Revised Draft

Phase II Remedial Investigation Report for Soil, Sediment, and Surface Water at RVAAP-44 Load Line 11

Former Ravenna Army Ammunition Plant Portage and Trumbull Counties, Ohio

Contract No. W912QR-15-C-0046

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U.S. Army Corps of Engineers Louisville District

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Leidos has completed the Revised Draft Phase II Remedial Investigation Report for Soil, Sediment, and Surface Water at RVAAP-44 Load Line 11 at the Former Ravenna Army Ammunition Plant, Portage and Trumbull Counties, Ohio. Notice is hereby given that an independent technical review has been conducted that is appropriate to the level of risk and complexity inherent in the project. During the independent technical review, compliance with established policy principles and procedures, utilizing justified and valid assumptions, was verified. This included review of data quality objectives; technical assumptions; methods, procedures, and materials to be used; the appropriateness of data used and level of data obtained; and reasonableness of the results, including whether the product meets the customer's needs consistent with law and existing U.S. Army Corps of Engineers (USACE) policy.

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Prepared for:

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June 22, 2016

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ACRONYMS AND ABBREVIATIONS

ACM Asbestos-containing Material

amsl Above Mean Sea Level

AOC Area of Concern

Army U.S. Department of the Army

AT123D Analytical Transient 1-, 2-, 3-Dimensional

bgs below ground surface
BHC Hexachlorocyclohexane

BHHRA Baseline Human Health Risk Assessment
Camp Ravenna Camp Ravenna Joint Military Training Center

CERCLA Comprehensive Environmental Response, Compensation, and Liability Act

CMCOC Contaminant Migration Chemical of Concern

CMCOPC Contaminant Migration Chemical of Potential Concern

COC Chemical of Concern

COPC Chemical of Potential Concern

COPEC Chemical of Potential Ecological Concern

CSM Conceptual Site Model
DAF Dilution Attenuation Factor

DERR Division of Environmental Response and Revitalization

DFFO Director's Final Findings and Orders

DNT Dinitrotoluene

DQO Data Quality Objective
DRO Diesel Range Organics

EcoSSLEcological Soil Screening LevelEPCExposure Point ConcentrationERAEcological Risk AssessmentESLEcological Screening LevelESVEcological Screening Value

EU Exposure Unit FA Functional Area

FPA Former Production Area

FS Feasibility Study

FWCUG Facility-wide Cleanup Goal

FWERWP Facility-wide Ecological Risk Work Plan

FWGWMP Facility-wide Groundwater Monitoring Program
FWHHRAM Facility-wide Human Health Risk Assessor Manual

FWSAP Facility-wide Sampling and Analysis Plan

gpm Gallons Per Minute
GRO Gasoline Range Organics
GSSL Generic Soil Screening Level

HELP Hydrologic Evaluation of Landfill Performance

HHRA Human Health Risk Assessment

HLC Henry's Law Constant

HMX Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine

HQ Hazard Quotient

ILCR Incremental Lifetime Cancer Risk

INRMP Integrated Natural Resources Management Plan

IRA Interim Removal Action

IRIS Integrated Risk Information System
 IRP Installation Restoration Program
 ISM Incremental Sampling Methodology
 K_d soil/water partitioning coefficient

K_ds distribution coefficients

K_{oc} water/organic carbon partition coefficient

K_{ow} octanol-water partition coefficient
MCL Maximum Contaminant Level
MDC Maximum Detected Concentration

MDL Method Detection Limit
NCP National Contingency Plan
NPA Non-Production Area

OHARNG Ohio Army National Guard

Ohio EPA Ohio Environmental Protection Agency

OMZA Outside Mixing Zone Average
OMZM Outside Mixing Zone Maximum
ORAM Ohio Rapid Assessment Method
PAH Polycyclic Aromatic Hydrocarbon

PBA08 RI Performance-Based Acquisition 2008 Remedial Investigation

PBA08 SAP Performance-Based Acquisition 2008 Supplemental Investigation Sampling and

Analysis Plan Addendum No. 1

PBT Persistent, Bioaccumulative, and Toxic

PCB Polychlorinated Biphenyl PETN Pentaerythritol tetranitrate

PP Proposed Plan
ppm Parts Per Million

PRG Preliminary Remediation Goals

QA Quality Assurance
QC Quality Control

QHEI Qualitative Habitat Evaluation Index

R Retardation Factor

RCRA Resource Conservation and Recovery Act

RDA Recommended Daily Allowance
RDI Recommended Daily Intake

RDX Hexahydro-1,3,5-trinitro-1,3,5-triazine

REIMS Ravenna Environmental Information Management System

RfD Reference Dose

RI Remedial Investigation

RM River Mile

ROD Record of Decision

RRSE Relative Risk Site Evaluation
RSL Regional Screening Level

RVAAP Ravenna Army Ammunition Plant

SAIC Science Applications International Corporation

SERA Screening Ecological Risk Assessment

SESOIL Seasonal Soil Compartment

SL Screening Level SOR Sum-of-Ratios

SRC Site-related Contaminant
SRV Sediment Reference Value
SSL Soil Screening Level

SSSL Site-specific Soil Screening Level SVOC Semi-volatile Organic Compound

TAL Target Analyte List

TEC Threshold Effects Concentration

TNT 2,4,6-Trinitrotoluene

TPH Total Petroleum Hydrocarbons

TR Target Risk

UCL Upper Confidence Limit

URF Unit Risk Factor

USACE U.S. Army Corps of Engineers

USEPA U.S. Environmental Protection Agency

USP&FO U.S. Property and Fiscal Officer

UXO Unexploded Ordnance

VOC Volatile Organic Compound

WOE Weight-of-Evidence WQC Water Quality Criteria THIS PAGE INTENTIONALLY LEFT BLANK.

EXECUTIVE SUMMARY

ES.1 INTRODUCTION AND SCOPE

This document has been revised by Leidos under U.S. Army Corps of Engineers (USACE), Louisville District Contract Number W912QR-15-C-0046. This report replaces the *Draft Phase II Remedial Investigation Report for Soil, Sediment, and Surface Water at RVAAP-44 Load Line 11*, dated January 19, 2012, originally submitted to Ohio Environmental Protection Agency (Ohio EPA). This Remedial Investigation (RI) Report addresses soil, sediment, and surface water at Load Line 11 within the former Ravenna Army Ammunition Plant (RVAAP) [now known as Camp Ravenna Joint

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Military Training Center (Camp Ravenna)] in Portage and Trumbull counties, Ohio.

This report has been prepared in accordance with the requirements of the Ohio EPA *Director's Final Findings and Orders* (DFFO) for RVAAP, dated June 10, 2004 (Ohio EPA 2004a). The DFFO requires conformance with the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) and National Contingency Plan (NCP) to implement an RI to characterize the area of concern (AOC); develop a Feasibility Study (FS) Report (if remediation is necessary); and evaluate remedial alternatives to address contamination presenting unacceptable risk to human health and the environment, present a preferred alternative in a proposed plan (PP), and document stakeholder selection and acceptance of the preferred final remedy in a record of decision (ROD). The following sections present the site history, scope of this report, and an explanation of the evaluation of future use.

ES.1.1 Site History

Load Line 11, formerly known as Booster Line #1, is an approximately 48-acre fenced AOC located immediately north and west of Fuze and Booster Spur Road and south of Newton Falls Road in the south-central portion of Camp Ravenna. Load Line 11 was formerly used for producing artillery primers and fuzes. Remaining features at Load Line 11 include a one-lane asphalt perimeter road that enters the AOC from the south and encircles 75% of the former production area (FPA) and an asphalt parking area remains by former Building AP-11.

From 1941–1945, Load Line 11 operated at full capacity to produce artillery primers. Load Line 8 was deactivated at the end of World War II, and the process equipment was placed on standby. From 1951-1957, Load Line 11 was reactivated to produce primers. From 1969-1971, it was reactivated to produce MR ZA4 fuzes. Load Line 11 was finally deactivated and all process equipment was removed in 1971. No historical information exists to indicate Load Line 11 was used for any other processes, other than what is presented above.

ES.1.2 Scope

The scope of this Phase II RI Report is to perform a CERCLA evaluation of soil, sediment, and surface water at Load Line 11 using available RI data to evaluate the nature and extent of contamination; fate and transport of contaminants in the environment; and risk assessments for

surface and subsurface soil, sediment, and surface water. This report includes sample results and information from the Phase I RI and Interim Removal Action (IRA), and also provides a summary of the Performance-Based Acquisition 2008 Remedial Investigation (PBA08 RI) at Load Line 11 that was performed to supplement data from previous sampling events.

1 2

The media of concern associated with Load Line 11 are surface soil [0–1 ft below ground surface (bgs)], subsurface soil (1–13 ft bgs), sediment, and surface water. This report does not include a full evaluation of groundwater or facility-wide sewers, as those will be evaluated as individual AOCs for the entire facility. However, the potential for soil contaminants to leach to and migrate in groundwater is evaluated in this RI Report to determine whether additional soil remedial actions to protect groundwater may be necessary.

ES.1.3 Evaluation of Future Use

In February 2014, the U.S. Department of the Army (Army) and Ohio EPA amended the risk assessment process to address changes in the RVAAP restoration program. The *Final Technical Memorandum: Land Uses and Revised Risk Assessment Process for the RVAAP Installation Restoration Program* (ARNG 2014) (herein referred to as the Technical Memorandum) identified three Categorical Land Uses and Representative Receptors to be considered during the RI phase of the CERCLA process. These three Land Uses and Representative Receptors are presented below.

1. Unrestricted (Residential) Land Use – Resident Receptor (Adult and Child) (formerly called Resident Farmer).

2. Military Training Land Use – National Guard Trainee.

3. Commercial/Industrial Land Use – Industrial Receptor [U.S. Environmental Protection Agency's (USEPA) Composite Worker].

Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then the AOC is also considered to have met the requirements of the other Land Uses (i.e., Industrial and Military Training), and those other Land Uses do not require evaluation.

As stated in Paragraph 6.d of the Technical Memorandum, if an AOC fails to meet Unrestricted (Residential) Land Use, then an FS will be completed that evaluates cleanup options for all three Land Uses [i.e., Unrestricted (Residential) Land Use, Military Training Land Use, and Commercial/Industrial Land Use]. Remedial alternatives for meeting each Land Use are to be evaluated per the current guidelines for selecting a remedy for the AOC. The preferred remedy is one that would meet Unrestricted (Residential) Land Use. RI/FS Reports in progress at the time of the Technical Memorandum's approval on February 11, 2014 will not be revised to include an evaluation of Commercial/Industrial Land Use as an Alternative if it achieves no further action for Unrestricted

41 (Residential) Land Use.

ES.2 INTERIM REMOVAL ACTION

In 2001, an IRA was completed at Load Line 11. The IRA was initiated following the Phase I RI activities as an early response action to remove the primary pathways for off-AOC contaminant migration. The IRA included removing sump water from production buildings, grouting selected sanitary sewer manholes, and performing limited excavations from open ditch systems draining the AOC (MKM 2004a).

Sumps located adjacent to Buildings AP-3, AP-5, AP-6, and AP-8 were excavated, removed, and disposed during the 2001 IRA. Approximately 15,000 gal of water were removed from sumps and sewer manholes downgradient of each sump and were filled with bentonite cement to prevent water from infiltrating back into the sumps during excavation and removal operations. Once the water was removed, effluent sewer lines were cut and plugged with mechanical packers and cement grout, which prepared the sumps for removal.

Six drainage ditch locations were excavated based on Phase I RI analytical data showing exceedances of background concentrations and/or Region 9 preliminary remediation goals (PRGs) for residential soil. Elevated concentrations of contaminants in the six ditch locations included metals in all six ditches, volatile organic compounds (VOCs) in two ditches, and semi-volatile organic compounds (SVOCs) and/or pesticides in only one ditch. A total of 230 cubic yards of contaminated soil were removed during the ditch excavation operations. Confirmation samples were collected to verify the removal of contaminated soil and sediment.

One 30 by 30 by 8 ft hot spot area, located in an open field north of Building AP-17, was excavated due to petroleum contamination encountered during RI soil boring activities (MKM 2004a). A total of 130 yd³ of petroleum-contaminated soil were excavated. Confirmation samples were used to verify the removal of the petroleum-contaminated soil and the excavations were backfilled with approved offsite soil. During the excavation of the hot spot, it was determined that the area needed to be assessed for unexploded ordnance (UXO). Upon excavating test trenches, cable, scrap metal, bolts, and magnetite-containing rock were found. Confirmation samples were then collected from the test trenches. After excavation and determining no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation.

ES.3 FINDINGS AND RECOMMENDATIONS OF THE REMEDIAL INVESTIGATION

This section presents a summary of data used in this RI, contaminant nature and extent, fate and transport, human health risk assessment (HHRA), and environmental risk assessment (ERA), followed by the conclusions of the RI.

ES.3.1 Data Use and Sample Selection Process

Quality-assured sample data from the Phase I RI (MKM 2005a), confirmation results from the IRA (MKM 2004a), and 2010 and 2012 PBA08 RI were used to evaluate nature and extent of contamination at Load Line 11. These investigations used discrete sampling methods.

All available sample data collected at the locations were evaluated to determine suitability for use in various key RI data screens and evaluations (i.e., nature and extent, fate and transport, risk assessment). Evaluation of the data's suitability for use in the PBA08 RI involved two primary considerations: (1) whether the data represented current AOC conditions, and (2) sample collection methods [e.g., discrete vs. incremental sampling methodology (ISM)].

1 2

Samples from the Phase I RI and IRA data sets were evaluated to determine if conditions had changed substantively between earlier characterization efforts and the 2010 and 2012 PBA08 RI, as building demolition activities occurred in 2004-2005 after the 2000 and 2001 Phase I RI and IRA confirmation sampling. The samples collected in 2000-2001 were collected within ditch lines adjacent to former buildings and in areas encompassing, but also extending substantially beyond the footprint of the former buildings. The 2001 confirmation sampling was conducted within the footprints of the removal areas after the contaminated soil was removed. Therefore, both data sets were considered representative of current conditions within and surrounding the footprints of the former buildings and removal areas at Load Line 11.

Data collected in 2010 and 2012 as part of the PBA08 RI focused on delineating the extent of contaminants identified in surface soil (0-1 ft bgs), subsurface soil (1-13 ft bgs), sediment, and surface water. The PBA08 RI sampled locations with the greatest likelihood of contamination (e.g., adjacent to production buildings or within sediment accumulation areas, such as ditches) and analyzed for chemicals identified in historical investigations. Additionally, sediment and surface water samples were collected from drainage ditches exiting the AOC.

ES.3.2 Summary of Nature and Extent of Contamination

Nature and extent of contamination in surface soil (0-1 ft bgs), subsurface soil (greater than 1 ft bgs), sediment, and surface water subsequent to the 2001 IRA was evaluated in this Phase II RI. Data from the Phase I RI, IRA, and 2010 and 2012 PBA08 RI effectively characterized the nature and extent of the contamination at the AOC. Figure ES-1 shows the sample locations used to conduct this RI. To support the evaluation of nature and extent of contamination, site-related contaminant (SRC) concentrations were compared to screening levels (SLs) corresponding to the lowest facility-wide cleanup goal (FWCUG) for the Resident Receptor (Adult and Child) and National Guard Trainee at a target hazard quotient (HQ) of 0.1 or target risk (TR) of 1E-06, as presented in the Facility-wide Human Health Cleanup Goals for the Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 2010a) (herein referred to as the FWCUG Report). It can be concluded that the vertical and horizontal extent of contamination is defined, and no further sampling is needed to evaluate Load Line 11.

ES.3.2.1 Soil

Locations where explosives were identified as potential contaminants from previous use were thoroughly evaluated, including around former process buildings and across the AOC as a whole. The maximum concentrations of the explosives and propellants were all below their respective SLs and were not considered chemicals of potential concern (COPCs) in the surface and subsurface soil at the FPA and non-production area (NPA).

Arsenic, chromium, lead, and mercury were identified as potential SRCs and as potentially related to the previous site use. When evaluating these chemicals against their SLs [using the trivalent chromium FWCUG for chromium and the regional screening level (RSL) of 400 mg/kg for lead], chromium, lead, and mercury concentrations were all below their SLs; therefore, these chemicals were not considered COPCs for the FPA or NPA in surface and subsurface soil. Arsenic was the only inorganic chemical potentially related to previous AOC operations that is considered a COPC in surface and subsurface soil at the FPA and NPA. Arsenic is considered a COPC in both surface and subsurface soil at the FPA and NPA, with a maximum detected concentration (MDC) of 44.1 mg/kg observed from 6–8 ft bgs at sample location LL11cs-045, which evaluated the effectiveness of the hot spot excavation efforts during the IRA.

With the exception of benzo(a)pyrene at four surface soil sample locations, none of the detected concentrations of SVOCs, VOCs, pesticides, or polychlorinated biphenyls (PCBs) in surface or subsurface soil were above the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. Benzo(a)pyrene is considered a COPC in the surface soil at the FPA and NPA, with an MDC of 0.45 mg/kg at sample location LL11sb-060, which evaluated the drainage ditch northeast of the Entrance Gate.

ES.3.2.2 Sediment and Surface Water

The East Ditch exposure unit (EU) was evaluated with two sediment samples and one surface water sample. No explosives or propellants were detected in the East Ditch sediment and surface water samples. Arsenic was the only inorganic chemical from previous site use that exceeded the SL in the sediment and is considered a COPC. No inorganic chemical concentrations detected in the surface water sample exceeded their respective SLs. One polycyclic aromatic hydrocarbon (PAH) [benzo(a)pyrene] concentration detected in sediment exceeded the SL and was identified as a COPC; however, the benzo(a)pyrene concentration was detected at a concentration below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. SVOCs were not detected in the surface water sample. In addition, no VOCs, pesticides, or PCBs were detected in sediment at the East Ditch. Two pesticides [beta-hexachlorocyclohexane (BHC) and gamma-chlordane] were identified as SRCs at low, estimated concentrations. No VOCs or PCBs were detected in the East Ditch surface water.

The West Ditch EU was evaluated with four sediment samples and one surface water sample. No explosives were detected in West Ditch sediment. The propellant nitrocellulose was detected in the sediment sample LL11sd-096 at a concentration below the SL. No explosives or propellants were detected in the surface water sample. No inorganic chemicals related to previous site use exceeded the SLs except for arsenic in the surface water sample. One PAH [benzo(a)pyrene] exceeded the SL in the West Ditch sediment and was identified as a COPC. Four PAHs [benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] exceeded the SLs in the surface water sample and were identified as COPCs. The PAH concentrations detected in the sediment and surface water were all above the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. No PCBs, pesticides, or VOCs were detected in sediment at the West Ditch. One pesticide

(beta-BHC) was identified as an SRC and was detected at a low, estimated concentration below laboratory detection limits. No VOCs or PCBs were detected in the West Ditch surface water.

The Sewer Outfall EU was evaluated with one sediment sample. The only SRCs identified were the inorganic chemicals sulfate and sulfide, neither of which has established background concentrations for comparison.

One sediment sample and one surface water sample were collected at the off-AOC location LL11sd/sw-082 located on Sand Creek. Inorganic chemicals were not detected above background concentrations in the sediment; however, six inorganic chemicals were detected above their respective background concentrations for surface water. All the detections were at concentrations below their respective SLs. No explosives, propellants, SVOCs, pesticides, or PCBs were detected in the off-AOC sediment. Two VOCs (carbon disulfide and toluene) were present at low, estimated concentrations below laboratory detection limits.

No explosives, propellants, SVOCs, VOCS, pesticides, or PCBs were observed at this Sand Creek location.

ES.3.3 Summary of Contaminant Fate and Transport

All SRCs identified in the surface soil, subsurface soil, and sediment at Load Line 11 were evaluated through the stepwise contaminant fate and transport evaluation. The evaluation included analyzing leaching and migration from soil and sediment to groundwater and determining whether contamination present in soil and sediment may potentially impact groundwater quality at the site.

Maximum concentrations of SRCs identified in surface and subsurface soil were evaluated using a series of generic screening steps to identify initial contaminant migration chemicals of potential concern (CMCOPCs). Initial CMCOPCs for soil were further evaluated using the Seasonal Soil Compartment (SESOIL) model to predict leaching concentrations and identify final CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria among USEPA maximum contaminant levels (MCLs), USEPA tap water RSLs, or RVAAP groundwater FWCUGs for the Resident Receptor Adult. A sediment screening analysis was performed for sediment samples at the AOC. Chemical-specific dilution attenuation factors (DAFs) were calculated using co-located surface water and sediment concentrations for identified sediment SRCs. These DAFs were used in the sediment screening analysis to identify the final CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria. Final CMCOPCs were evaluated using the Analytical Transient 1-, 2-, 3-Dimensional (AT123D) model to predict groundwater concentrations beneath source areas and at the nearest downgradient groundwater receptor to the AOC (e.g., stream).

The evaluation of modeling results with respect to current AOC groundwater data and model limitations identified the following CMCOPCs for soil and sediment:

• Arsenic, barium, cobalt, zinc, benzo(b)fluoranthene, naphthalene, and pentaerythritol tetranitrate (PETN) at the FPA and arsenic, cobalt, manganese, and naphthalene at the NPA were the soil CMCOPCs predicted to exceed the screening criteria in groundwater beneath the source area; however, none of these CMCOPCs were predicted to be above their respective groundwater criteria at the downgradient receptor location.

• Benz(a)anthracene and naphthalene at the East Ditch aggregate and antimony, benz(a)anthracene, benzo(b)fluoranthene, and dibenz(a,h)anthracene at the West Ditch aggregate were predicted to exceed the screening criteria in groundwater beneath the source area; however, none of these CMCOPCs were predicted to be above criteria at the downgradient receptor location.

A qualitative assessment of the sample results and considerations of the limitations and assumptions of the models were performed to identify if any contaminant migration chemicals of concern (CMCOCs) are present in soil and sediment at Load Line 11 that may impact the groundwater beneath the source or at the downstream receptor location. This qualitative assessment concluded that there were no CMCOCs present in soil and sediment that may impact the groundwater beneath the source or at the downstream receptor location. No further action is required of soil and sediment at Load Line 11 for the protection of groundwater.

ES.3.4 Summary of the Human Health Risk Assessment

The HHRA identified chemicals of concern (COCs) and conducted risk management analysis to determine if COCs pose unacceptable risk to the Resident Receptor (Adult and Child). Since the risk management analysis determined there were no unacceptable risks to the Resident Receptor (Adult and Child), it can be concluded that there is no unacceptable risk to the National Guard Trainee and Industrial Receptor.

Media of concern at Load Line 11 are surface soil, subsurface soil, sediment, and surface water. Soil data associated with Load Line 11 were aggregated into surface and subsurface soil at the FPA and NPA. Sediment and surface water were evaluated from ditches within and west of the FPA. The ditches include the East Ditch that flows north and the West Ditch that flows west/northwest and eventually to Sand Creek.

No COCs were identified for the Resident Receptor (Adult and Child) in subsurface soil or sediment. Five PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] were identified as COCs for the Resident Receptor (Adult and Child) in surface soil. The exposure point concentration (EPC) for benzo(a)pyrene in surface soil at the NPA (0.28 mg/kg) is approximately 1.3 times its FWCUG of 0.221 mg/kg. The EPC for this data aggregate is the MDC (at sample LL11sb-067); benzo(a)pyrene concentrations in the remaining NPA samples ranged from a non-detectable concentration to 0.11 mg/kg. Sample LL11sb-067 was collected in 2010 from the northeastern portion of Load Line 11. No load line operations were conducted in this area.

The EPC for benzo(a)pyrene in the FPA (0.219 mg/kg) was slightly less than the FWCUG (0.221 mg/kg). Other PAHs were identified as COCs because they contribute to sum-of-ratios (SORs) of two at the FPA and NPA. The EPCs of benzo(a)pyrene (0.28 mg/kg and 0.221) are slightly above and slightly below the FWCUG of 0.221 and the SORs are two, indicating cumulative cancer risks to a hypothetical future resident on the order of 1E-05. Reported benzo(a)pyrene concentrations above the FWCUG are scattered (i.e., no hot spots were identified); therefore, no COCs were identified for potential additional remediation.

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Four PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] were identified as COCs for the Resident Receptor (Adult and Child) in surface water collected from the West Ditch. The detected concentrations of benz(a)anthracene (0.00035 mg/L), benzo(b)fluoranthene (0.00041 mg/L), and indeno(1,2,3-cd)pyrene (0.00021) in surface water collected from the West Ditch range from 2-4 times their FWCUG of 0.0001 mg/L, and the detected concentration of benzo(a)pyrene (0.00033 mg/L) was 41 times the FWCUG of 0.000008 mg/L. All of the reported concentrations included a "J" qualifier, indicating they were estimated results below the detection limit of the analytical method. PAHs are present in the environment from natural and anthropogenic sources. Since surface runoff is the potential source of contamination to surface water in the West Ditch, the fact that PAH concentrations are not significantly elevated in soil or sediment indicates there is no identifiable source beyond normal levels of these chemicals to surface water due to runoff from roads and other traffic areas. In addition, incidental exposures of the Resident Receptor (Adult and Child) to surface water at the West Ditch that only intermittently holds water are much less than the incidental exposure that would occur due to ingesting drinking water (i.e., 2 L/day for an adult) and dermal contact while swimming and wading. These exposures were incorporated into the development of the FWCUGs. Thus, these low, estimated concentrations of PAHs were not identified as COCs for remediation in surface water.

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Based on the risk management analysis, no COCs were identified to be carried forth in an FS for the Resident Receptor (Adult and Child) in any of the media of concern; therefore, no other receptors were evaluated and no further action is recommended from a human health risk perspective.

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ES.3.5 Summary of the Ecological Risk Assessment

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The Level I ERA presents important ecological resources on or near the AOC and evaluates the potential for current contamination to impact ecological resources. There is chemical contamination present in surface soil, sediment, and surface water at Load Line 11. This contamination was identified using historical and PBA08 RI data. Dry, mid-successional, cold-deciduous shrubland (dominant vegetation type); dry, mid-successional, herbaceous field; and four types of forests were observed on the 48 acres in the AOC. There are important and significant ecological resources in the AOC. Specifically, wetlands and surface water (i.e., two ditches) are present and near contamination. These findings invoked a requirement of a Level II ERA.

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The Level II ERA evaluated integrated chemicals of potential ecological concern (COPECs) in soil, sediment, and surface water. Integrated COPECs are identified by screening PBA08 RI and historical data sets against ecological screening values. Twenty integrated COPECs were identified for soil.

Five integrated COPECs were identified for sediment. Five integrated COPECs were identified for surface water. The integrated soil, sediment, and surface water COPECs were further evaluated with technical and refinement factors in Step 3A. The factors in Step 3A showed there are no integrated COPECs that are of ecological concern requiring remediation or further evaluation. Consequently, the Level II Screening ERA for Load Line 11 concludes with a recommendation that no further action is required to be protective of important ecological resources. This supports the conclusion from the Phase I RI, which stated:

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"No additional action is recommended based on ecological risk because, based on the Step 3a refinement step, it was determined that site-related risks were not great enough to warrant proceeding further into the ecological risk assessment process."

ES.3.6 Conclusions of the Remedial Investigation

Based on the investigation results, Load Line 11 has been adequately characterized and the recommended path forward is no further action for soil, sediment, and surface water at Load Line 11 to attain Unrestricted (Residential) Land Use. Further investigation is not warranted at this AOC for the following reasons: (1) the IRA conducted in 2001 removed contamination from the primary pathways for off-AOC migration; (2) the current nature and extent of impacted media has been sufficiently characterized; (3) the fate and transport modeling did not identify soil CMCOCs requiring further evaluation or remediation to protect groundwater; (4) there are no CERCLA release-related human health COCs identified in soil, sediment, or surface water requiring further evaluation in an FS or additional remediation; and (5) remedial actions to protect ecological resources are not warranted.

The next step in the CERCLA process is to prepare a PP to solicit public input with respect to no further action for soil, sediment, and surface water. The PP will briefly summarize the history, characteristics, risks, and the basis for no further action. Comments on the PP received from state and federal agencies and the public will be considered in preparing a ROD to document the final remedy. The ROD will also include a responsiveness summary addressing comments received on the PP.

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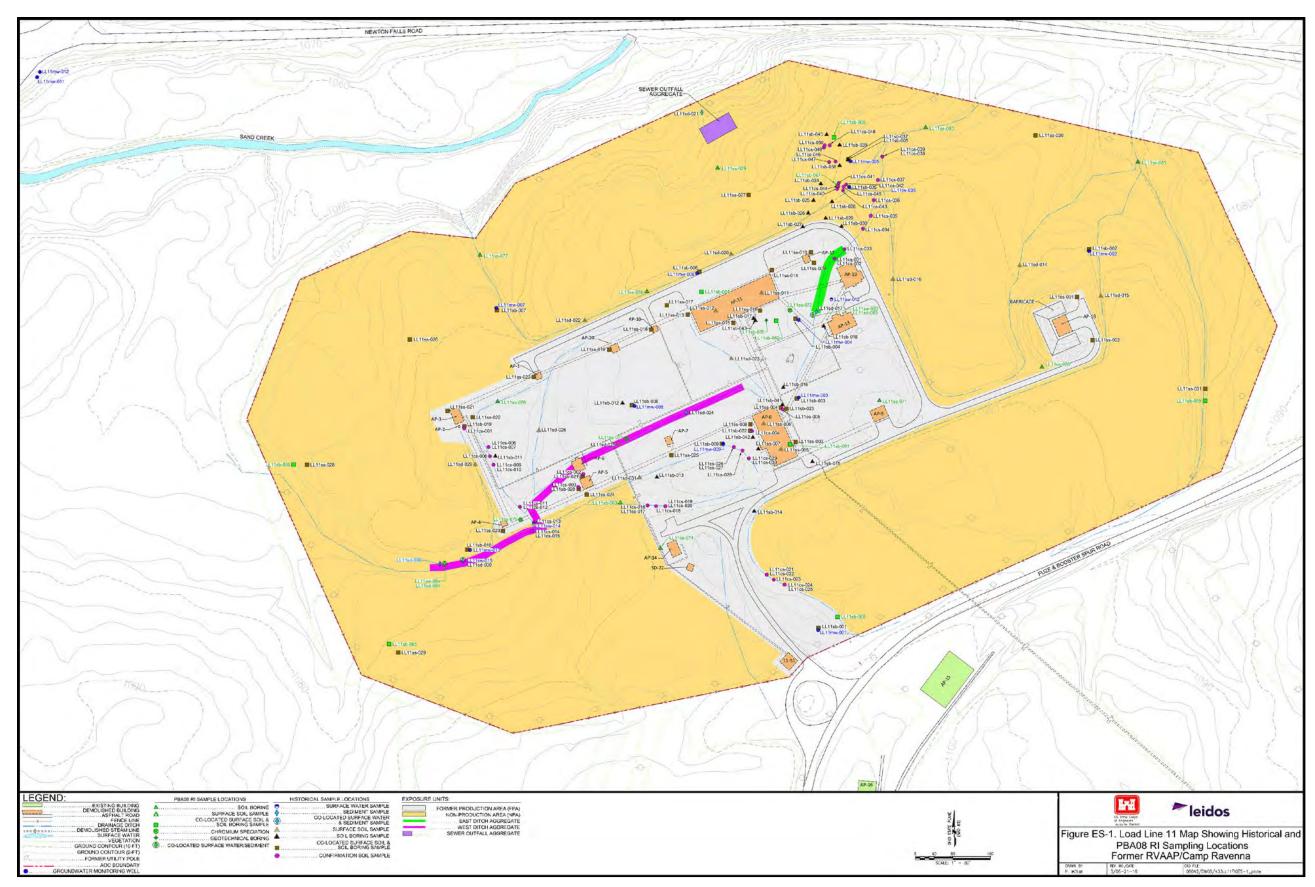


Figure ES-1. Load Line 11 Map Showing Historical and PBA08 RI Sampling Locations - Former RVAAP/Camp Ravenna

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1.0 INTRODUCTION

This document has been revised by Leidos under U.S. Army Corps of Engineers (USACE), Louisville District Contract Number W912QR-15-C-0046. This Phase II Remedial Investigation (RI) Report addresses soil, sediment, and surface water at Load Line 11 within the former Ravenna Army Ammunition Plant (RVAAP) [now known as Camp Ravenna Joint Military Training Center (Camp Ravenna)] in Portage and Trumbull counties, Ohio (Figures 1-1 and 1-2). Load Line 11 is designated as area of concern (AOC) RVAAP-44.

This report has been prepared in accordance with the requirements of the Ohio Environmental Protection Agency (Ohio EPA) *Director's Final Findings and Orders* (DFFO) for RVAAP, dated June 10, 2004 (Ohio EPA 2004a). The DFFO requires conformance with the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) and the National Contingency Plan (NCP) to implement an RI to characterize the AOC, develop a Feasibility Study (FS) Report (if remediation is necessary) and evaluate remedial alternatives to address contamination presenting unacceptable risk to human health and the environment, present a preferred alternative in a proposed plan (PP), and document stakeholder selection and acceptance of the preferred final remedy in a record of decision (ROD).

This Phase II RI Report includes the following components:

• A description of the operational history and environmental setting for the AOC.

A summary of all historical assessments, removal actions, and investigations at Load Line 11, including an Interim Removal Action (IRA) implemented to remove chemical contamination in 2001 and a Phase I RI performed in 2000 and 2001.

 A description of the nature and extent of contamination, including the identification of siterelated contaminants (SRCs) by screening applicable data against background concentrations, essential human nutrients, and frequency of detection/weight-of-evidence (WOE) screening.

• An evaluation of contaminant fate and transport by identifying contaminant migration chemicals of potential concern (CMCOPCs) and contaminant migration chemicals of concern (CMCOCs) that may pose a future threat to groundwater.

• A human health risk assessment (HHRA) to identify chemicals of potential concern (COPCs) and chemicals of concern (COCs).

• An ecological risk assessment (ERA) to identify chemicals of potential ecological concern (COPECs) and chemicals of ecological concern.

• Conclusions of the Phase II RI Report.

1.1 PURPOSE

The purpose of this Phase II RI Report is to use available IRA and RI data to evaluate the nature and extent of contamination; fate and transport of contaminants in the environment; and risk assessments for surface and subsurface soil, sediment, and surface water at Load Line 11. This report includes sample results and information from the Phase I RI and IRA, and also provides a summary of the Performance-Based Acquisition 2008 Remedial Investigation (PBA08 RI) at Load Line 11 that was

performed to supplement data from previous sampling events. Depending on the results of the evaluations contained in this report, a conclusion of no further action is provided or a recommendation to complete an FS to evaluate potential remedies and future actions will be made.

