Table H-32). The mean to SRV ratio was low at 2.03. Due to the low 22 mean concentration to ESV ratio and low mean to SRV ratio, chromium is eliminated from further 23 consideration and will not be a final COPEC.

24

Copper. Copper was detected in all eight samples. Four samples had detections above the background concentration and the ESV, and three samples had detections exceeding the SRV (Table 7-27). Because copper also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was fairly low; the mean concentration to ESV ratio was 1.79 (Appendix H, Table H-32). The mean to background ratio was low at 2.05. Due to the low mean concentration to ESV ratio and low mean to background ratio, copper is eliminated from further consideration and will not be a final COPEC.

32

33 Cyanide. Cyanide was detected in only one of six samples. PBA08 RI samples were not analyzed for 34 cyanide. Cyanide was detected in the one sample at a concentration above the background 35 concentration and ESV (Table 7-27). Because cyanide also had a mean concentration that exceeded 36 the ESV, it was evaluated further. The magnitude of ESV exceedance was high; the mean 37 concentration to ESV ratio was 2,250 (Appendix H, Table H-32). The second highest concentration of 38 cyanide (0.4 mg/kg at ULCPSD-024-0001-SD) in all three EUs occurred in the Backwater Area. The 39 overall frequency of detection for all three sediment EUs is small-there are only 4 detections in 27 40 samples, which includes QA/QC samples. Cyanide, being a negatively charged ion is relatively 41 mobile by itself; however, it forms very strong complexes with iron and other inorganic chemicals in 42 the environment that effectively immobilize it. These complexes dramatically reduce the availability 43 of cyanide to benthic organisms because the complexes are difficult to decompose (EC 1999).

1 Therefore, it is not likely the cyanide in sediment is causing a risk to aquatic receptors in Lower

2 Cobbs Pond. Due to the low frequency of detection (4 of 27) across the three sediment EUs and the

immobility of the COPEC, cyanide is eliminated from further consideration and will not be a finalCOPEC.

4 (5

6 *Silver*. Silver was detected in five of eight samples. Silver was not detected during the facility-wide 7 background investigation. Three samples had concentrations exceeding the ESV (Table 7-27). 8 Because silver also had a mean concentration that exceeded the ESV, it was evaluated further. The 9 magnitude of ESV exceedance was fairly low; the mean concentration to ESV ratio was 1.7 10 (Appendix H, Table H-32). The mean to SRV ratio was low at 1.98. Due to the low mean 11 concentration to ESV ratio and low mean to SRV ratio, silver is eliminated from further consideration 12 and will not be a final COPEC.

13

14 Acetone. Acetone was detected in both discrete samples collected. There was no background 15 concentration for comparison, but both samples had concentrations exceeding the ESV (Table 7-27). 16 Because acetone also had a mean concentration that exceeded the ESV, it was evaluated further. The 17 mean concentration to ESV ratio was 12.4 (Appendix H, Table H-32). Therefore, acetone will be 18 evaluated further.

19

The remaining sediment COPECs still requiring evaluation are silver, benz(a)anthracene,
benzo(a)pyrene, chrysene, pyrene, and acetone in the Backwater Area; silver and acetone in Upper
Cobbs Pond; and acetone in Lower Cobbs Pond.

23

24 Evaluation of Integrated COPECs in Surface Water

25

Integrated COPECs in surface water were subjected to some of the same evaluation and refinement factors applied to integrated COPECs for soil and sediment. Some factors, such as those concerning wetlands and off-site migration, apply to the whole AOC and are discussed after this section.

29

30 Backwater Area. Sulfate and chloromethane were conservatively identified as integrated COPECs 31 because they do not have an ESV or background concentration; these chemicals are discussed later in 32 Step 3A and in the uncertainty section (Section 7.3.3.9). Barium, manganese, selenium, silver, and 33 sulfide had an MDC that exceeded the ESV; these four COPECs are discussed below.

34

Barium. Barium was detected in all four surface water samples. Two of four samples had detections above the background concentration, but only one chemical detection was above the OMZA ESV; none were above the OMZM ESV (Appendix H, Tables H-16, H-23, and H-24; Table 7-28). Although the MDC for barium in surface water exceeds the OMZA ESV, the mean concentration is less than the OMZA ESV. Therefore, barium is eliminated from further consideration and will not be a final COPEC.

41

Manganese. Manganese was detected in all four surface water samples. All four samples had
 detections above the ESV, but only two chemical detections were above background concentrations

1 (Table 7-28). The background concentration is more than 3.2 times greater than the ESV; therefore, 2 the ESV can be considered conservative. Because manganese also had a mean concentration that 3 exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was fairly high; the 4 mean concentration to ESV ratio was 51.3 (Appendix H, Table H-33). Because the mean 5 concentration of manganese was above the ESV, this COPEC will be evaluated further.

6

Selenium. Selenium was detected in three of four surface water samples. All three samples had
detections above the background concentration, but only one chemical detection was above the ESV
(Table 7-28). Although the MDC for selenium in surface water exceeds the ESV, the mean
concentration is less than the ESV. Therefore, selenium is eliminated from further consideration and
will not be a final COPEC.

12

Silver. Silver was detected in only one of four surface water samples, and it was detected above its background concentration and ESV (Table 7-28). Although the MDC for silver in surface water exceeds the ESV, the mean concentration is less than the ESV. Therefore, silver is eliminated from further consideration and will not be a final COPEC.

17

Sulfide. Sulfide was detected in one of two surface water samples. There was no background concentration for comparison, but the sample had a concentration above the ESV (Table 7-28). Because sulfide also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was high; the mean concentration to ESV ratio was 640 (Appendix H, Table H-33). Because the mean concentration of sulfide was above the ESV, this COPEC will be evaluated further.

24

Upper Cobbs Pond. At Upper Cobbs Pond, there are four surface water integrated COPECs: manganese, nitrate/nitrite, sulfate, and sulfide. Nitrate/nitrite and sulfate were conservatively identified as integrated COPECs because they do not have an ESV or background concentration. Therefore, they are discussed later in Step 3A and in the uncertainty section (Section 7.3.3.9). Manganese and sulfide had an MDC that exceeded the ESV; these two COPECs are discussed below.

30

Manganese. Manganese was detected in all five surface water samples. Three of five samples had detections above the ESV, but only one chemical detection was above background concentration (Table 7-29). The background concentration is more than 3.2 times greater than the ESV; therefore, the ESV can be considered conservative. Although the MDC for manganese in surface water exceeds the background concentration, the mean concentration is less than the background concentration. Therefore, manganese is eliminated from further consideration and will not be a final COPEC.

37

Sulfide. Sulfide was detected in both surface water samples. There was no background concentration
 for comparison, but both samples had concentrations above the ESV (Table 7-29). Because sulfide

40 also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of

41 ESV exceedance was high; the mean concentration to ESV ratio was 350 (Appendix H, Table H-34).

- 42 Because the mean concentration of sulfide was above the ESV, this COPEC will be evaluated further.
- 43

Lower Cobbs Pond. At Lower Cobbs Pond, there are four surface water integrated COPECs: manganese, nitrate/nitrite, sulfate, and sulfide. Nitrate/nitrite and sulfate were conservatively identified as integrated COPECs because they do not have an ESV or background concentration. Therefore, they are discussed later in Step 3A and in the uncertainty section (Section 7.3.3.9). Manganese and sulfide had an MDC that exceeded the ESV; these two COPECs are discussed below.

6

Manganese. Manganese was detected in all seven surface water samples. Four samples had detections above the ESV, but only one chemical detection was above background concentration (Table 7-30). The background concentration is more than 3.2 times greater than the ESV; therefore, the ESV can be considered conservative. Although the MDC for manganese in surface water exceeds the background concentration, the mean concentration is less than the background concentration. Therefore, manganese is eliminated from further consideration and will not be a final COPEC.

13

Sulfide. Sulfide was detected in one of three surface water samples. There was no background concentration for comparison, but the sample had a concentration above the ESV (Table 7-30). Because sulfide also had a mean concentration that exceeded the ESV, it was evaluated further. The magnitude of ESV exceedance was high; the mean concentration to ESV ratio was 200 (Appendix H, Table H-35). Because the mean concentration of sulfide was above the ESV, this COPEC will be evaluated further.

20

21 22

Wetland Quality, Geographical Information, and On-site Migration of Chemicals

The next three evaluation and refinement factors are concerned with risk to wetlands. The threefactors are:

25 26

27

28

29

- Category of wetland quality inside the AOC;
- Geographical relationship of on-site wetlands to AOC exceedance area; and
- Information about on-site migration of chemicals to on-site wetlands.

30 If the wetland quality is low, it is distant from the AOC exceedance area (i.e., high concentration 31 area), or on-site migration is unlikely, it increases the likelihood that the remaining integrated 32 COPECs [cadmium, chromium, silver, benz(a)anthracene, benzo(a)pyrene, bis(2-33 ethylhexyl)phthalate, chrysene, di-n-butyl phthalate, pyrene, and acetone in sediment and manganese 34 and sulfide in surface water] will not be of ecological concern and do not need to be evaluated as final 35 COPECs.

36

There are three wetlands partially inside the AOC. They range in size from 0.34–4.75 acres, with between 0.17 acres and 3.9 acres inside the AOC boundary. Wetlands 1 and 2 are Category 2 wetlands, while Wetland 3 is a Category 3 wetland (Table 7-12). Category 2 indicates moderate wetland quality, with some degradation of wetland functions. Category 3 indicates high wetland quality that is often forested. Camp Ravenna contains about 1,970 acres of wetlands, and the 7.67 acres of wetland inside the habitat boundary at the AOC represent 0.39% of the total wetlands at Camp Ravenna (OHARNG 2014). The availability of many more wetland acres at Camp Ravenna
 lowers the importance of the wetlands at the AOC.

3

The three wetlands (Figure 7-2) are adjacent geographically to either the Backwater Area (Wetlands 2 and 3), Upper Cobbs Pond (Wetland 2), or to the unnamed tributary that drains Lower Cobbs Pond (Wetland 1). Wetland 3, the highest quality wetland, is associated with the Backwater Area, which contains the greatest number of sediment and surface water COPECs. This suggests integrated COPEC effects on wetland quality could be minimal.

9

10 Over a period of years, contaminants from Load Lines 3 and 12, as well as other upgradient AOCs, may have flowed northward through the Backwater Area, into Upper Cobbs Pond and Lower Cobbs 11 Pond, and finally out of the AOC through the unnamed tributary to Sand Creek. Soluble contaminants 12 13 are likely to be discharged, dispersed, and diluted. Insoluble (immobile) chemicals are more likely to be retained. The three wetlands associated with the AOC are considered an ecosystem sink during 14 15 most seasons of the year. Based on the number of sediment and surface COPECs in the three EUs, it 16 appears that COPEC migration becomes more limited moving downstream (i.e., northward) in the 17 AOC. Off-site migration to downstream sample locations also does not appear to be an issue (see 18 following section). These factors further reduce the concern about the remaining integrated COPECs.

19

20 Evaluation of Biological and Water Quality Sampling Stations

21

22 The last evaluation and refinement factor is:

- 23
- 24 25

• Evaluation of off-site migration of chemicals at biological/water quality stations.

26 As discussed in Section 7.3.2.3, various biological measurements of macroinvertebrates and fish, as 27 well as chemical and physical measurements of surface water and sediment, were taken and assessed for evidence of downgradient contamination from AOCs. The results indicated that Upper Cobbs 28 29 Pond and Lower Cobbs Pond are not currently affected by contamination from activities that occurred 30 at the former RVAAP when it was in operation. Review of the FWBWQS (USACE 2005a) data from 31 the downstream sampling location (Station S-8) showed many positive attribute ratings (e.g., good, 32 excellent, full attainment), no sign of aquatic impairment, and the station was rated at Full Use 33 Attainment Status, which indicated that all indices met the Ohio EPA biological criteria.

34

35 Therefore, COPECs at the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond have not adversely affected downstream ecological conditions. These findings greatly reduce concern for the 36 37 remaining COPECs [cadmium, chromium, silver, benz(a)anthracene, benzo(a)pyrene, bis(2-38 ethylhexyl)phthalate, chrysene, pyrene, di-n-butyl phthalate, and acetone in sediment and manganese 39 and sulfide in surface water] at the AOC. In essence, while there is some uncertainty associated with 40 dismissing these COPECs (because chemical data suggested the possibility for adverse effects), the 41 lack of obvious biological effects in the ponds and at the downstream station outweighs the chemical 42 findings. As a result, these remaining COPECs are not considered final COPECs and will not be evaluated further. 43

1 Evaluation of PBT Compounds and COPECs Without ESVs

2 3

4

As discussed in Level II, there are PBT compounds in soil and sediment as well as chemicals that are integrated COPECs because they did not have ESVs in surface soil, sediment, and surface water. These chemicals are evaluated below.

5 6

PBT Compounds. The Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008) includes a PBT compound screen in the Level II ERA. This screen is necessary because not all ESVs account for bioaccumulation; instead, they are derived based primarily on toxicity to endpoint receptors exposed by direct contact (e.g., plants, soil-dwelling invertebrates) or ingesting soil or water (e.g., mammals, birds). Ohio EPA allows PBTs to be screened out in Level II if the "method used to derive the screening value considered exposure to higher trophic level organisms in the development of the screening value" (Ohio EPA 2008).

14

15 For the first two sources of soil ESVs preferred by Ohio EPA (EcoSSLs and PRGs), bioaccumulation in higher trophic levels is considered when developing the ESV. According to the EcoSSL guidance, 16 17 "wildlife receptors may be exposed to contaminants in soil by two main pathways: incidental ingestion of soil while feeding, and ingestion of food items that have become contaminated due to 18 19 uptake from soil" (USEPA 2007a). Deriving EcoSSL values includes uptake equations that account 20 for both direct ingestion and food chain bioaccumulation (USEPA 2007a). The same is true of PRGs: 21 "the 90th percentile of the soil-to-biota uptake factor was used as a conservative estimate of the chemical concentrations in wildlife food types (earthworms, plants, or small mammals)," and "the 22 23 model accounts for the ingestion of soil as well as food" (DOE 1997). It is also important to note that 24 both sources often derive values for multiple receptors, and the most conservative (lowest) value is 25 chosen. Thus, for soil ESVs from these two sources, PBT compounds that have ratios less than one 26 can be dismissed as final COPECs.

27

28 The USEPA Region 5 ESLs (USEPA 2003b) are an Ohio EPA-approved source of soil, sediment, 29 and surface water ESVs (see Appendix H, Tables H-14 to H-16 for hierarchies) for some PBT 30 compounds not covered in the EcoSSLs or PRGs. The ESLs also account for bioaccumulation in the 31 food chain, as "development of ESLs focused on mammalian or avian species and identified those 32 chemicals that have the potential for significant bioaccumulation or biomagnification" (USEPA 1999b). Thus, development of ESLs included bioaccumulation in higher tropic levels as a standard 33 34 component in the equations, and PBT compounds with ESVs in soil, sediment, and surface water 35 from this source that have ratios less than one can be dismissed as final COPECs. For this ERA, this 36 includes beta-BHC in surface soil.

37

The other PBT compound detected in Pond Bank surface soil was mercury, which was detected above the ESV, which accounts for bioaccumulation (DOE 1997). This exceedance is to be expected, considering the background concentration is 70 times greater than the ESV. While the mean concentration of mercury (0.043 mg/kg) is above the background concentration (0.036 mg/kg) (Table 7-23), the similarity between the concentrations suggests exposures to mercury are not very different

43 from background concentrations. Mercury was evaluated in the food chain during the historical ERA

1 (MKM 2005). Because the conservative scenario and no observed adverse effect level (NOAEL) did

2 not result in HQs that were greater than 1, the risk was determined to be acceptable. Thus, mercury is

3 eliminated from further consideration as a PBT chemical and will not be a final COPEC.

4

5 For sediment, mercury and PCB-1254 were identified as PBT chemicals. Mercury was detected in 13 of 14 samples at an average concentration of 0.0816 mg/kg in the Backwater Area, in 5 of 5 samples 6 7 at an average concentration of 0.0662 mg/kg in Upper Cobbs Pond discrete sediment, in 1 of 1 8 samples at a concentration of 0.048 mg/kg in Upper Cobbs Pond ISM sediment, in 7 of 8 samples at 9 an average concentration of 0.0538 mg/kg in Lower Cobbs Pond discrete sediment, and in 1 of 1 10 samples at a concentration of 0.1 mg/kg in Lower Cobbs Pond ISM sediment. The background concentration of mercury in sediment is 0.059 mg/kg, and the SRV is 0.12 mg/kg. Based on the 11 similarity of the mean concentration and background concentration, mercury is not considered an 12 13 ecological concern for the AOC. PCB-1254 was detected in two of four samples in the Backwater 14 Area at a maximum concentration of 0.047 mg/kg and in one of two discrete samples collected in 15 Upper Cobbs Pond at a maximum concentration of 0.032 mg/kg. The ESV for PCB-1254 is 0.0598 16 mg/kg and does not account for bioaccumulation. PCB-1254 was not detected in Lower Cobbs Pond. 17 Both PCB-1254 and mercury were evaluated in the food chain during the historical ERA (MKM 2005). Because the NOAEL HOs were less than 1 for avian piscivores (heron and kingfisher; 18 19 Appendix H, Table H-9), NOAEL HQs were low for mink (1.02 using the average exposure scenario; 20 Appendix H, Table H-9). LOAEL HQs were less than 1 for all piscivores (Appendix H, Table H-10); 21 therefore, risks to piscivorous wildlife from PCB-1254 were acceptable. In addition, dietary exposure 22 doses based on mercury did not exceed dietary toxicity benchmarks for the heron, kingfisher, or 23 mink. Thus, wildlife exposure to mercury and PCB-1254 is not a concern at the AOC. PCB-1254 and 24 mercury are eliminated from further consideration as PBT chemicals and will not be final COPECs.