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1.2 SCOPE

The scope of this Phase II RI Report is to present: (1) the nature and extent of contamination, fate and transport of contaminants in the environment, and risk assessments for surface and subsurface soil at the AOC; (2) the results of the evaluation of remedial alternatives for meeting remedial action objectives for any CERCLA-related COCs identified in these media at the AOC; and (3) a conclusion of no further action or a preferred alternative to present to the public in a PP.

For the purposes of this report, the term "surface soil" includes dry sediment. Dry sediment refers to unconsolidated inorganic and organic material within conveyances, ditches, or low-lying areas that occasionally may be covered with water, usually following a precipitation event or due to snowmelt. Dry sediment is not covered with water for extended periods and typically is dry within seven days of a precipitation event. Dry sediment does not function as a permanent habitat for aquatic organisms, although it may serve as a natural medium for the growth of terrestrial organisms. Dry sediment is addressed the same as surface soil [0–1 ft below ground surface (bgs)] in terms of contaminant nature and extent, fate and transport, and risk exposure models. The term "sediment," as used in this report, refers to wet sediment within conveyances, ditches, wetlands, or water bodies that is inundated for extended periods. Sediment and associated surface water were not evaluated as part of this report as these media are not present at the AOC. These definitions and terminology usage are consistent with the *Facility-wide Human Health Cleanup Goals for the Ravenna Army Ammunition Plant, Ravenna, Ohio* (USACE 2010a) (herein referred to as the FWCUG Report).

Potential impacts to groundwater from soil (e.g., contaminant leaching) are evaluated in this report, as protectiveness to groundwater is included in the fate and transport analysis. However, groundwater will be evaluated as an individual AOC for the entire facility (designated as RVAAP-66) and addressed in a separate RI/FS Report. All buildings and structures at Load Line 11 have been demolished; therefore, they are not evaluated as continuing sources of contamination in this report.

As discussed in Section 2.4.1, only sanitary sewers are present at Load Line 11; no storm water sewer system exists within the AOC boundary. As part of the RVAAP-67 Facility-wide Sewers RI, sampling and evaluation of the sewer sediment, sewer water, outfall sediment, and pipe bedding material media was performed and identified inorganic chemicals and polycyclic aromatic hydrocarbons (PAHs) as the predominant sewer SRCs. Fate and transport modeling was performed, and an HHRA and ERA were conducted for sewers at Load Line 11. These evaluations recommend no further action with respect to the Facility-wide Sewers within Load Line 11. The full evaluation and conclusion are presented in the *Draft Remedial Investigation/Feasibility Study Report for RVAAP-67 Facility-wide Sewers* (USACE 2012a).

1.3 REPORT ORGANIZATION

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This report is organized in accordance with Ohio EPA and U.S. Environmental Protection Agency (USEPA) CERCLA RI/FS guidance and applicable USACE guidance. The following is a summary of the components of the report and a list of appendices:

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- Section 2.0 provides a description and history of the former RVAAP and Load Line 11, presents potential sources of contamination, presents potential receptors, and summarizes colocated or proximate sites.
- Section 3.0 describes the environmental setting at Camp Ravenna and Load Line 11, including the geology, hydrogeology, climate, and population.
- Section 4.0 summarizes previous assessments, investigations, and the IRA at Load Line 11, as well as the data used to support this Phase II RI.
- Section 5.0 discusses the occurrence and distribution of contamination at the AOC.
- Section 6.0 presents an evaluation of contaminant fate and transport.
 - Section 7.0 includes the methods and results of the HHRA and ERA.
 - Section 8.0 provides the conclusions and recommendations of the Phase II RI.
- Section 9.0 summarizes the framework for conducting the necessary agency and public involvement activities.
- Section 10.0 provides a list of references used to develop this report.
 - Appendices:
- 22 Appendix A: Field Sampling Logs,
- 23 Appendix B: Project Quality Assurance Summary,
- 24 Appendix C: Data Quality Control Summary Report,
- 25 Appendix D: Laboratory Analytical Results and Chains-of-Custody,
- 26 Appendix E: Fate and Transport Modeling Results,
- 27 Appendix F: Investigation-derived Waste Management Reports,
- Appendix G: Human Health Risk Assessment Tables,
- 29 Appendix H: Ecological Risk Assessment Information and Data, and
- 30 Appendix I: PBA08 RI Summary.

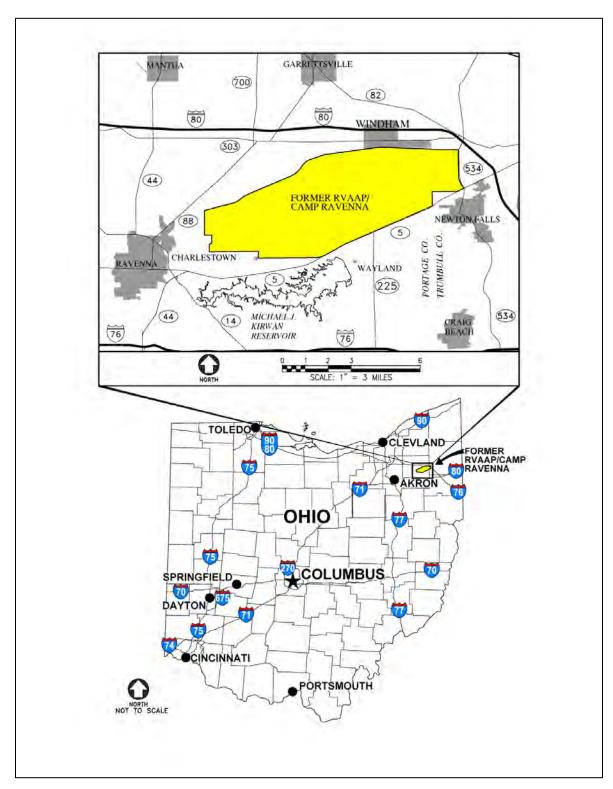


Figure 1–1. General Location and Orientation of Camp Ravenna

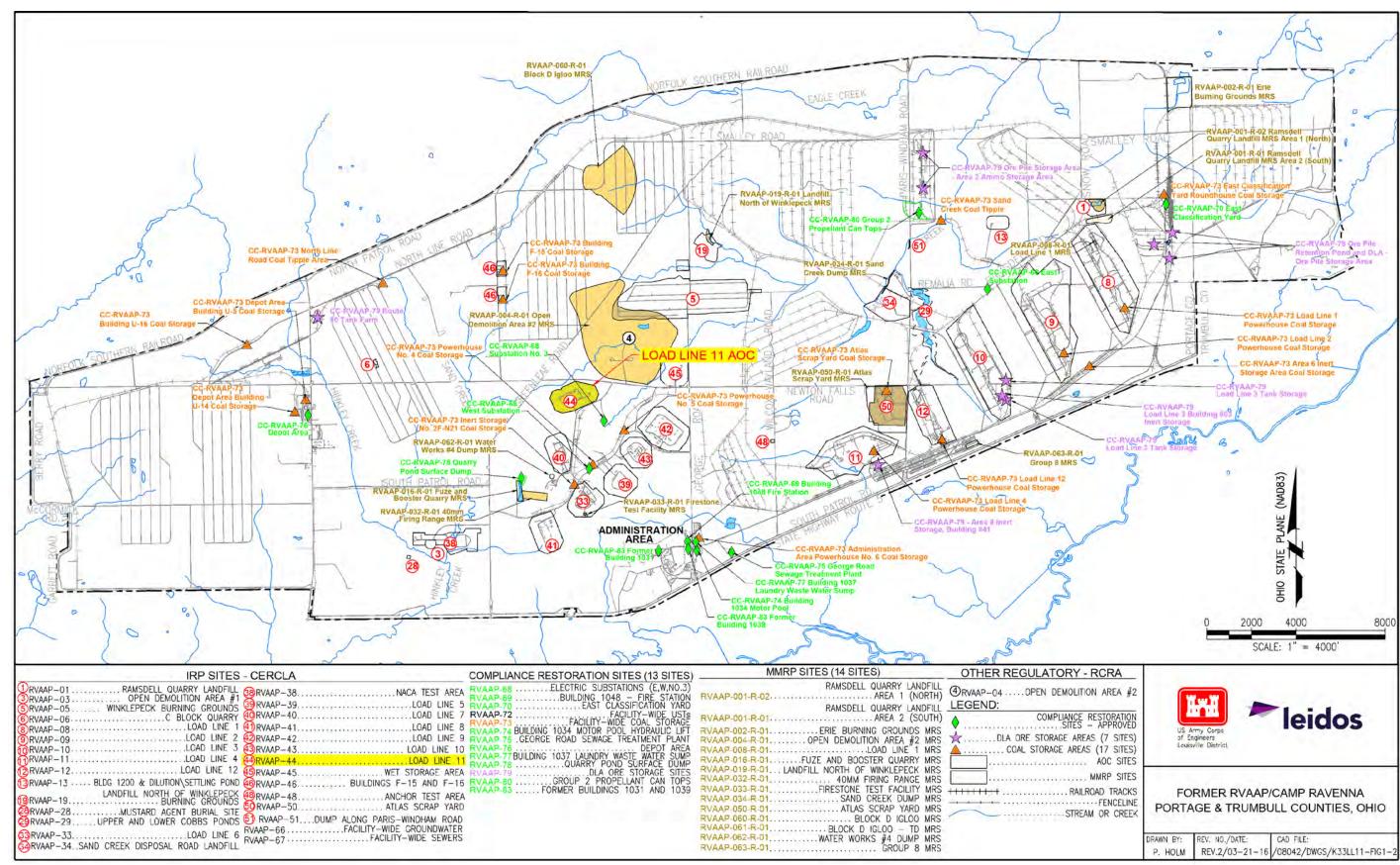


Figure 1–2. Location of AOCs and Munitions Response Sites at Camp Ravenna

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2.0 BACKGROUND

This section provides a description of the facility. In addition, it provides a summary of Load Line 11 operational history, potential sources, building demolition activities, potential human health and ecological receptors, co-located or proximate sites, and potential site-related releases.

2.1 FACILITY-WIDE BACKGROUND INFORMATION

2.1.1 General Facility Description

The facility, consisting of 21,683 acres, is located in northeastern Ohio within Portage and Trumbull counties, approximately 4.8 kilometers (3 miles) east/northeast of the city of Ravenna and approximately 1.6 kilometers (1 mile) northwest of the city of Newton Falls (Figure 1-1). The facility, previously known as RVAAP, was formerly used as a load, assemble, and pack facility for munitions production. As of September 2013, administrative accountability for the entire acreage of the facility has been transferred to the U.S. Property and Fiscal Officer (USP&FO) for Ohio and subsequently licensed to the Ohio Army National Guard (OHARNG) for use as a military training site (Camp Ravenna). References in this document to RVAAP relate to previous activities at the facility as related to former munitions production activities or to activities being conducted under the restoration/cleanup program.

In 1978, the U.S. Army Toxic and Hazardous Materials Agency conducted an Installation Assessment of RVAAP to review the potential for contaminant release at multiple former operations areas, as documented in Installation Assessment of Ravenna Army Ammunition Plant (USATHAMA 1978). The Installation Assessment indicated there is no evidence that bulk handling of the primary explosives lead azide or lead styphnate took place within the boundaries of Load Line 11. According to this assessment, from 1941–1945 Load Lines 5-11 produced 19,257,297 miscellaneous fuzes, 44,297,485 miscellaneous boosters, 79,580,576 miscellaneous detonators, 226,387,306 percussion elements, and 50,660,725 primers.

The former RVAAP received bulk 2,4,6-trinitrotoluene (TNT) product during operational activities and did not manufacture/produce dinitrotoluene (DNT) or TNT. A facility where DNT is manufactured will have the following isomers of DNT in the finished product: 2,4-DNT; 2,6-DNT; 2,5-DNT; 3,4-DNT; 2,3-DNT; and 3,5-DNT. This is not applicable to the former RVAAP. Degradation of TNT to 2,4-DNT occurs in soil; however, 2,4-DNT and 2,6-DNT do not degrade to the lesser isomers. It is the U.S. Department of the Army's (Army) position that testing DNT isomers other than 2,4- and 2,6-DNT is unnecessary and has no additional value of being protective to human health and the environment at the former RVAAP (RVAAP 2013).

2.1.2 Demography and Land Use

Camp Ravenna occupies east-central Portage County and southwestern Trumbull County. Census projections for 2010 indicated the populations of Portage and Trumbull counties are 161,419 and

210,312, respectively. Population centers closest to Camp Ravenna are Ravenna, with a population of 11,724, and Newton Falls, with a population of 4,795.

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- 4 The facility is located in a rural area and is not close to any major industrial or developed areas.
- 5 Approximately 55% of Portage County, in which the majority of Camp Ravenna is located, consists
- 6 of either woodland or farmland acreage. The closest major recreational area, the Michael J. Kirwan
- 7 Reservoir (also known as West Branch Reservoir), is located adjacent to the western half of Camp
- 8 Ravenna, south of State Route 5.

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- 10 Camp Ravenna is federally owned and is licensed to the OHARNG for use as a military training site.
- 11 Restoration activities at Camp Ravenna are managed by the Army National Guard and OHARNG.
- 12 Training and related activities at Camp Ravenna include field operations and bivouac training,
- 13 convoy training, maintaining equipment, C-130 aircraft drop zone operations, helicopter operations,
- and storing heavy equipment.

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2.2 LOAD LINE 11 BACKGROUND INFORMATION

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2.2.1 Operational History

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Load Line 11, formerly known as Booster Line #1, is an approximately 48-acre fenced AOC located near the southwest corner of the intersection of Fuze and Booster Spur Road and Newton Falls Road, north of Load Line 7, in the south-central portion of Camp Ravenna (Figure 1-2 and Figure 2-1). Load Line 11 was formerly used for producing artillery primers and fuzes.

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A description of the operational use at Load Line 11 is as follows (USATHAMA 1978):

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• 1941–1945 – The site operated at full capacity to produce artillery primers. Black powder was used to charge the primers. No bulk handling of other explosives occurred at Load Line 11. At the end of World War II, Load Line 11 was deactivated, and the process equipment remained on standby status.

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• 1951-1957 – The site was reactivated to produce primers. During reactivation, Load Line 11 produced 9,927,118 MK2A4 percussion primers, 24,482,465 MK2A4 primers, and 1,504,935 MK2A4 repack primers.

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- 1969–1971 The site was reactivated to produce approximately 7,000,000 MR ZA4 fuzes.
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- 1971 The site was deactivated, and process equipment was removed.

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No fuel storage tanks were present at Load Line 11 during operations. No historical information exists to indicate Load Line 11 was used for any other processes (including fuel storage and use, burning, etc.) other than what is presented above.

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As discussed in Section 4.2, an IRA was conducted in 2001 to remove building sumps, contaminated media within six ditch lines, and a hot spot with petroleum contamination. These activities were conducted by the Army in unison with the Phase I RI.

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As discussed in Section 2.2.3, all buildings (including slabs and foundations) were removed in 2004-2006. Remaining features at Load Line 11 include a one-lane asphalt perimeter road that enters the AOC from the south and encircles 75% of the former production area (FPA) and an asphalt parking area remains by former Building AP-11. The Load Line 11 perimeter fence is still in place, but it is not currently maintained. Small construction drainage ditches border the access road.

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2.2.2 Potential Sources

Historical facilities at Load Line 11 included 21 production and support buildings ranging in size from 34-8,862 ft². The locations of the former primary operational buildings are shown on Figure 2-1 and in the aerial photograph shown in Figure 2-2.

Former buildings and site features that are not expected to be potential sources of contamination are presented and described below.

- Buildings AP-15 and AP-16 Utilized for storage and shipping during the three production eras.
 - o Building AP-15 Inert Material Storage (located outside AOC boundary).
 - o Building AP-16 Shipping Building (located outside AOC boundary).
- Building AP-19 Employee Dining Hall.
- Building 11-51 Time Clock Utilized to clock employees in and out from shifts.
- Drainage ditches.

Table 2-1 presents a summary of potential sources of contamination at Load Line 11. This table identifies potential sources, previous uses, if there were documented releases to the environment, and potential contaminants associated with the previous use. As indicated in Section 2.4.1, Building SD-22, Ejector Station #4 will be evaluated as part of the Facility-wide Sewers RI.

Some of the potential contaminants are documented; however, some professional judgment was made to determine if additional contaminants should be considered as a product of historical use of the site.

The 1978 Installation Assessment identified the major contaminants of the former RVAAP to be TNT, composition B [a combination of TNT and hexahydro-1,3,5-trinitro-1,3,5-triazine (RDX)], sulfates, nitrates, lead styphnate, and lead azide. Additional site-specific contaminants include black powder; TNT; RDX; octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocane (HMX); nitroglycerine; nitrocellulose; nitroguanidine; and heavy metals (lead, chromium, mercury, and arsenic) from munitions assembly activities; volatile organic compounds (VOCs) were detected at former Building AP-17 that was utilized for storing solvents and at former Building AP-11 where lacquer sealing materials were used on finished primers; and polychlorinated biphenyl (PCB) detections were from on-site transformers. The evaluation of historical chemical contamination is not limited to these chemicals; rather, this evaluation is expanded to include all eligible chemical data that is available.

Load Line 11

In summary, the following chemicals were targeted to investigate these potential SRCs:

- Inorganic chemicals arsenic, chromium, lead, and mercury.
- Explosives TNT, RDX, HMX, nitroglycerine, nitrocellulose, nitroguanidine.
- Other PCBs and VOCs.

Many other chemicals were analyzed in the site investigations and are discussed in this report.

2.2.3 Building Decontamination and Demolition

Multiple activities were conducted at Load Line 11 to demolish buildings, remove footers, and remove remaining infrastructure. The following subsections summarize these activities.

2.2.3.1 2001 Sump Removal

To prevent contaminant migration outside the AOC boundary, building sumps adjacent to Buildings AP-3, AP-5, AP-6, and AP-8 were removed, and selected sanitary sewer manholes were grouted. These activities were conducted as part of the IRA conducted in 2001, as summarized in Section 4.2 and in the *Interim Removal Action at Load Line 11 (AOC-44)* (MKM 2004a).

Approximately 15,000 gal of water was pumped from the sumps through a filter directly into a holding tank and ground applied in an area just west of Buildings AP-3 and AP-4 as approved by Ohio EPA in January 2001. During the water removal at Building AP-3, it was determined that the infiltration rate of groundwater into the sewer-system was significant enough to impede the removal operations. Consequently, the sewer manholes downgradient of each sump were filled with bentonite cement to prevent water from infiltrating back into the sumps during the excavation and removal operations. Once the water was removed, effluent sewer lines were cut and plugged with mechanical packers and cement grout, which prepared the sumps for removal. Sewer lines were not excavated, as the lines exist below the groundwater table. Lead liners were removed from the sumps and recycled. Asbestos liners were removed, double bagged, and disposed as asbestos-containing material (ACM). The sump cavities were backfilled to grade using approximately 190 yd³ of approved, off-site soil.

2.2.3.2 2004-2005 Building Demolitions

- Demolition and removal of buildings at Load Line 11, including building slabs and foundations, were completed in 2004-2005 by the U.S. Army Tank-Automotive and Armaments Command. This was documented in the *Thermal Decomposition and Demolition of Load Line 11 and Buildings F15, 1200,* S. 4605 and T. 4602 (MVM 2005b).
- 38 S-4605 and T-4602 (MKM 2005b).

Prior to demolition and disposal, the following activities took place:

- The floors of each building were swept prior to removing transite roofs in order to remove loose paint chips and other potentially contaminated debris.
- All hazardous and non-hazardous material (e.g., fluorescent light fixtures and mercury switches) from the buildings were removed.
- ACM was removed from all buildings by a licensed asbestos removal contractor. ACM removal operations occurred in November 2004 to January 2005. A total of 63,343 ft² of transite roofing and 3,341 ft² of tar/shingle roofing material were removed from Load Line 11 and Buildings 1200, S4605, and T-4602 combined.
- Unexploded ordnance (UXO) personnel conducted building walkthroughs to confirm existing
 conditions, and inspected wall, floor slab, and structural steel surfaces for explosives
 contamination and other potential explosive hazards. Wall and floor penetrations,
 openings/cavities, and large cracks were inspected to determine if accumulated explosives
 were present. A hazard analysis building inspection form was used to document inspection at
 each building.
- Paint on the interior of each building was sampled to determine regulatory status of the applied dry paints. Sample results showed PCB concentrations in the applied paint at the buildings within Load Line 11 were in excess of 50 parts per million (ppm); therefore, building material was regulated as a PCB Bulk Product waste once demolition began.

During the demolition activities, sidewalks and footers were removed and the basements at Buildings AP-13 and AP-14 were demolished in place to 4 ft bgs. The footers at Buildings AP-3, AP-6, AP-7, AP-10, AP-11, AP-13, AP-14, AP-17, AP-19, and SD-22 were removed to 1 ft bgs. The footers were completely removed at Buildings AP-1, AP-4, AP-5, AP-9, AP-20, and 11-51. The footers at Building AP-18 were removed to 4 ft bgs. Building AP-8 was initially salvaged for use in future emissions testing relative to thermal decontamination operations but has since been demolished. Telephone poles and steam stanchions were also removed and staged north of former Building AP-18 during the demolition event.

Following demolition, all painted brick and structural steel were loaded for off-site disposal as PCB Bulk Product Waste at an approved facility. All unpainted concrete from floor slabs, footers, and blastwalls was deposited in the approved clean hard fill area at Load Line 1.

Soil in the vicinity of former production buildings was extensively disturbed during building demolition activities. The work areas were re-graded, cavities were filled with approved fill dirt as needed, and the area was vegetated in 2005 (MKM 2005b). The interior of the AOC is currently vegetated with grasses and shrub/scrub vegetation in unpaved areas and forested around the perimeter.

2.2.3.3 2006 Building Footer Removal

As part of demolition activities performed by Lakeshore Engineering Services, Inc., the footers of former Buildings AP-6, AP-7, AP-10, AP-13, AP-14, AP-17, and AP-19 were removed to a minimum

depth of 4 ft bgs, and stockpiled telephone poles and steam stanchions were removed from Load Line 11 for disposal in June-July 2006. These activities were documented in the *Munitions Response for the Demolition of Load Lines 5, 7, Building 1039, Transite Removal at Building T-1604, Removal of Remaining Concrete and Miscellaneous Debris at Load Lines 6, 9, and 11* (LES 2007). Soil near former production buildings was disturbed during building demolition activities, but the work areas were re-graded for drainage purposes and seeded following demolition activities.

1 2

2.2.4 AOC Boundary

Load Line 11 is an approximately 48-acre AOC bound by the existing perimeter fence, as presented in Figure 2-1. Load Line 11 is located near the southwest corner of the intersection of Fuze and Booster Spur Road and Newton Falls Road, north of Load Line 7.

As shown in Figure 2-3, the AOC boundary encompasses the FPA and non-production area (NPA). The FPA consists of approximately 11.4 acres and is encompassed within the gravel perimeter road surrounding 75% of the main production area. The buildings within the FPA were historically used to produce primers and fuzes for artillery projectiles, and contained the former production and storage buildings. The NPA is approximately 36.8 acres and includes the areas between the production area perimeter road and perimeter fence.

The surface and subsurface soil investigation at Load Line 11 generally focused on soil within the AOC boundary or fence line and includes samples collected from the FPA and NPA. Potential off-AOC contaminant migration has been evaluated by the samples collected in ditch lines leading to the northern, western, and eastern fence lines. In addition to the samples collected within the AOC boundaries, this Phase II RI Report discusses sample LL11sd/sw-082 that was collected north of Load Line 11 and north of Newton Falls Road to assess the potential transport beyond the AOC boundary.

2.2.5 Spatial Aggregates

Load Line 11 was divided into two soil spatial aggregates, three sediment spatial aggregates, and two surface water spatial aggregates for evaluation in this Phase II RI Report. The two soil spatial aggregates are the FPA and NPA. The FPA encompasses all production and non-production buildings. The NPA surrounds the FPA to the fence line, where no known activities occurred. The three sediment aggregates are the East Ditch, West Ditch, and Sewer Outfall. The East Ditch is represented by a discrete sediment sample collected at a ditch line which drained portions of the eastern part of the FPA. The West Ditch is represented by a discrete sediment sample collected at a ditch line west of former Building AP-4 which drains the western half of Load Line 11. The Sewer Outfall is represented by a sediment sample collected at the sanitary system overflow outfall north of the FPA. The two surface water aggregates are the East Ditch and West Ditch.

In addition to these aggregates, sediment and surface water samples were collected at off-AOC location LL11sd/sw-082. This location was included as part of the nature and extent evaluation for this Phase II RI.

2.3 POTENTIAL RECEPTORS AT LOAD LINE 11

1 2

3 The following sections discuss potential human and ecological receptors at Load Line 11.

2.3.1 Human Receptors

Camp Ravenna is a controlled-access facility. Load Line 11 is located in the south-central portion of the facility and is not currently used for training (Figure 1-2).

In February 2014, the Army and Ohio EPA amended the risk assessment process to address changes in the RVAAP restoration program. The *Final Technical Memorandum: Land Uses and Revised Risk Assessment Process for the RVAAP Installation Restoration Program* (ARNG 2014) (herein referred to as the Technical Memorandum) identified three Categorical Land Uses and Representative Receptors to be considered during the RI phase of the CERCLA process. These three Land Uses and Representative Receptors are presented below.

- 1. Unrestricted (Residential) Land Use Resident Receptor (Adult and Child) (formerly called Resident Farmer).
- 2. Military Training Land Use National Guard Trainee.
- 3. Commercial/Industrial Land Use Industrial Receptor (USEPA Composite Worker).

Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then the AOC is also considered to have met the requirements of the other Land Uses (i.e., Commercial/Industrial and Military Training), and the other Land Uses do not require evaluation.

2.3.2 Ecological Receptors

Camp Ravenna has a diverse range of vegetation and habitat resources. Habitats present within the facility include large tracts of closed-canopy hardwood forest, scrub/shrub open areas, grasslands, wetlands, open-water ponds and lakes, and semi-improved administration areas.

An abundance of wildlife is present on the facility: 35 species of land mammals, 214 species of birds, 41 species of fish, and 34 species of amphibians and reptiles have been identified. The northern long-eared bat (*Myotis septentrionalis*; federally threatened) exists at Camp Ravenna. There are no other federally listed species and no critical habitat occurs (OHARNG 2014). Ohio state-listed plant and animal species have been identified through confirmed sightings and/or biological inventories at the facility and are presented in Table 2-2. With exception of the access roads at Load Line 11, the AOC is vegetated with grasses and shrub vegetation in the central portion and has mature forest on the perimeter. Sand Creek runs along the northern perimeter, adjacent to the AOC. Additional information specific to ecological resources at Load Line 11 is included in Section 7.3.

Load Line 11

2.4 CO-LOCATED OR PROXIMATE SITES

The following subsections summarize sites that are co-located or proximate to Load Line 11 but are addressed separately.

2.4.1 Facility-wide Sewers

The defunct sanitary sewers within the perimeter of Load Line 11 are being investigated and assessed as part of the Facility-wide Sewers AOC (RVAAP-67). This includes an evaluation of Building SD-22, Sewer Ejector Station #4. Storm sewers are not present at Load Line 11. Sanitary sewer sediment, outfall sediment, pipe bedding material, and sewer water were evaluated as currently summarized in the *Draft Remedial Investigation/Feasibility Study Report for RVAAP-67 Facility-wide Sewers* (USACE 2012a). The sanitary sewers in the Load Line 11 functional area (FA) are part of the network that flowed to the George Road Sewage Treatment Plant network. When Ejector Station #4 was not functioning, drainage was discharged to the sanitary sewer overflow outfall at the northern portion of the FA (Figure 2-1). While cracks and joint separations within the sanitary sewer system have been observed due to the advanced age of the system, the pipes are generally intact, and some sewers associated with removed sumps were grouted associated with removed sumps, which limits vertical migration from sewer sediment to the exterior of the piping system.

The Load Line 11 IRA (MKM 2004a), Phase I RI (MKM 2005a), and Explosive Evaluation of Sewers (LES 2007, USACE-CERL 2007) are the historical investigations that specifically addressed the sewer system. Sewer water and sediment samples and one outfall sediment sample were collected from the sanitary sewer manholes at Load Line 11 in 2001 as part of the Phase I RI. The 2001 IRA consisted of removing lead/asbestos-lined sumps and lead-contaminated sediment. The IRA was initiated in unison with the Phase I RI activities as an early response action to remove the primary pathways for off-FA contaminant migration. This IRA included removing sump water from production buildings and grouting selected sanitary sewer manholes. Sewer samples were not collected from the sanitary sewer during this effort.

Samples of sewer and outfall sediment, sewer water, and pipe bedding materials (e.g., soil or backfill beneath the pipe) were also collected in 2009 and 2010 as part of the Facility-wide Sewers RI (USACE 2012a). The compiled data effectively characterized the nature and extent of the contamination at the Load Line 11 FA, and no further sampling was recommended. All SRCs found in the subsurface sewer media samples within the Load Line 11 FA and evaluated through the stepwise fate and transport screening evaluation were eliminated as posing future impacts to groundwater.

The HHRA did not identify COCs at the Load Line 11 FA for the National Guard Trainee or Resident Receptor. The point evaluation identified only manganese as a point evaluation COPC in outfall sediment. Manganese is not impacting downstream receptors; therefore, no further action was recommended from an ecological perspective. In summary, the Facility-wide Sewers RI recommended no further action for the Load Line 11 sanitary sewers. The Facility-wide Sewers RI predated the Technical Memorandum that specified evaluating Commercial/Industrial Land Use;

however, the Load Line 11 FA was determined to be protective for Unrestricted (Residential) Land Use. Therefore, an evaluation of Commercial/Industrial Land Use would not be required.

2.4.2 Facility-wide Groundwater

As part of the Installation Restoration Program, the Army implements the Facility-wide Groundwater Monitoring Program (FWGWMP) in accordance with previous agreements made with Ohio EPA. The FWGWMP was initiated in 2005 and involves quarterly sampling of selected wells within the former RVAAP.

In October 2006–2007, the FWGWMP collected groundwater samples from two wells at Load Line 11 (LL11mw-002 and LL11mw-007). In 2009, the FWGWMP collected groundwater data from 10 wells at Load Line 11 (wells LL11mw-001 to LL11mw-010). In October 2010, the FWGWMP collected additional groundwater data from wells LL11mw-001, LL11mw-002, LL11mw-007, and LL11mw-009. Most chemical concentrations in groundwater were below the maximum contaminant level (MCL) or regional screening level (RSL) [target risk (TR) of 1E-05, hazard quotient (HQ) of 1]. The exceptions are discussed below (EQM 2015).

- Arsenic in LL11mw-007 exceeded the MCL (10 $\mu g/L$) and RSL of 10⁻⁵ (0.052 $\mu g/L$) in all samples collected.
- Cadmium in LL11mw-004 exceeded the RSL of 10^{-5} (9.2 µg/L) in one sample in the first quarterly sample collected in 2009. The subsequent three samples collected in 2009 all had concentrations below the RSL of 10^{-5} , with a maximum concentration of 5.1 µg/L.
- Manganese in LL11mw-001 to LL11mw-004, LL11mw-009, and LL11mw-010 exceeded the RSL of 10^{-5} (430 $\mu g/L$).
- Bis(2-ethylhexyl)phthalate in LL11mw-001, LL11mw-003, and LL11mw-008 exceeded the MCL. All results in LL11mw-003 and LL11mw-008 had detected concentrations below the reporting limit (J-flagged).
- beta-Hexachlorocyclohexane (BHC) exceeded the RSL (TR of 1E-05, HQ of 1) in LL11mw-002 and LL11mw-009. However, all detected concentrations either had estimated concentrations below the reporting limit (J-flagged) or were only present in the laboratory blank (B-flagged).
- 2,6-DNT exceeded the RSL (TR of 1E-05, HQ of 1) in LL11mw-006 and LL11mw-010. However, all detected concentrations were below the reporting limit (J-flagged).

Wells LL11mw-007, LL11mw-011, and LL11mw-012 were added to the FWGWMP in 2012-2013 to evaluate the current levels of contaminants associated with former RVAAP operations, specifically arsenic, iron, and manganese (EQM 2015). A summary of data from these sampling events is provided below.

- LL11mw-007 All concentrations were below the MCL or RSL (TR of 1E-05, HQ of 1) except arsenic. Arsenic exceeded the MCL of 10 μ g/L in all three samples analyzed in 2012-2013.
- LL11mw-011 All concentrations were below the MCL or RSL (TR of 1E-05, HQ of 1) in the October 2012 sample except manganese at concentration of 570 μg/L, which is above the RSL (TR of 1E-05, HQ of 1) of 430 μg/L. All other samples analyzed in 2012-2013 for manganese had concentrations lower than the RSL (TR of 1E-05, HQ of 1).
- LL11mw-012 All concentrations were below the MCL or RSL (TR of 1E-05, HO of 1).

Additional groundwater level monitoring was performed in May 2014 at all 12 monitoring wells at Load Line 11; however, no samples were collected (EQM 2015). Facility-wide groundwater AOC is currently at the RI phase of the CERCLA process. Any future decisions or actions respective to groundwater at Load Line 11 will be addressed as part of that AOC.

2.4.3 Fuze and Booster Area Settling Tanks

RVAAP-26, the Fuze and Booster Area Settling Tanks AOC, is comprised of 15 tanks at various load lines located within the former RVAAP. These 15 tanks have also been referred to as "settling basins" or simply "basins" and were used for settling out explosive contamination from wastewater during production activities from 1941–1945 and 1969–1970. The settled sludge was periodically collected from the tanks (every one to three months) and transferred to one of the burning grounds for thermal destruction. The final method of disposition of the water from the tanks is not definitely known. It may have been pumped onto the ground surface or into the sewer systems (USAEHA 1994). The soil surrounding the process buildings sourcing the effluent to the settling tanks may also be of concern because building washout operations historically resulted in the release of wastewater on the ground adjacent to the building exits.

In 1971, all the tanks were emptied, cleaned, and covered. Three of the tanks are former concrete settling tanks located at Load Line 11 that were removed in November 1988. The locations of the concrete settling tanks included one at AP-3 (Black Powder Screening Facility) and two at Building AP-5 (Black Powder Dry House) at Load Line 11. These underground settling tanks (8 by 8 by 8 ft) had a 3,829-gal capacity (Jacobs Engineering 1989).

Regarding all settling tanks in the Fuze and Booster Area, the 1989 Resource Conservation and Recovery Act (RCRA) Facility Assessment states, "there is a high potential for releases to soil and groundwater from this unit. The integrity of the tanks and associated plumbing is not known. Extensive contamination may have occurred (Jacobs Engineering 1989)." Releases from these tanks were suspected, but not known.

As stated in the 2004 Installation Action Plan (RVAAP 2004), the RVAAP-26 Fuze and Booster Area Settling Tanks AOC was considered Response Complete under the Installation Restoration Program

| 1 | since each load line became its own AOC. Explosive contamination in surface and subsurface soil | | |
|----|---|--|--|
| 2 | was evaluated during the Phase I RI, IRA, and PBA08 RI, as discussed in Section 4.0. | | |
| 3 | | | |
| 4 | 2.4.4 | Munitions Response Sites | |
| 5 | | | |
| 6 | No munitions response sites have been documented within Load Line 11. | | |
| 7 | | | |
| 8 | 2.4.5 | Compliance Restoration Sites | |
| 9 | | | |
| 10 | There | are no compliance restoration sites, such as former or existing underground storage tanks, | |
| 11 | within or adjacent to the AOC boundary. | | |
| 12 | | | |
| 13 | 2.5 | POTENTIAL SITE-RELATED RELEASES | |
| 14 | | | |
| 15 | As pre | sented in Table 2-1, there have been no documented site-related releases at Load Line 11. | |
| 16 | Table 2-1 presents potential contaminants that may be present in Load Line 11 media from previous | | |
| 17 | use of the site. | | |

Table 2–1. Potential Source Area Description and Potential Impacts

| Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|---|---|-----------------------|------------------------------------|
| Building AP-1 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder service magazine for primer charging. | None | Metals, black powder |
| Building AP-2 | screening. | None | Metals, black powder |
| Building AP-3 | 1941-1945, 1951-1957, and 1969-1971: Utilized for black powder screening. Sump located on the southern side of AP-3. | None | Metals, black powder |
| Building AP-4 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. | None | Metals, black powder |
| Buildings AP-5 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder dry house. One sump was located between AP-5 and AP-6. | None | Metals, black powder |
| Buildings AP-6 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder service magazine for primer charging. 1941-1945, 1951-1957, and 1969-1971: Utilized as a motor house to support black powder screening. 1941-1945, 1951-1957, and 1969-1971: Utilized for black powder screening. Sump located on the southern side of AP-3. 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder dry house. One sump was located between AP-5 and AP-6. 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder dry house. Two sets of sumps were associated with Building AP-5 and AP-6: one set was located outside the southwest wall of AP-6 and one set was located between AP-5 and AP-6. 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. 1941-1945, 1951-1957, and 1969-1971: Utilized for artillery primer loading and served as an administrative building. Black powder charging operations were shifted from Building AP-8 to AP-11 in later years. Two sets of sumps, one on the east side and one on the west side, were associated with this building and were connected to the sewer mains of the facility. 1941-1945: Utilized as a final product shipping building during WWI. 1951-1957: Utilized as primer rest house. 1969-1971: Served as the percussion element storage and operations service magazine prior to primer charging. 1941-1945, 1951-1957, and 1969-1971: Utilized as percussion element service magazine prior to primer charging. 1941-1945, 1951-1957. Utilized for black powder charging operations. 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. 1941-1945, 1951-1957, and 1969-1971: Utilized as change houses where workers changed out of work clothing at end of shift. 1941-1945, 1951-1957, and 1969-1971: Utilized as | | |
| Buildings AP-7 | | None | Metals, black powder |
| charging. 1941-1945, 1951-1957, and 1969-1971: Utilized for artillery primer loading and served as an administrative building. Black powder charging operations were shifted from Building AP-8 to AP-11 in later years. Two sets of sumps, one on the east side and one on the west side, were associated with this building and were connected to the sewer mains of the facility. | | | Metals, black powder |
| Building AP-9 | 1941-1945: Utilized as a final product shipping building during WWII. 1951-1957: Utilized as primer rest house. | None | Metals, black powder |
| Building AP-10 | 1941-1945, 1951-1957, and 1969-1971: Utilized as percussion element service magazine | None | Metals, black powder |
| Building AP-11 | 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and | None | Metals, black powder, lacquers |
| Buildings AP-13 and AP-14 | out of work clothing at end of shift. | None | Metals, explosives |
| Building AP-17 | 1941-1945, 1951-1957, and 1969-1971: Utilized as a solvent storage facility. | None | Solvents |
| Building AP-18 | 1941-1945, 1951-1957, and 1969-1971: Utilized for storage of percussion elements. There was an earthen barricade around Building AP-18. Historical map indicated AP-18 used for black powder storage in 1951-1957. | None | Metals, black powder |

Table 2–1. Potential Source Area Description and Potential Impacts (continued)

| Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|---|---|-----------------------|------------------------------------|
| Building AP-20 | 1941-1945, 1951-1957, and 1969-1971: Served as a quality assurance (QA) primer sensitivity testing facility. | None | Metals, black powder |
| Transformers | Historical PCB Disposition Records indicated 3 transformers serviced all Load Line 11 buildings. All transformers were tested for PCBs and all contained PCBs (60, 165, and 663 ppm). All transformers was removed on November 3, 1992 and stored at Building 854 (RVAAP-27, PCB Storage) awaiting final disposition by the Defense Reutilization Material Office (DRMO). | None | PCBs |

Target metals = Lead, chromium, mercury, and arsenic.

Target explosives = Black powder; 2,4,6-trinitrotoluene; octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocane; and hexahydro-1,3,5-trinitro-1,3,5-triazine.

PCB = Polychlorinated biphenyl. WWII = World War II.

Table 2–2. Federal- and State-listed Species List

| | CAMP RAVENNA JOINT MILITARY TRAINING CENTER RARE SPECIES LIST | | | |
|----------|---|--|--|--|
| | December 2014 | | | |
| | | | | |
| I. S | pecies confirmed to be on Camp Ravenna property b | y biological inventories and confirmed sightings. | | |
| | A. Federal Threatened | | | |
| 1. N | Northern long-eared bat, Myotis septentrionalis | | | |
| | B. State Endangered | | | |
| 1. | American bittern, Botaurus lentiginosus (migrant) | 8. Tufted Moisture-loving Moss, <i>Philonotis Fontana</i> | | |
| 2. | Northern harrier, Circus cyaneus | var. caespitosa | | |
| 3. | Sandhill Crane, Grus Canadensis (probable | 9. Appalachian quillwort, Isoetes engelmannii | | |
| | nester) | 10. Handsome sedge, <i>Carex formosa</i> | | |
| 4. | Black bear, Ursus americanus | 11. Narrow-necked Pohl's Moss, <i>Pohlia elongata var</i> . | | |
| 5. | Mountain Brook Lamprey, Ichthyomyzon greeleyi | elongate | | |
| 6. | Brush-tipped emerald, Somatochlora walshii | 12. Philadelphia panic-grass, <i>Panicum</i> | | |
| 7. | Graceful Underwing, Catocala gracilis | philadelphicum | | |
| | | 13. Variegated scouring-rush, Equisetum variegatum | | |
| | C. State Threatened | | | |
| 1. | Barn owl, Tyto alba | 6. Northern long-eared bat, <i>Myotis septentrionalis</i> | | |
| 2. | Least bittern, Ixobrychus exilis | 7. Hobblebush, Viburnum alnifolium | | |
| 3. | Trumpeter swan, Cygnus buccinators (migrant) | 8. Simple willow-herb, <i>Epilobium strictum</i> | | |
| 4. | Bobcat, Felis rufus | 9. Lurking leskea, <i>Plagiothecium latebricola</i> | | |
| 5. | Caddis fly, Psilotreta indecisa | 10. Strict blue-eyed grass, Sisyrinchium montanum | | |
| 1 | D. State Potentially Threatened Plants | | | |
| 1. | Arborvitae, Thuja occidentalis | 6. Sharp-glumed manna-grass, Glyceria acutifolia | | |
| 2. | False hop sedge, Carex lupiliformis | 7. Straw sedge, Carex straminea | | |
| 3. | Greenwhite sedge, Carex albolutescens | 8. Water avens, Geum rivale | | |
| 4. | Long Beech Fern, Phegopteris connectilis | 9. Woodland Horsetail, Equisetum sylvaticum | | |
| 5 | (Thelypteris phegopteris) | 10. Shining ladies'-tresses, Spiranthes lucida | | |
| 5. | Pale sedge, Carex pallescens | | | |
| 1 | E. State Species of Concern | 17 Novel 1-11:4- C-1: | | |
| 1. | Big brown bat, Eptesicus fuscus | 17. Northern bobwhite, <i>Colinus virginianus</i> | | |
| 2. | Deer mouse, <i>Peromyscus maniculatus</i> Eastern red bat, <i>Lasiurus borealis</i> | 18. Common moorhen, Gallinula chloropus | | |
| 3. | | 19. Great egret, <i>Ardea alba</i> (migrant) | | |
| 4. 5. | Hoary bat, Lasiurus cinereus Little brown bat, Myotis lucifugus | 20. Sora, <i>Porzana carolina</i> 21. Virginia Rail, <i>Rallus limicola</i> | | |
| | Little brown bat, <i>Myotis lucifugus</i> Pygmy shrew, <i>Sorex hovi</i> | 22. Yellow-bellied Sapsucker, <i>Sphyrapicus varius</i> | | |
| 6. 7. | Southern bog lemming, Svnaptomys cooperi | 23. Creek heelsplitter, <i>Lasmigona compressa</i> | | |
| 8. | Star-nosed mole, <i>Condylura cristata</i> | 24. Eastern box turtle, <i>Terrapene carolina</i> | | |
| 9. | Tri-colored bat, <i>Perimyotis subflavus</i> | 25. Four-toed Salamander, <i>Hemidacrylium scutatum</i> | | |
| | Woodland jumping mouse, <i>Napaeozapus insignis</i> | 26. Eastern garter snake, <i>Thamnophis sirtalis</i> | | |
| | Sharp-shinned hawk, Accipiter striatus | 27. Smooth green snake, <i>Opheodrys vernalis</i> | | |
| | Marsh wren, Cistothorus palustris | 28. Eastern sand darter, <i>Ammocrypta pellucida</i> | | |
| | Henslow's sparrow, Ammodramus henslowii | 29. Mayfly, Stenonema ithica | | |
| | Cerulean warbler, <i>Dendroica cerulean</i> | 30. Moth, <i>Apamea mixta</i> | | |
| | Prothonotary warbler, <i>Protonotaria citrea</i> | 31. Moth, Brachylomia algens | | |
| | Bobolink, Dolichonyx oryzivorus | 32. Scurfy quaker, <i>Homorthodes furfurata</i> | | |
| 10. | , 2 000000000000000000000 | 33. Sedge wren, <i>Cistothorus platensis</i> | | |
| | | 55. Seage wien, Casomorus punchas | | |

Table 2–2. Federal- and State-listed Species List (continued)

| CAMP RAVENNA JOINT MILITARY TRAINING CENTER RARE SPECIES LIST December 2014 | | | |
|---|--|---|--|
| | F. State Special Interest | | |
| 1. | American black duck, Anas rubripes | 12. Pine siskit, Carduelis pinus | |
| 2. | Canada warbler, Wilsonia Canadensis | 13. Purple finch, Carpodacus purpureus | |
| 3. | Dark-eyed junco, Junco hyemalis (migrant) | 14. Red-breasted nuthatch, Sitta Canadensis | |
| 4. | Hermit thrush, Catharus guttatus (migrant) | 15. Golden-crowned kinglet, Regulus satrapa | |
| 5. | Least flycatcher, Empidonax minimus | 16. Blackburnian warbler, Dendroica fusca | |
| 6. | Magnolia warbler, Dendroica magnolia | 17. Gadwall, Anas strepera | |
| 7. | Northern waterthrush, Seiurus noveboracensis | 18. Green-winged teal, Anas crecca | |
| 8. | Winter wren, Troglodytes troglodytes | 19. Northern shoveler, Anas clypeata | |
| 9. | Back-throated blue warbler, Dendroica | 20. Redhead duck, Aytya Americana | |
| | caerulescens | 21. Ruddy duck, Oxyura jamaicensis | |
| 10. | Brown creeper, Certhia Americana | 22. Wilson's snipe, Gallinago delicata | |
| 11. | Mourning warbler, Oporornis Philadelphia | 23. Subflava sedge borer, Capsula subflava | |

Note: The Integrated Natural Resource Management Plan (OHARNG 2014) indicated that no federally listed species are known to reside at Camp Ravenna, and no critical habitat occurs. However, the northern long-eared bat exists at Camp Ravenna and was expected to be listed as an endangered species in mid-2015. Accordingly, this table indicates the northern long-eared bat is federally threatened (USFWS 2016) and state threatened (ODNR 2016).

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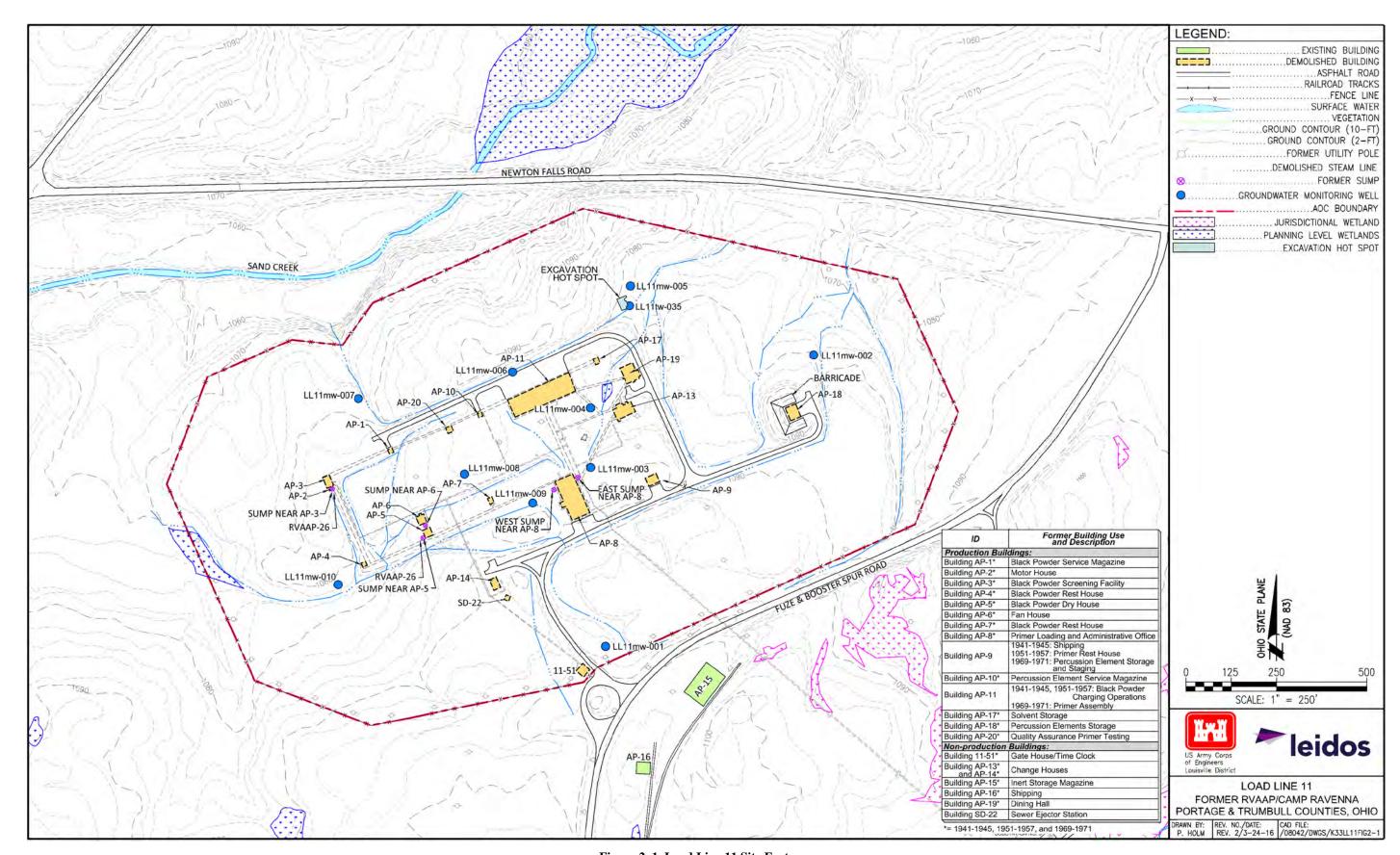


Figure 2–1. Load Line 11 Site Features

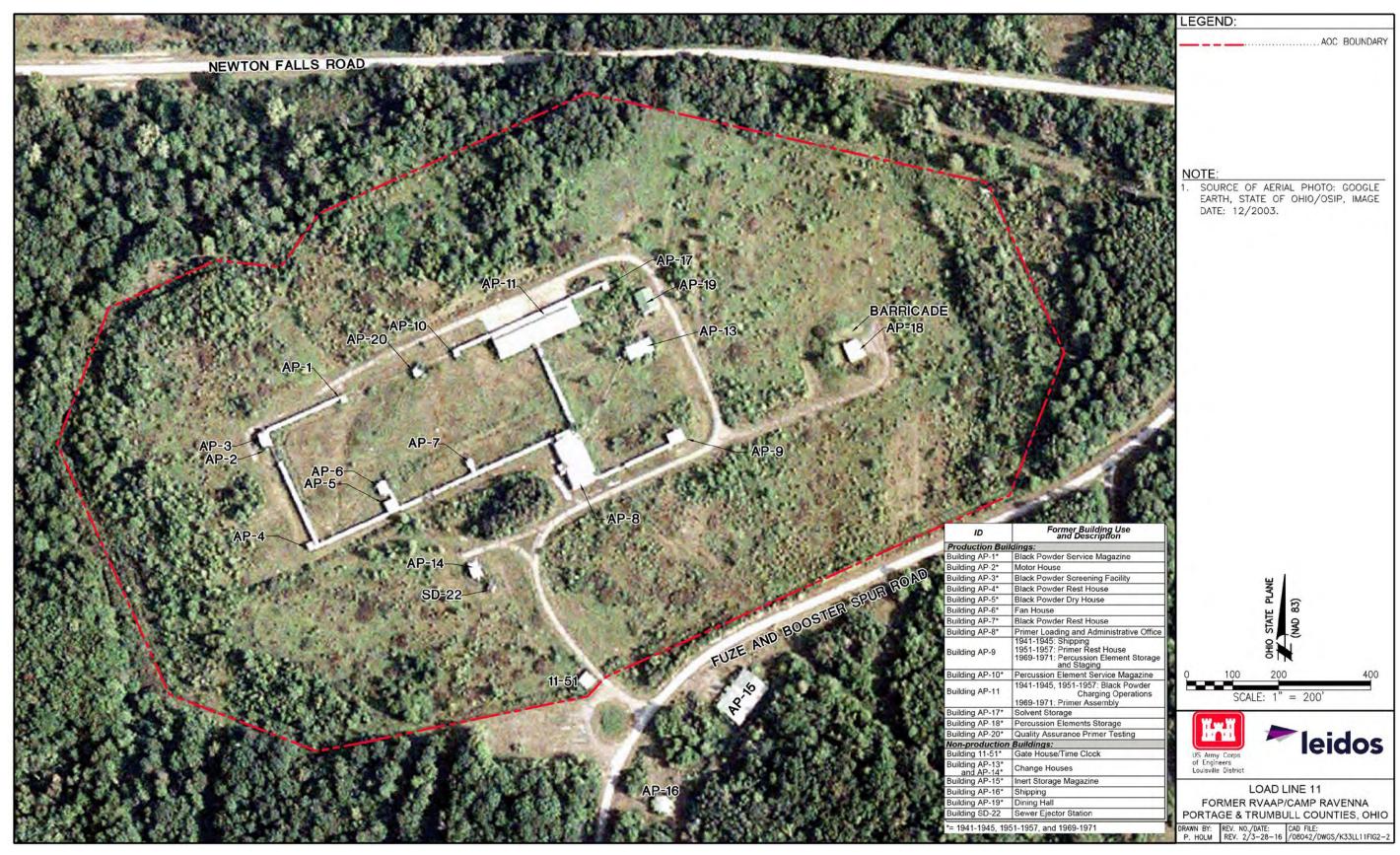


Figure 2–2. Load Line 11 Site Features Prior to Building Demolition (Aerial Photo dated 12/2003)

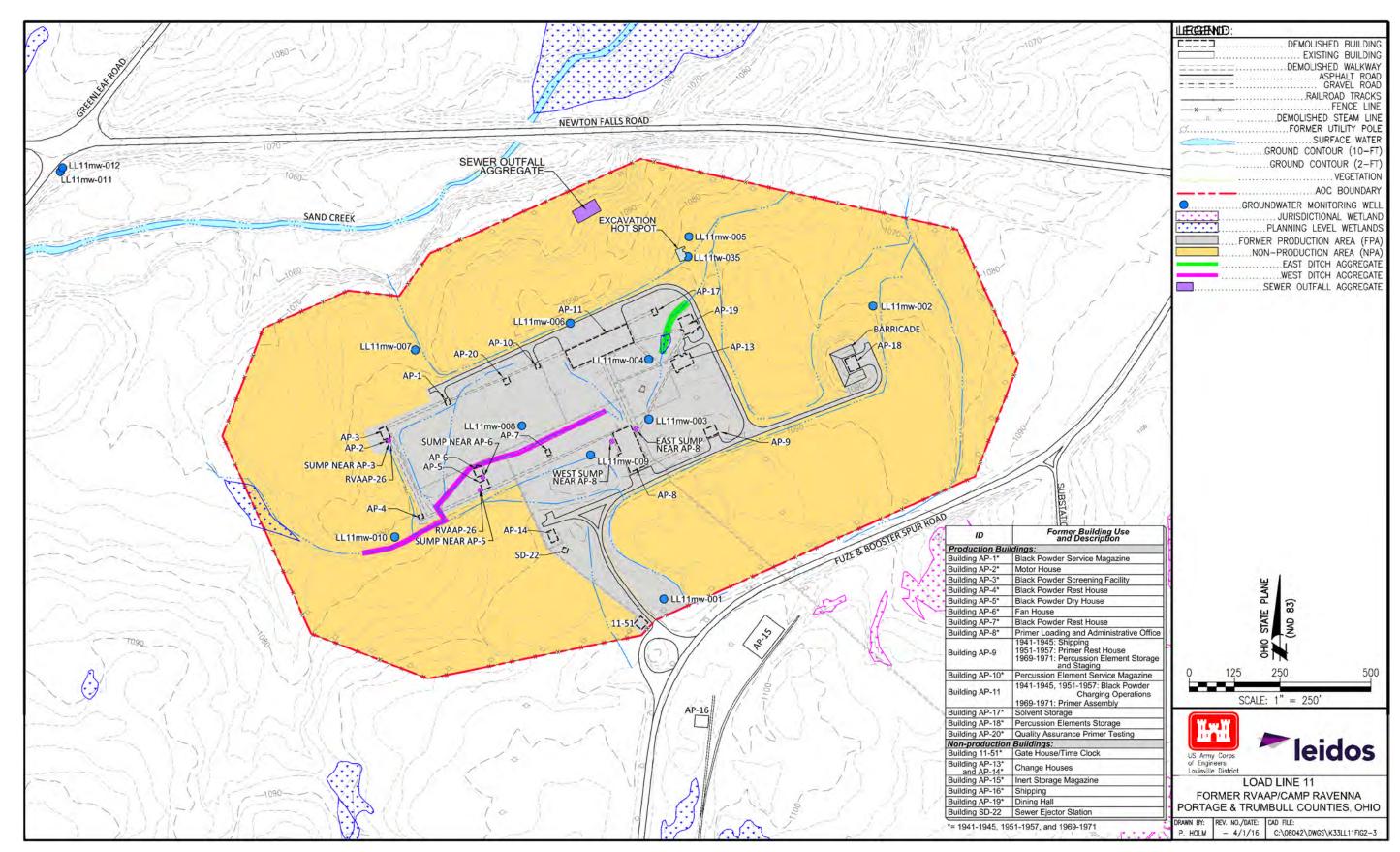


Figure 2–3. Load Line 11 Aggregates

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3.0 ENVIRONMENTAL SETTING

This section describes the physical features, topography, geology, hydrogeology, and environmental characteristics of Camp Ravenna and Load Line 11 that are factors in identifying the potential contaminant transport pathways, receptor populations, and exposure scenarios to evaluate human health and ecological risks.

3.1 CAMP RAVENNA PHYSIOGRAPHIC SETTING

Camp Ravenna is located within the southern New York section of the Appalachian Plateaus physiographic province (USGS 1968). This province is characterized by elevated uplands underlain primarily by Mississippian-age and Pennsylvanian-age bedrock units that are horizontal or gently dipping. The province is characterized by its rolling topography with incised streams having dendritic drainage patterns. The southern New York section has been modified by glaciation, which rounded ridges, filled major valleys, and blanketed many areas with glacially-derived unconsolidated deposits (e.g., sand, gravel, and finer-grained outwash deposits). As a result of glacial activity in this section, old stream drainage patterns were disrupted in many locales, and extensive wetland areas developed.

3.2 SURFACE FEATURES AND AOC TOPOGRAPHY

The topography of Camp Ravenna is gently undulating with an overall decrease in ground elevation from a topographic high of approximately 1,220 ft above mean sea level (amsl) in the far western portion of the facility to low areas at approximately 930 ft amsl in the far eastern portion of the facility. USACE mapped the facility topography in February 1998 using a 2-ft (60.1-cm) contour interval with an accuracy of 0.02 ft (0.61 cm). USACE based the topographic information on aerial photographs taken during the spring of 1997. The USACE survey is the basis for the topographical information illustrated in figures included in this report.

Load Line 11 is an approximately 48-acre AOC located near the southwest corner of the intersection of Fuze and Booster Spur Road and Newton Falls Road, north of Load Line 7, in the south-central portion of Camp Ravenna (Figure 1-2). A fence exists as the perimeter boundary of the AOC, although it is not currently maintained.

All buildings and structures within the Load Line 11 fence line have been demolished, and building slabs and footers have been removed. Soil in the vicinity of former production buildings was extensively disturbed during building demolition activities. The work areas were re-graded, cavities were filled with approved fill dirt as needed, and the area was vegetated following the building decontamination and demolition activities discussed in Section 2.2.3. Remaining features at Load Line 11 include a one-lane asphalt access road that enters the AOC from the south and encircles 75% the FPA (Figures 2-1 and 3-1). An asphalt parking area remains by former Building AP-11. Small drainage ditches border some portions of the access road, and drainage conveyances are located throughout the AOC boundary. Vacant, wooded land is located directly to the south and southwest.

 Topographic relief at the AOC is moderate, with a topographic high in the central portion of the AOC and gentle slopes to the north and northeast outside of the FPA. Ground elevations within Load Line 11 range from approximately 1,070–1,100 ft amsl (Figure 3-1). Surface water follows topographic relief and drains into ditches that exit the AOC. Sand Creek is immediately to the north of the AOC boundary.

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3.3 SOIL AND GEOLOGY

3.3.1 Regional Geology

The regional geology at Camp Ravenna consists of horizontal to gently dipping bedrock strata of Mississippian and Pennsylvanian age overlain by varying thicknesses of unconsolidated glacial deposits. The bedrock and unconsolidated geology at Camp Ravenna and the geology specific to Load Line 11 are presented in the following subsections.

3.3.2 Soil and Glacial Deposits

Bedrock at Camp Ravenna is overlain by deposits of the Wisconsin-aged Lavery Till in the western portion of the facility and the younger Hiram Till and associated outwash deposits in the eastern two-thirds of the facility (Figure 3-2). Unconsolidated glacial deposits vary considerably in their character and thickness across Camp Ravenna, from zero in some of the eastern portions of the facility to an estimated 150 ft in the south-central portion.

Thin coverings of glacial material have been completely removed as a consequence of human activities at locations such as Ramsdell Quarry. Bedrock is present at or near the ground surface in locations such as at Load Line 1 and the Erie Burning Grounds (USACE 2001a). Where this glacial material is still present, the distribution and character indicate their origin as ground moraine. These tills consist of laterally-discontinuous assemblages of yellow-brown, brown, and gray silty clays to clayey silts, with sand and rock fragments. Lacustrine sediment from bodies of glacial-age standing water also has been encountered in the form of deposits of uniform light gray silt greater than 50-ft thick in some areas (USACE 2001a).

Soil at Camp Ravenna is generally derived from the Wisconsin-age silty clay glacial till. Distributions of soil types are discussed and mapped in the *Soil Survey of Portage County, Ohio*, which describes soil as nearly level to gently sloping and poor to moderately well drained (USDA 1978). Much of the native soil at Camp Ravenna was disturbed during construction activities in former production and operational areas of the facility.

The Sharon Member of the Pennsylvanian Pottsville Formation is the primary bedrock beneath Camp Ravenna. In the western half of the facility, the upper members of the Pottsville Formation, including the Connoquenessing Sandstone (also known as the Massillon Sandstone), Mercer Shale, and uppermost Homewood Sandstone, have been found. The regional dip of the Pottsville Formation measured in the west portion of Camp Ravenna is between 5–11.5 ft per mile to the south.

3.3.3 Geologic Setting of Load Line 11

The bedrock formation underlying the unconsolidated deposits at Load Line 11, as inferred from existing geologic data, is the Pennsylvanian-age Pottsville Formation, Homewood Sandstone Member (Figure 3-3). The Homewood Sandstone Member is the uppermost unit of the Pottsville Formation. It typically occurs as a caprock on bedrock highs in the subsurface and ranges from well-sorted, coarse-grained, white quartzose sandstone to a tan, poorly sorted, clay-bonded, micaceous, medium- to fine-grained sandstone. Thin shale layers are prevalent in the Homewood member as indicated by a darker gray shade of color. Historical investigations have not encountered bedrock at Load Line 11 with an average borehole depth of 17 ft bgs and a maximum depth of 23 ft bgs. Bedrock was not encountered during PBA08 RI drilling activities where the maximum depth of boreholes was 13 ft bgs (Appendix A).

The primary soil type found at Load Line 11 is the Mahoning silt loam (MgB) (2-6% slopes) which covers over 85% of the AOC. The northern boundary of the AOC is comprised of Rittman silt loam (RsC2) (2-6% slopes) and the very eastern portion of the AOC is Wadsworth silt loam (WaB) (2-6% slopes) (USDA 2010). Mahoning silt loam is a gently sloping, poorly drained soil formed in silty clay loam or clay loam glacial till, generally where bedrock is greater than 6 ft bgs. The Mahoning silt loam has low permeability, with rapid runoff and seasonal wetness. The Rittman silt loam is a deep, moderately eroded, moderately well drained soil formed in silty clay or clay loam glacial till plains or moraines, generally where bedrock is greater than 6 ft bgs. The Wadsworth silt loam is a somewhat poorly drained, low permeability soil formed in silty clay loam glacial till.

The Phase I RI identified the Mahoning silt loam (2-6% slopes) as the primary soil type present at Load Line 11. The composition of unconsolidated deposits at Load Line 11 varies laterally and vertically across the AOC but generally consists of clay to sand-rich silt tills with interbedded sands and gravel lenses throughout. Deposits are generally stiff, moderately plastic, and tend to hold water. Cross-sections of the Load Line 11 subsurface were created from monitoring well lithology records to illustrate lateral distribution and variation of the discontinuous glacial sediment (MKM 2005a). Geotechnical analyses conducted during the Phase I RI indicated a grain size distribution of 0-20.9% gravel, 13.2-27.6% sand fractions, 17.1-60.3% silt fractions, and 27.2-69.5% clay fractions. Geotechnical classifications indicate the sample collected from 1-3 ft bgs was a lean clay with little sand, and the 4-6 ft bgs and 8-10 ft bgs samples were silts with increasing sand content with depth (MKM 2005a).

Geologic descriptions and geotechnical analyses of subsurface soil samples collected during the PBA08 RI are generally consistent with the conclusions from the Phase I RI. Overall, the PBA08 RI observed shallow clays followed by sandy silts and silty sands, and trace discontinuous gravel was observed. When encountered, groundwater was observed in soil borings from 4.6-9.2 ft bgs. Analyses of undisturbed geotechnical samples (Shelby tubes) collected during the PBA08 RI indicate a low permeability of 1.7E-06 to 1.7E-07 cm/sec for unconsolidated deposits at Load Line 11. A summary of the PBA08 RI geotechnical analyses, including porosity, density, and moisture content, is presented in Section 5.5. PBA08 RI boring logs containing geologic descriptions of unconsolidated deposits at Load Line 11 are presented in Appendix A.

3.4 HYDROGEOLOGY

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3.4.1 Regional Hydrogeology

Sand and gravel aquifers are present in the buried-valley and outwash deposits in Portage County, as described in the *Phase I Remedial Investigation Report for High-Priority Areas of Concern* (USACE 1998). Generally, these saturated zones are too thin and localized to provide large quantities of water for industrial or public water supplies; however, yields are sufficient for residential water supplies. Lateral continuity of these aquifers is unknown. Recharge of these units comes from surface water infiltrating precipitation and surface streams. Specific groundwater recharge and discharge areas at Camp Ravenna have not been delineated.

The thickness of the unconsolidated interval at Camp Ravenna ranges from thin to absent in the eastern and northeastern portion of Camp Ravenna to an estimated 150 ft (46 m) in the central portion of the facility. The groundwater table occurs within the unconsolidated zone in many areas of the facility. Because of the heterogeneous nature of the unconsolidated glacial material, groundwater flow patterns are difficult to determine with a high degree of accuracy. Vertical recharge from precipitation likely occurs via infiltration along root zones, desiccation cracks, and partings within the soil column. Laterally, most groundwater flow likely follows topographic contours and stream drainage patterns, with preferential flow along pathways (e.g., sand seams, channel deposits, or other stratigraphic discontinuities) having higher permeabilities than surrounding clay or silt-rich material. Figure 3-4 illustrates facility-wide potentiometric surface data in the unconsolidated interval from the January 2010 contemporaneous measurement event (EOM 2010).

Within bedrock units at Camp Ravenna, the principle water-bearing aquifer is the Sharon Sandstone/Conglomerate. Depending on the existence and depth of overburden, the Sharon Sandstone/Conglomerate ranges from an unconfined to a leaky artesian aquifer. Water yields from on-site water supply wells completed in the Sharon Sandstone/Conglomerate ranged from 30–400 gallons per minute (gpm) (USATHAMA 1978). Well yields of 5–200 gpm were reported for on-site bedrock wells completed in the Sharon Sandstone/Conglomerate (Kammer 1982). Other local bedrock units capable of producing water include the Homewood Sandstone, which is generally thinner and only capable of well yields less than 10 gpm, and the Connoquenessing Sandstone. Wells completed in the Connoquenessing Sandstone in Portage County have yields ranging from 5–100 gpm but are typically less productive than the Sharon Sandstone/Conglomerate due to lower permeabilities (Winslow et al. 1966).