25

26 **COPECs** without ESVs. The Guidance for Conducting Ecological Risk Assessments specifies 27 chemicals without screening benchmark values should be retained as COPECs (Ohio EPA 2008). 28 While Ohio EPA allows the use of additional screening benchmark values, such values need to be 29 approved prior to submitting the report. For the AOC, a search for (and subsequent approval of) 30 additional values was not deemed necessary. Rather, to mitigate concern for the uncertainties 31 associated with COPECs that lack ESVs, a limited additional evaluation was conducted for each 32 medium, focusing on frequency of detection, relationship to background concentration, and other 33 chemical-specific refinement factors.

34

For Pond Bank soil, the integrated COPECs without ESVs are nitrate/nitrite, nitrocellulose, and benzoic acid. Nitrate/nitrite was detected in 1 of 11 samples and benzoic acid was detected in 1 of 13 samples. Thus, exposure to these chemicals would be limited. While nitrocellulose was detected in three of three samples, it is essentially non-toxic (USEPA 1987), and this chemical is also not expected to be an ecological concern.

40

For Backwater Area sediment, antimony, beryllium, selenium, nitrate/nitrite, TNT, nitrocellulose, and tetryl were identified as integrated COPECs without an ESV. Antimony was detected in 11 of 14 samples. The mean concentration (0.607 mg/kg) is below the SRV (1.3 mg/kg). Sediment

1 concentrations can be compared to the Ohio specific SRVs in conjunction with site-specific 2 background concentrations to determine whether sediment has been potentially impacted by site-3 related activities. Because the mean concentration for antimony is below the SRV, antimony is not 4 considered of ecological concern in sediment. Beryllium was detected in all 14 samples collected; however, the mean concentration was 0.782 mg/kg which is below the SRV of 0.8 mg/kg. As a result, 5 beryllium is not considered of ecological concern for sediment. Selenium was detected in 8 of 14 6 7 samples collected; however, the mean concentration was 0.5 mg/kg, which is below the SRV of 2.1 8 mg/kg, indicating selenium is not of ecological concern in sediment. Nitrate/nitrite was detected in 1 9 of 12 samples, and tetryl was detected in 1 of 14 samples. Thus, exposure to these chemicals would 10 be limited. TNT was detected in 6 of 14 samples, with a mean concentration of 0.0811 mg/kg. The findings of the biological and water quality sampling stations discussed in the previous section 11 12 alleviate concerns about TNT in the Backwater Area. While nitrocellulose was detected in four of 13 four samples, it is essentially non-toxic (USEPA 1987) and not expected to be of ecological concern.

14

15 For Upper Cobbs Pond sediment, antimony; beryllium; hexavalent chromium; selenium; nitrate/nitrite; ammonia; total phosphorous; TNT; 4-amino-2,6-DNT; HMX; nitrocellulose; and tetryl 16 17 were identified as integrated COPECs without an ESV. Antimony was detected in all five discrete 18 samples collected. The mean concentration (0.78 mg/kg) is below the SRV (1.3 mg/kg). Because the mean concentration for antimony is below the SRV, antimony is not considered of ecological concern 19 20 in sediment. Beryllium was detected in all five discrete samples collected; however, the mean 21 concentration was 0.842 mg/kg, which is very close to the SRV of 0.8 mg/kg. As a result, beryllium 22 is not considered of ecological concern for sediment. Selenium was detected in three of five discrete 23 samples; however, the mean concentration was 0.839 mg/kg which is well below the SRV of 2.1 24 mg/kg, indicating selenium is not of ecological concern in sediment. Nitrate/nitrite, ammonia, and 25 total phosphorous were detected in the ISM sediment sample taken at Upper Cobbs Pond in the 2003 26 FWBWQS. Because nitrate/nitrate was only detected in 3 of 24 sediment samples collected across all 27 three EUs, it is not expected to be a concern. Although discrete samples at the AOC were not 28 analyzed for ammonia or phosphorous, the findings of the biological and water quality sampling 29 stations discussed in the previous section alleviate concerns about them in Upper Cobbs Pond. 30 Hexavalent chromium was detected in one of five discrete samples, and TNT was detected in two of five discrete samples. The findings of the biological and water quality sampling stations discussed in 31 32 the previous section alleviate concerns about hexavalent chromium and TNT in Upper Cobbs Pond. 33 4-Amino-2,6-DNT was detected in the one discrete sample collected at Upper Cobbs Pond and was 34 not detected in the ISM sample at Upper Cobbs Pond or in the sediment samples collected at other 35 EUs. 4-Amino-2,6-DNT was only detected in one sample in all three EUs and is not expected to be of 36 ecological concern in sediment. HMX was detected in one of five discrete sediment samples. Because 37 HMX was only detected in 2 of 27 discrete samples collected across all three EUs and was not 38 detected in the ISM sample at either pond, it is not expected to be a concern. Due to the infrequent 39 detection of this chemical, exposure would be limited, and HMX is not expected to be of ecological 40 concern. Nitrocellulose was detected in both discrete samples collected, but is essentially non-toxic 41 (USEPA 1987); this chemical is not expected to be of ecological concern. Tetryl was detected in one 42 of five discrete samples collected. Because tetryl was only detected in 3 of 27 discrete samples

1 collected across all three EUs and was not detected in the ISM sample at either pond, it is not

- 2 expected to be of ecological concern.
- 3

4 For Lower Cobbs Pond sediment, antimony, beryllium, hexavalent chromium, selenium, 5 nitrate/nitrite, ammonia, total phosphorous, HMX, and nitrocellulose were identified as integrated COPECs without an ESV. Antimony was detected in one of eight discrete samples collected. The 6 7 mean concentration (0.328 mg/kg) is below the SRV (1.3 mg/kg), and antimony is not considered an 8 ecological concern in sediment. Beryllium was detected in all eight discrete samples collected; 9 however, the mean concentration was 0.679 mg/kg, which is below the SRV of 0.8 mg/kg. For the 10 ISM sediment sample, the detected concentration was 0.83 mg/kg, which is essentially equal to the SRV. As a result, beryllium is not considered of ecological concern for sediment. Nitrate/nitrite, 11 12 ammonia, and total phosphorous were detected in the ISM sediment sample taken at Lower Cobbs 13 Pond in the FWBWOS. Because nitrate/nitrate was only detected in 3 of 24 sediment samples collected across all three EUs, it is not expected to be a concern. Although discrete samples at the 14 15 AOC were not analyzed for ammonia or phosphorous, the findings of the biological and water quality 16 sampling stations discussed in the previous section alleviate concerns about them in Lower Cobbs 17 Pond. Hexavalent chromium was detected in two of six discrete samples. The findings of the biological and water quality sampling stations discussed in the previous section alleviate concerns 18 19 about hexavalent chromium in Lower Cobbs Pond. Selenium was detected in four of eight discrete 20 samples collected; however, the mean concentration was 0.839 mg/kg, which is below the SRV of 2.1 21 mg/kg and indicates selenium is not of ecological concern in sediment. HMX was detected in one of 22 eight sediment samples. Because HMX was only detected in 2 of 27 discrete samples collected across 23 all three EUs and was not detected in the ISM sample at either pond, it is not expected to be a 24 concern. Nitrocellulose was detected in the one discrete sample collected, but it is essentially non-25 toxic (USEPA 1987) and not expected to be of ecological concern.

26

27 For Backwater Area surface water, sulfate and chloromethane were identified as integrated COPECs 28 without an ESV. For Upper Cobbs Pond and Lower Cobbs Pond, nitrate/nitrite and sulfate were 29 identified as integrated COPECs without an ESV. Nitrate/nitrite was detected in the four samples 30 taken for the FWBWOS at Upper Cobbs Pond and Lower Cobbs Pond but was not detected in the six 31 RI samples taken in the Backwater Area and Upper Cobbs Pond and Lower Cobbs Pond. Between the 32 lack of detection in the RI samples and the findings of the biological and water quality sampling 33 stations discussed in the previous section, nitrate/nitrite is not expected to be an ecological concern. 34 For surface water at the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond, sulfate was 35 detected in six of seven samples. The findings of the biological and water quality sampling stations 36 discussed in the previous section alleviate concerns about sulfate throughout the AOC. 37 Chloromethane was detected in the Backwater Area in one of three samples. Chloromethane was not 38 detected in any of the three samples at the other EUs; based on this information and chloromethane 39 having only one detection in the Backwater Area, it is not expected to be of ecological concern.

1 **Threatened and Endangered Species**

2

3 The Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008) specifies that 4 threatened and endangered species should be evaluated in a Level III Baseline ERA. The Least 5 Bittern (Ixobrychus exilis), a state-threatened species, has been observed at the AOC. This bird, the smallest in the heron family, feeds on insects and fish and has a diet similar to the Great Blue Heron. 6 7 The historical ERA evaluated the Great Blue Heron and other piscivores (kingfisher and mink) for 8 food web exposures to COPECs at the AOC. In particular, PCB-1254 and mercury (the two most 9 likely COPECs to be of bioaccumulative concern) were evaluated. The historical Level III work 10 reported NOAEL HOs for PCB-1254 were less than 1 for avian piscivores (heron and kingfisher; Appendix H, Table H-9), NOAEL HOs were low for mink (1.02 using the average exposure scenario; 11 Appendix H, Table H-9), and LOAEL HQs were less than 1 for all piscivores (Appendix H, Table H-12 13 10). In addition, dietary exposure doses based on mercury did not exceed dietary toxicity benchmarks 14 for the heron, kingfisher, or mink (Appendix H, Tables H-9 and H-10). The sediment concentrations 15 used in the historical ERA (0.042–0.047 mg/kg for PCB-1254 and 0.064-0.18 mg/kg for mercury) are similar to those used in the PBA08 RI ERA (0.032-0.047 mg/kg for PCB-1254 and 0.054-0.15 16 17 mg/kg for mercury); therefore, the doses used in the historical ERA are similar to current doses. Thus, 18 wildlife exposure and risk to PCB-1254 and mercury are similar to the historical findings and are not 19 a current concern at the AOC. Based on the results of the food web modeling in the historical ERA 20 (MKM 2005), another Level III ERA is not recommended for the Least Bittern (*Ixobrychus exilis*).

21

22 Summary of Findings in Step 3A

(aluminum, lead, and vanadium).

23

24 Of the 12 integrated COPECs in surface soil, 3 COPECs (nitrate/nitrite, nitrocellulose, and benzoic 25 acid) were eliminated because they did not have ESVs and had low frequency of detection or little to 26 no toxicity. One PBT compound (beta-BHC) was eliminated because it had a ratio less than one, 27 using an ESV that accounted for bioaccumulation. Additional integrated COPECs were eliminated 28 from further consideration because the mean concentration is smaller than the ESV (arsenic, 29 cadmium, and cobalt) or the mean concentration is smaller than the background concentration

30 31

32 The remaining two integrated COPECs in soil (mercury and zinc) have the following combination of 33 factors that together eliminated them from further consideration.

34

- 35 1. Mean concentrations are only slightly higher than background concentrations (mercury and 36 zinc).
 - 2. Mean concentration-to-ESV ratios of exceedance are near one (zinc).
- 38 3. Ohio EPA guidance allows alternative ESVs that are less conservative than the ESVs used in 39 this ERA. These alternate ESVs (unlike the preferred ESV) are above background 40 concentrations. If these alternate ESVs were used in lieu of the current ones, ratios for 41 mercury and zinc would be below one.
- 42
- 43 No final COPECs were identified for surface soil.

1 Of the 40 integrated COPECs in sediment, 12 were eliminated as COPECs because they did not have 2 ESVs and (1) they had mean concentrations below or slightly above the SRV (antimony, beryllium, 3 and selenium), (2) they had low frequency of detection (nitrate/nitrite; 4-amino-2,6-DNT; HMX; and 4 tetryl), (3) they are essentially non-toxic (nitrocellulose), or (4) the findings of the biological and water quality sampling stations alleviated concerns (ammonia, total phosphorous, hexavalent 5 chromium, and TNT). Mercury and PCB-1254 were identified as PBT chemicals in sediment. 6 7 Mercury was detected above the background concentration but below the SRV. Both PCB-1254 and 8 mercury were evaluated in the food chain during the historical ERA (MKM 2005) and determined to 9 not be a concern. For the 26 remaining integrated COPECs, most were eliminated because (1) the 10 mean concentration was less than the ESV, (2) the mean concentration was less than their background 11 concentrations and/or SRVs, (3) the frequency of detection was low, or (4) the frequency of detection 12 was low in combination with a low mean to ESV or low mean to background and/or SRV ratio. After 13 applying these factors, only silver, benz(a)anthracene, benzo(a)pyrene, bis(2-ethylhexyl)phthalate, di-14 n-butyl phthalate, chrysene, pyrene, and acetone remained as integrated COPECs in sediment. The 15 findings of the biological and water quality sampling stations alleviated concerns about these 16 remaining integrated COPECs.

17

18 No final COPECs were identified for sediment for the Backwater Area, Upper Cobbs Pond, or Lower19 Cobbs Pond.

20

21 Of the eight integrated COPECs in surface water throughout all three EUs, three were eliminated as 22 COPECs because they did not have ESVs and either had low frequency of detection (chloromethane), 23 or the findings of the biological and water quality sampling stations alleviated concerns about their 24 detection throughout the AOC (nitrate/nitrite and sulfate). No PBT compounds were detected in 25 surface water. Barium, selenium, and silver were eliminated from further consideration because the 26 mean concentration is smaller than the ESV. Manganese was eliminated in two of the three EUs 27 because its mean was below its background concentration. Sulfide at Upper Cobbs Pond, Lower 28 Cobbs Pond, and the Backwater Area and manganese in the Backwater Area were eliminated based 29 on the biological and water quality station sampling results.

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No final COPECs in surface water were identified for the Backwater Area, Upper Cobbs Pond, orLower Cobbs Pond.

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34 **7.3.3.8** Consideration of Human Health Driven Remediation

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The HHRA does not identify any COCs and recommends no further action for soil, sediment, or surface water at the AOC. Thus, there would be no remediation implemented to further protect human health that would also protect ecological resources.

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40 7.3.3.9 <u>Uncertainties and Mitigations</u>

- 42 Uncertainties or unknowns are present in both exposure data and effects data. To mitigate uncertainty
- 43 in exposure data, the MDCs of all available and appropriate data were used in Level I. In Level II, the

1 MDC and mean COPEC concentrations were used to mitigate uncertainty concerning exposure data 2 for receptors in the AOC. To mitigate uncertainty in effects information, a site visit for habitat 3 condition was conducted, and the latest INRMP of state- or federally-listed species sightings and 4 jurisdictional wetlands was used (OHARNG 2014). In addition, the ORAM was applied to the wetland complexes at the AOC. Conservative ESVs, which are typically based on concentrations 5 observed to have no effect on test species in laboratory studies, were used in Level II to mitigate 6 7 uncertainty concerning effects on receptors in the AOC. Some chemicals are COPECs because they 8 do not have ESVs. Even though all individual PAHs were eliminated as final COPECs, they 9 contribute to the combined (joint action) toxicity of PAHs on ecological receptors. As a result, the 10 average concentrations of all integrated PAH COPECs in Backwater Area sediment (the only EU and 11 medium with PAH COPECs) were summed (3.3 mg/kg) and compared to the total PAH ESV in 12 sediment (1.61 mg/kg), resulting in a ratio of 2.05. While the magnitude of the ratio suggests a limited 13 potential for concern to ecological receptors, there is other evidence that mitigates this concern. 14 Specifically, Wetland 3 (Category 3) is the highest quality wetland and is associated with the 15 Backwater Area. The high ORAM score suggests total PAH effects on wetland quality are minimal. 16

In Level II, to mitigate uncertainty concerning effects on receptors in the AOC, the ESVs for COPECs are compared to background concentrations. Using ESVs that are lower than background concentrations indicates the conservative nature of the evaluation. Conservative ESVs are appropriate for use as screening thresholds in Level I and II (i.e., soil constituents with an MDC below the ESV need no further consideration in Level II).

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7.3.3.10 <u>Summary and Recommendations of the Ecological Risk Assessment</u>

COPECs were identified in bank soil, sediment, and surface water at the Upper and Lower Cobbs
Ponds AOC. Those chemicals retained after screening historical and PBA08 RI data and data from
the FWBWQS were termed integrated COPECs.

28

A total of 12 integrated soil COPECs, 40 integrated sediment COPECs, and 8 integrated surface water COPECs were further evaluated in Step 3A with technical and refinement factors. All integrated soil, sediment, and surface water COPECs were determined to be of no ecological concern—none require remediation or further evaluation. Consequently, the ERA for the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond can conclude with a Level II that no further action is necessary to be protective of important ecological resources.