Figure 3-5 shows the potentiometric surface within bedrock strata at Camp Ravenna in January 2010 (EQM 2010). The bedrock potentiometric map shows a more uniform and regional eastward flow direction than the unconsolidated zone that is not as affected by local surface topography. Due to the lack of well data in the western portion of Camp Ravenna, general flow patterns are difficult to discern. For much of the eastern half of Camp Ravenna, bedrock potentiometric elevations are higher than the overlying unconsolidated potentiometric elevations, indicating an upward hydraulic gradient. This evidence suggests there is a confining layer that separates the two aquifers. In the far eastern

area, the two potentiometric surfaces are at approximately the same elevation, suggesting that hydraulic communication between the two aquifers is occurring.

3.4.2 Load Line 11 Hydrologic/Hydrogeologic Setting

Ten groundwater monitoring wells were installed at Load Line 11 during the Phase I RI. Initial depths to groundwater varied from 5–17 ft bgs. Monitoring wells at the AOC ranged in completion from 15.55–22.35 ft bgs. All monitoring wells at Load Line 11 were screened in the Homewood Sandstone to monitor groundwater in the unconsolidated zone. Two additional monitoring wells LL11mw-011 (unconsolidated) and LL11mw-012 (bedrock) were installed under the FWGWMP in 2012.

All monitoring well groundwater elevations were collected under the FWGWMP. The potentiometric surface of the AOC from the January 2010 monitoring event is shown in Figure 3-1. The estimated groundwater flow directions reflect the January 2010 facility-wide potentiometric data presented in the *Facility-wide Groundwater Monitoring Program Report on the January 2010 Sampling Event* (EQM 2010). The potentiometric surface shows the groundwater flow pattern to the north toward Sand Creek. The hydraulic gradient at the AOC is 0.017 ft/ft.

Water level elevations at the AOC range from 1,068.40-1,091.73 ft amsl (less than 0.08-13.6 ft below top of casing) with the highest elevation at the bedrock well LL11mw-001. Potentiometric data indicate the groundwater table occurs within the unconsolidated zone throughout the AOC.

Results of slug tests performed at 10 monitoring wells (wells LL11mw-001 to LL11mw-010) during the Phase I RI indicate an average hydraulic conductivity of 3.49E-05 cm/s for the uppermost bedrock interval (MKM 2005a). Table 3-1 presents the hydraulic conductivity for each well at Load Line 11.

3.4.3 Surface Water

The following sections describe the regional and AOC-specific surface water.

3.4.3.1 Regional Surface Water

the upland areas of the facility.

Camp Ravenna resides within the Mahoning River watershed, which is part of the Ohio River basin. The west branch of the Mahoning River is the main surface stream in the area. The west branch flows adjacent to the west end of the facility, generally in a north to south direction, before flowing into the Michael J. Kirwan Reservoir, located to the south of State Route 5 (Figure 1-1). The west branch flows out of the reservoir and parallels the southern Camp Ravenna boundary before joining the Mahoning River east of Camp Ravenna. The western and northern portions of Camp Ravenna display low hills and a dendritic surface drainage pattern. The eastern and southern portions are characterized by an undulating to moderately level surface, with less dissection of the surface drainage. The facility is marked with marshy areas and flowing and intermittent streams whose headwaters are located in

The three primary watercourses that drain Camp Ravenna are (Figure 1-2):

- South fork of Eagle Creek,
- Sand Creek, and
- Hinkley Creek.

These watercourses have many associated tributaries. Sand Creek, with a drainage area of 13.9 square miles, generally flows in a northeast direction to its confluence with the south fork of Eagle Creek. In turn, the south fork of Eagle Creek continues in a northerly direction for 2.7 miles to its confluence with Eagle Creek. The drainage area of the south fork of Eagle Creek is 26.2 square miles, including the area drained by Sand Creek. Hinkley Creek originates just southeast of the intersection between State Route 88 and State Route 303 to the north of the facility. Hinkley Creek, with a drainage area of 11 square miles, flows in a southerly direction through the facility, and converges with the west branch of the Mahoning River south of the facility (USACE 2001a).

Approximately one-third of Camp Ravenna meets the regulatory definition of a wetland, with the majority of the wetland areas located in the eastern portion of the facility. Wetland areas at RVAAP include seasonal wetlands, wet fields, and forested wetlands. Many of the wetland areas are the result of natural drainage or beaver activity; however, some wetland areas are associated with anthropogenic settling ponds and drainage areas.

Approximately 50 ponds are scattered throughout the facility. Many were constructed within natural drainage ways to function as settling ponds or basins for process effluent and runoff. Others are natural in origin, resulting from glacial action or beaver activity. Water bodies at Camp Ravenna could support aquatic vegetation and biota as described in Section 2.3.2. Storm water runoff is controlled primarily by natural drainage, except in former operations areas where an extensive storm sewer network helps to direct runoff to drainage ditches and settling ponds. In addition, the storm sewer system was one of the primary drainage mechanisms for process effluent while production facilities were operational.

3.4.3.2 Load Line 11 Surface Water

Surface water at Load Line 11 occurs intermittently as storm water runoff within constructed or natural drainage ditches or conveyances along the north, west, and central portions of the FPA (Figure 3-1). Surface water flow is the primary migration pathway for contamination to leave the AOC flowing through ditches or surface water drainage features that exit the AOC. A storm sewer system is not present at Load Line 11.

The primary drainage routes for surface water (Figure 3-1) are the East Ditch that flows north and the West Ditch that flows west-northwest, both leading to Sand Creek immediately north of the AOC boundary. Sand Creek drains to the northeast into the south fork of Eagle Creek. The ditches tend to hold water for extended periods due to the low permeability of soil at Load Line 11 (MKM 2005a). During the PBA08 RI, surface water was observed at the AOC, flowing through ditches in some areas.

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The general climate of Camp Rayenna is continental and is characterized by moderately warm and humid summers, reasonably cold and cloudy winters, and wide variations in precipitation from year to year. The climate data presented below for Camp Ravenna were obtained from available National Weather Service records for the 30-year period of record from 1981-2010 at the Youngstown Regional Airport, Ohio (http://www.weather.gov/climate/xmacis.php?wfo=cle). Wind speed data for Youngstown, Ohio, from the National Climatic Data are Center (http://www1.ncdc.noaa.gov/pub/data/ccd-data/wndspd14.txt) for the available 30-year period of record from 1984-2014.

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Average annual rainfall at Camp Ravenna is 38.86 inches, with the highest monthly average occurring in July (4.31 inches) and the lowest monthly average occurring in February (2.15 inches). Average annual snowfall totals approximately 62.9 inches, with the highest monthly average occurring in January (17.1 inches). Due to the influence of lake-effect snowfall events associated with Lake Erie (located approximately 35 miles to the northwest of Camp Ravenna), snowfall totals vary widely throughout northeastern Ohio.

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The average annual daily temperature in the Camp Ravenna area is 49.3°F, with an average daily high temperature of 70.9°F and an average daily low temperature of 26.1°F. The record high temperature of 100°F occurred in July 1988, and the record low temperature of -22°F occurred in January 1994. The prevailing wind direction at Camp Ravenna is from the southwest, with the highest average wind speed occurring in January (10.3 miles per hour) and the lowest average wind speed occurring in August (6.5 miles per hour). Thunderstorms occur approximately 35 days per year and are most abundant from April through August. Camp Ravenna is susceptible to tornadoes; minor structural damage to several buildings on facility property occurred as the result of a tornado in 1985.

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Table 3-1. Hydraulic Conductivities Measured at Load Line 11 during the Phase I RI

| Monitoring Well | Screened Interval | | Hydraulic Conductivity |
|------------------------|-------------------|--------------------------------------|------------------------|
| ID | (ft bgs) | Geologic Material Adjacent to Screen | (cm/s) |
| LL11-mw001 | 11.37-21.37 | Clayey Silt/Sand | 1.74E-05 |
| LL11-mw002 | 6.3–16.3 | Clayey Silt | 6.32E-05 |
| LL11-mw003 | 5.9-15.9 | Sandy Silt | 6.48E-05 |
| LL11-mw004 | 6.05-16.05 | Clayey Silt | 2.34E-07 |
| LL11-mw005 | 6.17–16.17 | Sandy Silt/Silty Clay/Sand | 4.69E-06 |
| LL11-mw006 | 5.55-15.55 | Sandy Silt | 6.81E-05 |
| LL11-mw007 | 12.35-22.35 | Sandy Silt/Silty Sand | 7.16E-06 |
| LL11-mw008 | 5.55–15.55 | Silt/Sand | 3.95E-05 |
| LL11-mw009 | 6.65-16.65 | Sandy Silt/Sand/Silty Clay | 3.23E-05 |
| LL11-mw010 | 10.85-20.85 | Sandy Silt/Silty Sand | 5.13E-05 |

Source: Report for the Remedial Investigation at Load Line 11 (AOC 44) (MKM 2005a).

AOC = Area of Concern.

bgs = Below ground surface.

cm/s = Centimeters per second.

ft = Feet.

ID = Identification.

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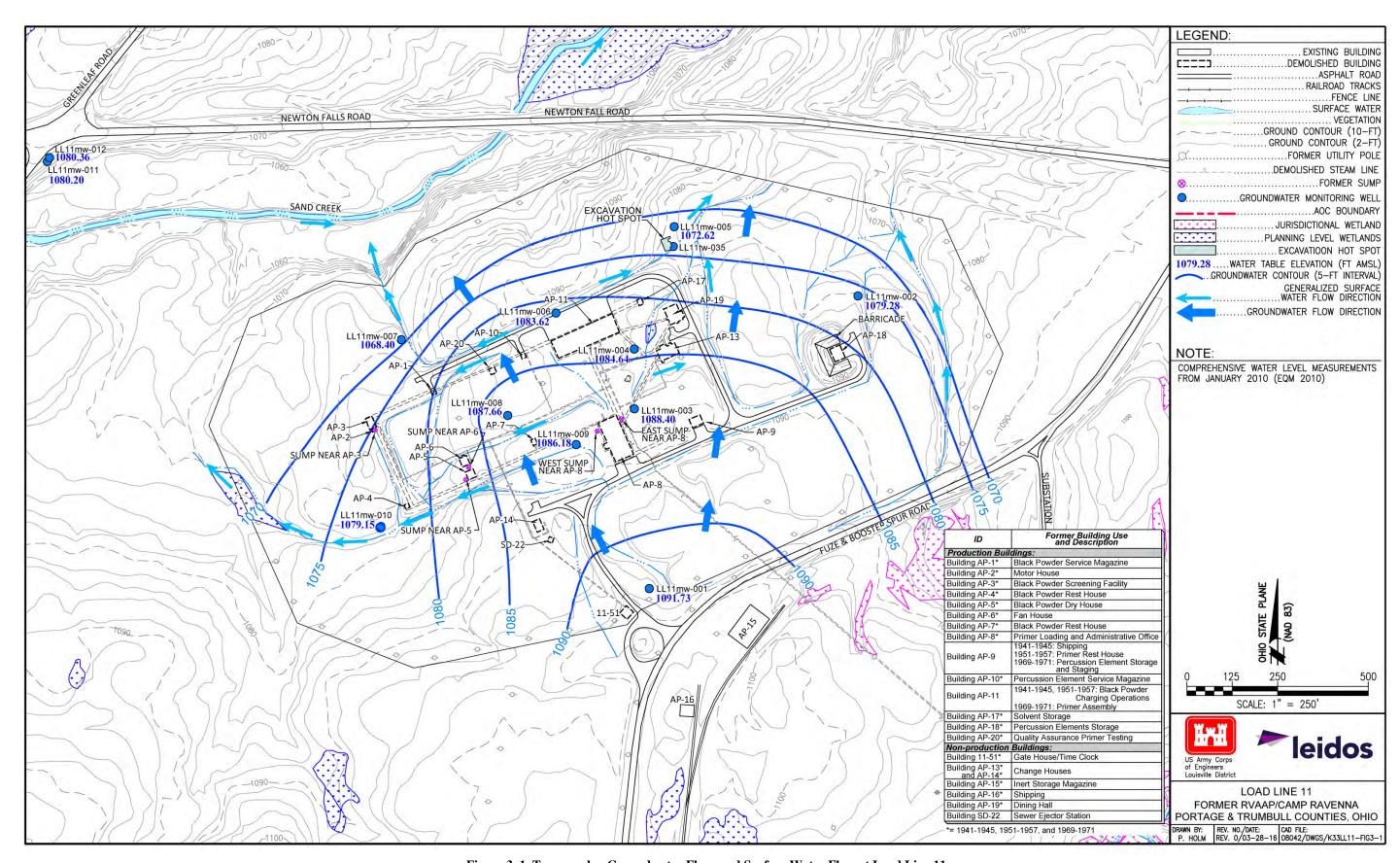


Figure 3–1. Topography, Groundwater Flow, and Surface Water Flow at Load Line 11

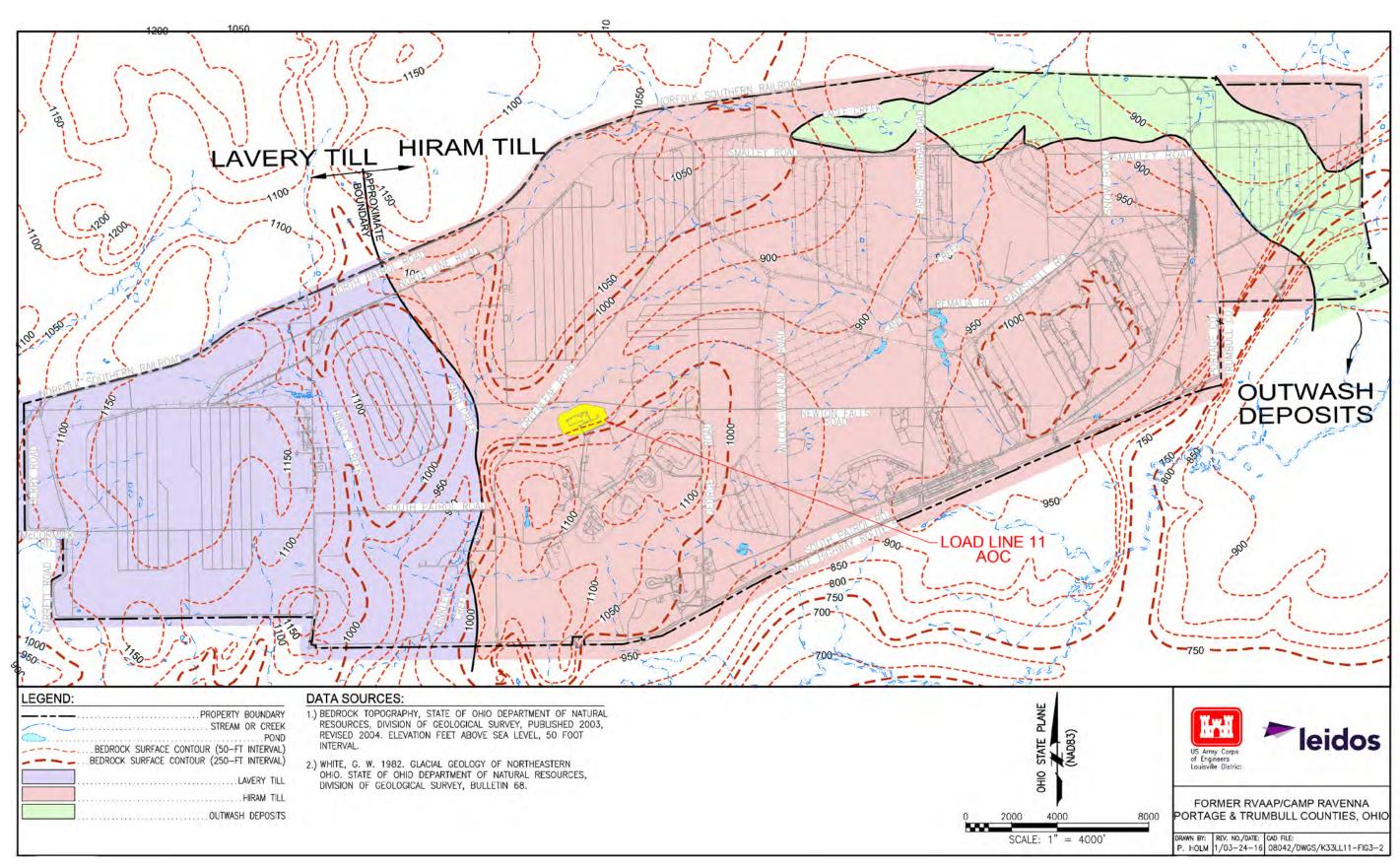


Figure 3-2. Geologic Map of Unconsolidated Deposits on Camp Ravenna

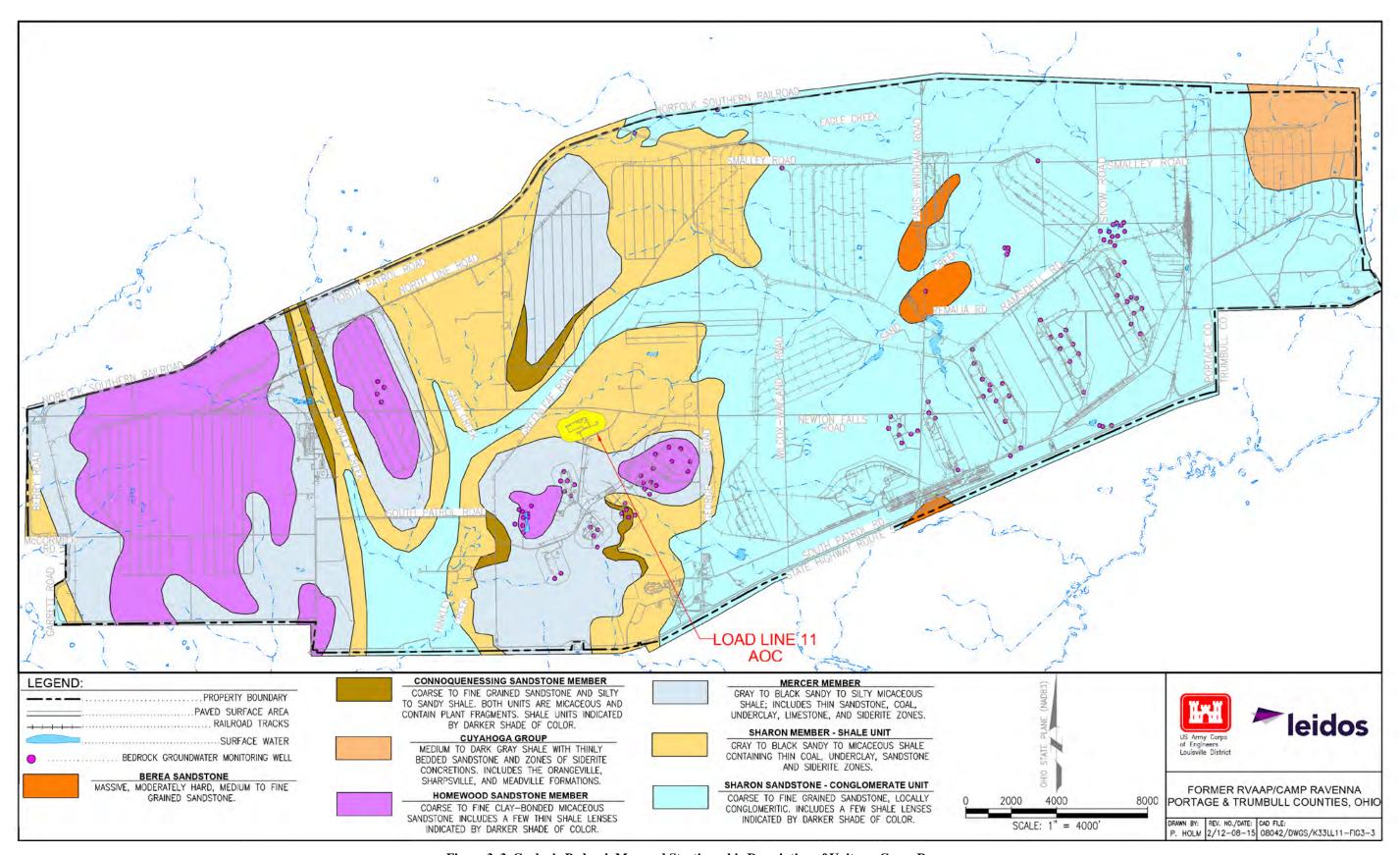


Figure 3–3. Geologic Bedrock Map and Stratigraphic Description of Units on Camp Ravenna

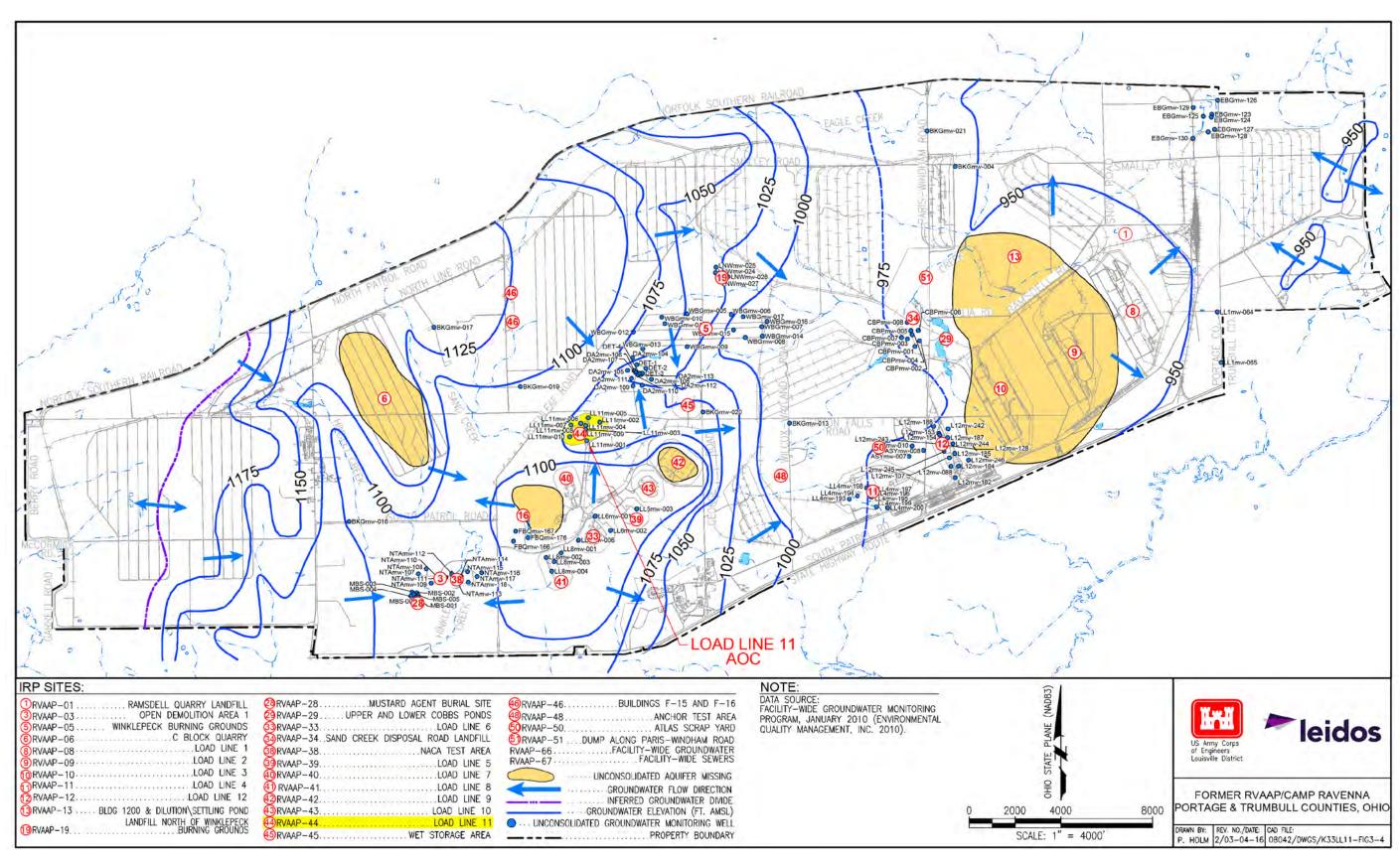


Figure 3-4. Potentiometric Surface of Unconsolidated Aquifer at Camp Ravenna

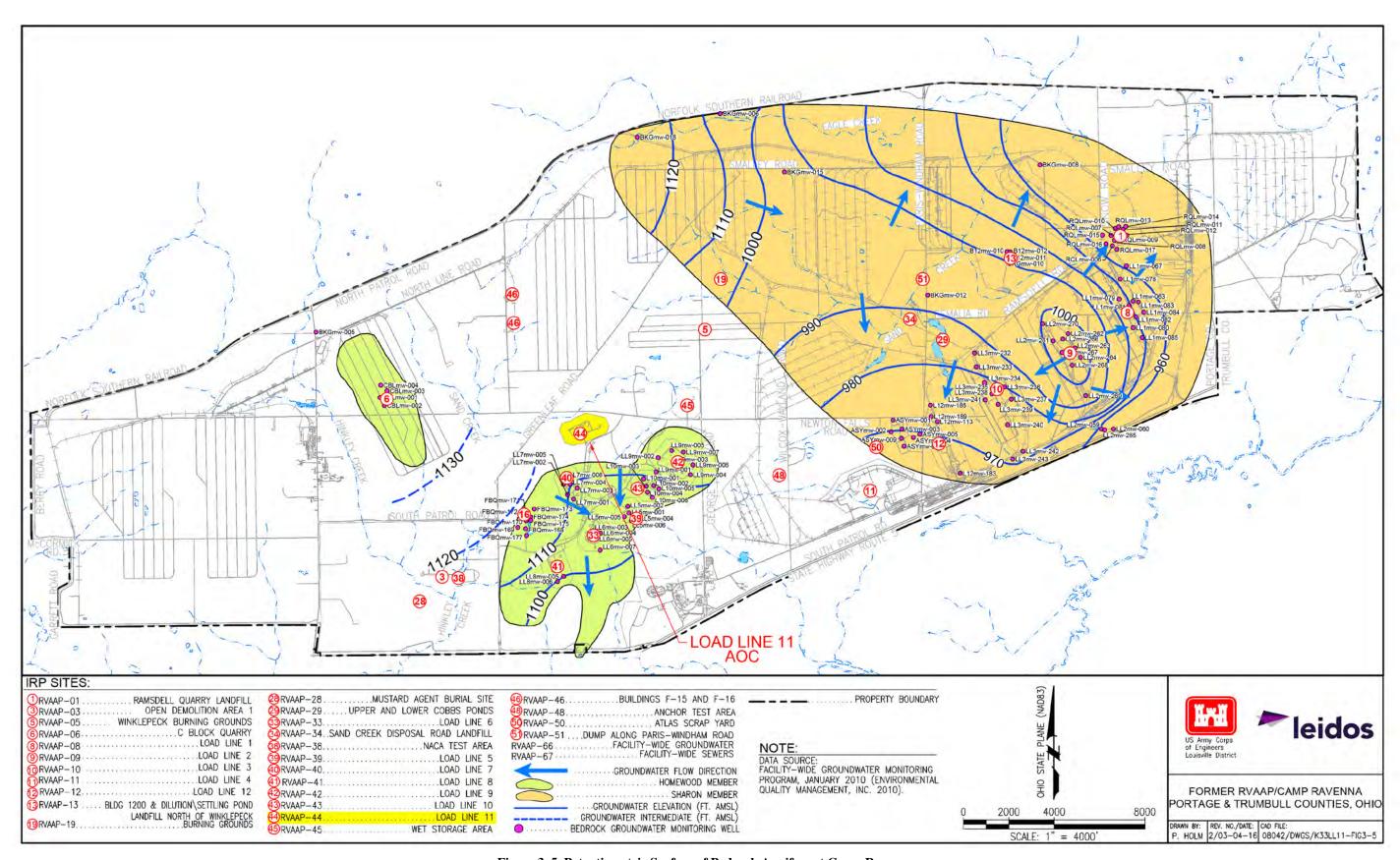


Figure 3–5. Potentiometric Surface of Bedrock Aquifers at Camp Ravenna

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4.0 SITE ASSESSMENTS, REMOVAL ACTIONS, INVESTIGATIONS, AND DATA ASSEMBLY

This section summarizes all previous site assessments and investigations conducted at Load Line 11. These previous activities include assessments to prioritize the AOC and investigations that collected data used in support of this RI.

4.1 LOAD LINE 11 PREVIOUS ASSESSMENTS AND EVALUATIONS

This section summarizes previous assessments and evaluations conducted at Load Line 11. These activities were generally performed to do an initial evaluation and/or prioritization assessment of the AOC. The data collected as part of these prioritization assessments and evaluations are not used in the nature and extent, fate and transport, HHRA, or ERA due to their age and lack of data quality documentation.

4.1.1 Installation Assessment of Ravenna Army Ammunition Plant

The *Installation Assessment of Ravenna Army Ammunition Plant* incorporated a review of historical operational information and available environmental data to assess the potential for contaminant releases from operational facilities. No sampling was performed at Load Line 11 as part of the assessment. The assessment identified the following conditions at RVAAP as applicable to Load Line 11 (USATHAMA 1978):

- Areas of RVAAP, including the production areas (i.e., Load Lines 5, 7, 8, 10, 11, and 12), burning grounds, test areas, and demolition areas were identified as AOCs contaminated with explosive waste, including TNT, composition B, lead azide, lead styphnate, and black powder;
- Black powder was the only explosive waste identified for Load Line 11;
- Load lines were rehabilitated in 1951; and
- No environmental stress was identified at RVAAP.

4.1.2 RCRA Facility Assessment

The purpose of the RCRA Facility Assessment was to perform a visual inspection of known AOCs and conduct historical documentation research to identify new AOCs and solid waste management units, as applicable (Jacobs 1989). The RCRA Facility Assessment did not include specific assessments or conclusions regarding Load Line 11, and no sampling was performed at Load Line 11 as part of the assessment. However, the RVAAP-26 Fuze and Booster Area Settling Tanks AOC was included in the assessment, as discussed in Section 2.4.3.

Load Line 11

4.1.3 RVAAP Preliminary Assessment

The *Preliminary Assessment for the Characterization of Areas of Contamination* researched RVAAP history, process operations, and historical data to identify AOCs (USACE 1996). Load Line 11 was not included in the assessment, with the exception of the RVAAP-26 Fuze and Booster Area Settling Tanks AOC, as discussed in Section 2.4.3.

The assessment provided the following general conclusions for RVAAP AOCs (all conclusions may not apply to Load Line 11; rather, these are general conclusions for all AOCs):

• COPCs at RVAAP AOCs were identified as explosives (TNT, RDX, HMX, composition B, and lead azide) and heavy metals (lead and mercury).

The primary sources of potential contamination were identified as wastewater effluent from the munitions assembly and demilitarization process, open burning and detonation of explosives, and landfill operations. Primary contaminant release mechanisms from load lines were process effluent discharges to surface water (drainage ditches, settling ponds, and streams) and process building wastewater wash-out to surface soil. Media of concern were identified as soil, sediment, groundwater, and surface water.

 The greatest potential for release of contaminants from load lines to groundwater likely occurred from wastewater effluent discharge to unlined earthen settling ponds. Concrete settling tanks, open drainage ditches, and storm sewers were also identified as concerns relative to groundwater.

 Known releases of contamination to surface water and soil have occurred from load line (assembly and demilitarization) operations.

The greatest potential for historical off-site contaminant migration during load line operations was identified as surface water. The greatest potential for current off-site contaminant migration was identified as groundwater and surface water.

4.1.4 Relative Risk Site Evaluation for Newly Added Sites

In 1998, the U.S. Army Center for Health Promotion and Preventative Medicine completed the *Relative Risk Site Evaluation for Newly Added Sites* (USACHPPM 1998) to "provide sufficient data to score RVAAP's newly discovered previously uninvestigated sites..." This document identified and provided a risk evaluation for 13 newly discovered and previously uninvestigated sites for the purpose of prioritizing future remedial or corrective activities. Of the 13 identified AOCs, 5 were assigned a Relative Risk Site Evaluation (RRSE) score of "high," and the remaining 8 were assigned a score of "medium."

The RRSE also included collecting surface soil, sediment, and subsurface soil samples at Load Line 11. The data collected at the site "...are minimal Level III data, as defined by U.S. EPA, and are not intended to be used as definitive evidence of contamination presence or absence or to support health risk assessment." This section summarizes the samples collected as part of the RRSE data, the

chemicals detected, and the associated prioritization recommendations, but the analytical results are not presented and are not used in subsequent evaluations in this Phase II RI Report.

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- 4 The RRSE evaluated the soil pathway (human receptor endpoint) using data from the six surface soil
- 5 samples collected near the former production buildings at Load Line 11 (RV-441 to RV-446). In
- 6 addition, sample RV-264A, collected during the 1997 RRSE, was included as part of the data set.
- 7 These samples were analyzed for explosives and inorganic chemicals. The groundwater pathway
- 8 (human receptor endpoint) was evaluated using 1996 RRSE data from one subsurface soil sample
- 9 (RV-264B) collected from 10-12 ft bgs, adjacent to the concrete settling tank east of Building AP-3.
- This sample was analyzed for explosives and inorganic chemicals.

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One sediment sample (RV-447) was collected during this evaluation; however, surface water was not evaluated as part of the RRSE as surface water was considered an intermittent source.

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Several inorganic chemicals and explosives were detected in surface soil, sediment, and groundwater (inferred from subsurface soil). Concentrations of analytes detected in surface soil, sediment, and groundwater are presented in Appendix D of the RRSE (USACHPPM 1998).

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The groundwater, surface soil, sediment, and surface water pathways were evaluated as follows:

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1. Groundwater

a. Contaminant Hazard Factor: Moderate.

- b. Migration Pathway Factor: Potential. There is no evidence that site contaminants are migrating. However, there are no physical barriers in place to prevent migration.
- c. Receptor Pathway Factor: Potential. It is unknown if any wells are directly downgradient from this AOC; however, groundwater near the AOC may be used for irrigation or drinking water.

2. Surface soil

- a. Contaminant Hazard Factor: Moderate.
- b. Migration Pathway Factor: Potential. There is no evidence that site contaminants are migrating. However, there are no physical barriers in place to prevent migration.
- c. Receptor Pathway Factor: Potential. While this area is surrounded by a fence with locked gates, hunters, scrappers, and firewood cutters may have access to the site.

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3. Sediment

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a. Contaminant Hazard Factor: Moderate.

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b. Migration Pathway Factor: Potential. There is no evidence that site contaminants are migrating. However, there are no physical barriers in place to prevent migration.

38 39 c. Receptor Pathway Factor: Potential. While this area is surrounded by a fence with locked gates, hunters, scrappers, and firewood cutters may have access to the site.

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4. Surface water

41 42 a. Not evaluated. While surface water is intermittently present at Load Line 11, it was not considered for RRSE scoring.

Human receptor endpoints were evaluated based on the available surface soil, sediment, and groundwater data. The RRSE scored Load Line 11 as a "high-priority" AOC due to potentially contaminated sediment potentially migrating to Sand Creek and affecting human and ecological receptors, including state endangered species (USACHPPM 1998).

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4.2 LOAD LINE 11 INTERIM REMOVAL ACTION

An IRA took place from January–June 2001 to remove contamination in primary migration pathways at Load Line 11 in accordance with the approved *Sampling and Analysis Plan Addendum for the Interim Removal Action at Load Line 11 (AOC 44)* (MKM 2004b). This section summarizes the IRA activities and information provided in the *LL-11 Interim Removal Action Report* (MKM 2004a).

The IRA was performed in parallel with the Phase I RI activities described in Section 4.3: removing sump water from production buildings, grouting selected sanitary sewer manholes, excavating contaminated ditch lines, and removing the petroleum-contaminated soil hot spot. Figure 4-1 presents the locations sampled under the IRA. Table 4-1 presents the sample locations, associated operations, and suite of chemicals analyzed as part of the IRA confirmation sampling. Tables 4-2 through 4-4 present the results of the analytes detected from samples collected during the IRA confirmation sampling.

The information and results from this IRA were included with the Phase I RI data when the Phase I RI HHRA and ERA were developed. As a result, all findings and recommendations relative to subsequent remedial efforts are provided in the Phase I RI Report (MKM 2005a) and used to support this Phase II RI Report.

4.2.1 Sump Excavation

Five sedimentation sumps were excavated, removed, and disposed during the IRA. These sumps were located adjacent to Buildings AP-3, AP-5, AP-6, and AP-8. During the initial sump removal operations, a fibrous liner material was found attached to the lead lining in each of the sumps. Bulk samples of the material were collected and submitted for asbestos analysis. Analytical results indicated the fibrous material was comprised of 60% chrysotile asbestos. As a result, the sumps were excavated and removed in accordance with federal (40 CFR Part 61, Subpart M) and state (OAC 3745-20) asbestos emission control regulations; the fibrous material was placed into plastic bags by asbestos abatement workers.

As part of the removal activities, approximately 15,000 gal of water were removed from sumps adjacent to process Buildings AP-3, AP-5, AP-6, and AP-8. During the water removal at Building AP-3, it was determined that the infiltration rate of groundwater into the sewer system was significant enough to impede removal operations. Consequently, the sewer manholes downgradient of each sump were filled with bentonite cement to prevent water from infiltrating back into the sumps during the excavation and removal operations. Once the water was removed, effluent sewer lines were cut and plugged with mechanical packers and cement grout, which prepared the sumps for removal.

Upon removing the sumps, one confirmation sample was collected from the bottom of each excavation [five total and one quality assurance (QA)/quality control (QC) sample]. The depths of confirmation samples LL11cs-001 to LL11cs-005 ranged from 5-6 ft bgs. The sump excavation confirmation samples were analyzed for explosives, target analyte list (TAL) metals, cyanide, sulfide, sulfate and nitrate. In addition, 10% of the samples were also analyzed for VOCs, semi-volatile organic compound (SVOCs), propellants, and pesticides/PCBs. Explosives and cyanide were not detected in any of the five sump excavations.

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4.2.2 Ditch Excavation

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Phase I RI sampling at six drainage ditch locations exhibited elevated concentrations of metals, VOCs, SVOCs, pesticides, and/or PCBs. These six drainage ditch locations were LL11sd-013, LL11sd-018, LL11sd-019, LL11sd-026, LL11sd-027, and LL11sd-033.

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Field screening for metals using X-ray fluorescence and nitrates using a Hach N Trak® soil test kit was performed to define lateral extent of the excavation. Figure 4-1 depicts the excavation extent of the ditches to a depth of 2 ft bgs. A total of 230 yds³ of contaminated soil was removed during the ditch excavation operations using a track-mounted excavator. The list below describes the drainage ditch locations that had a removal action performed, the targeted chemicals, and the associated confirmation samples.

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- Ditch southeast of Building AP-3 (at RI sample point LL11sd-027):
 - o Targeted metals and VOC contamination.
 - O Confirmation sample locations LL11cs-006 to LL11cs-010 are associated with this removal action. (Sample location LL11cs-010 was considered sediment in the IRA Report but was reclassified as soil in this Phase II RI Report.)
- Ditch east of Building AP-4 (at RI sample point LL11sd-028):
- o Targeted metals and VOC contamination.
 - O Confirmation soil sample locations LL11cs-011 to LL11cs-014 and sediment sample location LL11cs-015 are associated with this removal action.
 - Ditch north of Building AP-14 (at RI sample point LL11sd-032):
 - o Targeted metals, SVOC, and pesticide contamination.
 - O Confirmation soil sample locations LL11cs-016 to LL11cs-020 are associated with this removal action. (Sample locations LL11cs-016 and LL11cs-020 were considered sediment in the IRA Report but were reclassified as soil in this Phase II RI Report.)
 - Ditch east of Load Line 11 entrance road (at RI sample point LL11sd-013):
 - o Targeted metals contamination.
 - O Confirmation soil sample locations LL11cs-021 to LL11cs-025 are associated with this removal action. (Sample locations LL11cs-021 and LL11cs-025 were considered sediment in the IRA Report but were reclassified as soil in this Phase II RI Report.)
 - Ditch west of Building AP-8 (at RI sample point LL11sd-033):
 - o Targeted metals contamination.

- Confirmation soil sample locations LL11cs-026 to LL11cs-030 are associated with this removal action. (Sample locations LL11cs-026 and LL11cs-030 were considered sediment in the IRA Report but were reclassified as soil in this Phase II RI Report.)
- Ditch north of Building AP-19 (at RI sample points LL11sd-018 and LL11sd-019):
 - o Targeted metals contamination.
 - O Confirmation soil sample locations LL11cs-032 to LL11cs-039 and sediment sample location LL11cs-031 are associated with this removal action. (Sample location LL11cs-039 was considered sediment in the IRA Report but was reclassified as soil in this Phase II RI Report.)

4.2.3 Hot Spot Excavation

One 30 by 30 by 8 ft hot spot area, located in an open field north of Building AP-17, was excavated on March 22, 2001 due to petroleum contamination encountered during RI soil boring activities (MKM 2004a). This hot spot was identified from soil borings collected during the initial installation of monitoring well LL11mw-005. (The monitoring well was relocated approximately 50 ft northeast of the original location.)

To determine the lateral and vertical extent of contamination, a grid of soil borings were installed at 50 ft centers. Soil samples were collected from 2-4, 4-6, 6-8, or 8-10 ft bgs. Field assessments field (e.g., headspace readings, visual and olfactory observations) were used to determine if a sample was to be sent to an off-site laboratory. Samples sent to the off-site laboratory were analyzed for explosives, TAL metals, cyanide, VOCs, and SVOCs.

Field readings and laboratory results indicated that the contamination was between 4 and 8 ft bgs. In March 2001, a total of 130 yd³ of petroleum-contaminated soil was excavated at the hot spot area from 4-8 ft bgs. The upper 4 ft of the excavated soil was staged separately for reuse at Load Line 11. The hot spot removal location is presented on Figure 4-1.

Following excavation operations, six confirmation samples (LL11cs-040 to LL11cs-045) were collected at the bottom of the excavation (approximately 8 ft bgs) to verify that the petroleum-contaminated soil was removed. The confirmation samples were analyzed for VOCs, SVOCs, explosives, propellants, TAL metals, pesticides, PCBs, cyanide, sulfide, sulfate, nitrate, total petroleum hydrocarbons-diesel range organics (TPH-DRO), and total petroleum hydrocarbons-gasoline range organics (TPH-GRO). Upon obtaining approval to complete the hot spot removal, the excavation was backfilled with approved off-site soil.

4.2.4 Test Trench Excavations

UXO technicians encountered an area of high magnetometer readings north-northwest of the hot spot area while clearing the excavation area for the hot spot. During the excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavating test trenches, cable, scrap metal, bolts, and magnetite-containing rock were found.

- 1 Confirmation samples LL11cs-046 to LL11cs-050 were collected from the test trenches from 3-4 ft
- 2 bgs. As with the hot spot confirmation samples, these samples were analyzed for VOCs, SVOCs,
- 3 explosives, propellants, TAL metals, pesticides, PCBs, cyanide, sulfide, sulfate, nitrate, TPH-DRO,

4 and TPH-GRO.

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Excavation of test trenches and confirmation samples indicated that no UXO was present. Upon this determination, the test trenches were backfilled and leveled to the original ground surface elevation.

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4.3 LOAD LINE 11 REMEDIAL INVESTIGATIONS

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This section summarizes previous investigations conducted at Load Line 11. These investigations collected data of sufficient provenance and quality to be used to support the evaluations in this RI, including the nature and extent, fate and transport, HHRA, and/or ERA.

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The Report for the Remedial Investigation at Load Line 11 (AOC 44) (MKM 2005a) (herein referred to as the Phase I RI Report) presented SRCs and/or COPCs based on data evaluation protocols in use at the time the investigations were completed. The data and information is used in this Phase II RI Report; however, an updated screening process and the addition of new data and information may result in a different list of SRCs and/or COPCs.

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References to the "RVAAP full-suite analytes" generally include analyses of TAL metals, explosives, propellants (nitrocellulose and nitroguanidine), SVOCs, VOCs, PCBs, and pesticides.

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4.3.1 Phase I Remedial Investigation

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In 2000-2001, sampling was conducted at Load Line 11 in accordance with the *Sampling and Analysis Plan Addendum for the Remedial Investigation at Load Line 11* (MKM 2000; herein referred to as the Load Line 11 SAP Addendum). The Phase I RI was performed to accomplish the following:

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- Define the vertical and horizontal extent of contamination at Load Line 11,
- Gather sufficient data to conduct a Baseline Human Health Risk Assessment (BHHRA) and a Screening Ecological Risk Assessment (SERA),
 - Define potential transport pathways and receptor populations, and
 - Provide sufficient data for selecting remedial action alternatives.

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- The data from the RI sampling was used to direct IRA excavation/removal operations (discussed in Section 4.2) and was incorporated into an evaluation of risk for the AOC, followed by
- recommendations for remedial efforts, as necessary. Because the IRA confirmation results represent
- 39 existing site conditions, those confirmation results were included in the Phase I RI risk assessment.
- 40 Results of this Phase I RI are presented in the Phase I RI Report (MKM 2005a) and are summarized
- 41 in the following subsections.

4.3.1.1 Field Activities

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In addition to the activities performed under the IRA, the following field activities were conducted during the Phase I RI in June 2000-March 2001 at Load Line 11 to assess the potential impacts from former operations:

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- Collected discrete surface soil (0–1 ft bgs) samples,
- Collected subsurface soil (less than 1 ft bgs) samples,
- Collected surface water samples from ditches,
- Collected sediment samples from ditches,
- Collected subfloor samples in Buildings AP-8 and AP-11,
- Collected samples adjacent to sumps, and
- Surveyed sampling locations.

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The following activities also took place; however, these activities are pertinent to other co-located AOCs associated with Load Line 11 (e.g., Facility-wide Groundwater and Facility-wide Sewers) and are not discussed in this section:

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- Installed, developed, and sampled 10 monitoring wells;
- Conducted in-situ permeability testing (slug tests);
 - Collected sump sediment/sludge samples;
 - Collected sump water samples;
 - Collected sanitary sewer manhole water samples;
 - Collected sanitary sewer manhole sediment/sludge samples;
 - Collected geotechnical samples from monitoring well borings; and
- Surveyed monitoring well locations.

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Soil (surface and subsurface), sediment, and surface water samples were analyzed for TAL metals explosives, and anions. Additionally, 10% of samples were analyzed for RVAAP full-suite analytes. Hot spot locations near initial RI sampling locations with elevated VOC concentrations were analyzed for VOCs and SVOCs, and selected locations for TPH-DRO and TPH-GRO.

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Figure 4-2 presents the locations sampled under the Phase I RI. Table 4-5 presents the sample locations, associated operations, and suite of chemicals analyzed as part of the Phase I RI. Tables 4-6 through 4-9 present the results of the analytes detected from samples collected during the Phase I RI.

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Analytical laboratory procedures were completed in accordance with applicable professional standards, USEPA requirements, government regulations and guidelines, and specific project goals and requirements. Samples were analyzed as specified by the Facility-wide Sampling and Analysis Plan (FWSAP) current at the time of the investigation (USACE 1996) and the Load Line 11 SAP Addendum (MKM 2000). The data quality objectives (DQOs) were established for the Phase I RI and complied with USEPA Region 5 guidance. The requisite number of QA/QC samples was obtained during the investigation. The data validation determined that the data met the completeness

requirements for the project (100% complete), was usable, and that it satisfied the DQOs for the project.

4.3.1.2 Nature and Extent of Contamination

A total of 22 metals, 2 SVOCs, and 2 propellants were detected in surface soil (0–1 ft bgs), and 8 metals, 1 propellant, and 1 PCB were detected in subsurface soil (1–13 ft bgs) above RVAAP background concentrations (USACE 2001b) and/or Region 9 residential soil preliminary remediation goals (PRGs) at that time. The four metals and one SVOC detected in surface water were above RVAAP background concentrations and/or Region 9 tap water PRGs. A total of 20 metals, 1 propellant, and 1 SVOC were detected in sediment above RVAAP background concentrations and/or Region 9 residential soil PRGs. Figure 4-5 presents locations that exceed current screening criteria.

4.3.1.3 Baseline Human Health Risk Assessment

A BHHRA was included in the Phase I RI Report (MKM 2005a) for Load Line 11. The BHHRA utilized the screening process in effect at that time, based on the protocol established in the *RVAAP Facility-Wide Human Health Risk Assessor Manual* dated January 2004 to identify potential contaminants of concern (USACE 2005b). [This assessment predated the *RVAAP Facility-Wide Human Health Risk Assessor Manual – Amendment 1* (USACE 2005b) (herein referred to as the FWHHRAM)]. Risks were evaluated for the receptors available at that time [National Guard Trainee (the primary receptor of concern), National Guard Resident/Trainer, Hunter/Trapper, Security Guard/Maintenance Worker, and Resident Farmer (Adult and Child)]. The COPCs identified in the quantitative BHHRA are summarized in Table 4-10.

The BHHRA indicated that several significant uncertainties associated with the risk assessment were identified and must be considered carefully when making any risk management decisions, including the fact that the nature and extent of the contamination in soil, sediment, surface water, and groundwater at Load Line 11 had not been completely defined. Therefore, the results of the BHHRA are preliminary and subject to change based on the additional data and information presented in this Phase II RI Report.

4.3.1.4 Screening Ecological Risk Assessment

The SERA compared chemical concentrations detected in Load Line 11 environmental media to RVAAP facility-wide background concentrations for inorganic chemicals and ecological screening values (ESVs). The SERA followed screening methodology guidance presented in the 2003 RVAAP Facility-Wide Ecological Risk Work Plan (USACE 2003a) (herein referred to as the FWERWP) and Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2003). Chemicals were retained if they did not have screening values. Table 4-11 presents a summary of the COPCs identified per type of species based on the ecological risk calculations.

The SERA concluded no unacceptable ecological risks were identified at Load Line 11. No COPCs were retained for direct toxicity to terrestrial plants, soil invertebrates, sediment invertebrates

(benthos), and aquatic organisms. Refining conservative assumptions used during food chain modeling led to determining that the risks were acceptable. Refinement included using less conservative bioavailability, less conservative dietary benchmarks (LOAELs), more realistic average exposure scenarios (not maximum), and considering aquatic habitat quality in the site drainage ditches.

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4.3.1.5 Conclusions

The Phase I RI Report (MKM 2005a) provided the following conclusions applicable to Load Line 11:

• Potential risk exists to the National Guard Trainee due to manganese exposure based on conservative toxicity and exposure values used to calculate potential risk.

 Potential risk exists in using unconsolidated groundwater for domestic purposes due to arsenic concentrations. As arsenic is a naturally occurring metal, future groundwater monitoring was recommended to determine if arsenic concentrations was a background condition.

• There may be a need for remedial action or land use controls to limit future use of Load Line 11.

• No additional action is recommended based on ecological risk.

4.3.2 PBA08 Remedial Investigation

In November 2008, Science Applications International Corporation (SAIC) scientists performed a site walk of Load Line 11. The site walk was conducted to develop the *Performance-based Acquisition* 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1 (PBA08 SAP) (USACE 2009), which supplemented the Phase I RI in this Phase II RI Report and completed the RI phase of the CERCLA process. Numerous physical changes occurred at Load Line 11 between the Phase I RI and IRA sampling, and the development of the PBA08 SAP. As discussed in Section 2.2.3, building demolition activities took place in 2004-2005, and footer removal occurred in 2006. The PBA08 SAP considered the prior investigations and changes in AOC conditions when developing the DQOs and sampling scheme for completing the Load Line 11 RI. Section 4.5.4 discusses the suitability and use of samples collected to support this RI, with respect to changes in AOC conditions. The PBA08 SAP was reviewed and approved by representatives of the Army and Ohio EPA in January 2010.

As part of the PBA08 RI DQOs, an initial screening approach was used to help focus the investigation on specific chemicals and areas to be further evaluated by assessing the nature and extent of contamination observed in historical samples (Section 3.2.2 of the PBA08 SAP). Decision flowcharts for PBA08 RI surface and subsurface sampling are presented in Figures 4-3 and 4-4, respectively. The screening approach presented in the PBA08 SAP compared sample results from previous investigations at Load Line 11 to chemical-specific facility-wide cleanup goals (FWCUGs) at the 1E-06 cancer risk level and non-carcinogenic risk HQ of 0.1, as presented in the FWHHRAM (USACE 2005b). The most protective FWCUGs for the Resident Farmer (Adult and Child) and National Guard Trainee were referred to as "screening criteria." Previous results were also compared

to FWCUGs at the higher TR of 1E-05 and HQ of 1 to facilitate identifying potential source areas that may require additional sampling to refine the extent of contamination. Table 4-12 lists the chemicals with detected concentrations that exceeded screening criteria at the time of the PBA08 SAP in historical soil samples.

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In February-April 2010, the PBA08 RI was implemented by collecting surface and subsurface soil and surface water samples using discrete sampling techniques, and collecting sediment using composite sampling techniques. The results of the PBA08 RI sampling, combined with the results of the Phase I RI and IRA, were used to evaluate the nature and extent of contamination, assess potential future impacts to groundwater, conduct HHRAs and ERAs, and evaluate the need for remedial alternatives.

No groundwater samples were collected during the PBA08 RI, as the current conditions of groundwater will be evaluated as an individual AOC for the entire facility (designated as RVAAP-66) and addressed in a separate RI/FS Report.

A sample log for each sample and lithologic soil description for each soil boring collected during the PBA08 RI is included in Appendix A. The DQOs, field activities, sampling methodologies, QA/QC, and management of analytical data for the PBA08 RI are further expanded upon in Appendix I.

4.3.2.1 Surface Soil Sampling Rationale – Source Area Investigation

Samples were collected at the AOC to assess contaminant occurrence and distribution in surface soil. The PBA08 RI samples were designed to delineate the extent of areas previously identified as having the greatest likelihood of contamination (e.g., adjacent to production buildings or within sediment accumulation areas such as ditches). Table 4-13 presents the specific rationale for each surface soil sample collected during the PBA08 RI in April 2010. Table 4-14 presents the results of the analytes detected from surface soil samples collected during the PBA08 RI in April 2010. All PBA08 RI and historical sample locations used in this evaluation are presented on Figure 4-7.

Nine source area investigation discrete samples were collected in ditches not previously investigated, downgradient of former operational areas and former buildings, and at locations where historical screening criteria exceedances were observed to vertically delineate contamination (Figure 4-5). The sample locations in the PBA08 SAP were derived from the sample polygons provided during the development of the PBA08 SAP. One incremental sampling methodology (ISM) sample (10% of the total number of discrete samples collected) was analyzed for RVAAP full suite analytes.

4.3.2.2 Surface Soil Sampling Rationale – Chromium Speciation

As part of the PBA08 RI, three discrete chromium speciation samples were collected to evaluate the potential contribution of hexavalent chromium to the total chromium concentrations in soil. Samples from 0–1 ft bgs were collected in accordance with the bucket hand auger method described in Section 4.5.2.1.1 of the FWSAP (USACE 2001a). Two samples were collected from areas previously identified as having elevated total chromium concentrations (LL11sb-004 and LL11cs-011), and one

sample was collected from an area previously identified as having a total chromium concentration near background concentrations (LL11sd-025). The rationale for all samples collected as part of the PBA08 RI is summarized in Table 4-15. The locations of these samples are presented in Figure 4-5 and results are presented in Table 4-14.

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4.3.2.3 Subsurface Soil Sampling Rationale and Methods

The PBA08 RI used discrete samples from soil borings to complete the characterization of subsurface soil. Subsurface soil sampling was conducted according to the decision rules approved in the PBA08 SAP. The subsurface soil borings were located based on three objectives:

- Borings at locations where previous surface soil sampling results exceeded screening criteria and vertical delineation was warranted.
- Borings at locations where previous surface soil sampling results only slightly exceeded screening criteria to confirm that contaminant concentrations did not increase with depth.

Soil samples from 10 soil borings installed in areas with historical screening criteria exceedances were collected to further delineate the vertical extent of contamination in subsurface soil at the AOC (Figure 4-5). Table 4-16 presents the specific rationale for each subsurface soil sample collected for the PBA08 RI. Results of detected analytes are presented in Table 4-17.

The subsurface soil sampling depth intervals were documented in the PBA08 SAP. Each soil boring was sampled at 0–1, 1–4, 4–7, and 7–13 ft bgs. These sample intervals were selected to evaluate surface and subsurface exposure depths for the Resident Receptor (0–1 and 1–13 ft bgs) and National Guard Trainee (0–4 and 4–7 ft bgs). The sample collected from the 7–13 ft bgs interval was archived on site, while the 4–7 ft bgs interval sample was analyzed under an expedited five-day turnaround time. As specified in the PBA08 SAP, if any chemical concentration exceeded screening criteria in the 4–7 ft bgs sample, the 7-13 ft bgs sample was analyzed. One of the archived 7–13 ft bgs samples was analyzed because there were preliminary screening criteria exceedances for arsenic (24.7 mg/kg) in the 4–7 ft bgs sample interval from LL11sb-065. In addition, at least 10% of all subsurface samples from 7–13 ft bgs were submitted for laboratory analysis to ensure adequate characterization of the subsurface soil to 13 ft bgs. One sample collected from the 7–13 ft bgs sample interval was submitted for laboratory analysis for this purpose.

All samples were analyzed for TAL metals, explosives, PCBs, and PAHs; three samples were analyzed for RVAAP full-suite analytes to satisfy the PBA08 SAP sample requirements of a minimum of 10% frequency for full-suite analysis. Four QC field duplicates and four QA split samples were collected to satisfy the QA/QC sample requirements of 10% frequency for subsurface soil samples.

Two geotechnical samples were collected from one boring location to provide soil data for fate and transport modeling. Geotechnical data is provided in Appendix D. A pilot boring was installed with a Geoprobe at sample location LL11sb-085 to a depth of 6.5 ft bgs (where shallow bedrock was encountered) to allow lithological characterization of the soil and determine the appropriate

geotechnical sample intervals (Appendix A). The geotechnical sample location was offset from the pilot boring and drilled with hollow stem auger attachments. Geotechnical samples were collected beneath the hollow stem augers directly into the Shelby Tube. Undisturbed Shelby Tube samples were collected from 2–3.9 and 4-5.5 ft bgs. The Shelby Tubes were sealed with wax, capped, and submitted for laboratory geotechnical analysis for porosity, bulk density, moisture content, total organic carbon, grain size fraction analysis, and permeability.

4.3.2.4 Surface Water and Sediment Rationale and Methods

Surface water and sediment samples were collected to characterize current conditions and assess potential exit pathways from the AOC (Figure 4-5). Three co-located surface water and sediment samples were collected during the PBA08 RI from ditches and shallow conveyances within and exiting the AOC. The samples were collected in accordance with the following decision rules approved in the PBA08 SAP:

- At AOCs where overland flow of contaminants could occur to nearby perennial streams, those streams will be sampled. The sample locations may be outside of the AOC boundaries, but the samples represent the areas potentially impacted by the AOCs (Load Line 11).
- At points where contamination may migrate out of the AOC area, such as a ditch or a stream near the AOC boundary, samples were collected to characterize current conditions and determine whether contaminant migration may occur at surface water runoff exit points.

In addition to the samples collected within the AOC boundaries, this Phase II RI Report discusses the surface water and sediment samples collected at potential exit points from the Fuze and Booster Hill area (which includes Load Lines 5 through 11) and near the southern boundary of RVAAP/Camp Ravenna to determine nature and extent. Although not within the AOC boundary, this report evaluates sample LL11sd/sw-082 that was collected north of Load Line 11 and north of Newton Falls Road to assess the potential transport beyond the AOC boundary.

The surface water grab samples were collected by the handheld bottle method in accordance with Section 4.3 of the PBA08 SAP and analyzed for the RVAAP full-suite analytes. Additionally, water quality parameters for temperature, pH, conductivity, dissolved oxygen, and turbidity were collected using calibrated water quality meters (Hanna Instrument Models 9828 and 98703). The sediment samples were collected in accordance with Section 4.2 of the PBA08 SAP. The samples consisted of a multi-aliquot composite with 10 aliquots selected randomly within a 5-ft radius of the identified sample location. Each aliquot was collected by a push probe to a maximum depth of 0.5 ft bgs. For VOC analysis, one discrete sample collected from 0.5 ft bgs was collected within the 5-ft sampling radius of the discrete sample location and placed directly in the appropriate, labeled, sample container. Two sediment samples (and their associated QA/QC samples) were analyzed for explosives, TAL metals, PCBs, and SVOCs, while one sample was analyzed for RVAAP full-suite analytes. Table 4-18 presents the specific rationale for the surface water and sediment samples collected for the PBA08 RI. The locations of these samples are presented in Figure 4-5 and results are presented in Table 4-19 and Table 4-20 for surface water and sediment, respectively.

4.3.2.5 Changes from the Work Plan

Significant changes to the PBA08 SAP are documented in field change requests provided in Appendix B. Changes made in the field based on AOC-specific conditions are not documented on field change requests but on the field sampling logs (Appendix A). These changes are presented on Table 4-21. New coordinates for all station locations can be found on the field sampling logs.

4.3.2.6 August 2012 Sediment Sample

In August 2012, one sediment sample (LL11sd-096-5874-SD) was re-collected from the West Ditch for RVAAP full-suite analyses because the previous sediment sample LL11sd-084-5595-SD was analyzed only for TAL metals, explosives, SVOCs, and VOCs (Figure 4-5). The ditch was dry; therefore, no additional surface water sample was collected. The location of this sample is presented on Figure 4-5. Table 4-20 presents the analytes detected from the additional sediment sample collected in August 2012. There were several chemical concentrations in sediment exceeding background concentrations, and these results are included as part of the SRC screen for the West Ditch exposure unit (EU) and in the HHRA and ERA.

4.4 FACILITY-WIDE BACKGROUND EVALUATION

Facility-wide background values for inorganic constituents in soil, sediment, surface water, and groundwater were developed in 1998, as documented in the *Phase II Remedial Investigation Report for the Winklepeck Burning Grounds* (USACE 2001b). These facility-wide background values were employed in the data reduction and screening process described in Section 4.5.2 and the remainder of the evaluations in this RI (e.g., nature and extent, fate and transport). Background locations were selected using aerial photographs and site visits from areas believed to be unaffected by RVAAP activities. Soil, sediment, surface water, and groundwater samples were collected from those locations to determine the range of background concentrations that could be expected in these media. Results from the site-specific background data collection were used to determine if detected metals and potential anthropogenic compounds (such as PAHs) are site-related, naturally occurring, or from non-RVAAP-related anthropogenic sources.

A total of 14 wells were installed in established background locations to collect filtered and unfiltered samples from the bedrock and unconsolidated zones. These samples were analyzed for TAL metals and cyanide for determining background concentrations.

Soil samples were collected from each of the background monitoring well locations from three intervals: 0–1, 1–3, and greater than 3 ft bgs. Because boring locations were changed during sampling based on the lithological requirements for well screen intervals, all depth intervals for soil were not sampled for each boring. Background soil samples were analyzed for TAL metals, cyanide, SVOCs, total organic carbon, pesticides, PCBs, and VOCs.

Seven stream locations upstream of RVAAP activities were sampled for sediment and surface water to characterize background conditions. Seven locations for sampling sediment and surface water

representative of background conditions were selected along Hinkley, Sand, and Eagle Creeks. Background sediment samples were analyzed for TAL metals, cyanide, SVOCs, pesticides, PCBs, and VOCs. Surface water samples were analyzed for TAL metals and cyanide.

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Using the sampling results, an evaluation of outliers, data assessment, and statistical analysis were performed to determine background concentrations for each media. For surface soil samples, PAHs, in addition to metals, were elevated in four samples. PAHs are related to combustion products and could indicate human disturbance at the locations where they were detected. Visits to the sampling locations and a review of aerial photography prior to the establishment of RVAAP indicated that these sampling locations were near homes or farms and could have been influenced by activities associated with those structures.

During the finalization of background concentrations at the former RVAAP, the Army and Ohio EPA agreed that facility-wide background concentrations would only be applicable for inorganics. All organic analytes (e.g., PAHs, VOCs, explosives) were classified as anthropogenic and potentially related to RVAAP operations; therefore, no background values were established for these classes of compounds. The final, approved facility-wide background concentrations or inorganics are presented in Table 4-22.

4.5 DATA EVALUATION METHOD

Data evaluation methods for Load Line 11 are consistent with those established in the FWCUG Report and specified in the PBA08 SAP (USACE 2009). The processes used to evaluate the analytical data involved three general steps: (1) defining data aggregates; (2) conducting data verification, reduction, and screening; and (3) presenting data.

4.5.1 Definition of Aggregates

Load Line 11 data were grouped (aggregated) in three ways to evaluate contaminant nature and extent and complete the HHRA and ERA. The initial basic aggregation of data was by environmental medium: surface soil, subsurface soil, sediment, and surface water. For each medium-specific aggregate, an evaluation was conducted to determine if further aggregation was warranted with respect to AOC characteristics, historical operations, ecological habitat, and potential future remedial strategy and Land Use (e.g., spatial aggregates). Data for soil were further aggregated based on depth and sample type for consistency with RVAAP human health risk EUs and guidance established in the FWHHRAM and FWCUG Report. Data aggregates for evaluating the nature and extent of contamination at Load Line 11 are as follows:

• Surface Soil (0-1 ft bgs): Using the above data aggregation criteria, surface soil within the geographic area of Load Line 11 was subdivided into two spatial aggregates: FPA and NPA. The FPA aggregate is defined as the area encompassed within the perimeter road surrounding the Load Line 11 main production area. The NPA aggregate is defined as the area between the production area perimeter road and the fence line. The FPA contains all known or

- potential primary contaminant sources such as the former operational buildings. The former buildings located within the NPA were limited to administrative and storage functions.
 - Subsurface Soil (less than 1 ft bgs): Includes data from discrete sample intervals 1–4, 4–7, and 7–13 ft bgs. This medium is subdivided into two spatial aggregates, the FPA and NPA, on the same basis as surface soil.
 - Sediment: Sediment was subdivided into three spatial aggregates. The East Ditch aggregate is represented by a discrete sediment sample collected at a ditch line which drained portions of the eastern part of the FPA. The West Ditch aggregate is represented by a discrete sediment sample collected at a ditch line west of former Building AP-4 which drains the western half of Load Line 11. The Sewer Outfall aggregate is represented by a sediment sample collected at the sanitary system overflow outfall north of the FPA.
 - **Surface Water:** This medium is subdivided into the East Ditch and West Ditch spatial aggregates on the same basis as sediment. The drainage ditches only intermittently convey surface water. Although the Sewer Outfall is defined as a spatial aggregate for sediment, no surface water is available for this sample location, as it is predominantly dry.

The soil data aggregates were further subdivided to define HHRA and ERA, as discussed in Section 7.1.1.

4.5.2 Data Verification, Reduction, and Screening

4.5.2.1 <u>Data Verification</u>

Data verification was performed on 55 surface and subsurface soil samples (including QC duplicates) collected during the PBA08 RI in March 2010 and April 2011. Data from the Phase I RI and IRA were verified and completed as presented in these respective summary reports. Analytical results were reported by the laboratory in electronic format and loaded into the Ravenna Environmental Information Management System (REIMS) database. Data verification was performed to ensure all requested data were received and complete. Data qualifiers were assigned to each result based on the laboratory QA review and verification criteria.

Results were qualified as follows:

- "U" not detected.
- "UJ" not detected, reporting limit estimated.
- "J" indicates the analyte was positively identified, but the associated numerical value is an approximate concentration of the analyte in the sample.
- "R" result not usable.

In addition to assigning qualifiers, the verification process also selected the appropriate result to use when re-analyses or dilutions were performed. Where laboratory surrogate recovery data or laboratory QC samples were outside of analytical method specifications, the verification chemist determined whether laboratory re-analysis should be used in place of an original reported result. If the laboratory reported results for diluted and undiluted samples, diluted sample results were used for

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those analytes that exceeded the calibration range of the undiluted sample. A complete discussion of verification process results is contained in the data QC summary report (Appendix C). The data QC summary report also includes a summary table of the assigned data qualifiers and an accompanying rationale. Independent, third-party validation of 10% of the RI data and 100% of the USACE QA laboratory data was performed by a subcontractor to the USACE Louisville District.

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4.5.2.2 Data Reduction

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Calculating data summary statistics was the initial step in the data reduction process to identify SRCs. Eligible historic and current AOC data were extracted from the database. Results from OC splits and field duplicates, as well as rejected results, were excluded from the data screening process. As stated in Section 5.4.7 of the FWSAP, "The duplicate is submitted as 'blind' to the laboratory and is used to determine whether the field sampling technique is reproducible and to check the accuracy of reported laboratory results." Therefore, duplicates are not used in the data screening process. All analytes having at least one detected value were included in the data reduction process.

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Summary statistics calculated for each data aggregate included the minimum, maximum, and average (mean) values and the proportion of detected results to the total number of samples collected. For calculating mean values, non-detected results were addressed by using one-half of the reported detection limit as a surrogate value for each compound (USEPA 1989). Non-detected results with elevated detection limits (more than five times the contract-required detection limit) were excluded from the summary statistics in order to avoid skewing the mean value calculations.