35

36 7.3.4 Conclusions

37

There is chemical contamination present in soil, sediment, and surface water at the Upper and Lower Cobbs Ponds AOC. This contamination was identified using historical and PBA08 RI data and data from the FWBWQS. The terrestrial vegetation habitat at the AOC consists of four forest community types, one shrub community type, and three herbaceous community types. The dominant forest community at the AOC is the red maple (*A. rubrum*) successional forest. Upper Cobbs Pond, Lower Cobbs Pond, and the wetlands are important/significant ecological resources. In addition, the Least 1 Bittern (Ixobrychus exilis), which is a state of Ohio threatened species, uses the AOC. The Least

2 Bittern (Ixobrychus exilis) and its habitat at the AOC are important/significant ecological resources.

3 These findings invoked a Level II assessment.

4

5 The Level II assessment evaluated soil, sediment, and surface water using historical and PBA08 RI data, identified integrated COPECs, and evaluated the integrated COPECs using technical and 6 7 refinement factors in Step 3A. The factors in Step 3A showed there is no further evaluation necessary 8 for integrated COPECs, and there is no ecological concern requiring remediation. Consequently, the 9 ERA for the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond can conclude with a Level 10 II ERA that no further action is necessary to be protective of important ecological resources. This supports the conclusion from the Phase II RI, which states: 11

12

13 "No additional action is recommended based on ecological risk because, based on the

Т

- 14
- 15
- 16
- 17

| Location | Sample ID | Date | Depth (ft bgs) |
|------------|--------------------|-----------|----------------|
| CPCsb-030 | CPCsb-030-5105-SO | 3/29/2010 | 0 - 1 |
| CPCsb-031 | CPCsb-031-5109-SO | 3/24/2010 | 0 - 1 |
| CPCsb-032 | CPCsb-032-5113-SO | 3/24/2010 | 0 - 1 |
| CPCsb-034 | CPCsb-034-5119-SO | 3/29/2010 | 0 - 1 |
| CPCsb-035 | CPCsb-035-5123-SO | 3/29/2010 | 0 - 1 |
| CPCss-036 | CPCss-036-5014-SO | 2/23/2010 | 0 - 1 |
| CPCss-037 | CPCss-037-5015-SO | 2/23/2010 | 0 - 1 |
| CPCss-038 | CPCss-038-5016-SO | 2/23/2010 | 0 - 1 |
| CPCss-039 | CPCss-039-5017-SO | 2/23/2010 | 0 - 1 |
| CPCss-040 | CPCss-040-5018-SO | 2/23/2010 | 0 - 1 |
| CPCss-041 | CPCss-041-5019-SO | 2/23/2010 | 0 - 1 |
| CPCss-042 | CPCss-042-5020-SO | 2/23/2010 | 0 - 1 |
| CPCss-043 | CPCss-043-5021-SO | 2/23/2010 | 0 - 1 |
| ULCPsd-010 | ULCPsd-010-0001-SD | 8/21/2001 | 0 - 0.5 |
| ULCPss-001 | ULCPss-001-0001-SO | 7/17/2001 | 0 - 1 |
| ULCPss-002 | ULCPss-002-0001-SO | 7/17/2001 | 0 - 1 |
| ULCPss-003 | ULCPss-003-0001-SO | 7/18/2001 | 0 - 1 |
| ULCPss-004 | ULCPss-004-0001-SO | 7/17/2001 | 0 - 1 |
| ULCPss-005 | ULCPss-005-0001-SO | 7/18/2001 | 0 - 1 |
| ULCPss-006 | ULCPss-006-0001-SO | 7/19/2001 | 0 - 1 |
| ULCPss-007 | ULCPss-007-0001-SO | 7/18/2001 | 0 - 1 |
| ULCPss-008 | ULCPss-008-0001-SO | 7/18/2001 | 0 - 1 |
| ULCPss-009 | ULCPss-009-0001-SO | 7/19/2001 | 0 - 1 |
| ULCPss-010 | ULCPss-010-0001-SO | 7/19/2001 | 0 - 1 |

Table 7–1. Risk Assessment Data Set for Surface Soil (0–1 ft bgs) Discrete Samples Т

Step 3a refinement step, it was determined that site-related risks were not great

enough to warrant proceeding further into the ecological risk assessment process."

bgs = Below ground surface.

ft = Feet.

ID = Identification.

Table 7–2. Risk Assessment Data Set for Subsurface Soil Discrete Samples

| Location | Sample ID | Date | Depth (ft bgs) |
|-----------|-------------------|-----------|----------------|
| CPCsb-032 | CPCsb-032-5114-SO | 3/24/2010 | 1 - 4 |
| CPCsb-034 | CPCsb-034-5120-SO | 3/29/2010 | 1 – 4 |
| CPCsb-035 | CPCsb-035-5124-SO | 3/29/2010 | 1 – 4 |
| CPCsb-032 | CPCsb-032-5115-SO | 3/24/2010 | 4 - 7 |
| CPCsb-035 | CPCsb-035-5125-SO | 3/29/2010 | 4 - 7 |
| CPCsb-032 | CPCsb-032-5116-SO | 3/24/2010 | 7 – 10 |

bgs = Below ground surface. ft = Feet. ID = Identification.

| 2 |
|---|
| 3 |

| Table | 7–3. Risk | Assessment Data | Set for | Sediment |
|-------|-----------|------------------------|---------|----------|
| | | | | |

| Location | Location Sample ID Date | | Depth (ft bgs) |
|----------------|--------------------------|-------------|----------------|
| | Backwater Area – Discre | te Samples | |
| CPCsd-047 | CPCsd-047-5025-SD | 4/1/2010 | 0 - 0.5 |
| CPCsd-048 | CPCsd-048-5026-SD | 4/1/2010 | 0 - 0.5 |
| ULCPsd-004 | ULCPsd-004-0001-SD | 8/22/2001 | 0 - 1 |
| ULCPsd-005 | ULCPsd-005-0001-SD | 8/22/2001 | 0 - 1 |
| ULCPsd-006 | ULCPsd-006-0001-SD | 8/22/2001 | 0 - 1 |
| ULCPsd-007 | ULCPsd-007-0001-SD | 8/21/2001 | 0 - 0.75 |
| ULCPsd-008 | ULCPsd-008-0001-SD | 8/21/2001 | 0 - 0.5 |
| ULCPsd-009 | ULCPsd-009-0001-SD | 8/21/2001 | 0 - 0.85 |
| ULCPsd-011 | ULCPsd-011-0001-SD | 8/20/2001 | 0 - 0.75 |
| ULCPsd-012 | ULCPsd-012-0001-SD | 8/20/2001 | 0 – 1.75 |
| ULCPsd-013 | ULCPsd-013-0001-SD | 8/20/2001 | 0 – 1.75 |
| ULCPsd-014 | ULCPsd-014-0001-SD | 7/25/2001 | 0 - 0.5 |
| ULCPsd-015 | ULCPsd-015-0001-SD | 8/17/2001 | 0 - 0.5 |
| ULCPsd-015 | ULCPsd-015-0002-SD | 9/27/2001 | 0 - 1 |
| ULCPsd-016 | ULCPsd-016-0001-SD | 8/17/2001 | 0 – 1.5 |
| | Upper Cobbs Pond – Discr | ete Samples | |
| CPCsd-046 | CPCsd-046-5024-SD | 3/25/2010 | 0 - 0.5 |
| CPCsd-049 | CPCsd-049-5032-SD | 3/25/2010 | 0 - 0.5 |
| ULCPsd-017 | ULCPsd-017-0001-SD | 8/17/2001 | 0 - 1 |
| ULCPsd-018 | ULCPsd-018-0001-SD | 8/16/2001 | 0 - 0.5 |
| ULCPsd-019 | ULCPsd-019-0001-SD | 8/16/2001 | 0 - 0.5 |
| ULCPsd-020 | ULCPsd-020-0001-SD | 8/16/2001 | 0 - 0.75 |
| | Upper Cobbs Pond – ISN | | |
| Uppercobb Pond | FSW-SD-030-0000 | 6/23/2003 | 0 - 0.25 |
| | Lower Cobbs Pond – Discr | ete Samples | |
| CPCsd-044 | CPCsd-044-5022-SD | 3/29/2010 | 0 - 0.5 |
| CPCsd-045 | CPCsd-045-5023-SD | 4/1/2010 | 0 - 0.5 |
| ULCPsd-021 | ULCPsd-021-0001-SD | 8/15/2001 | 0 - 1 |
| ULCPsd-021 | ULCPsd-021-0002-SD | 9/27/2001 | 0 - 1 |
| ULCPsd-022 | ULCPsd-022-0001-SD | 8/15/2001 | 0 – 0.75 |
| ULCPsd-023 | ULCPsd-023-0001-SD | 8/15/2001 | 0 - 0.75 |
| ULCPsd-024 | ULCPsd-024-0001-SD | 8/14/2001 | 0 - 0.5 |
| ULCPsd-025 | ULCPsd-025-0001-SD | 7/25/2001 | 0 - 0.5 |
| ULCPsd-026 | ULCPsd-026-0001-SD | 8/14/2001 | 0 - 1 |

Table 7-3. Risk Assessment Data Set for Sediment (continued)

| Sample ID | Date | Depth (ft bgs) | | | | | |
|--------------------------------|------------------------|--------------------------------|--|--|--|--|--|
| Lower Cobbs Pond – ISM Samples | | | | | | | |
| FSW-SD-031-0000 | 6/23/2003 | 0 - 0.25 | | | | | |
| | Lower Cobbs Pond – ISA | Lower Cobbs Pond – ISM Samples | | | | | |

bgs = Below ground surface.

ft = Feet.

ID = Identification.

ISM = Incremental sample method.

2 3

| Table 7- | –4. Risk | Assessment | Data | Set for | Surface | Water |
|----------|----------|------------|------|---------|---------|-------|
|----------|----------|------------|------|---------|---------|-------|

| Location | Sample ID | Date |
|---------------------|--------------------|-----------|
| | Backwater Area | |
| CPCsw-047 | CPCsw-047-5030-SW | 4/1/2010 |
| CPCsw-048 | CPCsw-048-5031-SW | 4/1/2010 |
| ULCPsw-001 | ULCPsw-001-0001-SW | 7/26/2001 |
| ULCPsw-002 | ULCPsw-002-0001-SW | 7/25/2001 |
| | Upper Cobbs Pond | |
| CPCsw-046 | CPCsw-046-5029-SW | 3/25/2010 |
| Uppercobb Pond | FSW-SW-030-0000 | 6/23/2003 |
| Uppercobb Pond | FSW-SW-070-0000 | 6/23/2003 |
| ULCPsw-003 | ULCPsw-003-0001-SW | 7/24/2001 |
| ULCPsw-004 | ULCPsw-004-0001-SW | 7/24/2001 |
| | Lower Cobbs Pond | |
| CPCsw-044 | CPCsw-044-5027-SW | 3/29/2010 |
| CPCsw-045 | CPCsw-045-5028-SW | 4/1/2010 |
| Lowcobb Pond | FSW-SW-031-0000 | 6/23/2003 |
| Lowcobb Pond | FSW-SW-071-0000 | 6/23/2003 |
| ULCPsw-005 | ULCPsw-005-0001-SW | 7/24/2001 |
| ULCPsw-006 | ULCPsw-006-0001-SW | 7/24/2001 |
| ULCPsw-007 | ULCPsw-007-0001-SW | 7/24/2001 |
| D = Identification. | | |

4

Table 7–5. Summary of SRCs

| | Surface Soil | Subsurface Soil | Surface Water | | | Sediment | | |
|----------------------|---------------------|-----------------|---------------|---|----|----------|-----|---|
| SRC | (0–1 ft bgs) | (1-13 ft bgs) | BW UCP LCP | | BW | UCP | LCP | |
| | Inorganic Chemicals | | | | | | | |
| Aluminum | Х | | | | | Х | Х | Х |
| Antimony | | | Х | Х | Х | Х | Х | X |
| Arsenic | Х | | Х | | | | | Х |
| Barium | Х | | Х | | | Х | Х | Х |
| Beryllium | | | Х | | Х | Х | Х | X |
| Cadmium | Х | Х | Х | | Х | Х | Х | X |
| Chromium | Х | | Х | Х | | Х | Х | X |
| Chromium, hexavalent | Х | | | | | | Х | X |
| Cobalt | Х | | Х | | Х | Х | Х | X |
| Copper | Х | | Х | | | Х | Х | X |
| Cyanide | Х | | | | | Х | | X |
| Lead | Х | Х | Х | Х | Х | Х | Х | X |
| Manganese | | | Х | Х | Х | | | |
| Mercury | Х | | | | | Х | Х | Х |
| Nickel | Х | | Х | Х | Х | Х | Х | Х |

Table 7–5. Summary of SRCs (continued)

| | Surface Soil | Subsurface Soil | Sur | face W | ater | Sediment | | |
|----------------------------|---------------|------------------|------------|--------|------|----------|-----|----------|
| SRC | (0–1 ft bgs) | (1–13 ft bgs) | BW UCP LCP | | | BW | LCP | |
| Nitrate/Nitrite | X | | | Х | Х | Х | | |
| Selenium | Х | X | Х | Х | Х | Х | Х | Х |
| Silver | Х | X | Х | | | Х | Х | X |
| Sulfate | | | Х | Х | Х | | | |
| Sulfide | | | Х | Х | Х | | | |
| Thallium | Х | | Х | | Х | | | |
| Vanadium | Х | | Х | Х | Х | Х | Х | Х |
| Zinc | Х | | Х | | | | | |
| | I | Explosives | | | | | | <u> </u> |
| 4-Amino-2,6-Dinitrotoluene | | | Х | Х | | | Х | |
| 1,3-Dinitrobenzene | | | | | | | Х | |
| 2,6-Dinitrotoluene | | | | | | Х | | |
| 2,4,6-Trinitrotoluene | | | | | | Х | Х | |
| HMX | | | | | | | Х | Х |
| Nitrocellulose | Х | X | | | | Х | X | X |
| Tetryl | | | | | | X | | X |
| | Pe | sticide/PCB | | l | | | | |
| beta-BHC | Х | | | | | | | |
| delta-BHC | | | | | | | | |
| PCB-1254 | | | | | | Х | Х | |
| | Semi-volatile | organic Compound | ds | l | | | | L |
| 2-Methylnaphthalene | Х | X | | | | | | |
| Acenaphthene | | | | | | | | |
| Acenaphthylene | Х | | | | | | | |
| Anthracene | | | | | | Х | | |
| Benz(a)anthracene | Х | Х | | | | Х | Х | Х |
| Benzenemethanol | Х | | | | Х | | | |
| Benzo(a)pyrene | Х | Х | | | | Х | Х | Х |
| Benzo(b)fluoranthene | Х | Х | | | | Х | Х | Х |
| Benzo(ghi)perylene | Х | Х | | | | Х | Х | Х |
| Benzo(k)fluoranthene | Х | Х | | | | Х | Х | Х |
| Benzoic acid | Х | | | | | | | |
| Bis(2-ethylhexyl)phthalate | Х | | | Х | | Х | | |
| Butyl benzyl phthalate | | | | | Х | | | |
| Chrysene | Х | X | | | | Х | Х | Х |
| Dibenz(a,h)anthracene | | X | | | | X | | |
| Di-n-butyl phthalate | Х | X | | | | | | X |
| Fluoranthene | X | X | | | | Х | Х | X |
| Fluorene | | | | | | X | | |
| Indeno(1,2,3-cd)pyrene | Х | X | | | | X | Х | X |
| Naphthalene | | | | | | | | |
| Phenanthrene | Х | X | | | | Х | Х | Х |
| Pyrene | X | X | | | | X | X | X |

Table 7–5. Summary of SRCs (continued)

| | Surface Soil | Subsurface Soil | Surface Water | | | Sediment | | | |
|--------------------|----------------------------|-----------------|---------------|-----|-----|----------|-----|-----|--|
| SRC | (0–1 ft bgs) | (1-13 ft bgs) | BW | UCP | LCP | BW | UCP | LCP | |
| | Volatile Organic Compounds | | | | | | | | |
| 2-Butanone | | | | | | Х | Х | Х | |
| Acetone | | | Х | | | Х | Х | Х | |
| Carbon disulfide | | | | | | | | Х | |
| Chloromethane | | | Х | | | | | | |
| Methylene chloride | | | | | | Х | | | |
| Toluene | | | Х | | | | | | |

Exposure Units:

BW = Backwater.

LCP = Lower Cobbs Pond.

UCP = Upper Cobbs Pond.bgs = Below ground surface.

BHC = Hexachlorocy clohexane.

EU = Exposure unit.

ft = Feet.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

X = Chemical is an SRC at this EU.

--= Chemical is not an SRC at this EU.

Table 7-6. Summary of COPCs

| | Surface Soil | Subsurface Soil | Surface Water | | | | Sedimen | t | |
|------------------------------|--------------|--------------------|---------------|--------|-----|----|---------|-----|--|
| COPC | (0-1 ft bgs) | (1–13 ft bgs) | BW | UCP | LCP | BW | UCP | LCP | |
| Inorganic Chemicals | | | | | | | | | |
| Aluminum | Х | | - | - | | Х | Х | Х | |
| Ammonia ^a | | | - | - | | | Х | Х | |
| Arsenic | Х | | Х | - | | | | Х | |
| Chromium | Х | | - | | | Х | Х | Х | |
| Chromium, hexavalent | | | | | | | Х | Х | |
| Cobalt | Х | | Х | | | Х | Х | Х | |
| Cyanide | Х | | | | | Х | | Х | |
| Manganese | | | Х | | Х | | | | |
| Nitrate/Nitrite ^a | | | | Х | Х | | | | |
| Selenium | | | Х | | | | | | |
| Sulfate ^a | | | Х | Х | Х | | | | |
| Sulfide ^a | | | Х | Х | Х | | | | |
| | S | Semi-volatile Orga | nic Com | pounds | | | | | |
| Benz(a)anthracene | | | | | | Х | | | |
| Benzo(a)pyrene | Х | Х | | | | Х | Х | Х | |
| Benzo(b)fluoranthene | | | | | | Х | | | |
| Dibenz(a,h)anthracene | | | | | | Х | | | |
| Indeno(1,2,3-cd)pyrene | | | | | | X | | | |

^aAnalyte identified as COPC because no screening criteria are available, these parameters are reported as water quality parameters for protection of aquatic organisms and have very low toxicity to humans.

Exposure Units:

BW = Backwater.

LCP = Lower Cobbs Pond.

UCP = Upper Cobbs Pond.

bgs = Below ground surface.

COPC = Chemical of potential concern.

ft = Feet.

X = Chemical is a COPC at this depth interval.

-- = Chemical is not a COPC in this exposure unit.

| | | Screening Level | | |
|---------------------|----------------------------|-----------------|--------------------------|--|
| | Critical Effect or | Resident Rec | eptor ^a FWCUG | |
| COPC | Target Organ | HQ=1 | TR=1E-05 | |
| Aluminum | Neurotoxicity in offspring | 73,798 | | |
| Arsenic | Skin | 20.2 | 4.25 ^b | |
| Chromium, trivalent | NOAEL | 81,473 | | |
| Cobalt | NS | 1,313 | 8,030 | |
| Cyanide | Reproductive | 2.7 | | |
| Benzo(a)pyrene | NA | | 0.221 | |

Table 7–7. Screening Levels for Soil Corresponding to an HQ of 1 and TR of 1E-05

All units are milligrams/kilogram (mg/kg).

^a Resident Receptor FWCUGs are the smaller of the Adult or Child values for each COPC and endpoint (non-cancer and cancer).

^b Risk-based FWCUG is less than the background screening values for arsenic in surface soil (15.4 mg/kg) and subsurface soil (19.8 mg/kg).

COPC = Chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

HQ = Hazard quotient.

NA = Not applicable.

NOAEL = No observed adverse effect level.

NS = Not specified. TR = Target risk. -- = No value available.

Table 7-8. Screening Levels for Sediment Corresponding to an HQ of 1 and TR of 1E-05

| | | Screening Level | | |
|------------------------|----------------------------|--------------------------------------|----------|--|
| | Critical Effect or | Resident Receptor ^a FWCUG | | |
| COPC | Target Organ | HQ=1 | TR=1E-05 | |
| Aluminum | Neurotoxicity in offspring | 73,798 | | |
| Arsenic | Skin | 20.2 4.25 ^b | | |
| Chromium, trivalent | NOAEL | 81,473 | | |
| Chromium, hexavalent | Stomach, liver/kidney | 199 268 ^d | | |
| Cobalt ^c | NS | 23 4200 | | |
| Cyanide | Reproductive | 2.7 | | |
| Benz(a)anthracene | NA | | 2.21 | |
| Benzo(a)pyrene | NA | | 0.221 | |
| Benzo(b)fluoranthene | NA | 2.21 | | |
| Dibenz(a,h)anthracene | NA | - 0.221 | | |
| Indeno(1,2,3-cd)pyrene | NA | - 2.21 | | |

COPC = Chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

HQ = Hazard quotient.

NA = Not applicable.

 $NOAEL = \hat{No}$ observed adverse effect level.

NS = Not specified.

TR = Target risk.

-- = No value available.

| | Critical Effect | Screening Level | |
|-----------------------|------------------------|-------------------------------------|----------|
| | or | Resident Receptor ^a FWCU | |
| COPC | Target Organ | HQ=1 | TR=1E-05 |
| Arsenic | Skin | 0.0463 | 0.011 |
| Cobalt ^b | NS | 0.006 | |
| Manganese | CNS | 6.326 | |
| Selenium ^b | | 0.1 | |

 Table 7–9. Screening Levels for Surface Water Corresponding to an HQ of 1 and TR of 1E-05

CNS = Central nervous system.

COPC = Chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

HQ = Hazard quotient.

NS = Not specified.

TR = Target risk.

-- = No value available.

| Table | 7–10. Concentrations | of Benzo(a)pyrene | in Soil from | Various | Environmental | Studies |
|-------|----------------------|-------------------|--------------|---------|----------------------|---------|
| | | | | | | |

| | Number of | Geometric Mean or | | 95 th | |
|---|-----------|----------------------|---------|-------------------------|---------|
| Study | Samples | Median | Minimum | Percentile ^a | Maximum |
| CA/T Project ^b | 873 | 0.3 | 0.031 | 17 | 230 |
| LSPA Project ^b | 489 | 0.44 | ND | | 222 |
| Watertown ^b | 17 | 0.95 | 0.6 | 4.77 | 6.08 |
| Worcester ^b | 67 | | ND | 3.3 | 9.7 |
| New England ^c | 62 | 0.686 | ND | 1.82 | 13 |
| Illinois ^d Urban | | | | 2.1 | |
| Illinois ^d Rural | | | | 0.98 | |
| ATSDR ^e Urban | | | 0.165 | | 0.22 |
| ATSDR ^e Rural | | | 0.002 | | 1.3 |
| ATSDR ^e Agricultural | | | 0.0046 | | 0.9 |
| NYSDEC Rural Near Roads ^f | 28 | | ND | 1.1 | 2.4 |
| NYSDEC Rural Distant Roads ^f | 118 | | ND | 0.12 | 3.4 |

^a Lognormal 95th percentile value for all studies except: (1) New England value is 95% upper confidence limit and (2) NYSDEC values are distribution-free 95th percentile.

^b Data reported by Massachusetts Department of Environmental Protection (MADEP 2002) are from the following data sets: CA/T = Data collected by Mass Highway Department as part of the Central Artery/Tunnel (CA/T) project.

LSPA = Preliminary data compiled by the Massachusetts Licensed Site professional Association (LSPA) from data submitted by its members in 2001.

Water Town and Worcester Site-specific samples.

^c Data from three New England locations from Bradley et al. 1994.

^d Concentrations of polycyclic aromatic hydrocarbons (PAHs) in Illinois metropolitan statistical areas (urban) and nonmetropolitan statistical areas (rural) as reported by Illinois Environmental Protection Agency (IEPA 2005).

^e Data published by the Agency for Toxic Substances and Disease Registry (ATSDR) in PAHs August 1995.

^f Distribution-free 95th percentile values for near roads (less than 10 ft from roads and pavement) and not near roads (more than 15 ft from roads and pavement) from New York State Brownfield Cleanup program Development of Soil Cleanup Objectives Technical Support Document (September 2006), Appendix D.

ND = Not detected.

NYSDEC = New York State Department of Environmental Conservation.

-- = No value reported for this source.

| Group | COPEC | Shallow Soil | Sediment | Surface Water |
|------------|------------------------|-----------------|----------|------------------|
| | Aluminum | Х | | |
| | Antimony | | Q | |
| | Arsenic | Х | X | |
| | Beryllium | | Q | |
| | Cadmium | Х | X | |
| | Chromium | Х | Х | |
| | Copper | Х | Х | |
| | Cyanide | | Х | |
| | Iron | Х | | Х |
| Inorganic | Lead | | Х | |
| Chemicals | Manganese | | | Q |
| | Mercury | Х | Q | |
| | Nickel | Х | X | |
| | Selenium | | | Х |
| | Silver | Х | Х | Х |
| | Vanadium | Х | | |
| | Zinc | Х | | |
| | Nitrate | Q | Q | |
| | Sulfate | | | Q |
| | Sulfide | | | Х |
| PCB | Aroclor-1254 | | Q | |
| | 2,6-Dinitrotoluene | | Х | |
| | Anthracene | | Х | |
| | Benz(a)anthracene | | Х | |
| | Benzo(a)pyrene | | Х | |
| | Benzo(ghi)perylene | | Х | |
| SVOCs | Benzo(k)fluoranthene | | Х | |
| 57003 | Chrysene | | Х | |
| | Fluoranthene | | Х | |
| | Indeno(1,2,3-cd)pyrene | | Х | |
| | Phenanthrene | | Х | |
| | Pyrene | | Х | |
| | Total PAHs | | Х | |
| VOCs | Acetone | | | Q |
| | Chloromethane | | | Q |
| Explosives | 2,4,6-Trinitrotoluene | | Q | |
| - | Nitrocellulose | Q | Q | |

Adapted from Tables 7-1, 7-2, and 7-3 from the Phase II Remedial Investigation Report for Upper and Lower Cobbs Ponds (MKM 2005)

COPEC = Chemical of potential ecological concern.

PCB = Polychlorinated biphenyl.

SVOC = Semi-volatile organic compound.

VOC = Volatile organic compound.

--= Chemical not identified as a COPEC in this data set.

Q = Qualitative COPEC; persistent, bioaccumulative, and toxic chemical; or no ecological screening value (ESV).

X = Quantitative COPEC, exceeds ESV.

| Wetland | | Total Wetland | | |
|------------------------|-----------------------------|------------------------------------|-------------------|----------------------|
| ID | Wetland Type ¹ | Size | ORAM Score | ORAM Category |
| Wetland 1 ² | PFO1E | 0.34 acres (0.17 acres inside AOC) | 34 | 2 |
| Wetland 2 ³ | PUBH/PEM1F/PSS1F/PFO1E | 4.75 acres (3.60 acres inside AOC) | 55 | 2 |
| Wetland 3 ⁴ | PUBH/PABH/PEM1F/PSS1F/PFO1E | 4.62 acres (3.90 acres inside AOC) | 61 | 3 |

Table 7-12. Wetlands Associated with the Upper and Lower Cobbs Ponds AOC

¹Cowardin Classification: PUBH = palustrine, unconsolidated bottom, permanently flooded; PABH = palustrine, aquatic bed, permanently flooded; PEM 1E = palustrine persistent emergent, seasonally flooded/saturated; PEM 1F = palustrine persistent emergent, semi-permanently flooded; PSS1E = palustrine, broad-leaved deciduous, scrub-shrub, seasonally flooded; PSS1F = palustrine, broad-leaved deciduous, scrub-shrub, semi-permanently flooded; PFO1E = palustrine, broad -leaved deciduous, forested, seasonally flooded/saturated.

²Solitary wetland.

³Wetland complex between Track 33 and Track FA.

⁴Wetland complex south of Track FA.

AOC = Area of concern. ID = Identification.

ORAM = Ohio Rapid Assessment Method.

| Natural Resource | Inside Habitat Area | Near the AOC | Distances to Nearest Resources and Comments ^a |
|---|---|--|---|
| Wetlands (planning level survey and jurisdictional) | No jurisdictional wetlands in AOC. Several small- to moderate- size planning level wetlands associated with Lower Cobbs Pond spillway, shoreline fringe, and shallow embayments and shoreline fringe of Backwater Area | Jurisdictional and planning level wetlands along northern, eastern, southern, and western border. ORAM Category 2 wetland complex on Load Line 12 drainage ditch, southwest of AOC | Numerous other wetlands are in vicinity of AOC (Figure 7-2) |
| State-listed or federally- listed species ^b | Least Bittern in Backwater Area | Yellow-bellied Sapsucker and Butternut ^b | <500 ft to the south- southwest and ~1,700 ft to the north-northwest (Yellow-bellied Sapsucker) ~2,200 ft to the north (Butternut) |
| Beaver dams One at Track FA culvert in Backwater Area | | Two: one on ditch from Load Line 12 and one west of Paris-Windham Road | <200 ft west and <700 ft southwest |
| 100-year floodplain | Floodplain along unnamed tributaries to the south and west | butaries to the south and northwest | |
| Stream sampling ^c | None | Stream sampling station <900 ft to the northwest on unnamed tributary to Sand Creek (S-8) | Additional stream sampling station approximately 4,000 ft upstream of the AOC (S- 9) |
| Pond sampling ^c | Upper Cobbs Pond and Lower Cobbs Pond | None | None |

^a Measurements of distance and direction are taken from the nearest boundary of the AOC to the resource being measured.

^bThe butternut is a federal species of concern, which is not an officially listed status with legal protection.

^cStream and pond sampling refers to Facility-wide Biological and Water Quality Study (USACE 2005a).

AOC = Area of concern.

FA = Functional area.

ft = Feet.

ORAM = Ohio Rapid Assessment Method.

| COPEC | MDC (mg/kg) | ESV (mg/kg) | Ratio of MDC to ESV | Comments |
|-----------------|-------------|----------------|------------------------|---|
| Aluminum | 19200 | 50 | 384 | Highest ratio at 384x |
| Arsenic | 28.4 | 18 | 1.58 | None |
| Cadmium | 0.49 | 0.36 | 1.36 | None |
| Cobalt | 16.4 | 13 | 1.26 | None |
| Lead | 31.3 | 11 | 2.85 | None |
| Mercury | 0.074 | 0.00051 | 145.10 | Second highest ratio at about 145x, PBT compound |
| Vanadium | 35.3 | 7.8 | 4.53 | None |
| Zinc | 121 | 46 | 2.63 | None |
| Nitrate/Nitrite | 1.4 | No ESV | | None |
| Nitrocellulose | 1.5 | No ESV | | None |
| Benzoic acid | 17 | No ESV | | None |
| beta-BHC | 0.0035 | 0.00398 | 0.88 | PBT compound |

Table 7-14. Summary of Integrated COPECs for Surface Soil

Table excludes nutrients.

BHC = Hexachlorocyclohexane.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.MDC = Maximum detected concentration.

mg/kg= Milligrams per kilogram.

PBT = Persistent, bioaccumulative, and toxic.

x = M ultiplier.

| | | ESV | Ratio of | |
|--------------------------------------|-------------|---------|------------|-----------------------------|
| COPEC | MDC (mg/kg) | (mg/kg) | MDC to ESV | Comments |
| Antimony | 2.1 | No ESV | | Exceeded SRV |
| Beryllium | 1.1 | No ESV | | Exceeded SRV |
| Cadmium | 2.3 | 0.99 | 2.32 | None |
| Chromium | 56.5 | 43.4 | 1.30 | None |
| Copper | 62.3 | 31.6 | 1.97 | Exceeded ESV and SRV |
| Cyanide | 0.55 | 0.0001 | 5500 | Highest ratio at 5500x |
| Lead | 57.9 | 35.8 | 1.62 | Exceeded ESV and SRV |
| Mercury | 0.11 | 0.18 | 0.61 | PBT Compound |
| Nickel | 35.5 | 22.7 | 1.56 | Exceeded ESV and SRV |
| Selenium | 2.7 | No ESV | | None |
| Silver | 23 | 0.5 | 46 | Second highest ratio at 46x |
| Nitrate/Nitrite | 2.2 | No ESV | | None |
| TNT | 0.32 | No ESV | | None |
| 2,6-Dinitrotoluene | 0.16 | 0.0398 | 4.02 | None |
| Nitrocellulose | 10.4 | No ESV | | None |
| Tetryl | 0.024 | No ESV | | None |
| Anthracene | 0.22 | 0.0572 | 3.85 | None |
| Benz(a)anthracene | 0.83 | 0.108 | 7.69 | None |
| Benzo(a)pyrene | 0.89 | 0.15 | 5.93 | None |
| Benzo(ghi)perylene | 0.49 | 0.17 | 2.88 | None |
| Benzo(k)fluoranthene | 0.8 | 0.24 | 3.33 | None |
| Chrysene | 0.93 | 0.166 | 5.60 | None |
| Dibenz(a,h)anthracene | 0.066 | 0.033 | 2 | None |
| Fluoranthene | 1.8 | 0.423 | 4.26 | None |
| Indeno(1,2,3-cd)pyrene | 0.48 | 0.2 | 2.40 | None |
| Phenanthrene | 0.91 | 0.204 | 4.46 | None |
| Pyrene | 1.9 | 0.195 | 9.74 | None |
| PCB-1254 | 0.047 | 0.0598 | 0.79 | PBT Compound |
| 2-Butanone | 0.055 | 0.0424 | 1.30 | None |
| Acetone Table excludes nutrients. | 0.19 | 0.0099 | 19.19 | None |

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg= Milligrams per kilogram. PBT = Persistent, bioaccumulative, and toxic.

PCB = Polychlorinated biphenyl.

SRV = Sediment reference value.

TNT = 2,4,6-Trinitrotoluene.

x = M ultiplier.

| | | ESV | Ratio of | |
|----------------------------|-------------|-------------|------------|----------------------------------|
| COPEC | MDC (mg/kg) | (mg/kg) | MDC to ESV | Comments |
| | D | iscrete Sed | iment | |
| Antimony | 1.9 | No ESV | | Exceeded SRV |
| Beryllium | 1.3 | No ESV | | Exceeded SRV |
| Cadmium | 2.3 | 0.99 | 2.32 | None |
| Chromium | 91.6 | 43.4 | 2.11 | None |
| Chromium, hexavalent | 10.6 | No ESV | | None |
| Copper | 106 | 31.6 | 3.35 | Exceeded ESV and SRV |
| Lead | 52.4 | 35.8 | 1.46 | Exceeded ESV and SRV |
| Mercury | 0.15 | 0.18 | 0.83 | PBT Compound |
| Nickel | 41.1 | 22.7 | 1.81 | Exceeded ESV and SRV |
| Selenium | 2.9 | No ESV | | None |
| Silver | 11 | 0.5 | 22 | Highest ratio at 22x |
| 1,3-Dinitrobenzene | 0.036 | 0.00861 | 4.18 | None |
| 2,4,6-Trinitrotoluene | 0.15 | No ESV | | None |
| 4-Amino-2,6- | 0.12 | No ESV | | None |
| Dinitrotoluene | •••= | | | None |
| HMX | 0.083 | No ESV | | None |
| Nitrocellulose | 5.7 | No ESV | | None |
| Tetryl | 0.019 | No ESV | | None |
| PCB-1254 | 0.032 | 0.0598 | 0.54 | PBT Compound |
| Acetone | 0.091 | 0.0099 | 9.19 | Second highest ratio at about 9x |
| | | ISM Sedin | nent | |
| Mercury | 0.048 | 0.18 | 0.27 | PBT Compound |
| Silver | 1.1 | 0.5 | 2.2 | None |
| Nitrate/Nitrite | 5.3 | No ESV | | None |
| Ammonia | 69 | No ESV | | None |
| Total Phosphorous | 270 | No ESV | | None |
| Bis(2-ethylhexyl)phthalate | 0.84 | 0.182 | 4.62 | None |
| Di-n-butyl phthalate | 2.7 | 1.114 | 2.42 | None |

Table 7-16. Summary of Integrated COPECs for Sediment at Upper Cobbs Pond

Table excludes nutrients.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine. ISM = Incremental sampling methodology.

MDC = M aximum detected concentration.

mg/kg= Milligrams per kilogram.

PBT = Persistent, bioaccumulative, and toxic.

PCB = Polychlorinated biphenyl.