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4.5.2.3 Data Screening

27 28 After reduction, the data were screened to identify SRCs using the processes outlined below. All surface and subsurface soil samples collected under the PBA08 RI were discrete samples and screened for SRCs.

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Additional screening of identified SRCs against applicable criteria (e.g., EPA RSLs, FWCUGs, and ESVs) was conducted: (1) in the fate and transport evaluation (Section 6.0) to identify CMCOPCs, (2) in the HHRA to identify human health COPCs and COCs (Section 7.2), and (3) in the ERA to evaluate COPECs (Section 7.3). Figure 4-6 illustrates the screening process to identify SRCs and COPCs at Load Line 11 in accordance with the FWCUG Report. The steps involved in the SRC screening process are summarized below. All chemicals that were not eliminated during the screening steps were retained as SRCs.

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- Data quality assessment: Review the usability of the RI data set with respect to established DOOs as discussed in Section I.4.5 of Appendix I.
- 40 • Background screening: The maximum detected concentrations (MDCs) of naturally 41 42 43
 - occurring inorganic chemicals were compared to background concentrations. If background concentrations were exceeded, the respective inorganic chemicals were retained as SRCs. No background concentrations were established for organic chemicals. As such, all detected organic chemicals were retained as SRCs.

- Screening of essential human nutrients: Chemicals that are considered essential nutrients (e.g., calcium, chloride, iodine, iron, magnesium, potassium, phosphorous, and sodium) are an integral part of the human food supply and are often added to foods as supplements. USEPA recommends these chemicals not be evaluated unless they are grossly elevated relative to background concentrations or would exhibit toxicity at the observed concentrations at an AOC (USEPA 1989). Recommended daily allowance (RDA) and recommended daily intake (RDI) values are available for all of these chemicals (Table 4-23). Screening values were calculated for receptors ingesting 100 mg of soil per day or 1 L of groundwater per day to meet their RDA/RDI. In the case of calcium, magnesium, phosphorous, potassium, and sodium, a receptor ingesting 100 mg of soil per day would receive less than the RDA/RDI value, even if the soil consisted of the pure mineral (i.e., soil concentrations at 1,000,000 mg/kg). Essential nutrients detected at or below their RDA/RDI-based screening levels (SLs) were eliminated as SRCs.
- **Frequency of detection/WOE screening:** The FWCUG Report and *Final (Revised) USACE* RVAAP Position Paper for the Application and Use of Facility-Wide Human Health Cleanup Goals (USACE 2012b) (hereafter referred to as the Position Paper for Human Health CUGs) establish the protocol for frequency of detection and WOE screening. These guidance documents denote that analytes (with exception of explosives and propellants) detected in less than 5% of the discrete samples are screened out from further consideration if the sample population consists of 20 or more samples and evidence exists that the analyte is not AOC related. The WOE evaluated magnitude and location (clustering) of detected results and if the distribution of detected results indicated a potential source of the chemical. If the detected results for a chemical showed: (1) no clustering; (2) concentrations were not substantially elevated relative to detection limit; and (3) the chemical did not have an evident source, the results were considered spurious, and the chemical was eliminated from further consideration. This screening was applied to all organic and inorganic chemicals with the exception of explosives and propellants; all detected explosives and propellants were considered SRCs regardless of frequency of detection. The frequency-of-detection screening was performed for surface soil and subsurface soil, because these data sets were comprised of 20 or more samples.

4.5.3 Data Presentation

Data summary statistics and screening results for SRCs in surface and subsurface soil, sediment, and surface water at Load Line 11 are presented for each medium and spatial aggregate. Analytical results for SRCs are also presented in data summary tables: Tables 4-24 and 4-25 for surface soil, Tables 4-26 and 4-27 for subsurface soil, Tables 4-28 through and 4-30 for sediment, and Tables 4-31 and 4-32 for surface water.

The complete laboratory analytical data packages are included in Appendix D. In order to maximize efficiency for laboratory reporting and data management activities, all of the samples received at the laboratory on a given day were reported in a single data package. Therefore, results may be present in data packages associated with different AOCs. All samples for Load Line 11 have sample identifications beginning with "LL11."

The tables in Appendix D present the analytical results for samples collected during the PBA08 RI, Phase I RI, and IRA. Sample locations from the IRA and two RIs are presented on Figure 4-7. Analytical results are grouped by media (e.g., surface soil, subsurface soil, sediment, surface water), spatial aggregate (i.e., FPA, NPA, West Ditch, East Ditch, Sewer Outfall), and class of analyte (e.g., explosives, inorganic chemicals) for ease of reference.

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4.5.4 Data Evaluation

All quality-assured sample data were further evaluated to determine suitability for use in the RI under two primary considerations: (1) representativeness with respect to current AOC conditions; and (2) sample collection methods (e.g., discrete vs. ISM). Table 4-33 presents the designated use for all available Load Line 11 samples.

4.5.4.1 Surface and Subsurface Soil

Surface and subsurface soil samples at Load Line 11 were collected during the 2000–2001 Phase I RI, 2001 IRA, and the PBA08 RI. Samples were collected during the Phase I RI from June 2000-March 2001, and the results indicated the need for an IRA to remove contaminated soil and infrastructure from Load Line 11. Soil excavations were conducted at building sumps in the FPA and the hot spot northeast of former Building AP-17. Samples collected within these excavated areas prior to the removal actions were eliminated from evaluation in the screening in this Phase II RI, as these data are no longer representative of media located at the Load Line 11. These samples are not carried forward in the nature and extent evaluation or risk assessments to evaluate and characterize Load Line 11. These sample locations include:

- LL11sb-031 from 4-6 ft bgs,
- LL11sb-031A from 6-8 ft bgs,
- LL11sb-032 from 4-6 ft bgs,
- LL11sb-033 from 6-8 ft bgs,
- LL11sb-034 from 2-4 ft bgs, and
- LL11sb-036 from 4-6 ft bgs.

Confirmatory samples collected following the removal actions were retained for SRC screening purposes, as these results are considered representative of current conditions at Load Line 11. These samples are presented in Table 4-33. No disturbance activities were conducted in the intervening time between collection and the PBA08 RI activities. Historical surface soil samples collected beneath building slabs prior to removal were retained for nature and extent evaluation only. These locations were likely disturbed during slab removal, but the soil itself was not removed. Historical subsurface soil samples were collected below a depth of 13 ft bgs and retained for evaluating contaminant migration and nature and extent, but these samples were not utilized in the SRC screening.

The RRSE samples RV-264A, RV-264B, and RV-441 to RV-446 were not included in this RI evaluation, as the RRSE data collected at the site "...are minimal Level III data, as defined by U.S.

EPA, and are not intended to be used as definitive evidence of contamination presence or absence or to support health risk assessment."

Additionally, two subsurface soil samples (from two soil borings) were collected during the PBA08 RI for geotechnical analysis only.

4.5.4.2 Sediment

Sediment samples were collected during the Phase I RI in 2000-2001, and the results indicated the need for an IRA to remove contaminated sediment and infrastructure (e.g. sumps) from Load Line 11. Excavations were conducted within contaminated ditch lines. Samples collected within these excavated areas prior to the removal actions were eliminated from evaluation in the PBA08 RI screening, as these data are no longer representative of media located at the Load Line 11. These samples are not carried forward in the nature and extent evaluation or risk assessments to evaluate and characterize Load Line 11. These sample locations include:

- LL11sd-002 at 3 ft bgs,
- LL11sd-003 at 3 ft bgs,
- LL11sd-005 at 7 ft bgs,
- LL11sd-013 from 0-1 ft bgs,
- LL11sd-018 from 0-0-5 ft bgs,
 - LL11sd-019 from 0-1 ft bgs,
 - LL11sd-027 from 0-0-5 ft bgs,
 - LL11sd-028 from 0-0-5 ft bgs,
 - LL11sd-032 from 0-1 ft bgs, and
 - LL11sd-033 from 0-1 ft bgs.

Confirmatory samples collected following the removal actions were retained for SRC screening purposes, as these results are considered representative of current conditions at Load Line 11.

For sediment data from the West Ditch and East Ditch, if a PBA08 RI sample was obtained from a historical sample location, the PBA08 RI result was considered to represent current conditions and was screened for SRCs. If the drainage had changed in a production area after IRA excavations or building demolition, the historical sample was not used in the screening. All historical sediment and surface water samples superseded in the SRC screen by PBA08 RI data were used only for evaluating contaminant nature and extent (e.g., temporal trends) and contaminant transport.

Sediment samples collected from the sanitary sewer system during the Phase I RI and Facility-wide Sewers RI are not included in the evaluation within this report, as those samples are evaluated as part of the Facility-wide Sewers AOC (designated as RVAAP-67).

One sediment sample (RV-447) was collected during 1998 RRSE The RRSE sample RV-447 was not included in this RI evaluation, as the RRSE data collected at the site "...are minimal Level III data, as

defined by U.S. EPA, and are not intended to be used as definitive evidence of contamination presence or absence or to support health risk assessment."

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4.5.4.3 Surface Water

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As noted in Sections 2.2.3 and 4.2.1, sump water was sampled then removed, filtered, and discharged to the ground during IRA sump excavation and removal activities at former Buildings AP-3, AP-5, AP-6, and AP-8. Surface water samples were collected during the Phase I RI in 2000-2001, where sediment was later excavated from contaminated ditch lines during the IRA. Samples collected within these excavated areas prior to the removal actions were eliminated from evaluation in the PBA08 RI screening, as these data are no longer representative of media located at the Load Line 11. These samples are not carried forward in the nature and extent evaluation or risk assessments to evaluate and characterize Load Line 11. These sample locations include:

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- LL11sw-001,
- LL11sw-002,
- LL11sw-003,
 - LL11sw-004,
 - LL11sw-005,
- 20 LL11sw-006, and
 - LL11sw-013.

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Surface water samples collected from the sanitary sewer system during the Phase I RI are not included in the evaluation within this report, as those samples are evaluated as part of the Facility-wide Sewers AOC (designated as RVAAP-67).

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All previous surface water samples from the Phase I RI were collected using discrete sample methods, as were those collected under the PBA08 RI. Both the historical and PBA08 RI data were utilized in screening for SRCs and COPCs and carried forward into the risk assessment, with the exception of samples considered not representative of current conditions, or collected from contaminated ditch lines prior to IRA activities.

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Table 4–1. IRA Confirmation Sample Locations

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|-------------------------------|---|---|-----------------------|------------------------------------|
| LL11cs-001 | 5–6 | Metals, explosives, anions | Sump Removal near Building AP-3 | 1941-1945, 1951-1957, and 1969-1971: Building AP-3 utilized for black powder screening. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-002 | 5–6 | Full suite, anions | Sump Removal near Building AP-6 | 1941-1945, 1951-1957, and 1969-1971: Building AP-6 utilized as black powder rest house. Evaluated effectiveness of excavation efforts during the IRA activities. QC sample collected. | None | Metals, black powder |
| LL11cs-003 | 5–6 | Metals, explosives, anions | Sump Removal at Building AP-5 | 1941-1945, 1951-1957, and 1969-1971: Building AP-5 utilized as a black powder dry house. Evaluate effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-004 | 5–6 | Full suite, anions | West Sump Removal at Building AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-005 | 5–6 | Full suite, anions | East Sump Removal at Building AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-006 | 0–1 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch southeast of Building AP-3 (Black Powder Screening) excavated at LL11sd-027. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-007 | 1–2 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch southeast of Building AP-3 (Black Powder Screening) excavated at LL11sd-027. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-008 | 1–2 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch southeast of Building AP-3 (Black Powder Screening) excavated at LL11sd-027. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-009 | 1–2 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch southeast of Building AP-3 (Black Powder Screening) excavated at LL11sd-027. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-010 | 0-1 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch southeast of Building AP-3 (Black Powder Screening) excavated at LL11sd-027. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-011 | 0-1 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch east of Building AP-4 (Black Powder Rest House) excavated at LL11sd-026. Evaluated effectiveness of excavation efforts during the IRA activities. QC sample collected. | None | Metals, black powder |
| LL11cs-012 | 1–2 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch east of Building AP-4 (Black Powder Rest House) excavated at LL11sd-026. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-013 | 1–2 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch east of Building AP-4 (Black Powder Rest House) excavated at LL11sd-026. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-014 | 1–2 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch east of Building AP-4 (Black Powder Rest House) excavated at LL11sd-026. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-015 | 0–1 | Metals, VOCs | Drainage Ditch Excavation | Drainage ditch east of Building AP-4 (Black Powder Rest House) excavated at LL11sd-026. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-016 | 0–1 | VOCs, pesticides, PCBs, SVOCs | Drainage Ditch Excavation | Drainage ditch just north of Building AP-14 (Change House) excavated at LL11sd-032. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-017 | 1–2 | VOCs, pesticides, PCBs, SVOCs | Drainage Ditch Excavation | Drainage ditch just north of Building AP-14 (Change House) excavated at LL11sd-032. Evaluated effectiveness of excavation efforts during the IRA activities. QC sample collected. | None | Metals, explosives |
| LL11cs-018 | 1–2 | VOCs, pesticides, PCBs, SVOCs | Drainage Ditch Excavation | Drainage ditch just north of Building AP-14 (Change House) excavated at LL11sd-032. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-019 | 1–2 | VOCs, pesticides, PCBs, SVOCs | Drainage Ditch Excavation | Drainage ditch just north of Building AP-14 (Change House) excavated at LL11sd-032. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-020 | 0–1 | VOCs, pesticides, PCBs, SVOCs | Drainage Ditch Excavation | Drainage ditch, on the south end of the AOC just east of the entrance road, excavated at LL11sd-013. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-021 | 0–1 | Metals | Drainage Ditch Excavation | Drainage ditch, on the south end of the AOC just east of the entrance road, excavated at LL11sd-013. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-022 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch, on the south end of the AOC just east of the entrance road, excavated at LL11sd-013. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-023 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch, on the south end of the AOC just east of the entrance road, excavated at LL11sd-013. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-024 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch, on the south end of the AOC just east of the entrance road, excavated at LL11sd-013. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |
| LL11cs-025 | 0-1 | Metals | Drainage Ditch Excavation | Drainage ditch, on the south end of the AOC just east of the entrance road, excavated at LL11sd-013. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives |

Table 4–1. IRA Confirmation Sample Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|-------------------------------------|---|--|-----------------------|------------------------------------|
| LL11cs-026 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch west Building AP-8 (artillery primer loading) excavated at LL11sd-033. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-027 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch west Building AP-8 (artillery primer loading) excavated at LL11sd-033. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-028 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch west Building AP-8 (artillery primer loading) excavated at LL11sd-033. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-029 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch west Building AP-8 (artillery primer loading) excavated at LL11sd-033. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-030 | 0–1 | Metals | Drainage Ditch Excavation | Drainage ditch west Building AP-8 (artillery primer loading) excavated at LL11sd-033. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, black powder |
| LL11cs-031 | 0–1 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. QC sample collected. | None | Metals, explosives, solvents |
| LL11cs-032 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-033 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-034 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-035 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-036 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-037 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-038 | 1–2 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-039 | 0-1 | Metals | Drainage Ditch Excavation | Drainage ditch north of Buildings AP-19 (dining hall) (also east of Building AP-17 [solvent storage], excavated at LL11sd-018 and LL11sd-019. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Metals, explosives, solvents |
| LL11cs-040 | 3–7 | Full suite, anions, TPH- DRO/GRO | Hot Spot Excavation | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. As a result of the contamination, LL11mw-005 was relocated approximately 50 ft northeast of the original location. Evaluated effectiveness of excavation efforts during the IRA activities. QC sample collected. | None | Solvents, petroleum |
| LL11cs-041 | 3–7 | Full suite, anions, TPH- DRO/GRO | Hot Spot Excavation | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. As a result of the contamination, LL11mw-005 was relocated approximately 50 ft northeast of the original location. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Solvents, petroleum |
| LL11cs-042 | 3–7 | Full suite, anions, TPH- DRO/GRO | Hot Spot Excavation | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. As a result of the contamination, LL11mw-005 was relocated approximately 50 ft northeast of the original location. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Solvents, petroleum |
| LL11cs-043 | 3–7 | Full suite, anions, TPH- DRO/GRO | Hot Spot Excavation | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. As a result of the contamination, LL11mw-005 was relocated approximately 50 ft northeast of the original location. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Solvents, petroleum |
| LL11cs-044 | 7–8 | Full suite, anions, TPH- DRO/GRO | Hot Spot Excavation | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. As a result of the contamination, LL11mw-005 was relocated approximately 50 ft northeast of the original location. Evaluated effectiveness of excavation efforts during the IRA activities. | None | Solvents, petroleum |
| LL11cs-045 | 7–8 | Full suite, anions, TPH- DRO/GRO | Hot Spot Excavation | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at | None | Solvents, petroleum |

Table 4–1. IRA Confirmation Sample Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|---|---|---|---|--|-----------------------|------------------------------------|
| | | · | | LL11sb-005 for installing LL11mw-005. As a result of the contamination, LL11mw-005 was relocated approximately 50 ft northeast of the original location. Evaluated effectiveness of excavation efforts during the IRA activities. QC sample collected. | | |
| LL11cs-046 | 3–4 | Full suite, anions, TPH- DRO/GRO | Test Pit Excavation | An area of high magnetometer readings was encountered by the UXO technicians north-northwest of the hot spot; while clearing the excavation area for the hot spot. During excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavation, cable, scrap metal, bolts, and magnetite-containing rock were found. After excavation, and determining no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation. Confirmation samples collected. | None | Metals, explosives |
| LL11cs-047 | 3–4 | Full suite, anions, TPH- DRO/GRO | Test Pit Excavation | An area of high magnetometer readings was encountered by the UXO technicians north-northwest of the hot spot; while clearing the excavation area for the hot spot. During excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavation, cable, scrap metal, bolts, and magnetite-containing rock were found. After excavation, and determining no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation. Confirmation samples collected. | None | Metals, explosives |
| LL11cs-048 | 3–4 | Full suite, anions, TPH- DRO/GRO | Test Pit Excavation | An area of high magnetometer readings was encountered by the UXO technicians north-northwest of the hot spot; while clearing the excavation area for the hot spot. During excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavation, cable, scrap metal, bolts, and magnetite-containing rock were found. After excavation, and determining no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation. Confirmation samples collected. | None | Metals, explosives |
| LL11cs-049 | 3–4 | Full suite, anions, TPH- DRO/GRO | Test Pit Excavation | An area of high magnetometer readings was encountered by the UXO technicians north-northwest of the hot spot; while clearing the excavation area for the hot spot. During excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavation, cable, scrap metal, bolts, and magnetite-containing rock were found. After excavation, and determining no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation. Confirmation samples collected. | | Metals, explosives |
| LL11cs-050 | 3–4 | Metals, explosives, pesticides, PCBs, SVOCs, anions, TPH- DRO/GRO | Test Pit Excavation | An area of high magnetometer readings was encountered by the UXO technicians north-northwest of the hot spot; while clearing the excavation area for the hot spot. During excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavation, cable, scrap metal, bolts, and magnetite-containing rock were found. After excavation, and determining no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation. Confirmation samples collected. | None | Metals, explosives |
| AOC = Area of concern. bgs = Below ground surf ft = Feet. IRA = Interim remedial a PCB = Polychlorinated b QC= Quality control. RI = Remedial investiga SVOC = Semi-volatile o TPH-GRO/DRO = Total UXO = Unexploded orde | action. iphenyl. ion. rganic compo petroleum hy | ound. rdrocarbons/gasoline range organics/die | esel range organics. | | | |

VOC = Volatile organic compound.

Table 4–2. Analytes Detected in IRA Confirmation Surface Soil Samples

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | NPA |
|------------------------|------------|------------------|------------------|------------------|------------------|---------------------|------------------|------------------|------------------|-----------------------|------------------|
| Station | | LL11cs-006 | LL11cs-010 | LL11cs-011 | LL11cs-011 | LL11cs-016 | LL11cs-020 | LL11cs-021 | LL11cs-025 | LL11cs-030 | LL11cs-039 |
| 2 1111-1 | | LL11cs-006-0001- | LL11cs-010-0001- | LL11cs-011-0001- | LL11cs-011-0001- | LL11cs-016-0001- | LL11cs-020-0001- | LL11cs-021-0001- | LL11cs-025-0001- | LL11cs-030-0001- | LL11cs-039-0001- |
| Sample ID | | SD | SD | FD | SD | SD | SO | SD | SD | SD | SD |
| Date | | 03/21/01 | 03/21/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/21/01 | 03/21/01 | 03/21/01 | 03/20/01 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | | | | | Pesticides/PCBs, | Pesticides/PCBs, | | | | |
| Analyte | Criteria | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals, VOCs | SVOCs, VOCs | SVOCs, VOCs | TAL Metals | TAL Metals | TAL Metals | TAL Metals |
| | | | | | | ls (mg/kg) | 1 | 1 | | | |
| Aluminum | | 14500 | 9340 | 18300* | 23600* | NR | NR | 13400 | 23800* | 13600 | 7940 |
| Antimony | | <0.32U | <0.23U | 1.1* | 0.5 | NR | NR | <0.37U | 0.46 | <0.26U | <0.32U |
| Arsenic | 15.4 | | 15 | 14.9 | 22.6* | NR | NR | 8.2 | 8.8 | 10.9 | <4.3U |
| Barium | 88.4 | | 52 | 62.5 | 70.6 | NR | NR | 68.5 | 94.7* | 65.8 | 61.1 |
| Beryllium | 0.88 | | 0.49 | 0.69 | 0.84 | NR | NR | 0.7 | 1.1* | 0.56 | 0.61 |
| Cadmium | | 0.71* | 0.33* | <0.19U | <0.16U | NR | NR | 0.44* | 0.58* | <0.2U | 0.57* |
| Calcium | 15800 | | 1790 | 1740 | 1490 | NR | NR | 3220 | 3660 | 3610 | 2160 |
| Chromium | | 20.4* | 13.9 | 23.2* | 28.4* | NR | NR | 15.7 | 26.3* | 18.1* | 10.8 |
| Cobalt | 10.4 | | 10.6* | 7.4 | 7.4 | NR | NR | 9.9 | 16.6* | 5.6 | 10.7* |
| Copper | | 21.9* | 16.1 | 19.2* | 23.8* | NR | NR | 14.6 | 27.9* | 9.8 | 9.3 |
| Cyanide | | NR | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Iron | 23100 | | 21300 | 25600* | 38300* | NR | NR | 16400 | 27000* | 25100* | 20600 |
| Lead | | 102* | 31.2* | 26.1 | 17 | NR | NR | 26.8* | 42.4* | 12.4 | <8.6U |
| Magnesium | | 3600* | 2740 | 3800* | 4450* | NR | NR | 1830 | 2940 | 3550* | 1500 |
| Manganese | 1450 | | 356 | 211 | 147 | NR | NR | 410 | 342 | 260 | 443 |
| Mercury Nickel | | 0.074* | <0.042U | 0.049* | <0.043U | NR NR | NR NR | 0.07 * | 0.08* | <0.043U | <0.048U |
| Potassium | | 23.7* 1950* | 20.9 1480* | 20.5 2150* | 22.6* 2910* | NR | NR NR | 1500* | 33.2* 2660* | 15.4 1620 * | 12 675 |
| Selenium | | <0.64U | <0.48U | 0.61 | 0.94 | NR | NR | 0.63 | 1.3 | 0.55 | 0.86 |
| Sodium | | 232* | 172* | 827* | 943* | NR | NR | 189* | 278* | 185* | 655* |
| Thallium | 123 | <0.21U | <0.15U | <0.22U | 0.21* | NR | NR | <0.25U | <0.28U | <0.17U | <0.21U |
| Vanadium | 31.1 | | 16.8 | 30.3 | 38.8* | NR | NR | 22 | 32* | 25.2 | 20.9 |
| Zinc | 61.8 | | 71.3* | 75.5* | 63.8* | NR | NR | 219* | 465* | 55.2 | 176* |
| Zine | 01.0 | 137 | 71.5 | 13.3 | 1 | ns (mg/kg) | TVIC | 21) | 103 | 33.2 | 170 |
| Nitrate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Sulfide | None | | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | ı | ı | Explosives and | Propellants (mg/kg) | • | | l | | ı |
| Nitrocellulose | None | NR | NR | NR | | NR | NR | NR | NR | NR | NR |
| Nitroguanidine | None | | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | | Cs (mg/kg) | | | | | |
| Benz(a)anthracene | None | | NR | NR | NR | <0.53U | 0.27J* | NR | NR | NR | NR |
| Benzo(a)pyrene | None | | NR | NR | NR | 0.16J* | 0.34J* | NR | NR | NR | NR |
| Benzo(b)fluoranthene | None | | NR | NR | NR | <0.53U | 0.39J* | NR | NR | NR | NR |
| Benzo(ghi)perylene | None | | NR | NR | NR | <0.53U | 0.24J* | NR | NR | NR | NR |
| Benzo(k)fluoranthene | None | | NR | NR | NR | <0.53U | <0.51U | NR | NR | NR | NR |
| Chrysene | None | | NR | NR | NR | 0.17J* | 0.34J* | NR | NR | NR | NR |
| Fluoranthene | None | | NR | NR | NR | 0.29J* | 0.6* | NR | NR | NR | NR |
| Indeno(1,2,3-cd)pyrene | None | | NR | NR | NR | <0.53U | 0.27J* | NR | NR | NR | NR |
| Phenanthrene | None | | NR | NR | NR | <0.53U | 0.26J* | NR | NR | NR | NR |
| Pyrene | None | NR | NR | NR | NR | <0.53U | 0.67* | NR | NR | NR | NR |

Table 4–2. Analytes Detected in IRA Confirmation Surface Soil Samples (continued)

| Aggregate | | FPA | NPA |
|---------------------|-----------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|
| Station | | LL11cs-006 | LL11cs-010 | LL11cs-011 | LL11cs-011 | LL11cs-016 | LL11cs-020 | LL11cs-021 | LL11cs-025 | LL11cs-030 | LL11cs-039 |
| | | LL11cs-006-0001- | LL11cs-010-0001- | LL11cs-011-0001- | LL11cs-011-0001- | LL11cs-016-0001- | LL11cs-020-0001- | LL11cs-021-0001- | LL11cs-025-0001- | LL11cs-030-0001- | LL11cs-039-0001- |
| Sample ID | | SD | SD | FD | SD | SD | SO | SD | SD | SD | SD |
| Date | | 03/21/01 | 03/21/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/21/01 | 03/21/01 | 03/21/01 | 03/20/01 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | | | | | Pesticides/PCBs, | Pesticides/PCBs, | | | | |
| Analyte | Criteria | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals, VOCs | SVOCs, VOCs | SVOCs, VOCs | TAL Metals | TAL Metals | TAL Metals | TAL Metals |
| | | | | | Pesticides | /PCBs (mg/kg) | | | | | |
| PCB-1254 | None | NR |
| gamma-Chlordane | None | NR | NR | NR | NR | <0.0064U | <0.0064U | NR | NR | NR | NR |
| | $VOCs\ (mg/kg)$ | | | | | | | | | | |
| Acetone | None | 0.008* | <0.006U | <0.007U | <0.007U | <0.008U | <0.008U | NR | NR | NR | NR |

FPA = Former Production Area.

ft = Feet.

ID = Identification.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.

NPA= Non-production Area.

NR = Not reported/not analyzed.

PCB = Polychlorinated biphenyl.

SVOC = Semi-volatile organic compound.

TAL = Target analyte list.

U = Not detected.

VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

< = I ess than

< = Less than.

Table 4–3. Analytes Detected in IRA Confirmation Subsurface Soil Samples

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|-------------------|----------------------------|--------------------|--------------------|--------------------|---------------------------------|--------------------|------------------|------------------|--------------------|--------------------|
| Station | | LL11cs-001 | LL11cs-002 | LL11cs-002 | LL11cs-003 | LL11cs-004 | LL11cs-005 | LL11cs-007 | LL11cs-008 | LL11cs-009 |
| Sample ID | | LL11cs-001-0001-SO | LL11cs-002-0001-FD | LL11cs-002-0001-SO | LL11cs-003-0001-SO | LL11cs-004-0001-SO | | | LL11cs-008-0001-SO | LL11cs-009-0001-SO |
| Date | | 01/24/01 | 01/26/01 | 01/26/01 | 01/26/01 | 02/21/01 | 02/21/01 | 03/21/01 | 03/21/01 | 03/21/01 |
| Depth (ft) | | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 |
| Parameters | | 210 010 | 2.0 0.0 | 2.0 0.0 | 2.0 0.0 | 2.0 0.0 | 2.0 0.0 | 110 210 | 1.0 2.0 | 1.0 2.0 |
| Analyzed | | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | | | |
| Analyte | Background Criteria | Explosives | analytes | analytes | Explosives | analytes | analytes | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals, VOCs |
| V | <u> </u> | <u>*</u> | | | Metals (mg/kg) | · | V | , | , | , |
| Aluminum | 19500 | | 12800 | 15000 | 12400 | 7450 | 12500 | 6910 | 13600 | 11700 |
| Antimony | 0.96 | | 0.24 | 0.91 | <0.22U | 0.28 | 0.35 | <0.21U | <0.24U | <0.23U |
| Arsenic | 19.8 | | 11.2 | 6.3 | 13.8 | 17.5 | 17.4 | 12.5 | 13.7 | 16.6 |
| Barium | | 40.4 | 73.1 | 74.6 | 65.8 | 27.7 | 81.9 | 32.1 | 62.4 | 55.2 |
| Beryllium | 0.88 | 0.38 | 0.57 | 0.6 | 0.49 | 0.45 | 0.7 | <0.3U | 0.58 | 0.54 |
| Cadmium | 0 | 0.47* | 0.26* | <0.16U | 0.19* | <0.15U | <0.15U | <0.15U | <0.17U | <0.19U |
| Calcium | | 24000 | 11000 | 10300 | 7610 | 1190 | 1770 | 1080 | 1870 | 13800 |
| Chromium | 27.2 | | 16.2 | 17.6 | 13.9 | 11.3 | 16.8 | 11 | 16.7 | 17.5 |
| Cobalt | 23.2 | | 10.1 | 9.2 | 8 | 7.9 | 7.6 | 5.6 | 9.4 | 8.6 |
| Copper | 32.3 | 18.5 | 20 | 18.2 | 17.2 | 19.7 | 17.6 | 15.1 | 17.4 | 22.7 |
| Cyanide | 0 | <0.54U | <0.4U | <0.5U | <0.47U | <0.0059U | <0.0058U | NR | NR | NR |
| Iron | 35200 | 17000 | 23000 | 23100 | 21600 | 21300 | 20800 | 18300 | 22500 | 25400 |
| Lead | | 31.7* | 39* | 61.4* | 28.1* | 11.1 | 12.7 | 11.5 | 16 | 20.2* |
| Magnesium | 8790 | | 4040 | 4160 | 3160 | 2490 | 2930 | 1890 | 3300 | 5730 |
| Manganese | 3030 | 422 | 353 | 336 | 369 | 222 | 313 | 294 | 300 | 249 |
| Mercury | 0.044 | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.039U | <0.043U | <0.04U |
| Nickel | 60.7 | 14.2 | 21 | 21.4 | 20 | 19.4 | 22.5 | 13.7 | 20.6 | 23.8 |
| Potassium | 3350 | 1370 | 1850 | 2330 | 1580 | 1470 | 1800 | 977 | 1800 | 1820 |
| Selenium | 1.5 | | <0.04U | 0.44 | <0.34U | 0.48 | <0.38U | <0.38U | <0.43U | <0.47U |
| Sodium | | 91.7 | 115 | 117 | 83.6 | <73.4U | 108 | 121 | 207* | 198* |
| Thallium | 0.91 | | 0.17 | <0.16U | <0.15U | <0.14U | <0.15U | <0.14U | <0.16U | 0.19 |
| Vanadium | | 15.3 | 23.2 | 25.9 | 22.9 | 13.4 | 20.4 | 11.2 | 22.8 | 19 |
| Zinc | 93.3 | 78.9 | 85.9 | 63.5 | 66.7 | 79.2 | 59 | 40.1 | 51.9 | 57.3 |
| 77'. | NT. | 0.5311 | T = 0.1 | | Anions (mg/kg) | 0.05511 | 1 | l vin | l vin | 3.10 |
| Nitrate | None | <0.73U | 5.8* | 1* | 4.5* | <0.075U | 15* | NR | NR | NR |
| Sulfate | None | <16.1U | <17.9U | <17.8U | <15.4U | 21* | 19.9* | NR | NR | NR |
| Sulfide | None | 15.1* | 43.9* | 48.4* | 18.7* | <8.6U | <8.5U | NR | NR | NR |
| TRU DRO | None | ND | ND | ND | Miscellaneous (mg/kg) | ND | ND | ND | ND | ND |
| TPH-DRO | None None | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR |
| TPH-GRO | INOHE | NIV | INK | | NK plosives/Propellants (mg | | JAK | INK | INK | NK |
| Nitrocellulose | None | NR | NR | NR | piosives/Propellants (mg. NR | 1.7* | 1.7* | NR | NR | NR |
| Minocentinose | None | INIX | INIX | INIX | SVOCs (mg/kg) | 1./ | 1./ | INIX | INIX | INIX |
| Benz(a)anthracene | None | NR | <0.42U | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| Benzo(a)pyrene | None | NR | 0.15J* | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| | None | NR | <0.42U | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| Chrysene | None | NR | <0.42U | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| Fluoranthene | None | NR | <0.42U | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| Phenanthrene | None | NR | <0.42U | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| Pyrene | None | NR | <0.42U | <0.42U | NR | <0.38U | <0.38U | NR | NR | NR |
| 1 /10110 | 110110 | 1110 | -0.720 | -0.720 | 1111 | -0.500 | -0.500 | 1111 | 1110 | 1117 |

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|----------------------|----------------------------|--------------------|--------------------|--------------------|-------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11cs-001 | LL11cs-002 | LL11cs-002 | LL11cs-003 | LL11cs-004 | LL11cs-005 | LL11cs-007 | LL11cs-008 | LL11cs-009 |
| Sample ID | | LL11cs-001-0001-SO | LL11cs-002-0001-FD | LL11cs-002-0001-SO | LL11cs-003-0001-SO | LL11cs-004-0001-SO | LL11cs-005-0001-SO | LL11cs-007-0001-SO | LL11cs-008-0001-SO | LL11cs-009-0001-SO |
| Date | | 01/24/01 | 01/26/01 | 01/26/01 | 01/26/01 | 02/21/01 | 02/21/01 | 03/21/01 | 03/21/01 | 03/21/01 |
| Depth (ft) | | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 5.0 - 6.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 |
| Parameters | | | | | | | | | | |
| Analyzed | | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | | | |
| Analyte | Background Criteria | Explosives | analytes | analytes | Explosives | analytes | analytes | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals, VOCs |
| | | | | | Pesticides/PCBs (mg/kg) | | | | | |
| PCB-1254 | None | NR | 0.16* | 0.023* | NR | <0.0097U | <0.0096U | NR | NR | NR |
| | | | | | VOCs (mg/kg) | | | | | |
| Acetone | None | NR | 0.019* | <0.006U | NR | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |
| Carbon tetrachloride | None | NR | <0.006U | <0.006U | NR | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |
| Toluene | None | NR | <0.006U | <0.006U | NR | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |

Table 4–3. Analytes Detected in IRA Confirmation Subsurface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|-------------------------------------|---------------------|--------------------|------------------|--------------------|--------------------------------|------------------|--------------------|------------|--------------------|------------|
| Station State | | LL11cs-012 | LL11cs-017 | LL11cs-017 | LL11cs-018 | LL11cs-019 | LL11cs-022 | LL11cs-023 | LL11cs-024 | LL11cs-026 |
| Sample ID | | LL11cs-012-0001-SO | | LL11cs-017-0001-SO | | | LL11cs-022-0001-SO | 1 | LL11cs-024-0001-SO | |
| Date | | 03/23/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/21/01 | 03/21/01 | 03/21/01 | 03/21/01 |
| Depth (ft) | | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 |
| Parameters | | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 |
| Analyzed | | | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | | | | |
| Analyte | Background Criteria | TAL Metals, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | TAL Metals | TAL Metals | TAL Metals | TAL Metals |
| | | , | | , | Metals (mg/kg) | | | | | |
| Aluminum | 19500 | 12100 | NR | NR | NR | NR | 6260 | 15700 | 13200 | 12900 |
| Antimony | 0.96 | <0.32U | NR | NR | NR | NR | <0.23U | <0.23U | <0.23U | 0.5 |
| Arsenic | 19.8 | 7.5 | NR | NR | NR | NR | 15.7 | 11.5 | 11.3 | 11.6 |
| Barium | 124 | 127* | NR | NR | NR | NR | 35.9 | 82.8 | 58.8 | 89.4 |
| Beryllium | 0.88 | | NR | NR | NR | NR | <0.32U | 0.99* | 0.47 | 1.2* |
| Cadmium | 0 | <0.15U | NR | NR | NR | NR | <0.16U | <0.19U | <0.19U | 0.24* |
| Calcium | 35500 | 1030 | NR | NR | NR | NR | 408 | 1020 | 745 | 12300 |
| Chromium | 27.2 | | NR | NR | NR | NR | 8.8 | 21.9 | 16 | 14.1 |
| Cobalt | 23.2 | | NR | NR | NR | NR | 6.4 | 13.9 | 6.9 | 6.8 |
| Copper | 32.3 | | NR | NR | NR | NR | 19.4 | 26 | 16.1 | 12.6 |
| Cyanide | 0 | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Iron | | 26900 | NR | NR | NR | NR | 18700 | 47600* | 20800 | 18500 |
| Lead | 19.1 | 9 | NR | NR | NR | NR | 17.3 | 10.8 | 7.8 | 26.2* |
| Magnesium | 8790 | | NR | NR | NR | NR | 1650 | 3780 | 2310 | 5430 |
| Manganese | 3030 | 673 | NR | NR | NR | NR | 592 | 1030 | 203 | 654 |
| Mercury | 0.044 | | NR | NR | NR | NR | <0.039U | <0.039U | <0.041U | <0.048U |
| Nickel | 60.7 | 36 | NR | NR | NR | NR | 13.9 | 27.9 | 15.7 | 13.4 |
| Potassium | 3350 | | NR | NR | NR | NR | 883 | 2700 | 1640 | 1430 |
| Selenium | 1.5 | | NR | NR | NR | NR | <0.4U | <0.47U | <0.48U | 0.9 |
| Sodium | 145 | 1010* | NR | NR | NR | NR | 122 | 245* | 155* | 241* |
| Thallium | 0.91 | | NR | NR | NR | NR | <0.16U | <0.15U | <0.16U | <0.19U |
| Vanadium | 37.6 | | NR | NR | NR | NR | 10.2 | 26.5 | 20.5 | 18.1 |
| Zinc | 93.3 | 62.9 | NR | NR | NR | NR | 54.8 | 67.5 | 51.8 | 65.8 |
| | | | 1 | | Anions (mg/kg) | ı | 1 | 1 | 1 | |
| Nitrate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Sulfate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Sulfide | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TDU DDO | NT. | NB | 1 3 ID | N.T.D. | Miscellaneous (mg/kg) | N.D. | l vin | 1.10 | 1 3 ID |) ID |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| NT' 11 1 | NT. | NID | ND | | plosives/Propellants (mg/ | |) ID | ND | ND | ND |
| Nitrocellulose | None | NR | NR | NR | NR SVOCa (ma/la) | NR | NR | NR | NR | NR |
| Dang(a)anthrasans | None | NR | <0.39U | <0.39U | <i>SVOCs (mg/kg)</i> <0.39U | 0.17J* | NR | NR | NR | NR |
| Benz(a)anthracene | None | NR NR | <0.39U <0.39U | <0.39U | <0.39U | 0.17J* 0.2J* | NR NR | NR NR | NR NR | NR NR |
| Benzo(a)pyrene Benzo(b)fluoranthene | None | NR NR | <0.39U | <0.39U | <0.39U | 0.2J* 0.22J* | NR NR | NR NR | NR NR | NR NR |
| | None | NR NR | <0.39U <0.39U | <0.39U | <0.39U | 0.22J* 0.21J* | NR NR | NR NR | NR NR | NR NR |
| Chrysene Fluoranthene | None | NR NR | <0.39U <0.39U | <0.39U | <0.39U | 0.21J* 0.33J* | NR NR | NR NR | NR NR | NR NR |
| Phenanthrene | None | NR NR | <0.39U <0.39U | <0.39U | <0.39U | 0.33J* 0.17J* | NR NR | NR NR | NR NR | NR NR |
| Prienantifrene Pyrene | None | NR NR | <0.39U | <0.39U | <0.39U | 0.1/J* 0.41J* | NR NR | NR NR | NR NR | NR NR |
| 1 yrene | TNOHE | INIX | \U.J7U | ~∪. J7U | \U.J7U | n.411 . | 111/ | 111/ | 111/ | IVIX |

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|----------------------|----------------------------|--------------------|--------------------|--------------------|-------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11cs-012 | LL11cs-017 | LL11cs-017 | LL11cs-018 | LL11cs-019 | LL11cs-022 | LL11cs-023 | LL11cs-024 | LL11cs-026 |
| Sample ID | | LL11cs-012-0001-SO | LL11cs-017-0001-FD | LL11cs-017-0001-SO | LL11cs-018-0001-SO | LL11cs-019-0001-SO | LL11cs-022-0001-SO | LL11cs-023-0001-SO | LL11cs-024-0001-SO | LL11cs-026-0001-SD |
| Date | | 03/23/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/23/01 | 03/21/01 | 03/21/01 | 03/21/01 | 03/21/01 |
| Depth (ft) | | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 |
| Parameters | | | | | | | | | | |
| Analyzed | | | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | | | | |
| Analyte | Background Criteria | TAL Metals, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | TAL Metals | TAL Metals | TAL Metals | TAL Metals |
| | | | | | Pesticides/PCBs (mg/kg) |) | | | | |
| PCB-1254 | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | VOCs (mg/kg) | | | | | |
| Acetone | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.007U | NR | NR | NR | NR |
| Carbon tetrachloride | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.007U | NR | NR | NR | NR |
| Toluene | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.007U | NR | NR | NR | NR |

Table 4–3. Analytes Detected in IRA Confirmation Subsurface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | NPA | NPA | NPA | NPA |
|----------------------|---------------------|--------------------|--------------------|--------------------|---------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11cs-027 | LL11cs-028 | LL11cs-029 | LL11cs-032 | LL11cs-033 | LL11cs-013 | LL11cs-014 | LL11cs-034 | LL11cs-035 |
| Sample ID | | LL11cs-027-0001-SO | LL11cs-028-0001-SO | LL11cs-029-0001-SO | LL11cs-032-0001-SO | LL11cs-033-0001-SO | LL11cs-013-0001-SO | LL11cs-014-0001-SO | LL11cs-034-0001-SO | LL11cs-035-0001-SO |
| Date | | 03/21/01 | 03/21/01 | 03/21/01 | 03/20/01 | 03/20/01 | 03/23/01 | 03/23/01 | 03/20/01 | 03/20/01 |
| Depth (ft) | | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 |
| Parameters | | | | | | | | | | |
| Analyzed | | | | | | | | | | |
| Analyte | Background Criteria | TAL Metals | TAL Metals | TAL Metals | TAL Metals | TAL Metals | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals | TAL Metals |
| | | | | | Metals (mg/kg) | | | | T | |
| Aluminum | 19500 | | 14500 | 10500 | 9150 | 13900 | 24500* | 13800 | 13400 | 12200 |
| Antimony | 0.96 | | 0.51 | 0.26 | <0.29U | <0.28U | <0.35U | <0.32U | <0.28U | <0.28U |
| Arsenic | 19.8 | | 15.1 | 15 | <3.8U | <1.9U | <2.3U | 6 | <1.9U | <3.8U |
| Barium | 124 | | 95.3 | 68.7 | 72 | 52.8 | 167* | 71 | 87.8 | 41.6 |
| Beryllium | 0.88 | | 0.7 | 0.53 | 0.51 | 0.5 | 1.3* | 0.8 | 0.73 | 0.69 |
| Cadmium | | <0.2U | <0.16U | <0.18U | <0.19U | <0.18U | <0.19U | 0.37* | 0.38* | 0.22* |
| Calcium | 35500 | | 1440 | 2680 | 392 | 1240 | 2880 | 3200 | 1970 | 930 |
| Chromium | 27.2 | | 18.5 | 14.8 | 10.5 | 19.5 | 24.3 | 17.1 | 15.8 | 16.5 |
| Cobalt | 23.2 | | 12.1 | 9.5 | 12.6 | 5.5 | 3.3 | 11.7 | 10 | 8.9 |
| Copper | 32.3 | | 20.5 | 21.8 | 19.4 | 18.6 | 25.2 | 21.6 | 13.1 | 25.6 |
| Cyanide | | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Iron | 35200 | | 31200 | 25100 | 22300 | 23100 | 9560 | 20300 | 25800 | 29800 |
| Lead | 19.1 8790 | | 9.2 | 13.2 | <1.9U | <1.9U | 12.3 | 26.4* | <1.9U | <1.9U |
| Magnesium | 3030 | | 2850 | 3770 508 | 1920 488 | 2950 | 2380 | 2890 | 2380 861 | 2940 242 |
| Manganese | | | 1140 | | | 102 | | 201 | | |
| Mercury Nickel | 60.7 | <0.04U | <0.04U 23.8 | <0.039U 27.1 | <0.043U 15.7 | <0.041U 17.2 | <0.043U 12.2 | 0.054* 24.7 | <0.044U 17.1 | <0.039U 21.4 |
| Potassium | 3350 | | 2260 | 1600 | 872 | 1330 | 2600 | 1680 | 1480 | 1470 |
| Selenium | | <0.49U | <0.4U | <0.45U | 0.6 | <0.44U | <0.48U | 1.1 | 0.85 | <0.43U |
| Sodium | | 172* | 182* | 193* | 623* | 928* | 951* | 831* | 923* | 1050* |
| Thallium | | <0.16U | <0.16U | <0.15U | <0.19U | <0.19U | <0.23U | 0.22 | <0.19U | <0.19U |
| Vanadium | 37.6 | | 24.9 | 16.9 | 16.3 | 22.6 | 30.2 | 22.7 | 25.6 | 20.3 |
| Zinc | 93.3 | | 62.3 | 59.5 | 47.2 | 57.3 | 77.8 | 201* | 147* | 61.8 |
| Zinc | 75.5 | 37.7 | 02.3 | 37.3 | Anions (mg/kg) | 37.3 | 77.0 | 201 | 177 | 01.0 |
| Nitrate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Sulfate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Sulfide | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | • | | Miscellaneous (mg/kg) | • | • | • | • | • |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | Ex | plosives/Propellants (mg/ | /kg) | | | | |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Benzo(a)pyrene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Benzo(b)fluoranthene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Chrysene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Fluoranthene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Phenanthrene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Pyrene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |

| Aggregate | | FPA | FPA | FPA | FPA | FPA | NPA | NPA | NPA | NPA |
|----------------------|----------------------------|--------------------|--------------------|--------------------|-------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11cs-027 | LL11cs-028 | LL11cs-029 | LL11cs-032 | LL11cs-033 | LL11cs-013 | LL11cs-014 | LL11cs-034 | LL11cs-035 |
| Sample ID | | LL11cs-027-0001-SO | LL11cs-028-0001-SO | LL11cs-029-0001-SO | LL11cs-032-0001-SO | LL11cs-033-0001-SO | LL11cs-013-0001-SO | LL11cs-014-0001-SO | LL11cs-034-0001-SO | LL11cs-035-0001-SO |
| Date | | 03/21/01 | 03/21/01 | 03/21/01 | 03/20/01 | 03/20/01 | 03/23/01 | 03/23/01 | 03/20/01 | 03/20/01 |
| Depth (ft) | | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 |
| Parameters | | | | | | | | | | |
| Analyzed | | | | | | | | | | |
| Analyte | Background Criteria | TAL Metals | TAL Metals | TAL Metals | TAL Metals | TAL Metals | TAL Metals, VOCs | TAL Metals, VOCs | TAL Metals | TAL Metals |
| | | | | | Pesticides/PCBs (mg/kg) |) | | | | |
| PCB-1254 | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | VOCs (mg/kg) | | | | | |
| Acetone | None | NR | NR | NR | NR | NR | <0.007U | <0.008U | NR | NR |
| Carbon tetrachloride | None | NR | NR | NR | NR | NR | <0.007U | <0.008U | NR | NR |
| Toluene | None | NR | NR | NR | NR | NR | <0.007U | <0.008U | NR | NR |

Table 4–3. Analytes Detected in IRA Confirmation Subsurface Soil Samples (continued)

| Aggregate Station | | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|--------------------|---------------------|--------------------|----------------------|--------------------|---------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| | | NPA LL11cs-036 | NPA LL11cs-037 | LL11cs-038 | LL11cs-040 | LL11cs-040 | LL11cs-041 | LL11cs-042 | LL11cs-043 | LL11cs-044 |
| Sample ID | | LL11cs-036-0001-SO | | LL11cs-038-0001-SO | | LL11cs-040-0001-SO | | | LL11cs-043-0001-SO | |
| Date | _ | 03/20/01 | 03/20/01 | 03/20/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 |
| Depth (ft) | | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 3.0 - 7.0 | 3.0 - 7.0 | 3.0 - 7.0 | 3.0 - 7.0 | 3.0 - 7.0 | 7.0 - 8.0 |
| Parameters | | 1.0 2.0 | 1.0 2.0 | 1.0 2.0 | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyzed | | | | | Misc., Explosives, | Misc., Explosives, | Misc., Explosives, | Misc., Explosives, | Misc., Explosives, | Misc., Explosives, |
| | | | | | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, |
| Analyte B | Background Criteria | TAL Metals | TAL Metals | TAL Metals | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs |
| | | | | | Metals (mg/kg) | | | | | |
| Aluminum | 19500 | | 8960 | 6670 | 11400 | 12100 | 8170 | 9500 | 9380 | 4900 |
| Antimony | 0.96 | <0.29U | <0.25U | <0.27U | <0.27U | <0.28U | <0.27U | <0.26U | <0.27U | <0.28U |
| Arsenic | 19.8 | <1.9U | <1.7U | <1.8U | 19.7 | 16.2 | 17.3 | 16.1 | 15.6 | 13.6 |
| Barium | 124 | | 33.7 | 27.1 | 30.9 | 32.2 | 26.7 | 33.3 | 30.6 | 28.2 |
| Beryllium | | 0.58 | 0.47 | <0.36U | 0.41 | 0.42 | 0.34 | 0.38 | 0.38 | <0.37U |
| Cadmium | | 0.31* | <0.18U | <0.18U | <0.18U | <0.2U | <0.17U | <0.18U | <0.17U | <0.18U |
| Calcium | | 823 | 508 | 620 | 688 | 762 | 216 | 489 | 435 | 162 |
| Chromium | | 13.3 | 12.5 | 8.5 | 12.1 | 12.6 | 9.3 | 12 | 11.5 | 5.9 |
| Cobalt | 23.2 | | 6.7 | 3.4 | 7.2 | 7.8 | 6.7 | 7 | 6.2 | 4.9 |
| Copper | 32.3 | | 18.4 | 6 | 20.5 | 22.4 | 18.4 | 18 | 17.7 | 14.4 |
| Cyanide | | NR | NR | NR | 0.018* | 0.015* | <0.0053U | <0.0044U | <0.0052U | <0.0052U |
| Iron | | 24300 | 22400 | 14800 | 22400 | 22700 | 17800 | 19900 | 18600 | 13700 |
| Lead | | <1.9U | <1.7U | <1.8U | 11.4 | 15.4 | 9.3 | 9.5 | 9.9 | 8.2 |
| Magnesium | | 2730 | 2190 | 1800 | 2020 | 2120 | 1580 | 2220 | 1930 | 1060 |
| Manganese | | 1930 | 218 | 165 | 304 | 341 | 324 | 274 | 222 | 228 |
| Mercury | 0.044 | <0.038U | <0.039U | <0.04U | <0.04U | <0.04U | <0.04U | <0.041U | <0.04U | <0.041U |
| Nickel | | 33.5 | 16.9 | 10.6 | 14.2 | 15 942 | 12.2 | 15.8 | 13.5 | 10.6 |
| Potassium | | 1090 <0.43U | 1020 | 688 0.78 | 831 <0.46U | <0.51U | 873 | 1010 <0.45U | 1080 <0.43U | 691 <0.46U |
| Selenium Sodium | | 807* | 0.63 724 * | 608* | <18.2U | <0.310 <20.4U | <0.42U <16.7U | <0.43U <18.1U | <0.43U <17.2U | <0.46U <18.3U |
| Thallium | | <0.19U | <0.17U | <0.18U | <0.18U | <0.19U | <0.18U | <0.17U | <0.18U | <0.18U |
| Vanadium | | 14.8 | 15.1 | 13.5 | 19.4 | 20.6 | 13.6 | 14.8 | 15.1 | 9.1 |
| Zinc | 93.3 | | 50.9 | 29 | 48.7 | 51.9 | 48.3 | 53.8 | 49.7 | 48.7 |
| Zinc | 73.3 | 00.0 | 30.9 | 2) | Anions (mg/kg) | 31.9 | 1 40.3 | 33.6 | 77.1 | 70.7 |
| Nitrate N | None | NR | NR | NR | 0.44* | 0.88* | 0.72* | 0.48* | 0.4* | 0.62* |
| | None | NR | NR | NR | 27.8* | 23.1* | 40* | 19.8* | 18.1* | 29.9* |
| | | NR | NR | NR | 16.3* | <8.9U | <8.9U | <9U | <9.3U | <8.9U |
| | . (6110 | 1,11 | 1.11 | L | Miscellaneous (mg/kg) | 0.5 0 | 0.50 | , , , | 7.00 | 0.5 0 |
| TPH-DRO N | None | NR | NR | NR | <7.2U | <7.4U | <7.2U | <7.6U | 34* | <7.4U |
| | | NR | NR | NR | <0.061U | <0.061U | <0.06U | <0.062U | 0.054J* | <0.061U |
| | - | | | | olosives/Propellants (mg/ | /kg) | • | | | |
| Nitrocellulose N | None | NR | NR | NR | 0.46* | 1.5* | 0.77* | 0.86* | 0.83* | 0.97* |
| | | | | | SVOCs (mg/kg) | | | | | |
| | | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | <0.4U | <0.4U |
| 710 | | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | 0.12J* | <0.4U |
| | | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | <0.4U | <0.4U |
| | None | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | <0.4U | <0.4U |
| | None | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | 0.16J* | <0.4U |
| | | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | <0.4U | <0.4U |
| Pyrene N | None | NR | NR | NR | <0.39U | <0.39U | <0.39U | <0.4U | <0.4U | <0.4U |

| Aggregate | | NPA |
|-------------------------|----------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11cs-036 | LL11cs-037 | LL11cs-038 | LL11cs-040 | LL11cs-040 | LL11cs-041 | LL11cs-042 | LL11cs-043 | LL11cs-044 |
| Sample ID | | LL11cs-036-0001-SO | LL11cs-037-0001-SO | LL11cs-038-0001-SO | LL11cs-040-0001-FD | LL11cs-040-0001-SO | LL11cs-041-0001-SO | LL11cs-042-0001-SO | LL11cs-043-0001-SO | LL11cs-044-0001-SO |
| Date | | 03/20/01 | 03/20/01 | 03/20/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 |
| Depth (ft) | | 1.0 - 2.0 | 1.0 - 2.0 | 1.0 - 2.0 | 3.0 - 7.0 | 3.0 - 7.0 | 3.0 - 7.0 | 3.0 - 7.0 | 3.0 - 7.0 | 7.0 - 8.0 |
| Parameters | | | | | Misc., TAL Metals, |
| Analyzed | | | | | Misc., Explosives, |
| | | | | | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, |
| Analyte | Background Criteria | TAL Metals | TAL Metals | TAL Metals | SVOCs, VOCs |
| Pesticides/PCBs (mg/kg) | | | | | | | | | | |
| PCB-1254 | None | NR | NR | NR | <0.01U | <0.01U | <0.0099U | <0.01U | <0.01U | <0.01U |
| VOCs (mg/kg) | | | | | | | | | | |
| Acetone | None | NR | NR | NR | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |
| Carbon tetrachloride | None | NR | NR | NR | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |
| Toluene | None | NR | NR | NR | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |

Table 4–3. Analytes Detected in IRA Confirmation Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | | | |
|-----------------------|---------------------|------------------------------|---------------------------|----------------------------------|---------------------------|---------------------------|---------------------------|---------------------------|--|--|--|
| Aggregate Station | | LL11cs-045 | LL11cs-045 | LL11cs-046 | LL11cs-047 | LL11cs-048 | LL11cs-049 | LL11cs-050 | | | |
| Sample ID | | LL11cs-045-0001-FD | LL11cs-045-0001-SO | LL11cs-046 LL11cs-046-0001-SO | LL11cs-047-0001-SO | LL11cs-048-0001-SO | LL11cs-049-0001-SO | LL11cs-050-0001-SO | | | |
| Date | | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | | | |
| Depth (ft) | | 7.0 - 8.0 | 7.0 - 8.0 | 3.0 - 4.0 | 3.0 - 4.0 | 3.0 - 4.0 | 3.0 - 4.0 | 3.0 - 4.0 | | | |
| Parameters Analyzed | _ | 7.0 - 8.0 | Misc., TAL Metals, Misc., | Misc., TAL Metals, Misc., | Misc., TAL Metals, Misc., | Misc., TAL Metals, Misc., | Misc., TAL Metals, Misc., | 3.0 - 4.0 | | | |
| 1 at ameters Analyzeu | _ | Misc., TAL Metals, Misc., | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Misc., TAL Metals, Misc., | | | |
| | | Explosives, Pesticides/PCBs, | Pesticides/PCBs, SVOCs, | Pesticides/PCBs, SVOCs, | Pesticides/PCBs, SVOCs, | Pesticides/PCBs, SVOCs, | Pesticides/PCBs, SVOCs, | Explosives, | | | |
| Analyte | Background Criteria | SVOCs, VOCs | VOCs | VOCs | VOCs | VOCs | VOCs | Pesticides/PCBs, SVOCs | | | |
| Metals (mg/kg) | | | | | | | | | | | |
| Aluminum | 19500 | 7620 | 7030 | 11500 | 9980 | 10500 | 12600 | 13200 | | | |
| Antimony | 0.96 | <0.25U | <0.28U | <0.26U | <0.24U | <0.26U | <0.26U | 0.39 | | | |
| Arsenic | 19.8 | 36.4* | 44.1* | 9.6 | 13.9 | 13.6 | 10.4 | 33.6* | | | |
| Barium | 124 | 83.7 | 82.3 | 80 | 36.2 | 40.6 | 57.4 | 4190* | | | |
| Beryllium | 0.88 | 0.39 | <0.38U | 0.61 | 0.37 | 0.35 | 0.44 | 3.2* | | | |
| Cadmium | 0 | <0.19U | <0.19U | <0.19U | <0.16U | <0.17U | <0.17U | <0.25U | | | |
| Calcium | 35500 | | 246 | 1500 | 682 | 655 | 983 | 4730 | | | |
| Chromium | 27.2 | | 10.1 | 12.3 | 10.6 | 11.4 | 14.7 | 14.1 | | | |
| Cobalt | 23.2 | | 17.4 | 8.1 | 6.8 | 7.5 | 8.4 | 5.5 | | | |
| Copper | 32.3 | 20.2 | 18.5 | 9.2 | 11.4 | 10.8 | 9.6 | 62.3* | | | |
| Cyanide | 0 | <0.0045U | <0.0044U | 0.1* | 0.0097* | 0.046* | 0.051* | 0.25* | | | |
| Iron | | I. | 27500 | 16800 | 17400 | 16200 | 18900 | 54800* | | | |
| Lead | 19.1 | 17.9 | 9.5 | 17.7 | 10 | 8.2 | 17.3 | 160* | | | |
| Magnesium | 8790 | 1970 | 1970 | 1840 | 1890 | 1970 | 2440 | 786 | | | |
| Manganese | 3030 | 1420 | 1630 | 1040 | 411 | 394 | 598 | 173 | | | |
| Mercury | 0.044 | <0.041U | <0.04U | 0.047* | <0.037U | <0.039U | <0.04U | <0.045U | | | |
| Nickel | 60.7 | 23.5 | 25.1 | 13.5 | 12.4 | 11.3 | 13.5 | 10.9 | | | |
| Potassium | 3350 | 1190 | 1060 | 896 | 834 | 1050 | 1050 | 1130 | | | |
| Selenium | 1.5 | <0.48U | <0.47U | 0.56 | <0.41U | <0.43U | 0.68 | <0.61U | | | |
| Sodium | 145 | <19.2U | <18.9U | <19.2U | <16.3U | <17.1U | 53.1 | 1530* | | | |
| Thallium | 0.91 | <0.16U | <0.18U | 0.2 | <0.16U | <0.17U | <0.17U | 2.5* | | | |
| Vanadium | 37.6 | | 9.7 | 20.3 | 18.4 | 18.7 | 24.8 | 21.8 | | | |
| Zinc | 93.3 | 64./ | 66.6 | 49.3 | 37.8 | 37.5 | 46.7 | 390* | | | |
| Nitrate | None | 0.47* | 0.55* | Anions (mg/kg) 0.79* | 0.61* | 0.71* | 0.83* | 2.9* | | | |
| Sulfate | None | 27* | 35* | 13.6* | 19.5* | 44.9* | 25.7* | 29.7* | | | |
| Sulfide | None | 14.3* | <9.1U | 12.2* | <8.7U | <9U | <8.4U | <9U | | | |
| Bullide | TAOHC | 17.0 | \J.10 | Miscellaneous (mg/kg | | | \0.7U | 1 -90 | | | |
| TPH-DRO | None | <7.7U | <7.3U | <7.6U | <6.9U | <7.4U | <7.3U | <8.5U | | | |
| TPH-GRO | None | <0.062U | <0.061U | <0.061U | 0.026J* | <0.06U | <0.06U | NR | | | |
| 1111 0110 | 110110 | 1 0.0020 | 0.0010 | Explosives/Propellants (m | | 3.000 | 3.000 | | | | |
| Nitrocellulose | None | 0.9* | 0.89* | 0.84* | 0.93* | 0.87* | 0.9* | NR | | | |
| | | 1 | 1 | SVOCs (mg/kg) | 1 | 1 | | · · · | | | |
| Benz(a)anthracene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| Benzo(a)pyrene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| Benzo(b)fluoranthene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| Chrysene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| Fluoranthene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| Phenanthrene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| Pyrene | None | <0.41U | <0.4U | <0.4U | <0.37U | <0.38U | <0.4U | <0.45U | | | |
| * | 1 | | | 1 | | 1 | | | | | |

Table 4–3. Analytes Detected in IRA Confirmation Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|----------------------|---------------------|------------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|
| Station | | LL11cs-045 | LL11cs-045 | LL11cs-046 | LL11cs-047 | LL11cs-048 | LL11cs-049 | LL11cs-050 |
| Sample ID | | LL11cs-045-0001-FD | LL11cs-045-0001-SO | LL11cs-046-0001-SO | LL11cs-047-0001-SO | LL11cs-048-0001-SO | LL11cs-049-0001-SO | LL11cs-050-0001-SO |
| Date | | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 | 03/22/01 |
| Depth (ft) | | 7.0 - 8.0 | 7.0 - 8.0 | 3.0 - 4.0 | 3.0 - 4.0 | 3.0 - 4.0 | 3.0 - 4.0 | 3.0 - 4.0 |
| Parameters Analyzed | | | Misc., TAL Metals, Misc., | |
| | | Misc., TAL Metals, Misc., | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Misc., TAL Metals, Misc., |
| | | Explosives, Pesticides/PCBs, | Pesticides/PCBs, SVOCs, | Explosives, |
| Analyte | Background Criteria | SVOCs, VOCs | VOCs | VOCs | VOCs | VOCs | VOCs | Pesticides/PCBs, SVOCs |
| | | | | Pesticides/PCBs (mg/kg | g) | | | |
| PCB-1254 | None | <0.01U | <0.01U | <0.0099U | <0.0093U | <0.0098U | <0.01U | <0.011U |
| | | | | VOCs (mg/kg) | | | | |
| Acetone | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | NR |
| Carbon tetrachloride | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | NR |
| Toluene | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | NR |

FPA = Former production area.

ft = Feet.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.

NPA= Non-production area.
NR = Not reported/not analyzed.
PCB = Polychlorinated biphenyl.
RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

TAL = Target analyte list.

TPH-GRO/DRO = Total petroleum hydrocarbons-gasoline range organics/diesel range organics.

U = Not detected.

VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

Table 4-4. Analytes Detected in IRA Confirmation Sediment Samples

| Aggregate | | East Ditch EU | East Ditch EU | East Ditch EU |
|----------------------------|----------------------------|----------------------|--------------------|--------------------|
| Station | | LL11cs-031 | LL11cs-031 | LL11cs-031 |
| Sample ID | | LL11cs-031-0001-FD | LL11cs-031-0001-SD | LL11cs-031-0001-FD |
| Date | | 03/20/01 | 03/20/01 | 03/20/01 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | | | | |
| Analyte | Background Criteria | TAL Metals | TAL Metals | TAL Metals |
| | | Metals (mg/kg) | | |
| Aluminum | 13900 | 21000* | 10600 | 17700* |
| Antimony | 0 | <0.37U | <0.36U | 0.41* |
| Arsenic | 19.5 | <2.5U | <2.4U | 11.6 |
| Barium | 123 | 87.5 | 85.2 | 86.3 |
| Beryllium | 0.38 | 0.71* | 0.61* | 0.82* |
| Cadmium | 0 | 0.29* | 0.47* | <0.17U |
| Calcium | 5510 | 1980 | 2390 | 2170 |
| Chromium | 18.1 | 27.1* | 16.5 | 20.5* |
| Cobalt | 9.1 | 8.7 | 8.2 | 7.9 |
| Copper | 27.6 | 20.6 | 19.7 | 26.3 |
| Iron | 28200 | 33100* | 19200 | 30000* |
| Lead | 27.4 | <9.8U | <9.6U | 14.7 |
| Magnesium | 2760 | 3880* | 2140 | 3080* |
| Manganese | 1950 | 882 | 754 | 193 |
| Nickel | 17.7 | 22.3* | 15.2 | 19.8* |
| Potassium | 1950 | 2000* | 856 | 1980* |
| Selenium | 1.7 | 0.64 | 2* | 0.5 |
| Sodium | 112 | 1180* | 817* | 938* |
| Thallium | 0.89 | <0.25U | <0.24U | 0.23 |
| Vanadium | 26.1 | 43.9* | 24.2 | 26.1 |
| Zinc | 532 | 81 | 234 | 65.3 |
| | | Anions | | |
| Sulfate | None | NR | NR | NR |
| Sulfide | None | NR | NR | NR |
| | | Explosives (mg/kg | | |
| Nitrocellulose | None | NR | NR | NR |
| | | Pesticides/PCBs (mg/ | - 07 | |
| PCB-1254 | None | NR | NR | NR |
| | | VOCs (mg/kg) | | |
| Acetone | None | NR | NR | <0.006U |
| Acetone EU = Exposure unit | None | NR | NR | <0.006U |

EU = Exposure unit. ft = Feet.

tt = Feet.

ID = Identification.

mg/kg = Milligrams per kilogram.

NR = Not reported/not analyzed.

PCB = Polychlorinated biphenyl.

TAL = Target analyte list.

VOC = Volatile organic compound.

U = Not detected.

* = Result exceeds background criteria or no background criteria was available.