SRV = Sediment reference value.

x = M ultiplier.

| | | ESV | Ratio of | |
|----------------------|-------------|-------------|------------|-----------------------------------|
| COPEC | MDC (mg/kg) | (mg/kg) | MDC to ESV | Comments |
| | | Discrete Se | diment | |
| Antimony | 1.4 | No ESV | | Exceeded SRV |
| Arsenic | 34.3 | 9.79 | 3.50 | Exceeded ESV and SRV |
| Beryllium | 0.93 | No ESV | | Exceeded SRV |
| Cadmium | 1.5 | 0.99 | 1.52 | None |
| Chromium | 150 | 43.4 | 3.46 | None |
| Chromium, hexavalent | 5.7 | No ESV | | None |
| Copper | 149 | 31.6 | 4.72 | Exceeded ESV and SRV |
| Cyanide | 0.4 | 0.0001 | 4000 | Highest ratio at 4000x |
| Lead | 60.4 | 35.8 | 1.69 | Exceeded ESV and SRV |
| Mercury | 0.093 | 0.18 | 0.52 | PBT Compound |
| Selenium | 2.2 | No ESV | | None |
| Silver | 2.4 | 0.5 | 4.80 | None |
| HMX | 0.017 | No ESV | | None |
| Nitrocellulose | 7.8 | No ESV | | None |
| Tetryl | 0.022 | No ESV | | None |
| 2-Butanone | 0.047 | 0.0424 | 1.11 | None |
| Acetone | 0.17 | 0.0099 | 17.17 | Second highest ratio at about 17x |
| | | ISM Sedi | ment | |
| Beryllium | 0.83 | No ESV | | Exceeded SRV |
| Cadmium | 1.2 | 0.99 | 1.21 | None |
| Chromium | 121 | 43.4 | 2.79 | None |
| Mercury | 0.1 | 0.18 | 0.56 | PBT Compound |
| Silver | 1.5 | 0.5 | 3.00 | None |
| Nitrate/Nitrite | 8.5 | No ESV | | None |
| Ammonia | 61 | No ESV | | None |
| Total Phosphorous | 360 | No ESV | | None |

Table 7-17. Summary of Integrated COPECs for Sediment at Lower Cobbs Pond

Table excludes nutrients.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

HMX = Octahy dro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.ISM = Incremental sampling methodology. MDC = Maximum detected concentration.

mg/kg= Milligrams per kilogram.

PBT = Persistent, bioaccumulative, and toxic.

SRV = Sediment reference value.

x = M ultiplier.

| | MDC | ESV | Ratio of MDC | |
|---------------|---------|-------------------|--------------|---------------------------------|
| COPEC | (mg/kg) | (mg/kg) | to ESV | Comments |
| Barium | 0.382 | 0.22ª | 1.7 | Ratio is MDC/OMZA (MDC/OMZM<1) |
| Manganese | 15.8 | 0.12 ^b | 131.7 | ESV is Tier II |
| Selenium | 0.0104 | 0.005 | 2.1 | ESV is OMZA (no OMZM available) |
| | | | | Ratios are MDC/OMZM and |
| Silver | 0.0017 | 0.0016/0.0013° | 1.1/1.3 | MDC/OMZA |
| Sulfate | 29 | No ESV | | None |
| Sulfide | 2.3 | 0.002 | 1150 | ESV is NAWQC |
| Chloromethane | 0.0007 | No ESV | | None |
| Chloromethane | | No ESV | | None |

Table 7-18. Summary of Integrated COPECs for Surface Water at the Backwater Area

^aESV shown is OMZA. The analyte is not a COPEC when screened against OMZM (MDC<OMZM). See Appendix H, Table H-23 for OMZM screening of this analyte.

^bNo Ohio Environmental Protection Agency Administrative Code ESV (OMZM or OMZA) is available; ESV is next available in hierarchy of sources.

^cESVs shown are OMZM/OMZA.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg= Milligrams per kilogram.

NAWQC = National ambient water quality criteria.

OMZA = Outside mixing zone average.

OMZM = Outside mixing zone maximum.

-- = No ratio could be calculated because there is no ESV.

Table 7-19. Summary of Integrated COPECs for Surface Water at Upper Cobbs Pond

| COPEC | MDC (mg/kg) | ESV (mg/kg) | Ratio of MDC to ESV | Comments |
|-----------------|-------------|--------------------|------------------------|----------------|
| Manganese | 0.528 | 0.12 ^a | 4.4 | ESV is Tier II |
| Nitrate/Nitrite | 0.18 | No ESV | | None |
| Sulfate | 13 | No ESV | | None |
| Sulfide | 0.7 | 0.002 ^a | 350 | ESV is NAWQC |

^aNo Ohio Environmental Protection Agency Administrative Code ESV (OMZM or OMZA) is available; ESV is next available in hierarchy of sources.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg= Milligrams per kilogram.

NAWQC = National ambient water quality criteria.

-- = No ratio could be calculated because there is no ESV.

| Table 7–20. Summary | of Integrated | COPECs | for Surface | Water at Lower | Cobbs Pond |
|---------------------|---------------|--------|-------------|----------------|------------|
| | | | | | |

| COPEC | MDC(mg/kg) | ESV (mg/kg) | Ratio of MDC to ESV | Comments |
|-----------------|------------|--------------------|------------------------|----------------|
| Manganese | 0.654 | 0.12 ^a | 5.45 | ESV is Tier II |
| Nitrate/Nitrite | 0.13 | No ESV | | None |
| Sulfate | 19 | No ESV | | None |
| Sulfide | 0.7 | 0.002 ^a | 350 | ESV is NAWQC |

^aNo Ohio Environmental Protection Agency Administrative Code ESV (OMZM or OMZA) is available; ESV is next available in hierarchy of sources.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg= Milligrams per kilogram.

NAWQC = National ambient water quality criteria.

Table 7-21. Application and Decisions of Selected Evaluation Factors to Integrated COPECs for Surface Soil from Level II

| | Condition for Decision | |
|-----------------------------------|---|--|
| Action | to Dismiss or Retain COPEC | Outcome |
| Compare mean concentration to ESV | (A) Mean concentration smaller than or equal to the ESV | COPEC is not a candidate to be a final COPEC. Evaluation ceased. |
| | (B) Mean concentration larger than the ESV | Continue evaluation of chemical. |
| Compare mean concentration above | (A) Mean concentration smaller than the background concentration | COPEC is not a candidate to be a final COPEC. Evaluation ceased. |
| ESV to background concentration | (B) Mean concentration larger than background concentration | Continue evaluation of chemical. |

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

Table 7-22. Summary of Data for Step 3A Refinement of Integrated COPECs in Surface Soil at the AOC

| COPEC | Mean Concentration (mg/kg) | ESV (mg/kg) | Mean Concentration > ESV? (Yes/No) | Background Concentration (mg/kg) | Mean Concentration > Background Concentration? (Yes/No) | ESV > Background Concentration? (Yes/No) | Frequency of Detections ^a > ESV | Frequency of Detections ^a > Background Concentration | Further Evaluation in Level II Required? (Yes/No) | |
|----------|--|----------------|---|--|--|---|---|--|---|--|
| | | | | COPEC with | Mean Concentration < E | SV | | | | |
| Arsenic | 11.1 | 18 | No | 15.4 | No | Yes | 2/24 | 4/24 | No | |
| Cadmium | 0.135 | 0.36 | No | 0 | Yes | Yes | 1/19 | 19/19 | No | |
| Cobalt | 8.9 | 13 | No | 10.4 | No | Yes | 2/24 | 8/24 | No | |
| | | COPEC | with Mean Conce | entration > ESV | and Mean Concentration | n < Background | Concentration | | | |
| Aluminum | 11900 | 50 | Yes | 17,700 | No | No | 24/24 | 2/24 | No | |
| Lead | 18.7 | 11 | Yes | 26.1 | No | No | 23/24 | 3/24 | No | |
| Vanadium | 20.9 | 7.8 | Yes | 31.1 | No | No | 24/24 | 1/24 | No | |
| | COPEC with Mean Concentration > ESV and Mean Concentration > Background Concentration | | | | | | | | | |
| Mercury | 0.043 | 0.00051 | Yes | 0.036 | Yes | No | 21/21 | 10/21 | Yes | |
| Zinc | 63.9 | 46 | Yes | 61.8 | Yes | No | 24/24 | 10/24 | Yes | |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken. AOC = Area of concern.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg= Milligrams per kilogram.

Table 7-23. Summary of Mean Concentrations and Background Concentrations of Remaining Integrated Soil COPECs in the Refinement Factors

| COPEC | Mean Concentration (mg/kg) | Background Concentration (mg/kg) | Ratio of Mean Concentration to Background Concentration | Qualitative Assessment of Mean to Background Concentration |
|---------|----------------------------------|--|--|---|
| | | | Surface Soil | |
| Mercury | 0.043 | 0.036 | 1.19 | Concentrations are similar |
| Zinc | 63.9 | 61.8 | 1.03 | Concentrations are similar |

COPEC = Chemical of potential ecological concern.

mg/kg= Milligrams per kilogram.

Table 7-24. Comparison of Alternative ESV to Mean Concentration for Remaining Integrated Soil COPECs

| CODEC | AOC Mean Concentration | Background Concentration | Preferred ESV | Alternative ESV ^a | Ratio of Mean Concentration to Preferred | Ratio of Mean Concentration to Alternative |
|---------|---------------------------|-----------------------------|------------------|---------------------------------|---|---|
| COPEC | (mg/kg) | (mg/kg) | (mg/kg) | (mg/kg) | ESV | ESV |
| Mercury | 0.043 | 0.036 | 0.00051 | 0.1 | 84 | 0.43 |
| Zinc | 63.9 | 61.8 | 46 | 79 | 1.4 | 0.81 |

^aThe Alternative ESV is the ESV with the closest concentration to the preferred ESV that is above the background concentration.

AOC = Area of concern.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg= Milligrams per kilogram.

Table 7-25. Summary of Data for Step 3A Refinement of Integrated COPECs in Sediment at the Backwater Area

| COPEC | Mean Concentration (mg/kg) | ESV or (SRV) (mg/kg) | Mean Concentration > ESV? (Yes/No) | Background Concentration (mg/kg) | Mean Concentration > Background Concentration? (Yes/No) | ESV> Background Concentration? (Yes/No) | Frequency of Detections ^a > ES V | Frequency of Detections ^a > Background Concentration | Further Evaluation in Level II Required? (Yes/No) |
|------------------------|----------------------------------|----------------------------|--|--|---|--|---|--|---|
| | | | СОРЕ | C with Mean Co | ncentration < ESV | , | | | |
| Cadmium | 0.972 | 0.99 | No | 0 | Yes | Yes | 5/14 | 14/14 | No |
| Chromium | 23.3 | 43.4 | No | 18.1 | Yes | Yes | 1/14 | 9/14 | No |
| Copper | 28.2 | 31.6 | No | 27.6 | Yes | Yes | 4/14 (4) | 4/14 | No |
| Lead | 29.6 | 35.8 | No | 27.4 | Yes | Yes | 4/14 (1) | 7/14 | No |
| Nickel | 20.5 | 22.7 | No | 17.7 | Yes | Yes | 6/14 (1) | 8/14 | No |
| Indeno(1,2,3-cd)pyrene | 0.192 | 0.2 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | No |
| 2-Butanone | 0.0256 | 0.0424 | No | 0 | Yes | Yes | 1/3 | 3/3 | No |
| | | | COPEC with 1 | Mean Concentra | tion < Background | or SRV | | | |
| Antimony | 0.607 | (1.3) | <srv< td=""><td>0</td><td>Yes</td><td></td><td>(2<srv)< td=""><td>11/11</td><td>No</td></srv)<></td></srv<> | 0 | Yes | | (2 <srv)< td=""><td>11/11</td><td>No</td></srv)<> | 11/11 | No |
| Beryllium | 0.782 | (0.8) | <srv< td=""><td>0.38</td><td>Yes</td><td></td><td>(8<srv)< td=""><td>13/14</td><td>No</td></srv)<></td></srv<> | 0.38 | Yes | | (8 <srv)< td=""><td>13/14</td><td>No</td></srv)<> | 13/14 | No |
| Selenium | 0.5 | (1.7) | <srv< td=""><td>1.7</td><td>No</td><td></td><td>(7<srv)< td=""><td>1/14</td><td>No</td></srv)<></td></srv<> | 1.7 | No | | (7 <srv)< td=""><td>1/14</td><td>No</td></srv)<> | 1/14 | No |
| | CO | PEC with M | lean Concentration | > ESV and Mea | n Concentration > | Background Con | centration | | |
| Cyanide | 0.224 | 0.0001 | Yes | 0 | Yes | Yes | 3/3 | 3/3 | No |
| Silver | 5.7 | 0.5 | Yes | 0 | Yes | Yes | 13/13 | 13/13 | Yes |
| 2,6-Dinitrotoluene | 0.0483 | 0.0398 | Yes | 0 | Yes | Yes | 1/1 | 1/1 | No |
| Anthracene | 0.0956 | 0.0572 | Yes | 0 | Yes | Yes | 2/2 | 2/2 | No |
| Benz(a)anthracene | 0.274 | 0.108 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | Yes |
| Benzo(a)pyrene | 0.323 | 0.15 | Yes | 0 | Yes | Yes | 3/4 | 4/4 | Yes |
| Benzo(ghi)pyrene | 0.22 | 0.17 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | No |
| Benzo(k)fluoranthene | 0.265 | 0.24 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | No |
| Chrysene | 0.338 | 0.166 | Yes | 0 | Yes | Yes | 3/4 | 4/4 | Yes |
| Dibenz(a,h)anthracene | 0.095 | 0.033 | Yes | 0 | Yes | Yes | 1/1 | 1/1 | No |
| Fluoranthene | 0.61 | 0.423 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | No |
| Phenanthrene | 0.285 | 0.204 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | No |
| Pyrene | 0.604 | 0.195 | Yes | 0 | Yes | Yes | 2/3 | 3/3 | Yes |
| Acetone | 0.0145 | 0.0099 | Yes | 0 | Yes | Yes | 5/5 | 5/5 | Yes |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg= Milligrams per kilogram.

SRV = Sediment reference value.

-- = No value available.

Table 7–26. Summary of Data for Step 3A Refinement of Integrated COPECs in Discrete Sediment at Upper Cobbs Pond

| COPEC | Mean Concentration (mg/kg) | ESV (SRV) (mg/kg) | Mean Concentration > ES V? (Yes/No) | Background Concentration (mg/kg) | Mean Concentration > Background Concentration? (Yes/No) | ESV> Background Concentration? (Yes/No) | Frequency of Detections ^a > ES V | Frequency of Detections ^a > Background Concentration | Further Evaluation in Level II Required? (Yes/No) |
|--------------------|----------------------------------|----------------------|---|--|---|---|---|--|---|
| | | | COF | PEC with Mean C | Concentration < ESU | 1 | | | |
| Cadmium | 0.95 | 0.99 | No | 0 | Yes | Yes | 2/3 | 3/3 | No |
| Lead | 27.7 | 35.8 | No | 27.4 | Yes | Yes | 2/5 | 2/5 | No |
| | | | COPEC with | h Mean Concenti | ration < Background | d or SRV | | | |
| Antimony | 0.78 | (1.3) | (<srv)< td=""><td>0</td><td>Yes</td><td>Yes (SRV)</td><td>(1>SRV)</td><td>5/5</td><td>No</td></srv)<> | 0 | Yes | Yes (SRV) | (1>SRV) | 5/5 | No |
| Beryllium | 0.842 | (0.8) | (>SRV) | 0.38 | Yes | Yes (SRV) | (2>SRV) | 5/5 | No |
| Nickel | 28.7 | 22.7 (33) | Yes (<srv)< td=""><td>17.7</td><td>Yes</td><td>Yes (<srv)< td=""><td>4/5 (2>SRV)</td><td>4/5</td><td>No</td></srv)<></td></srv)<> | 17.7 | Yes | Yes (<srv)< td=""><td>4/5 (2>SRV)</td><td>4/5</td><td>No</td></srv)<> | 4/5 (2>SRV) | 4/5 | No |
| Selenium | 0.839 | (1.7) | (<srv)< td=""><td>1.7</td><td>No</td><td>No (SRV)</td><td>(1>SRV)</td><td>1/5</td><td>No</td></srv)<> | 1.7 | No | No (SRV) | (1>SRV) | 1/5 | No |
| | СОРЕ | C with Mean | Concentration > 1 | ESV (or >SRV) at | nd Mean Concentra | tion > Background | d Concentration | | |
| Chromium | 50.8 | 43.4 | Yes | 18.1 | Yes | Yes | 3/6 | 6/6 | No |
| Copper | 57.1 | 31.6 | Yes | 27.6 | Yes | Yes | 3/5 | 4/5 | No |
| Silver | 3.81 | 0.5 | Yes | 0 | Yes | Yes | 3/3 | 3/3 | Yes |
| 1,3-Dinitrobenzene | 0.0144 | 0.00861 | Yes | 0 | Yes | Yes | 1/1 | 1/1 | No |
| Acetone | 0.0895 | 0.0099 | Yes | 0 | Yes | Yes | 2/2 | 2/2 | Yes |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken. COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg= Milligrams per kilogram. SRV = Sediment reference value.

Table 7-27. Summary of Data for Step 3A Refinement of Integrated COPECs in Discrete Sediment at Lower Cobbs Pond

| COPEC | Mean Concentration (mg/kg) | ESV (SRV) (mg/kg) | Mean Concentration > ES V? (Yes/No) | Background Concentration (mg/kg) | Mean Concentration > Background Concentration? (Yes/No) | ESV> Background Concentration? (Yes/No) | Frequency of Detections ^a > ES V | Frequency of Detections ^a > Background Concentration | Further Evaluation in Level II Required? (Yes/No) | | |
|---|---|----------------------|---|--|---|--|---|--|---|--|--|
| COPEC with Mean Concentration < ESV (or <srv)< td=""></srv)<> | | | | | | | | | | | |
| Antimony | 0.328 | (1.3) | No (<srv)< td=""><td>0</td><td>Yes</td><td>Yes (SRV)</td><td>(1>SRV)</td><td>1/1</td><td>No</td></srv)<> | 0 | Yes | Yes (SRV) | (1>SRV) | 1/1 | No | | |
| Arsenic | 14.9 | 9.79 (25) | Yes (<srv)< td=""><td>19.5</td><td>Yes</td><td>No</td><td>5/8 (1>SRV)</td><td>2/8</td><td>No</td></srv)<> | 19.5 | Yes | No | 5/8 (1>SRV) | 2/8 | No | | |
| Beryllium | 0.679 | (0.8) | No (<srv)< td=""><td>0.38</td><td>Yes</td><td>Yes (SRV)</td><td>(2>SRV)</td><td>8/8</td><td>No</td></srv)<> | 0.38 | Yes | Yes (SRV) | (2>SRV) | 8/8 | No | | |
| Cadmium | 0.536 | 0.99 | No | 0 | Yes | Yes | 2/6 | 6/6 | No | | |
| Lead | 29.4 | 35.8 | No | 27.4 | Yes | Yes | 3/8 | 4/8 | No | | |
| Selenium | 0.72 | (1.7) | No (<srv)< td=""><td>1.7</td><td>No</td><td>No (SRV)</td><td>(1>SRV)</td><td>1/4</td><td>No</td></srv)<> | 1.7 | No | No (SRV) | (1>SRV) | 1/4 | No | | |
| 2-Butanone | 0.036 | 0.0424 | No | 0 | Yes | Yes | 1/2 | 2/2 | No | | |
| | COPEC with Mean Concentration > ESV and Mean Concentration > Background Concentration | | | | | | | | | | |
| Chromium | 59 | 43.4 | Yes | 18.1 | Yes | Yes | 3/8 | 5/8 | No | | |
| Copper | 56.6 | 31.6 (32) | Yes | 27.6 | Yes | Yes | 4/8 (3>SRV) | 4/8 | No | | |
| Cyanide | 0.225 | 0.0001 | Yes | 0 | Yes | Yes | 1/1 | 1/1 | No | | |
| Silver | 0.852 | 0.5 | Yes | 0 | Yes | Yes | 3/5 | 5/5 | No | | |
| Acetone | 0.123 | 0.0099 | Yes | 0 | Yes | Yes | 2/2 | 2/2 | Yes | | |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken. COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg= Milligrams per kilogram.

SRV = Sediment reference value.

Mean Further ESV >Mean Concentration > **Frequency of** Evaluation in Detections^a > Mean Concentration > Background Background Background Level II **Frequency of** ESV ESV? **Concentration** Concentration? **Concentration?** Detections^a> Background **Required?** Concentration COPEC (mg/kg) (mg/kg) (Yes/No) (mg/kg) (Yes/No) (Yes/No) ESV Concentration (Yes/No) **COPEC** with Mean Concentration < ESV 0.0475 Barium 0.141 0.22 No Yes Yes 1/4 2/4 No Selenium 0.00335 0.005 No 0 Yes Yes 1/33/3 No 0.000603 Silver 0.0016 No 0 Yes Yes 1/11/1No **COPEC** with Mean Concentration > ESV and Mean Concentration > Background Concentration 2/4 Manganese 6.16 0.12 Yes 0.391 Yes No 4/4 Yes Sulfide 1.28 0.002 Yes Yes Yes 1/11/1Yes 0

Table 7–28. Summary of Data for Step 3A Refinement of Integrated COPECs in Surface Water at the Backwater Area

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken. COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg= Milligrams per kilogram.

Table 7-29. Summary of Data for Step 3A Refinement of Integrated COPECs in Surface Water at Upper Cobbs Pond

| | | | | | Mean | | | | Further | |
|--|---------------|---------|-----------------|---------------|-----------------------|-----------------------|---------------------------|---------------------------|----------------------|--|
| | | | Mean | | Concentration > | ESV > | | Frequency of | Evaluation in | |
| | Mean | | Concentration > | Background | Background | Background | Frequency of | Detections ^a > | Level II | |
| | Concentration | ES V | ES V? | Concentration | Concentration? | Concentration? | Detections ^a > | Background | Required? | |
| COPEC | (mg/kg) | (mg/kg) | (Yes/No) | (mg/kg) | (Yes/No) | (Yes/No) | ES V | Concentration | (Yes/No) | |
| COPEC with Mean Concentration < Background Concentration | | | | | | | | | | |
| Manganese | 0.216 | 0.12 | Yes | 0.391 | No | No | 3/5 | 1/5 | No | |
| COPEC with Mean Concentration > ESV and Mean Concentration > Background Concentration | | | | | | | | | | |
| Sulfide | 0.7 | 0.002 | Yes | 0 | Yes | Yes | 2/2 | 2/2 | Yes | |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken. COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value. mg/kg= Milligrams per kilogram.

| | | | | | Mean | | | | Further | | |
|--|--|---------|-----------------|---------------|-----------------------|-----------------------|---------------------------|---------------------------|----------------------|--|--|
| | | | Mean | | Concentration > | ESV > | | Frequency of | Evaluation in | | |
| | Mean | | Concentration > | Background | Background | Background | Frequency of | Detections ^a > | Level II | | |
| | Concentration | ES V | ES V? | Concentration | Concentration? | Concentration? | Detections ^a > | Background | Required? | | |
| COPEC | (mg/kg) | (mg/kg) | (Yes/No) | (mg/kg) | (Yes/No) | (Yes/No) | ES V | Concentration | (Yes/No) | | |
| | COPEC with Mean Concentration < Background Concentration | | | | | | | | | | |
| Manganese | 0.215 | 0.12 | Yes | 0.391 | No | No | 4/7 | 1/7 | No | | |
| COPEC with Mean Concentration > ESV and Mean Concentration > Background Concentration | | | | | | | | | | | |
| Sulfide | 0.4 | 0.002 | Yes | 0 | Yes | Yes | 1/1 | 1/1 | Yes | | |

Table 7-30. Summary of Data for Step 3A Refinement of Integrated COPECs in Surface Water at Lower Cobbs Pond

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken. COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value. mg/kg= Milligrams per kilogram.

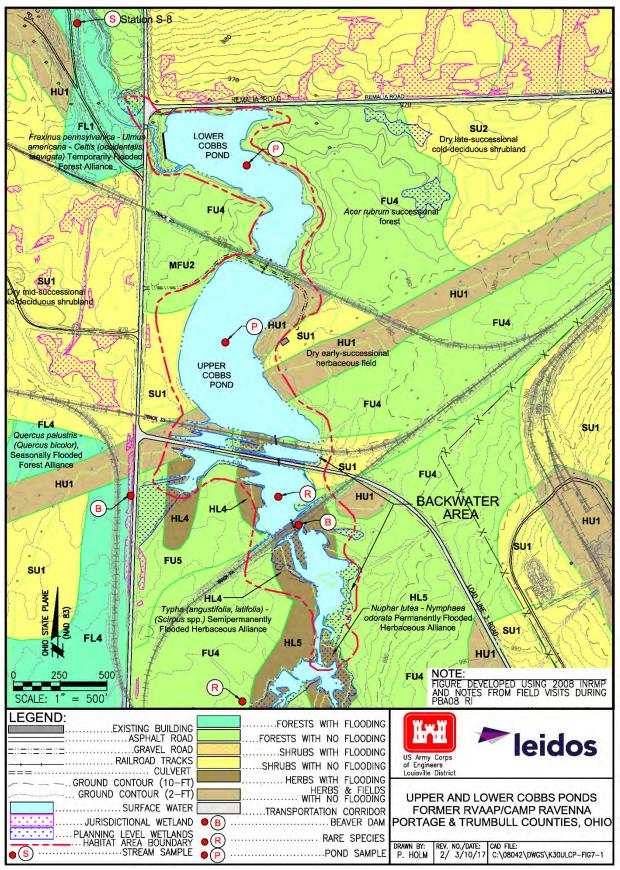


Figure 7-1. Natural Resources Inside and Near Habitat Area at the Upper and Lower Cobbs Ponds

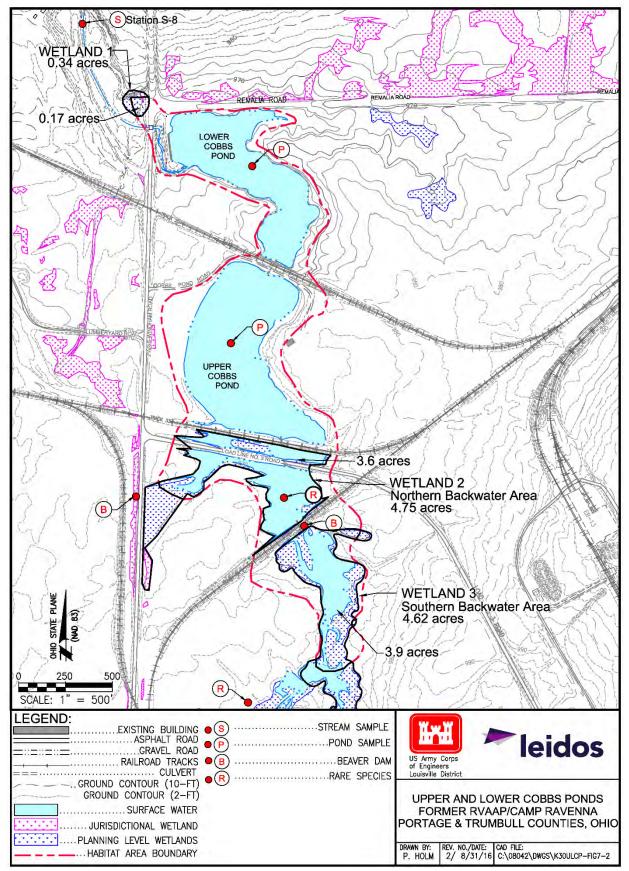


Figure 7-2. Wetland and Aquatic Resources Inside and Near Habitat Area at the Upper and Lower Cobbs Ponds

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1 8.0 REMEDIAL INVESTIGATION CONCLUSIONS AND 2 RECOMMENDATIONS

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8.1 INTRODUCTION

6 This Phase III RI Report for the Upper and Lower Cobbs Ponds AOC presents a detailed analysis of 7 historical and newly acquired environmental data. The following sections summarize the major 8 findings of the nature and extent of contamination, modeling of contaminant fate and transport, 9 HHRA, and ERA. A CSM incorporating all available information is also presented to integrate results 10 of prior investigations and the PBA08 RI. The CSM denotes, based on available data where source 11 areas occur, the mechanisms for contaminant migration from source areas to receptor media (e.g., 12 surface water and groundwater), exit pathways from the AOC, and if COCs occur that may require 13 further evaluation in an FS. This section presents the need for any further characterization of the 14 media evaluated under the RI phase of work and whether to proceed to the FS phase of the CERCLA 15 RI/FS process.

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8.2 SUMMARY OF DATA USED IN THE REMEDIAL INVESTIGATION

19 Quality-assured sample data from the RIs (1996 Phase I RI, 2001 Phase II RI, and 2010 PBA08 RI) 20 and the 2003 FWBWQS were used to evaluate nature and extent of contamination at the Upper and 21 Lower Cobbs Ponds AOC. All available sample data collected at the locations were evaluated to 22 determine suitability for use in various key RI data screens and evaluations (i.e., nature and extent, 23 fate and transport, and risk assessment). Evaluating data suitability for use in the PBA08 RI involved 24 two primary considerations: (1) whether the data represented current AOC conditions, and (2) sample 25 collection methods (e.g., discrete vs. ISM).

26

27 Data collected during the RIs (1996 Phase I RI, 2001 Phase II RI, and 2010 PBA08 RI) and the 2003 28 FWBWOS were deemed usable for this report. Pond Bank surface soil data collected during the Phase II RI and PBA08 RI (with exception of field duplicate samples) were included in data 29 30 screening to identify SRCs, contaminant nature and extent, contaminant fate and transport, and risk 31 assessments. For subsurface soil, only PBA08 RI sample data are available for evaluations performed 32 in this report. For sediment, Phase II RI, 2003 FWBWQS, and PBA08 RI samples were deemed most 33 representative of current AOC conditions and were used for SRCs screening, fate and transport 34 evaluation, and risk assessment. Surface water data collected during the Phase II RI, 2003 FWBWQS, 35 and the PBA08 RI were deemed representative of current AOC conditions and were used to screen 36 for SRCs and for risk assessment. All available historical and current sediment and surface water are 37 included in contaminant nature and extent to evaluate temporal trends for identified SRCs and the 38 potential for off-AOC transport of contaminants via the surface water migration pathway.

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8.3 SUMMARY AND CONCLUSIONS OF NATURE AND EXTENT OF CONTAMINATION

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Nature and extent of contamination in surface soil (0–1 ft bgs), subsurface soil (greater than 1 ft bgs),
sediment, and surface water was evaluated in the RI. Data from the RIs (1996 Phase I RI, 2001 Phase

II RI, and 2010 PBA08 RI) and the 2003 FWBWQS effectively characterized the nature and extent of the contamination at the AOC. To support the evaluation of nature and extent of contamination, SRC concentrations were compared to SLs corresponding to the lowest FWCUG for the Resident Receptor (Adult and Child) and National Guard Trainee at a target HQ of 0.1 or TR of 1E-06, as presented in the FWCUG Report. It can be concluded that the vertical and horizontal extent of contamination is defined, and no further sampling is needed to evaluate the Upper and Lower Cobbs Ponds AOC.

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8 8.3.1 Surface Soil

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10 The predominant SRCs for surface soil at the Pond Bank were inorganic chemicals and SVOCs; the 11 majority of which were PAHs. A total of 18 inorganic chemicals (16 metals, cyanide, and 12 nitrate/nitrite) were identified as SRCs in surface soil. Elevated inorganic chemical concentrations 13 above background concentrations were observed in surface soil throughout the Pond Bank, and only 1 14 of 24 surface soil samples collected exhibited no inorganic chemicals above background 15 concentrations (ULCPss-009). However, the majority of elevated inorganic chemical concentrations 16 occurred at levels marginally in excess of background concentrations, as all but two concentrations 17 occurred at magnitudes less than twice their respective background concentrations. As such, no 18 apparent spatial trend is evident in the distribution of inorganic chemicals throughout the surface soil 19 in the Pond Bank.

20

Sixteen SVOCs, predominantly PAHs, were identified as SRCs in surface soil at the Pond Bank. The highest frequency and magnitudes of detection were observed at CPCsb-035, located on the western bank of Upper Cobbs Pond, immediately south of Cobbs Pond Road. The pesticide beta-BHC and propellant nitrocellulose were also identified as SRCs in surface soil at the Pond Bank. PAH concentrations are an order of magnitude higher in sediment than those observed in the adjacent Pond Bank surface soil.

27

One propellant (nitrocellulose) and one pesticide (BHC) were identified as SRCs in surface soil at the Pond Bank. Neither exceeded their respective SLs. No explosives, VOCs, or PCBs were detected or identified as SRCs in Pond Bank surface soil.

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32 8.3.2 Subsurface Soil

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Four inorganic chemicals (cadmium, lead, selenium, and silver) were identified as SRCs in subsurface soil samples at the Pond Bank. Lead and selenium were detected above background concentrations at only one sample location. Selenium was detected at 1.6 mg/kg, slightly above the background concentration of 1.5 mg/kg, and lead was detected at 23.3 mg/kg above the background concentration of 19.1 mg/kg. As such, no apparent spatial trend in the distribution of inorganic chemicals throughout the subsurface soil in the Pond Bank is evident.

40

A total of 13 SVOCs, 12 of which were PAHs, were detected and identified as SRCs in subsurface
soil at the Pond Bank. Of the 13 SVOCs, 12 were detected from the 1–4 ft bgs interval at CPCsb-035;
the majority of which were observed at low, estimated concentrations below laboratory reporting
limits. Sample location CPCsb-035 is on the western bank of Upper Cobbs Pond and near Cobbs

Pond Road. PAH concentrations in the subsurface soil samples were generally comparable to those observed in the corresponding surface soil sample (i.e., within the same order of magnitude). One propellant (nitrocellulose) was identified as an SRC in subsurface soil at the Pond Bank. No explosives, VOCs, PCBs, or pesticides were detected or identified as SRCs in Pond Bank subsurface soil.

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8.3.3 Sediment

Backwater Area

8.3.3.1

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11 The Backwater Area was evaluated using 14 surface sediment samples and 2 subsurface sediment 12 samples. Three explosives (TNT; 2,6-DNT; and tetryl) and one propellant (nitrocellulose) were 13 detected in surface sediment. All of the detected concentrations were below the Resident Receptor 14 (Adult and Child) FWCUG and RSL at a TR of 1E-06, HO of 0.1. No explosives or propellants were 15 detected in subsurface sediment in the Backwater Area. In surface sediment, aluminum, cobalt, and 16 cyanide concentrations exceeded a TR of 1E-06, HQ of 0.1, but not a TR of 1E-05, HQ of 1. Only the 17 cobalt concentration exceeded a TR of 1E-06, HQ of 0.1 in subsurface sediment in one subsurface 18 soil location, but not at a TR of 1E-05, HQ of 1. Five PAHs [benzo(a)anthracene, benzo(a)pyrene, 19 benzo(b)fluoranthene, dibenzo(a,h)anthracene, and indeno(1,2,3-cd)pyrene] exceeded their respective 20 SLs in surface sediment. Four of the five PAHs were detected at their maximum concentrations in the 21 Phase II RI sample ULCPsd-015. The detected concentrations in the adjacent PBA08 RI sample 22 CPCsd-047 indicate that concentrations have declined since the 2001 Phase II RI sampling event. 23 Benzo(a)pyrene concentrations at ULCPsd-006, ULCPsd-015, and CPCsd-047 exceeded the Resident 24 Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. Only benzo(a)pyrene was detected 25 above its respective SL (0.022 mg/kg) at PBA08 RI sample location CPCsd-047 in the 0.5-2 ft bgs 26 interval. The concentration was below the Resident Receptor (Adult and Child) FWCUG at a TR of 27 1E-05, HQ of 1. Three VOCs (acetone, methylene chloride, and 2-butanone) were detected below 28 their respective SLs in surface and subsurface sediment, respectively. PCB-1254 also was detected in 29 the Phase II RI surface sediment samples ULCPsd-006 and ULCPsd-011; however, it was not 30 detected in the PBA08 RI samples. No pesticides were detected in surface sediment. Pesticides and 31 PCBs were not detected in the Backwater Area subsurface sediment.

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33 8.3.3.2 <u>Upper Cobbs Pond</u>

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35 Five surface discrete sediment samples, one surface ISM sample, and one subsurface sediment sample 36 were used to evaluate Upper Cobbs Pond. Five explosives (1,3-dinitrobenzene; TNT; 4-amino-2,6-37 DNT; HMX; and tetryl) and one propellant (nitrocellulose) were detected in surface sediment at 38 PBA08 RI sample location CPCsd-046. All detected concentrations were below the Resident 39 Receptor (Adult and Child) FWCUG and RSL at a TR of 1E-06, HQ of 0.1. No explosives or 40 propellants were detected in subsurface sediment. In surface sediment, concentrations of aluminum, 41 hexavalent chromium, and cobalt exceeded a TR of 1E-06, HQ of 0.1, but not a TR of 1E-05, HQ of 42 1. Only the cobalt concentration exceeded a TR of 1E-06, HQ of 0.1 in subsurface sediment, but not 43 at a TR of 1E-05, HQ of 1. Ten SVOCs were detected in surface and subsurface sediment at PBA08 44 RI sample location CPCsd-046. Two additional SVOCs (acenaphthylene and anthracene) were

1 detected in subsurface sediment. The 10 SVOCs occurred at higher concentrations in surface 2 sediment than in the corresponding subsurface sediment sample. Benzo(a)pyrene was detected above 3 its respective SL in surface and subsurface sediment at CPCsd-046; however, the concentrations were 4 below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. Two VOCs 5 (acetone and 2-butanone) were detected in surface sediment below their respective SLs, both were 6 detected at their maximum concentrations at PBA08 RI location CPCsd-046. These VOCs also were 7 detected at the 2001 Phase II RI sample location ULCPsd-020 at lower concentrations than CPCsd-8 046. PCB-1254 also was detected in surface sediment at Upper Cobbs Pond in ULCPsd-020 below its 9 respective SL. PCB-1254 was not detected in the PBA08 RI sample. Pesticides were not detected in 10 surface sediment at Upper Cobbs Pond. The VOC 2-butanone was detected in subsurface sediment below the TR of 1E-05, HO of 1. The pesticide delta-BHC does not have an SL. PCBs were not 11 12 detected in subsurface sediment.

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14 8.3.3.3 Lower Cobbs Pond

16 Lower Cobbs Pond was evaluated using eight discrete surface sediment samples, one surface ISM 17 sample, and one subsurface sediment sample. Two explosives (HMX and tetryl) and one propellant 18 (nitrocellulose) were detected in surface sediment, and the explosive HMX and propellant 19 nitrocellulose were detected in subsurface sediment at Lower Cobbs Pond. The concentrations were 20 below the Resident Receptor (Adult and Child) FWCUG and RSL at a TR of 1E-06, HQ of 0.1. In 21 surface sediment, aluminum, arsenic, hexavalent chromium, cobalt, and cyanide concentrations 22 exceeded a TR of 1E-06, HO of 0.1, but only arsenic exceeded a TR of 1E-05, HO of 1. Only the 23 cobalt concentrations exceeded a TR of 1E-06, HO of 0.1 in the subsurface sediment, but not at a TR 24 of 1E-05, HQ of 1. A total of 11 SVOCs were detected in surface sediment in Lower Cobbs Pond and 25 18 SVOCs were detected in subsurface sediment. The 11 SVOC SRCs that were identified for surface 26 sediment also were observed in subsurface sediment and occurred at their highest concentrations in 27 the PBA08 RI subsurface sediment sample CPCsd-045. The benzo(a)pyrene concentration exceeded 28 its respective SL in surface and subsurface sediment at CPCsd-045; however, only the subsurface 29 sediment concentration exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, 30 HQ of 1. Three VOCs (2-butanone, acetone, and carbon disulfide) were detected in surface sediment 31 at Lower Cobbs Pond. All three VOCs were detected at PBA08 RI location CPCsd-045, with 2-32 butanone and carbon disulfide observed at their maximum concentrations in this sample location. 2-33 Butanone also was detected in the subsurface sediment. All of the VOCs detected in surface and 34 subsurface soil were below their respective SLs. No PCBs or pesticides were detected in surface or 35 subsurface sediment in Lower Cobbs Pond.

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- 37 8.3.4 Surface Water
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39 8.3.4.1 <u>Backwater Area</u> 40

Four surface water samples were used to evaluate the Backwater Area. One explosive (4-amino-2,6-DNT) was detected in surface water at a low, estimated concentration below laboratory reporting limits. No propellants were detected in the Backwater Area surface water. Four inorganic chemical concentrations (arsenic, cobalt, manganese, and selenium) exceeded the RSL at a TR of 1E-06, HQ of 1 0.1. Manganese concentrations at Phase II RI sample locations ULCPsw-001 and ULCPsw-002 and 2 cobalt at ULCPsw-001 exceeded the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, 3 HQ of 1. Three VOCs (acetone, chloromethane, and toluene) were detected in surface water for the 4 Backwater Area, all of which were detected in ULCPsw-002. No propellants, SVOCs, PCBs, or 5 pesticides were detected in surface water at the Backwater Area. Elevated inorganic chemical 6 concentrations and 4-amino-2,6-DNT are likely attributable to continuing upstream sources as these 7 constituents are observed in the PBA08 RI locations immediately downstream of the confluence of 8 the drainage channels from Load Lines 3 and 12.

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8.3.4.2 Upper Cobbs Pond

12 Surface water in Upper Cobbs Pond was evaluated using five surface water samples. One explosive 13 (4-amino-2,6-DNT) was detected at a low, estimated concentration in two surface water samples. The 14 concentrations were below the Resident Receptor (Adult and Child) FWCUG and RSL at a TR of 1E-15 06, HO of 0.1. No surface water concentrations for inorganic chemicals in Upper Cobbs Pond 16 exceeded the SL at a TR of 1E-06, HQ of 0.1. The SVOCs bis(2-ethylhexyl)phthalate and di-n-butyl 17 phthalate were detected at low, estimated concentrations below the laboratory reporting limit at 18 PBA08 RI sample location CPCsw-046 and the August 2003 FWBWQS location, respectively. No VOCs, PCBs, or pesticides were detected in surface water at Upper Cobbs Pond. 19

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21 8.3.4.3 Lower Cobbs Pond

- 23 Lower Cobbs Pond surface water was evaluated using seven surface water samples. The explosive 4-24 amino-2,6-DNT was identified as an SRC in surface water at Lower Cobbs Pond. 4-Amino-2,6-DNT 25 was detected in the August 2003 sample from the FWBWQS Lower Cobbs Pond station (0.00029J 26 mg/L). 4-Amino-2,6-DNT was not detected in the subsequent PBA08 RI samples taken at CPCsw-27 044 and CPCsw-045 in 2010. No propellants were detected. Only manganese was detected at a 28 concentration which exceeded the Resident Receptor (Adult and Child) FWCUG and RSL at a TR of 29 1E-06, HQ of 0.1 at one sample location. Two SVOCs (benzenemethanol and butyl benzyl phthalate) 30 were detected at low, estimated concentrations at PBA08 RI location CPCsw-045, located at the 31 center of the aggregate. No VOCs, PCBs, or pesticides were detected in surface water in Lower 32 Cobbs Pond.
- 33

34 8.4 SUMMARY OF CONTAMINANT FATE AND TRANSPORT

35

All SRCs identified in surface soil, subsurface soil, and sediment at the Upper and Lower Cobbs Ponds AOC were evaluated through the stepwise contaminant fate and transport evaluation. The evaluation included analyzing leaching and migration from soil and sediment to groundwater and determining whether contamination present in soil and sediment may potentially impact groundwater quality at the site.

41

42 Maximum SRC concentrations identified in surface and subsurface soil were evaluated using a series 43 of generic screening steps to identify initial CMCOPCs. Initial CMCOPCs for soil were further 44 evaluated using the SESOIL model to predict leaching concentrations and identify final CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria among USEPA MCLs, USEPA tap water RSLs, or RVAAP groundwater FWCUGs for the Resident Receptor Adult. These final CMCOPCs were evaluated using the AT123D model to predict groundwater concentrations beneath source areas and at the nearest downgradient groundwater receptor to the AOC (e.g., Cobbs Ponds) and identify CMCOCs.

6

A sediment screening analysis was performed for sediment samples at the AOC. Chemical-specific DAFs were calculated using co-located surface water and sediment concentrations for identified sediment SRCs. These DAFs were used in the sediment screening analysis to identify the initial CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria. The initial CMCOPCs for sediment were further evaluated by comparing their maximum detected surface water concentrations against their respective MCL/RSLs to identify the final sediment CMCOPCs.

- 14
- 15 The modeling results identified the following CMCOCs for soil and final CMCOPCs for sediment:
- 16 17
- 18 19
- Arsenic, nickel, selenium, and thallium in soil were predicted to exceed the screening criteria in groundwater beneath the source area, and only arsenic was predicted to exceed the screening criteria in groundwater at the downgradient receptor location.
- Hexavalent chromium in Upper Cobbs Pond and Lower Cobbs Pond sediment was predicted to exceed the screening criteria in groundwater beneath the source areas.
- 21 22

20

A qualitative assessment of the sample results and considerations of the limitations and assumptions of the models were performed to identify if any soil CMCOCs may impact the groundwater beneath the source or at the downstream receptor location (Cobbs Ponds) or if any CMCOPCs in sediment may impact groundwater beneath Cobbs Ponds. This qualitative assessment concluded that the remaining soil CMCOCs and sediment CMCOPCs are not expected to adversely impact groundwater at the AOC. No further action is required of soil and sediment at the Upper and Lower Cobbs Ponds AOC for the protection of groundwater.

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- 31 32

8.5 SUMMARY AND CONCLUSIONS OF THE HUMAN HEALTH RISK ASSESSMENT

The HHRA identified COCs and conducted risk management analysis to determine if COCs pose unacceptable risk to the Resident Receptor. If there is no unacceptable risk to the Resident Receptor, it can be concluded that there is no unacceptable risk to the National Guard Trainee and Industrial Receptor.

37

Media of concern at the Upper and Lower Cobbs Ponds AOC are surface soil, subsurface soil, surface water, and sediment. Soil data were aggregated into surface and subsurface soil. Surface water and sediment data were aggregated into three EUs (Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond).

- 42
- No COCs were identified in soil. Several PAHs were identified as COCs in sediment at the backwater
 area. The MDC of benzo(a)pyrene (0.89 mg/kg) at the Backwater Area exceeds the FWCUG (0.221

1 mg/kg). The MDC is at ULCPsd-015 collected in 2001. Sample ULCPsd-047 was collected at 2 approximately the same location in 2010 and showed a benzo(a)pyrene concentration of 0.39 mg/kg. 3 Detected benzo(a)pyrene concentrations in the other Backwater Area samples ranged from 0.063-4 0.23 mg/kg. Three additional PAHs, present below FWCUGs, were identified as COCs based on the 5 SOR analysis. Benzo(a)pyrene, and by association the other PAHs, were not identified as COCs to be carried forward for potential remediation because the elevated PAH concentrations are located near a 6 7 culvert between Load Line 3 Road and Track 33. The source of the PAHs in sediment is most likely 8 runoff from the road and railroad bed. The FWCUGs for exposure to sediment are the same as the 9 FWCUGs for soil (i.e., exposure is assumed to be the same for soil in a residential yard and sediment 10 in a pond). However, the MDC at the backwater area (0.89 mg/kg) is below a more realistic (but still 11 conservative) calculated RSL of 9 mg/kg.

12

The EPC of arsenic in sediment at Lower Cobbs Pond (26.9 mg/kg) exceeds the FWCUG for the Resident Receptor Adult of 4.25 mg/kg, the facility-wide background concentration of 19.5 mg/kg, and the Ohio EPA SRV of 25 mg/kg. The MDC of arsenic (34.3 mg/kg) at ULCPsd-026, collected in 2001, is the only sample concentration greater than the facility-wide background concentration and Ohio EPA SRV. Reported arsenic concentrations in the other Lower Cobbs Pond samples ranged from 5.1–20.2 mg/kg. Arsenic was not identified as a COC for potential remediation because the reported concentrations appear to represent naturally occurring levels in sediment.

20

The EPC of cobalt (0.01 mg/L) is less than two times the tap water RSL of 0.006 mg/L. The Backwater Area is a shallow pond created through beaver dam construction activity upstream of Upper Cobbs Pond. This limited shallow surface water is not a potential source of residential drinking water; therefore, this low exceedance of the tap water RSL does not warrant potential remediation in this EU.

26

The EPC of manganese in surface water at the Backwater Area EU exceeds the FWCUG for the Resident Receptor (Adult and Child) of 6.326 mg/L. The EPC is strongly influenced by the elevated concentration reported in a single sample collected in 2001. The manganese concentrations in all samples collected in 2010 are below the facility-wide background concentration of 0.391 mg/L. Because recent sample data indicate manganese is present at naturally occurring background concentrations, manganese does not warrant potential remediation in this EU.

33

Based on the risk management analysis, no COCs were identified to be carried forth in an FS for the
Resident Receptor (Adult and Child) in any of the media of concern at the Upper and Lower Cobbs
Pond AOC; therefore, no other receptors were evaluated and no further action is recommended from a
human health risk perspective.

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- 39 40

8.6 SUMMARY AND CONCLUSIONS OF THE ECOLOGICAL RISK ASSESSMENT

The Level I ERA presents important ecological resources on or near the AOC and evaluates the potential for current contamination to impact ecological resources. There is chemical contamination present in surface soil, sediment, and surface water at Upper and Lower Cobbs Ponds AOC. This contamination was identified using historical and PBA08 RI data and data from the 2003 FWBWQS.

| 1 | Although the ponds constitute much of the AOC, red maple successional forest (dominant vegetation |
|----------|--|
| 2 | type), dry, mid-successional, cold-deciduous shrubland, three types of herbaceous communities, and |
| 3 | three additional types of forests were observed on the 39 acres in the AOC. There are important and |
| 4 | significant ecological resources in the AOC. Specifically, a state-listed threatened species (Least |
| 5 | Bittern) and its habitat, wetlands, and surface water (i.e., ponds) are present and near contamination. |
| 6 | These findings invoked a requirement of a Level II ERA. |
| 7 | |
| 8 | The Level II ERA evaluated integrated COPECs in soil, sediment, and surface water. Integrated |
| 9 | COPECs are identified by screening PBA08 RI, FWBWQS, and historical data sets against ESVs. |
| 10 11 | Twelve integrated COPECs were identified for soil. Forty integrated COPECs were identified for sediment. Eight integrated COPECs were identified for surface water. The integrated soil, sediment, |
| 12 | and surface water COPECs were further evaluated with technical and refinement factors in Step 3A. |
| 12 | The factors in Step 3A showed there are no integrated COPECs that are of ecological concern |
| 14 | requiring remediation or further evaluation. Consequently, the Level II Screening ERA for the Upper |
| 15 | and Lower Cobbs Ponds AOC concludes with a recommendation that no further action is required to |
| 16 | be protective of important ecological resources. |
| 17 | |
| 18 | 8.7 UPDATED CONCEPTUAL SITE MODEL |
| 19 | |
| 20 | The CSM is presented in this section to incorporate results of this RI. Elements of the CSM include: |
| 21 | |
| 22 | • Primary and secondary contaminant sources and release mechanisms, |
| 23 | • Contaminant migration pathways and discharge or exit points, |
| 24 | • Potential receptors with unacceptable risk, and |
| 25 | • Data gaps and uncertainties. |
| 26 | |
| 27 | The following sections describe each of the above elements of the CSM for the Upper and Lower |
| 28 | Cobbs Ponds AOC, and the CSM is presented on Figure 8-1. In addition, figures contained in earlier |
| 29 | sections of the report that illustrate AOC features, topography, groundwater and surface water flow |
| 30 | directions, and nature and extent of SRCs are cited to assist in visualizing key summary points of the |
| 31 | revised CSM. |
| 32 | |
| 33 | 8.7.1 Primary and Secondary Contaminant Sources and Release Mechanisms |
| 34 | |
| 35 | No primary contaminant sources (e.g., operational facilities) were located at the AOC. The ponds |
| 36 | themselves functioned as sedimentation basins receiving storm water runoff and effluent discharges |
| 37 | during historical operations at Load Line 3 and Load Line 12. The Installation Assessment identified |
| 38 | the pink wastewater and washout of residue, dusts, and spills at Load Lines 3 and 12, which were |
| 39 40 | discharged through surface drainage channels towards the AOC (USATHAMA 1978). Remnant |
| 40 41 | contamination in Pond Bank soil and sediment within the AOC is considered as a secondary source of contamination. |
| 41 42 | |
| 42 43 | A fish kill occurred on November 15, 1966 at Lower Cobbs Pond as a result of improper handling of |
| | The second of th |

1 12 was removed and disposed in Ramsdell Quarry. The pond was settled and drained, and the 2 contaminants were moved to Ramsdell Quarry.

3

In soil collected from the pond banks and surface water, contamination is not evident. Remnant contamination in sediment within the AOC is considered a secondary source of contamination. RI evaluations indicate sediment contamination is present in the Backwater Area, Upper Cobbs Pond, and Lower Cobbs Pond. Shallow sediment (0–0.5 ft bgs) in the Backwater Area and Upper Cobbs Pond contained generally higher chemical concentrations than the deeper sediment interval (0.5–2 ft bgs).

10

11 The primary mechanisms for release of contaminants from secondary sources at the AOC include:

12 13

14

- Eroding soil matrices with sorbed contaminants and mobilization in overland surface water storm runoff during heavy rainfall conditions,
- Dissolving soluble contaminants and transport in perennial surface water conveyances and
 intermittent surface water runoff,
- Re-suspending contaminated sediment during periods of high flow with downstream
 transport within the surface water system, and
 - Leaching contaminants leaching to groundwater.
- 19 20 21
- 8.7.2 Contaminant Migration Pathways and Exit Points
- 23 24

22

8.7.2.1 <u>Surface Water Pathways</u>

25 Contaminant migration from soil sources via surface water occurs primarily through: (1) particle-26 bound contaminants moving in surface water runoff, and (2) dissolved constituents being transported 27 in surface water. Upon reaching portions of surface water conveyances where flow velocities 28 decrease, particle-bound contaminants will settle out as sediment accumulation. Bank soil data 29 indicate few SRCs and minimal dispersal of contaminants from secondary sources into the AOC via 30 this pathway. Sediment-bound contaminants may become re-suspended and migrate during storm 31 events or may partition to dissolved phase in surface water. As noted in Section 3.0, surface water 32 flows into the AOC via drainage conveyances from Load Line 3, Load Line 12, and miscellaneous 33 small ditches alongside roads and railroad trackbeds. Flow through the surface water system is 34 constrained at four culverts beneath several railroad trackbeds and Load Line No. 3 Road (Figure 3-35 1). Surface water exits the system via a stand pipe and overflow spillway at the north end of Lower 36 Cobbs Pond and discharges into a tributary to Sand Creek. The culverts and Lower Cobbs Pond dam 37 serve to decrease flow velocities at multiple points within the system and promote deposition of any 38 suspended sediment. Beaver dams and heavy aquatic vegetation within the upper reaches of the 39 system also promote deposition of sediment, sorption of dissolved phase contaminants into the 40 sediment zone, and decay of organic biomass and bio-uptake.

1 8.7.2.2 **Groundwater Pathways**

2

3 The estimated direction of groundwater flow at the AOC is from southeast to northwest. This reflects 4 the January 2010 facility-wide potentiometric data presented in the Facility-wide Groundwater 5 Monitoring Program Report on the January 2010 Sampling Event (EQM 2010a). Water level elevations at the AOC range from 957.38-973.40 ft amsl. Potentiometric data indicate the 6 7 groundwater table occurs within unconsolidated zone throughout the AOC. Groundwater discharge to 8 surface water features (e.g., via base flow to the ponds) occur within the AOC boundary. Surface 9 water exits Lower Cobbs Pond via a dam overflow spillway, which leaves the AOC and flows into an 10 unnamed tributary to Sand Creek.

11

12 The FWGWMP was initiated in 2005 and involves quarterly sampling of selected wells within the 13 former RVAAP. Between 2009-2013, several sampling events under the FWGWMP collected 14 groundwater data at the Upper and Lower Cobbs Ponds AOC.

15

16 Contaminant leaching pathways from soil to the water table are through unconsolidated soil 17 representing silty loam with low permeability. The overall average hydraulic conductivity of the 18 unconsolidated soil is approximately 5.64E-06 cm/s. Conservative transport modeling indicated four 19 chemicals (arsenic, nickel, selenium, and thallium) may leach from soil and migrate to the 20 groundwater table at concentrations exceeding MCLs/RSLs beneath their respective sources; 21 however, only arsenic was predicted to exceed the screening criteria in groundwater at the 22 downgradient receptor location. None of these chemicals except arsenic (which is determined to be 23 background related) were detected in AOC groundwater samples collected from 2009–2013 above 24 their respective groundwater criteria; therefore, this evaluation concludes that the model-predicted 25 concentrations are conservative. For sediment, the screening indicated that hexavalent chromium may 26 be a leachability concern; however, this assessment was conservative because hexavalent chromium 27 was not analyzed in surface water or groundwater and did not allow for a chemical-specific DAF to 28 be calculated. A qualitative assessment of the sample results was performed and the limitations and 29 assumptions of the models were considered to identify if any CMCOCs are present in soil or sediment 30 at the Upper and Lower Cobbs Ponds AOC that may potentially impact groundwater. This qualitative 31 assessment concluded that CMCOCs are not adversely impacting groundwater quality based on 32 current data and are not predicted to have future impacts. No further action is required of soil or 33 sediment to be protective of groundwater.

34

35 8.7.3 **Potential Human Receptors and Ecological Resources**

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- 37

In February 2014, the Army and Ohio EPA amended the risk assessment process to address changes 38 in the RVAAP restoration program. The Technical Memorandum identified three Categorical Land 39 Uses and Representative Receptors to be considered during the RI phase of the CERCLA process.

1 These three Land Uses and Representative Receptors are presented below.

2. Military Training Land Use – National Guard Trainee.

- 2 3
- 1. Unrestricted (Residential) Land Use Resident Receptor (Adult and Child) (formerly called Resident Farmer).
- 4 5
- 6
- 7

8 Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp 9 Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then 10 the AOC is also considered to have met the requirements of the other Land Uses (i.e., 11 Commercial/Industrial and Military Training), and the other Land Uses do not require evaluation. The 12 HHRA did not identify Resident Receptor COCs to be carried forward for potential remediation; 13 therefore, the Upper and Lower Cobbs Ponds AOC is considered protective for all potential human 14 health receptors.

3. Commercial/Industrial Land Use – Industrial Receptor (USEPA Composite Worker).

15

16 Camp Ravenna has a diverse range of vegetation and habitat resources. Habitats present within the facility include large tracts of closed-canopy hardwood forest, scrub/shrub open areas, grasslands, 17 18 wetlands, open-water ponds and lakes, and semi-improved administration areas. An abundance of 19 wildlife is present on the facility: 35 species of land mammals, 214 species of birds, 41 species of 20 fish, and 34 species of amphibians and reptiles have been identified. Ecological resources at the 21 Upper and Lower Cobbs Ponds AOC were compared to the list of important ecological places and 22 resources. The ponds and wetlands are important/significant ecological resources. In addition, the 23 Least Bittern (*Ixobrychus exilis*), a state of Ohio threatened species, has been observed at the AOC.

24

25 8.7.4 Uncertainties

26

Uncertainties are inherent in the CSM depending on the density and availability of data. The CSM for the Upper and Lower Cobbs Ponds AOC is overall well defined using existing data, and major data gaps do not remain to be resolved. However, some uncertainties for the CSM include the following:

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- The potential for a continuing flux of contamination from sources in the Upper Cobbs Pond and Lower Cobbs Pond drainage basin headwater areas – from Load Line 3 in particular. Remedial actions for portions of Load Line 3 soil (outside of building footprints) have been completed, and Load Line 3 building demolition actions have been conducted. Disturbed portions of the Load Line 3 were re-vegetated following building demolition, and the vegetation is in early stages of recovery. Soil remediation actions at Load Line 12 were also completed 2010, and Load Line 12 building demolition actions were completed in 1999.
- A fish kill occurred on November 15, 1966 at Lower Cobbs Pond as a result of improper handling of aluminum chloride during manufacturing operations. The bulk of the aluminum chloride at Load Line 12 was removed and disposed in Ramsdell Quarry. The pond was settled and drained, and the contaminants were moved to Ramsdell Quarry. Documentation of this fish kill and subsequent cleanup is limited.
- 43

The lack of established RVAAP-specific background concentrations for identifying SRCs for
 PAHs is a source of uncertainty. Evaluating potential former RVAAP process-related sources
 and other common anthropogenic sources using available PAH environmental data minimizes
 the impact of this uncertainty on the conclusions of the RI.

While this RI addresses soil, sediment, and surface water, additional ongoing investigations

5 6

7

8.8 **RECOMMENDATION OF THE REMEDIAL INVESTIGATION**

are being conducted for the Facility-wide Groundwater AOC.

8 9

10 Based on the investigation results, the Upper and Lower Cobbs Ponds AOC has been adequately 11 characterized and the recommended path forward is no further action for soil, sediment, and surface 12 water to attain Unrestricted (Residential) Land Use. Further investigation is not warranted at this 13 AOC for the following reasons: (1) the current nature and extent of impacted media has been 14 sufficiently characterized; (2) the fate and transport modeling did not identify soil CMCOCs requiring 15 further evaluation or remediation to protect groundwater; (3) there are no CERCLA release-related 16 human health COCs identified in soil, sediment, or surface water requiring further evaluation in an FS 17 or additional remediation; and (4) remedial actions to protect ecological resources are not warranted.

18

19 The next step in the CERCLA process is to prepare a PP to solicit public input with respect to no 20 further action for soil, sediment, and surface water. The PP will briefly summarize the history, 21 characteristics, risks, and the basis for no further action. Comments on the PP received from state and 22 federal agencies and the public will be considered in preparing a ROD to document the final remedy.

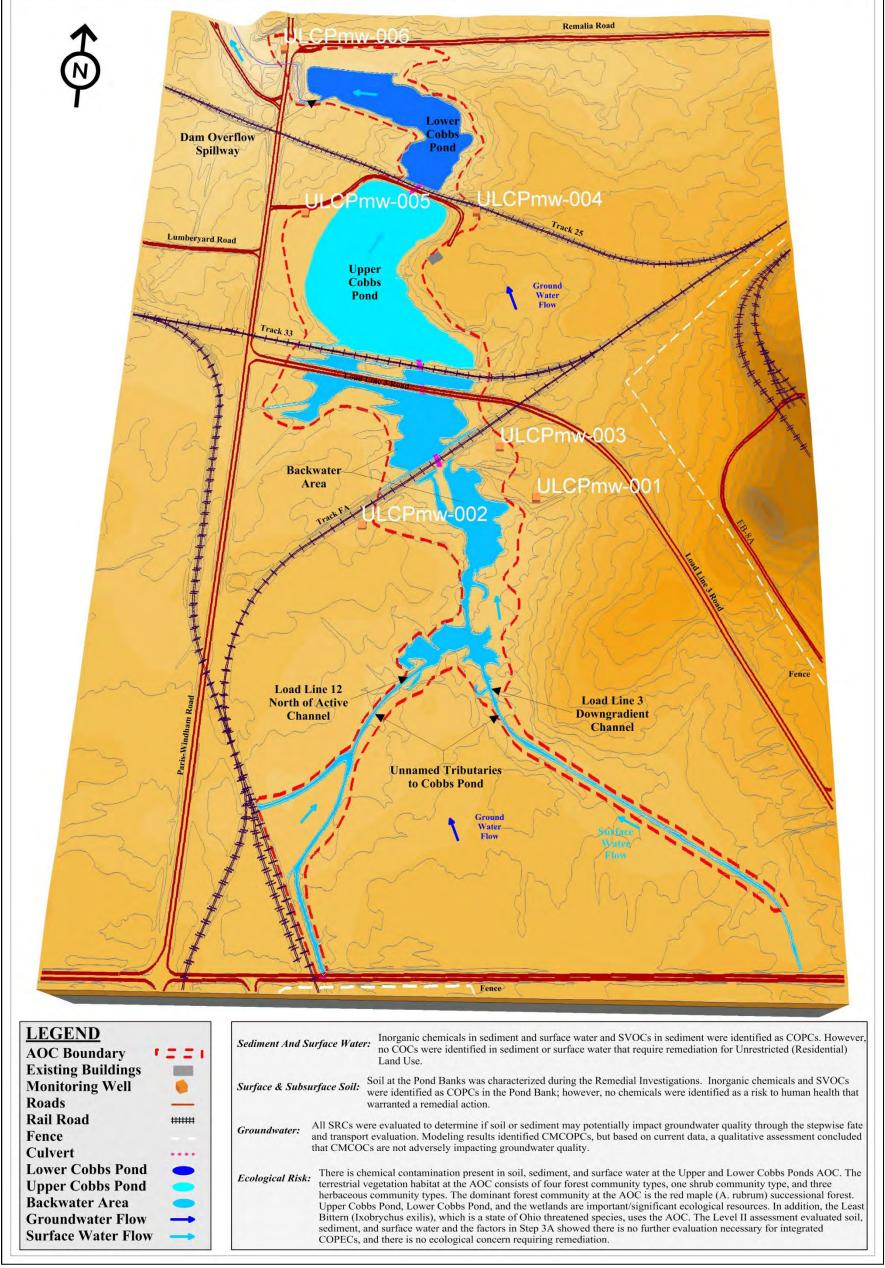
23 The ROD will also include a responsiveness summary addressing comments received on the PP.

Groundwater:

Potentiometric data indicate the groundwater flow direction is to the northwest with an average hydraulic gradient of 0.011 ft/ft. The water table in the unconsolidated bank soil occurs at approximately 5 ft bgs (ranging from 1 ft bgs to greater than 13 ft bgs) or around 965 ft amsl as observed in subsurface borings installed during the PBA08 RI.

Surface Water:

The topography at the AOC is dominated by the ponds and their associated drainage conveyances. Perennial surface water is present within Upper Cobbs Pond, Lower Cobbs Pond, and the Backwater Area. Surface water flows south to north and is received at the Upper and Lower Cobbs Ponds AOC from the Load Line 12 North of Active Area Channel and Load Line 3 Dowgradient Channel. Surface water then flows through the Backwater Area to Upper Cobbs Pond the to Lower Cobbs Pond, ultimately exiting the AOC through a dam overflow spillway located near the intersection of Remalia and Paris-Windham Roads. Surface water then enters an unnamed tributary that discharges into Sand Creek, northwest of the AOC.



Sewers: There are no storm or sanitary sewers present in the Upper and Lower Cobbs Ponds AOC.

Figure 8-1. Upper and Lower Cobbs Ponds Conceptual Site Model

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9.0 AGENCY COORDINATION AND PUBLIC INVOLVEMENT

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The Army is the lead agency responsible for executing the CERCLA process and ultimately completing an approved ROD for soil, sediment, and surface water at the Upper and Lower Cobbs Ponds AOC. This section reviews actions that have been conducted and presents activities that are planned to ensure the regulatory agencies and members of the public have been provided with appropriate opportunities to stay informed of the progress of the environmental investigation, restoration efforts, and final selection of a remedy.

10 11

12

9.1 STATE ACCEPTANCE

State acceptance considers comments received from agencies of the state of Ohio on the proposed remedial alternatives. Ohio EPA is the lead regulatory agency for supporting the remedy for soil, sediment, and surface water at the Upper and Lower Cobbs Ponds AOC. This Phase III RI Report has been prepared in consultation with Ohio EPA.

17

Ohio EPA provided input during the ongoing investigation and report development to ensure the remedy ultimately selected is protective of human health and the environment and fulfills the requirements of the DFFO (Ohio EPA 2004). Ohio EPA will provide comments on this report and the subsequent PP and ROD. The Army will obtain Ohio EPA concurrence prior to the final selection of the remedy for soil, sediment, and surface water at the AOC.

23

24 9.2 COMMUNITY ACCEPTANCE

25

26 Community acceptance considers comments provided by community members for each proposed remedial alternative. CERCLA 42 U.S.C. 9617(a) emphasizes early, constant, and responsive 27 28 community relations. The Army has prepared a Community Relations Plan for the Ravenna Army 29 Ammunition Plant Restoration Program (Vista 2016) to facilitate communication between the former 30 RVAAP and the community surrounding Ravenna, Ohio, during environmental investigations and 31 potential remedial action. The plan was developed to ensure the public has convenient access to 32 information regarding project progress. The community relations program interacts with the public 33 through news releases, public meetings, public workshops, and Restoration Advisory Board meetings 34 with local officials, interest groups, and the general public.

35

36 CERCLA 42 U.S.C. 9617(a) requires an Administrative Record to be established "at or near the
 37 facility at issue." Relevant documents regarding the former RVAAP have been made available to the
 38 public for review and comment.

| 1 | The Administrative Record for this project is available at the following location: |
|----|--|
| 2 | |
| 3 | Camp Ravenna |
| 4 | Environmental Office |
| 5 | 1438 State Route 534 SW |
| 6 | Newton Falls, OH 44444 |
| 7 | |
| 8 | Access to Camp Ravenna is restricted but can be obtained by contacting the environmental office at |
| 9 | (614) 336-6136. In addition, an Information Repository of current information and final documents is |
| 10 | available to any interested reader at the following libraries: |
| 11 | |
| 12 | Reed Memorial Library |
| 13 | 167 East Main Street |
| 14 | Ravenna, Ohio 44266 |
| 15 | |
| 16 | Newton Falls Public Library |
| 17 | 204 South Canal Street |
| 18 | Newton Falls, Ohio 44444-1694 |
| 19 | |
| 20 | Additionally, there is an online resource for restoration news and information. This website is |
| 21 | available at: www.rvaap.org. |
| 22 | |
| 23 | Comments will be received from the community upon issuing the PP. As required by the CERCLA |
| 24 | regulatory process and Community Relations Plan (Vista 2016), the Army will hold a public meeting |
| 25 | and request public comments on the PP for the Upper and Lower Cobbs Ponds AOC. These |
| 26 | comments will be considered mice to the final selection of a nemedy. Despenses to these comments |

comments will be considered prior to the final selection of a remedy. Responses to these comments 26 will be addressed in the responsiveness summary of the ROD. 27

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