Table 4–5. Phase I RI Sampling Locations

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|------------------------------|---|---|--|------------------------------|---------------------------------------|
| LL11sb-001 | 0–1 | Full suite, anions | Southern portion of NPA, northeast of Entrance Gate | Characterize an area not previously sampled. Installed monitoring well LL11mw-002. QC sample collected. | None | Metals, explosives |
| LL11ss-001 | 0–1 1–3 | Full suite, anions Metals, explosives, anions | Building AP-18 | 1941-1945, 1951-1957, and 1969-1971: Utilized for storage of percussion elements. QC sample collected. | None | Metals, explosives |
| LL11ss-002 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Building AP-18 | 1941-1945, 1951-1957, and 1969-1971: Utilized for storing percussion elements. | None | Metals, explosives |
| LL11sb-002 | 0–1 6–8 | Metals, explosives, anions Metals, explosives, anions | Eastern end of NPA | Characterize an area not previously sampled, downgradient of Building AP-18 (1941-1945, 1951-1957, and 1969-1971: Utilized for storing percussion elements). Installed monitoring well LL11mw-002. | None | Metals, explosives |
| LL11sb-003 | 0–1 | Metals, explosives, anions | East of Building AP-8 | Characterize an area not previously sampled, downgradient of Building AP-8 (1941-1945, 1951-1957, and 1969-1971: Utilized for storing percussion elements). Installed monitoring well LL11mw-003. | None | Metals, black powder |
| LL11ss-003 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Buildings AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. | None | Metals, black powder |
| LL11ss-004 | 0_1 Metals explosives anions | | Buildings AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. QC sample (0-1) collected. | None | Metals, black powder |
| | 0–1 | Metals, explosives, anions | | Characterize an area not previously sampled, downgradient of Building AP-11 (1941-1945, 1951-1957: Utilized | | |
| LL11sb-004 | 10-12 | Metals, explosives, anions | Southeast of Building AP-11 | for black powder charging operations. 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials). Installed monitoring well LL11mw-004. | None | Metals, black powder, lacquers |
| LL11ss-005 | 0-1 | Metals, explosives, anions | Buildings AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. Subfloor sample collected inside AP-8. | None | Metals, black powder |
| | 0–1 | Full suite, anions | | Characterized area downgradient of petroleum-contaminated soil hot spot. Installed monitoring well LL11mw- | Petroleum- | |
| LL11sb-005 | 6–8 | Metals, explosives, anions | Downgradient of Hot Spot | 005. | contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11ss-006 | 0–1 | Metals, explosives, anions | Buildings AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. Sample collected inside AP-8. | None | Metals, black powder |
| I I 11-1- 006 | 0–1 | Metals, explosives, anions | North of Duilding AD 11 | Across the road from Building AP-11 near drainage ditch. 1941-1945, 1951-1957: Utilized for black powder | Nama | Metals, black powder, |
| LL11sb-006 | 2–4 | Metals, explosives, anions | North of Building AP-11 | charging operations. 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. Installed monitoring well LL11mw-006. | None | lacquers |
| LL11sb-007 | 0–1 | Metals, explosives, anions | Adjacent to Northern Ditch | Drainage ditch leading north prior to exiting the AOC. Installed monitoring well LL11mw-007. | None | Metals, explosives |
| LL11ss-007 | 1–3 | Metals, explosives, anions | Buildings AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. | None | Metals, black powder |
| LL11ss-008 | 0-1 | Metals, explosives, anions | Buildings AP-8 | 1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an | None | Metals, black powder |
| | 1–3 | Metals, explosives, anions | | administrative building. | | , 1 |
| LL11sb-008 | 0–1 12-13 | Metals, explosives, anions Metals, explosives, anions | Central area of FPA | Characterize an area not previously sampled. Installed monitoring well LL11mw-008. QC sample (12-13) collected. | None | Metals, explosives |
| T. 1.1 | 0–1 | Full suite, anions | D 1111 1 D 15 | | 27 | G 1 |
| LL11ss-009 | 1–3 | Metals, explosives, anions | Building AP-17 | 1941-1945, 1951-1957, and 1969-1971: Utilized as a solvent storage facility. | None | Solvents |
| LL11sb-009 | 0–1 | Metals, explosives, anions | Adjacent to duringes dital west of | Characterized area adjacent to drainage ditch west of Building AP-8 not previously sampled. 1941-1945, 1951- | | |
| | 4–6 | Metals, explosives, anions | Adjacent to drainage ditch west of Building AP-8 | 1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building. Installed monitoring well LL11mw-009. | None | Metals, black powder |
| LL11sb-010 | 0–1 | Metals, explosives, anions | Southwest of Building AP-4 | Characterize an area downgradient of Building AP-4 not previously sampled. 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. Installed monitoring well LL11mw-008. | None | Metals, black powder |
| LL11ss-010 | 0–1 | Metals, explosives, anions | Building AP-17 | 1941-1945, 1951-1957, and 1969-1971: Utilized as a solvent storage facility. QC sample (0-1) collected. | None | Solvents |
| | 1–3 | Full suite, anions Building AP-17 | | | 110110 | Sorveito |
| LL11ss-011 | 0–1 | | | 1941-1945, 1951-1957: Utilized for black powder charging operations. 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. Subfloor sample collected inside building (east end). | None | Metals, black powder, lacquers |
| LL11sb-011 | 4–6 | Metals, explosives, anions | Sewer Line in Western End of FPA | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |

Table 4–5. Phase I RI Sampling Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|---|---|---|--------------------|---------------------------------------|
| LL11ss-012 | 0-1 | Full suite, anions | Building AP-11 | 1941-1945, 1951-1957: Utilized for black powder charging operations. 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. Subfloor sample collected inside building (west end). | None | Metals, black powder, lacquers |
| LL11sb-012 | 8-10 | Full suite, anions | Sewer Line in Center of FPA | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. QC sample collected. | None | Metals, explosives |
| LL11ss-013 | 0–1 | Metals, explosives, anions | | 1941-1945, 1951-1957: Utilized for black powder charging operations. | | Metals, black powder, |
| | 1–3 | Full suite, cyanide | Building AP-11 | 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. | None | lacquers |
| LL11sb-013 | 8-10 | Metals, explosives, anions | Sewer Line southwest of Building AP-7 | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |
| LL11ss-014 | 0–1 | Metals, explosives, anions | | 1941-1945, 1951-1957: Utilized for black powder charging operations. | | Metals, black powder, |
| | 1–3 | Metals, explosives, anions | Building AP-11 | 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. | None | lacquers |
| LL11sb-014 | 11-13 | Metals, explosives, anions | Sewer Line along entrance road | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |
| LL11ss-015 | 0–1 | Metals, explosives, anions | | 1941-1945, 1951-1957: Utilized for black powder charging operations. | | Metals, black powder, |
| | 1–3 | Metals, explosives, anions | Building AP-11 | 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. | None | lacquers |
| LL11sb-015 | 10-12 | Metals, explosives, anions | Sewer Line northeast of Building AP- | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |
| LL11ss-016 | 0–1 | Metals, explosives, anions | | 1941-1945, 1951-1957: Utilized for black powder charging operations. | | Metals, black powder, |
| | 1–3 | Full suite, anions | Building AP-11 | 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. QC sample (1-3 ft) collected. | None | lacquers |
| LL11sb-016 | 8.5-11.5 | Metals, explosives, anions | Sewer line southeast of Building AP-8 | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |
| LL11ss-017 | 0–1 | Metals, explosives, anions | | 1941-1945, 1951-1957: Utilized for black powder charging operations. | | Metals, black powder, |
| | 1–3 | Metals, explosives, anions | Northwest of Building AP-11 | 1969-1971: Served as the major assembly and shipping building for MK2A4 primers, and also used to apply lacquer sealing materials. | None | lacquers |
| LL11sb-017 | 4–6 | Metals, explosives, anions | Sewer Line south of Building AP-11 | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |
| LL11ss-018 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Building AP-10 | 1941-1945, 1951-1957, and 1969-1971: Utilized as percussion element service magazine prior to primer charging. | None | Metals, explosives |
| LL11sb-018 | 4–6 | Metals, explosives, anions | Sewer line west of Building AP-13 | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | None | Metals, explosives |
| LL11sb-019 | 0–1 | Full suite, anions | | 1941-1945, 1951-1957, and 1969-1971: Utilized for black powder screening. Sump located on the southern side | | |
| | 6–8 | Metals, explosives, anions | Sump near Building AP-3 | of AP-3. Evaluated sump integrity by collecting two samples, zero to two ft interval, and from the six to eight ft interval, which straddled the bottom elevation of the sump. QC sample (0-1 ft) collected. | None | Metals, black powder |
| LL11ss-019 | 0–1 | Metals, explosives, anions | Building AP-20 | 1941-1945, 1951-1957, and 1969-1971: Served as a quality assurance (QA) primer sensitivity testing facility. | None | Metals, black powder |
| TT 11 000 | 1–3 | Metals, explosives, anions | 25.10.119711 20 | 25.12.15.15, 25.12.15.15, and 15.05.15.11. Self-ed as a quality assurance (Q11) printer sensitivity testing mentity. | 1.510 | include, order powder |
| LL11ss-020 | 0–1 1–3 | Full suite, anions Metals, explosives, anions | Building AP-1 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder service magazine for primer charging. | None | Metals, black powder |
| LL11sb-020 | 0–1 | Metals, explosives, anions | Sump near Building AP-5 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder dry house. One sump was located between AP-5 | None | Metals, black powder |

Table 4–5. Phase I RI Sampling Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|--|---|--|---|---------------------------------------|
| | 6–8 | Metals, explosives, anions | g | and AP-6. Evaluated sump integrity by collecting two samples, zero to two ft interval, and from the six to eight ft interval, which straddled the bottom elevation of the sump. | | |
| LL11ss-021 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Building AP-3 | 1941-1945, 1951-1957, and 1969-1971: Utilized for black powder screening. | None | Metals, black powder |
| LL11sb-021 | 0–1 6–8 | Metals, explosives, anions Metals, explosives, anions | Sump near Building AP-6 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder dry house. Two sets of sumps were associated with Building AP-5 and AP-6: one set was located outside the southwest wall of AP-6 and one set was located between AP-5 and AP-6. Evaluated sump integrity by collecting two samples, zero to two ft interval, and from the six to eight ft interval, which straddled the bottom elevation of the sump. | None | Metals, black powder |
| LL11ss-022 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Building AP-3 | 1941-1945, 1951-1957, and 1969-1971: Utilized for black powder screening. | None | Metals, black powder |
| LL11sb-022 | 0–1 6–8 | Metals, explosives, anions Metals, explosives, anions | Sump near Building AP-8 | 1941-1945, 1951-1957, and 1969-1971: Utilized for artillery primer loading and served as an administrative building. Two sets of sumps, one on the east side and one on the west side, were associated with this building and were connected to the sewer mains of the facility. Evaluated sump integrity by collecting two samples, zero to two ft interval, and from the six to eight ft interval, which straddled the bottom elevation of the sump. | None | Metals, black powder |
| LL11ss-023 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Building AP-4 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. QC sample collected. | None | Metals, black powder |
| LL11sb-023 | 0-1 6-8 | Metals, explosives, anions Metals, explosives, anions | Sump near Building AP-8 | 1941-1945, 1951-1957, and 1969-1971: Utilized for artillery primer loading and served as an administrative building. Two sets of sumps, one on the east side and one on the west side, were associated with this building and were connected to the sewer mains of the facility. Evaluated sump integrity by collecting two samples, zero to two ft interval, and from the six to eight ft interval, which straddled the bottom elevation of the sump. | None | Metals, black powder |
| LL11ss-024 | 0-1 1-3 | Metals, explosives, anions Metals, explosives, anions | Buildings AP-5 and AP-6 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder dry house and fan house. Two sets of sumps were associated with Building AP-5 and AP-6: one set was located outside the southwest wall of AP-6 and one set was located between AP-5 and AP-6. | None | Metals, black powder |
| LL11sb-024 | 7–9 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | None | Solvents, petroleum |
| LL11ss-025 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Building AP-7 | 1941-1945, 1951-1957, and 1969-1971: Utilized as black powder rest house prior to primer charging. | None | Metals, black powder |
| LL11sb-025 | 5–7 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11ss-026 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | North of Building AP-3 within NPA | Characterize an area not previously sampled. | None | Metals, explosives |
| LL11sb-026 | 4–6 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installingLL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11ss-027 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | North of Buildings AP-11 and AP-17 within NPA | Characterize an area not previously sampled. | None | Metals, explosives |
| LL11sb-027 | 6–8 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. QC sample collected. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11ss-028 | 0-1 1-3 | Metals, explosives, anions Metals, explosives, anions | West of Building AP-3 within NPA | Characterize an area not previously sampled. | None | Metals, explosives |
| LL11sb-028 | 4–6 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11ss-029 | 0-1 1-3 | Metals, explosives, anions Metals, explosives, anions | Southwest of Building AP-4 within NPA | Characterize an area not previously sampled. QC sample (0-1) collected. | None | Metals, explosives |
| LL11sb-029 | 6–8 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil | Petroleum- contaminated soil (Hot | Solvents, petroleum |

Table 4–5. Phase I RI Sampling Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|---|---|---|---|---------------------------------------|
| | | | | boring operations at LL11sb-005 for installing LL11mw-005. | Spot) | |
| LL11ss-030 | 0–1 1–3 | Metals, explosives, anions Metals, explosives, anions | Northeast portion of NPA | Characterize an area not previously sampled. QC sample (1-3) collected. | None | Metals, explosives |
| LL11sb-030 | 8-10 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11ss-031 | 0-1 | Metals, explosives, anions | Southeast portion of NPA (Drainage Ditch) | Characterize an area not previously sampled. | None | Metals, explosives |
| | 1–3 | Metals, explosives, anions | | | | |
| LL11sb-031 | 4–6 | Full suite, anions | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| | 6–8 | Metals, SVOCs, VOCs, TPH-DRO/GRO | Hot Spot hear ELITHW-003 | Resampling soil boring location LL11sb-031. An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-032 | 4–6 | Full suite, anions | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-033 | 6–8 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-034 | 2–4 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-035 | 4–6 | Full suite, anions | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. QC sample collected. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-036 | 4–6 | VOCs, SVOCs | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-037 | 6–8 | Full Suite, anions, TPH- DRO/GRO | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-038 | 6–8 | Full Suite, anions, TPH- DRO/GRO | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-039 | 6–8 | Full Suite, anions, TPH- DRO/GRO | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-040 | 6–8 | Full Suite, anions, TPH- DRO/GRO | Hot Spot near LL11mw-005 | An area of petroleum-contaminated soil (Hot Spot) approximately 30 ft long by 30 ft wide and 8 ft deep was identified just north of Building AP-17 (Solvent Storage) during the RI. The hot spot was encountered during soil boring operations at LL11sb-005 for installing LL11mw-005. | Petroleum- contaminated soil (Hot Spot) | Solvents, petroleum |
| LL11sb-041 | 6–8 | Full suite, anions | Sump near Building AP-8 | 1941-1945, 1951-1957, and 1969-1971: Utilized for artillery primer loading and served as an administrative building. When sump sediment/water samples were collected from the two AP-8 sumps, a sheen was noted on the water. Data from the AP-8 sediment/water samples raised the possibility that VOC/SVOC contaminants might be present in the soil surrounding the sump. Because the original sump soil boring samples had not been analyzed for VOCs/SVOCs, an additional boring was advanced at each of the AP-8 sumps and the soil samples were analyzed for VOCs/SVOCs. | None | Metals, black powder |
| LL11sb-042 | 6–8 | Full suite, anions | Sump near Building AP-8 | 1941-1945, 1951-1957, and 1969-1971: Utilized for artillery primer loading and served as an administrative building. When sump sediment/water samples were collected from the two AP-8 sumps, a sheen was noted on the | None | Metals, black powder |

Table 4–5. Phase I RI Sampling Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|-----------------------------|---|--|--------------------|---------------------------------------|
| Location | (It bgs) | Tillary tes | Investigation | water. Data from the AP-8 sediment/water samples raised the possibility that VOC/SVOC contaminants might be | Documented Release | CSC |
| | | | | present in the soil surrounding the sump. Because the original sump soil boring samples had not been analyzed for | | |
| | | | | VOCs/SVOCs, an additional boring was advanced at each of the AP-8 sumps and the soil samples were analyzed | | |
| | | | | for VOCs/SVOCs. | | |
| LL11sb-043 | 6–8 | Full suite, anions | Sewer Line south of Building AP-11 | Re-collected subsurface soil sample near sewer line boring LL11sb-017, to delineate vertical extent of contamination indicated. | None | Metals, explosives |
| LL11sb-017 | 4–6 | | | Evaluated Load Line 11 sewer line adjacent to sewer thrust point or manhole where sewer integrity was | | |
| | | Metals, explosives, anions | Sewer Line south of Building AP-11 | potentially suspect. Sewer soil sample was collected at elevations representing the elevation of the sewer system, | None | Metals, explosives |
| 7744 1040 | 0.1 | | | starting at the elevation equal to the total manhole depth and extending to two ft below the total manhole depth. | | |
| LL11sd-013 | 0-1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch, on the south end of the AOC just east of the entrance road. QC sample collected. | None | Metals, explosives |
| LL11sd-014 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch north of Building AP-18 (1941-1945, 1951-1957, and 1969-1971: Utilized for storage of percussion elements) | None | Metals, explosives |
| LL11sd-015 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch east of Building AP-18 (1941-1945, 1951-1957, and 1969-1971: Utilized for storage of percussion elements). | None | Metals, explosives |
| LL11sd-016 | 0–1 | Metals, explosives, anions | East Ditch | Drainage ditch northeast and downgradient of Building AP-13 (Change House), leading to toward AOC boundary. | None | Metals, explosives |
| | | ivietais, explosives, amons | East Ditch | Location was moved east across the road in the ditch line | None | wictais, explosives |
| LL11sd-017 | 0–1 | Metals, explosives, anions | East Ditch | Drainage ditch (East Ditch) northwest of Building AP-13 (Change House); however, it was moved to the pond | None | Metals, explosives |
| | - | · • · | | influent (due to presence of thick cattails and algae-rich pond water. | | |
| LL11sd-018 | 0–1 | Metals, explosives, anions | East Ditch | Drainage ditch north of Building AP-19 (dining hall). | None | Metals, explosives |
| LL11sd-019 | 0-1 | Metals, explosives, anions | East Ditch | Drainage ditch northeast and downgradient of Building AP-19 (dining hall), leading toward AOC boundary. | None | Metals, explosives |
| LL11sd-020 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch along northern road, directly across from Building AP-11 (black powder charging operations; assembly) | None | Metals, explosives |
| LL11sd-021 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch in northern portion of NPA, which sample location was moved to an outfall located 30 to 35 ft east of the fence line | None | Metals, explosives |
| LL11sd-022 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch along northern road, northwest of Building AP-20 (quality assurance primer sensitivity testing facility). | None | Metals, black powder |
| LL11sd-023 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch south of Building AP-11 (black powder charging operations; assembly) within the FPA. | None | Metals, black powder |
| LL11sd-024 | 0–1 | Metals, explosives, anions | West Ditch | Drainage ditch northeast of Building AP-7 (black powder rest house), within the FPA. | None | Metals, black powder |
| LL11sd-025 | 0–1 | Metals, explosives, anions | West Ditch | Drainage ditch northwest of Building AP-7 (black powder rest house), within the FPA. | None | Metals, black powder |
| LL11sd-026 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch in western portion of FPA. | None | Metals, explosives |
| LL11sd-027 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch within western end of FPA, south of Building AP-3 (black powder screening). | None | Metals, black powder |
| LL11sd-028 | 0–1 | Metals, explosives, anions | West Ditch (Drainage Ditch East of | Drainage ditch east/southeast of Building AP-4 (black powder rest house). Location moved approximately 10 ft | None | Metals, black powder |
| | | wictars, explosives, amons | Building AP-4) | northwest to a culvert that had an open pool. | TVOIC | wictars, black powder |
| LL11sd-029 | 0–1 | Full suite, anions | Drainage Ditch | Drainage ditch south of Building AP-3 (black powder screening), along western end of FPA. QC sample collected. | None | Metals, black powder |
| LL11sd-025 | 0–1 | Metals, explosives, anions | West Ditch | Drainage ditch northwest of Building AP-7 (black powder rest house), within the FPA. | None | Metals, black powder |
| LL11sd-026 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch in western portion of FPA. | None | Metals, explosives |
| LL11sd-027 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch within western end of FPA, south of Building AP-3 (black powder screening). | None | Metals, black powder |
| LL11sd-028 | 0–1 | Metals, explosives, anions | West Ditch (Drainage Ditch East of Building AP-4) | Drainage ditch east/southeast of Building AP-4 (black powder rest house). Location moved approximately 10 ft northwest to a culvert that had an open pool. | None | Metals, black powder |
| LL11sd-029 | 0–1 | Full suite, anions | Drainage Ditch | Drainage ditch south of Building AP-3 (black powder screening), along western end of FPA. QC sample collected. | None | Metals, black powder |
| LL11sd-030 | 0–1 | Metals, explosives, anions | West Ditch | Drainage ditch southwest of Building AP-4 (black powder rest house), prior to exiting the AOC. | None | Metals, black powder |
| LL11sd-031 | 0–1 | Metals, explosives, anions | Drainage Ditch | Drainage ditch along southern area of FPA. | None | Metals, explosives |
| LL11sd-032 | 0–1 | Full suite, anions | Drainage Ditch North of Building AP-14 | Drainage ditch north of Building AP-14 (Change House). | None | Metals, explosives |
| LL11sd-033 | 0–1 | Metals, explosives, anions | Drainage Ditch West of Building AP-8 | Drainage ditch just west of Building AP-8 (1941-1945, 1951-1957, and 1969-1971: Building AP-8 utilized for artillery primer loading and served as an administrative building.) | None | Metals, explosives |
| LL11sw-012 | NA | Full suite, anions | East Ditch | Drainage ditch north of Building AP-13 and southwest of Building AP-19 (dining hall) | None | Metals, explosives |
| LL11sw-013 | NA | Full suite, anions | Drainage Ditch | Drainage ditch south of AP-3 (black powder screening). When sediment sample LL11sd-027 was collected, water | None | Metals, explosives |

Table 4–5. Phase I RI Sampling Locations (continued)

| Phase I RI Sample Location | Sample Depth (ft bgs) | Analytes | Potential Sources or Areas for Investigation | Previous Use and/or Description | Documented Release | Potential Contaminants from Use |
|-------------------------------|-----------------------------|--------------------|---|--|--------------------|---------------------------------------|
| | | | | was present in the ditch. Therefore, a surface water sample was collected from that ditch rather than from the | | |
| | | | | sewer. | | |
| LL11sw-014 | NA | Full suite, anions | Drainage Ditch | Drainage ditch southwest of Buildings AP-5 and AP-6 (black powder dry house and fan house). Due to lack of water, sewer sample LL11sw-014 was collected instead near sediment location LL11sd-028. | None | Metals, black powder |
| LL11sw-015 | NA | Full suite, anions | West Ditch | Drainage ditch southwest of Building AP-4 (black powder rest house) prior to exiting the AOC. Due to lack of water, sewer water Sample LLsw-015 was collected near LL11sd-030 | None | Metals, explosives |

Note: Interim Removal Action confirmation samples are not included in this table, which are presented in Table 4-2. Anions include cyanide, sulfide, sulfate, and nitrate.

Anons include cyanide, surface,
AOC = Area of concern.
bgs = Below ground surface.
DRO = Diesel range organics.
FPA = Former production area.
ft = Feet.

R = Feet.
GRO = Gasoline range organics.
NPA = Non-production area.
QC= Quality control.
RI = Remedial investigation.
SVOC = Semi-volatile organic compound.

TPH = Total petroleum hydrocarbons.

VOC = Volatile organic compound.

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|------------------------|------------|--------------------|--------------------|--------------------|----------------------------|--------------------|--------------------|--------------------|--------------------|----------------------------------|
| Aggregate Station | - | LL11sb-001 | LL11sb-001 | LL11sb-003 | LL11sb-004 | LL11sb-008 | LL11sb-009 | LL11sb-019 | LL11sb-019 | LL11sb-020 |
| Sample ID | | LL11sb-001-0001-FD | LL11sb-001-0001-SO | LL11sb-003-0001-SO | LL11sb-004-0001-SO | LL11sb-008-0001-SO | | LL11sb-019-0001-FD | LL11sb-019-0001-SO | LL11sb-020 LL11sb-020-0001-SO |
| Date | - | 10/31/00 | 10/31/00 | 10/30/00 | 10/31/00 | 10/31/00 | 10/30/00 | 08/21/00 | 08/21/00 | 08/21/00 |
| Depth (ft) | - | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | RVAAP Full-suite | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | Misc., TAL Metals, |
| Analyte | Criteria | analytes | analytes | Explosives | Explosives | Explosives | Explosives | analytes | analytes | Explosives |
| 111tily to | Criteria | unurytes | unarytes | LAPIOSIVES | Metals (mg/kg) | LAPIOSIVES | Lapiosives | unarytes | unarytes | LAPIOSIVES |
| Aluminum | 17700 | 14000 | 10600 | 12000 | 18500* | 14000 | 15500 | 3970 | 4070 | 12800 |
| Antimony | 0.96 | <0.26U | 0.45 | 0.37 | 0.57 | 0.32 | 0.42 | 0.32 | <0.18U | 0.25 |
| Arsenic | 15.4 | 13.2 | 24.8* | 14.5 | 18.1* | 10.3 | 15.5* | 8.8 | 5.9 | 10.3 |
| Barium | 88.4 | 64.4 | 50.8 | 67.4 | 65.1 | 66.5 | 66.2 | 22.1 | 24.8 | 68 |
| Beryllium | | 0.45 | 0.35 | 0.44 | 0.48 | 0.43 | 0.5 | 0.31 | 0.35 | 0.68 |
| Cadmium | 0 | <0.14U | <0.13U | <0.13U | <0.13U | <0.14U | <0.14U | 0.21* | 0.17* | 0.21* |
| Calcium | 15800 | 5090 | 3490 | 1740 | 1140 | 1770 | 976 | 11700 | 10300 | 2680 |
| Chromium | | 17.1 | 13.7 | 15.4 | 20.7* | 15.9 | 18.8* | 6.2 | 6.8 | 16.7 |
| Cobalt | 10.4 | 7.7 | 6.5 | 9.4 | 7.5 | 8.9 | 9 | 3.9 | 4.2 | 8.6 |
| Copper | | 16.9 | 13.6 | 17.6 | 22.7* | 9.3 | 14.6 | 15.6 | 14.9 | 19.3* |
| Cyanide | 0 | <0.53U | <0.49U | <0.35U | <0.52U | <0.52U | <0.48U | <0.36U | <0.44U | <0.3U |
| Iron | 23100 | 23100 | 19000 | 22600 | 28600* | 18600 | 29200* | 13600 | 14400 | 22200 |
| Lead | 26.1 | 37.7* | 32.9* | 22 | 17.2 | 17.7 | 19.9 | 28.8* | 21.7 | 31.5* |
| Magnesium | 3030 | 3360* | 2930 | 2490 | 3070* | 2540 | 2770 | 2910 | 3210* | 3300* |
| Manganese | 1450 | 354 | 272 | 666 | 246 | 532 | 413 | 300 | 337 | 366 |
| Mercury | 0.036 | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.03U | 0.03 | <0.04U |
| Nickel | 21.1 | 16.9 | 15.2 | 17.2 | 18.7 | 13.6 | 16.5 | 10.2 | 11.2 | 20.8 |
| Potassium | 927 | 1970* | 1320* | 1600* | 2230* | 1610* | 1540* | 684 | 673 | 1790* |
| Selenium | 1.4 | <0.43U | <0.42U | <0.42U | <0.45U | <0.42U | <0.43U | <0.59U | <1.5U | <0.35U |
| Sodium | 123 | 825* | 700* | 896* | 892* | 878* | 847* | 402* | 397* | 763* |
| Thallium | 0 | <0.17U | 0.17* | <0.17U | <0.18U | <0.17U | <0.17U | 2.9* | <0.12U | 0.21* |
| Vanadium | 31.1 | 25.2 | 19.7 | 20.8 | 29.2 | 23.9 | 27.8 | 8.5 | 8.6 | 23.1 |
| Zinc | 61.8 | 66.3* | 52.3 | 62.1* | 58.9 | 51.3 | 65.1* | 54 | 55.9 | 68.5* |
| | | | | | Anions (mg/kg) | | | | | |
| Nitrate | None | <1.1U | <1.1U | 1.4* | <1.1U | <1.1U | <1.2U | 1.5* | <1.1U | <1.2U |
| Sulfide | None | <29U | <28.3U | <29U | 37* | <27.7U | <29.3U | <23U | <26.6U | <31U |
| | | | T | | Explosives and Propellants | | T | | | |
| Nitrocellulose | | 0.81* | 0.95* | NR | NR | NR | NR | NR | NR | NR |
| Nitroguanidine | None | <0.3U | <0.33U | NR | NR | NR | NR | NR | NR | NR |
| | | | T | | SVOCs (mg/kg) | | T | | | |
| Benz(a)anthracene | | 0.1J* | 0.14J* | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Benzo(a)pyrene | | 0.1J* | <0.39U | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Benzo(b)fluoranthene | | <0.39U | <0.39U | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Benzo(ghi)perylene | | <0.39U | <0.39U | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Benzo(k)fluoranthene | | <0.39U | <0.39U | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Chrysene | | 0.11J* | 0.17J* | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Fluoranthene | | 0.24J* | 0.32J* | NR | NR | NR | NR | 0.1J* | <0.36U | NR |
| Indeno(1,2,3-cd)pyrene | | <0.39U | <0.39U | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Phenanthrene | | 0.1J* | 0.12J* | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| Pyrene | None | 0.16J* | 0.31J* | NR | NR | NR | NR | <0.35U | <0.36U | NR |
| DCD 1254 | ът 1 | <0.000/II | <0.000(11 | ND | Pesticides/PCBs (mg/ | | ND | <0.000011 | <0.000011 | ND |
| PCB-1254 | None | <0.0096U | <0.0096U | NR | NR NB | NR NB | NR NB | <0.0089U | <0.0089U | NR |
| gamma-Chlordane | None | <0.0048U | <0.0048U | NR | NR VOCa (ma/ka) | NR | NR | 0.0018* | <0.0018U | NR |
| Agatons | NT | <0.006U | <0.006U | NID | VOCs (mg/kg) | NR | ND | <0.005U | <0.005U | ND |
| Acetone | none | ~∪.UU0U | <u> </u> | NR | NR | INV | NR | \u0.003U | ~0.003U | NR |

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| | I | TID A | TD 4 | TID A | TID A | TID 4 | TID A | ED. | ED. | TID 4 |
|------------------------|------------|--------------------|--------------------|--------------------|----------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Aggregate | | FPA TITLE 021 | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
| Station | | LL11sb-021 | LL11sb-022 | LL11sb-023 | LL11sd-023 | LL11sd-025 | LL11sd-026 | LL11sd-031 | LL11ss-001 | LL11ss-001 |
| Sample ID | | LL11sb-021-0001-SO | LL11sb-022-0001-SO | LL11sb-023-0001-SO | LL11sd-023-0001-SD | LL11sd-025-0001-SD | | LL11sd-031-0001-SD | LL11ss-001-0001-FD | LL11ss-001-0001-SO |
| Date | | 08/21/00 | 08/22/00 | 08/22/00 | 11/15/00 | 11/16/00 | 11/16/00 | 11/16/00 | 11/07/00 | 11/07/00 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite |
| Analyte | Criteria | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes | analytes |
| | 15500 | 12000 | 10500 | 6700 | Metals (mg/kg) | 12500 | 11700 | 1,5500 | (500 | 0110 |
| Aluminum | | 13000 | 10700 | 6780 | 13300 | 13700 | 11700 | 15500 | 6520 | 9110 |
| Antimony | 0.96 | 0.28 | 0.34 | <0.19U | 0.3 | 0.6 | 0.63 | 0.5 | <0.26U | 0.29 |
| Arsenic | | 9.9 | 10.2 | 12.3 | 12.7 | 21* | 13.4 | 21.1* | 10.4 | 9.9 |
| Barium | 88.4 | 70.8 | 63.3 | 53.5 | 51.8 | 67.6 | 59.4 | 66.9 | 35.2 | 48.3 |
| Beryllium | 0.88 | 0.7 | 0.68 | 0.52 | 0.46 | 0.68 | 0.49 | 0.69 | 0.41 | 0.52 |
| Cadmium | | 0.34* | <0.13U | <0.13U | <0.2U | 0.2* | <0.23U | <0.15U | <0.17U | <0.18U |
| Calcium | 15800 | 6690 | 1820 | 6960 | 830 | 1290 | 1810 | 2030 | 14300 | 53100* |
| Chromium | 17.4 | 17 | 14.4 | 10 | 17 | 17.5* | 15.1 | 19.9* | 8.7 | 10.3 |
| Cobalt | 10.4 | 8.6 | 9.6 | 6.8 | 3.8 | 33.8* | 6.3 | 9.6 | 5.2 | 6.3 |
| Copper | 17.7 | 23.2* | 16.7 | 18.8* | 13.2 | 12.1 | 13.4 | 18* | 13.6 | 15.7 |
| Cyanide | 0 | <0.39U | <0.36U | <0.43U | <0.44U | <0.69U | <0.54U | <0.57U | <0.48U | <0.36U |
| Iron | | 23400* | 24400* | 19600 | 25800* | 25400* | 20200 | 25400* | 16200 | 16400 |
| Lead | | 70.9* | 28.5* | 33.4* | 13 | 33.8* | 29.2* | 20.7 | 18.2 | 17.9 |
| Magnesium | | 3720* | 2440 | 2470 | 2220 | 2510 | 2400 | 3300* | 2220 | 3420* |
| Manganese | | 378 | 982 | 785 | 63.9 | 1160 | 285 | 294 | 365 | 442 |
| Mercury | 0.036 | <0.04U | <0.04U | <0.03U | <0.04U | 0.05* | <0.06U | <0.04U | <0.04U | <0.04U |
| Nickel | 21.1 | 21 | 18.6 | 14.8 | 11.5 | 15.1 | 14.6 | 20 | 10.9 | 12.2 |
| Potassium | 927 | 1930* | 1400* | 945* | 831 | 1320* | 1130* | 1830* | 926 | 1330* |
| Selenium | 1.4 | <2U | <0.33U | <0.31U | 1.1 | 1 | 0.68 | 0.55 | <2.1U | <0.92U |
| Sodium | | 804* | 668* | 528* | <97.7U | 721* | 702* | 796* | 456* | 599* |
| Thallium | | 0.26* | 0.13* | 0.14* | 0.21* | 0.21* | <0.21U | 0.21* | <0.17U | <0.18U |
| Vanadium | | 23.3 | 21.6 | 13.1 | 29.1 | 28 | 21.8 | 30.7 | 12 | 14.7 |
| Zinc | 61.8 | 94.6* | 54.5 | 58.4 | 36.6 | 71.2* | 90.6* | 60.8 | 51 | 52.4 |
| | | | | | Anions (mg/kg) | T | T | T | | |
| Nitrate | None | <1.2U | <1.1U | <1.1U | <1.3U | <0.69U | <0.77U | 0.92* | 3* | 3.2* |
| Sulfide | None | <26.9U | <28.6U | <25U | 76* | 45.3* | <29.5U | 42.8* | 133* | 244* |
| | | | | | Explosives and Propellants | | T | T | | |
| Nitrocellulose | | NR | NR | NR | NR | NR | NR | NR | 0.74* | 0.89* |
| Nitroguanidine | None | NR | NR | NR | NR | NR | NR | NR | <0.29U | 0.077J* |
| | | | 1 | | SVOCs (mg/kg) | T | T | T | 1 | |
| Benz(a)anthracene | | NR | NR | NR | NR | NR | 0.17J* | 0.078J* | NR | NR |
| Benzo(a)pyrene | None | | NR | NR | NR | NR | 0.15J* | 0.07J* | NR | NR |
| Benzo(b)fluoranthene | None | | NR | NR | NR | NR | 0.16J* | <0.38U | NR | NR |
| Benzo(ghi)perylene | | NR | NR | NR | NR | NR | <0.37U | <0.38U | NR | NR |
| Benzo(k)fluoranthene | | NR | NR | NR | NR | NR | 0.15J* | <0.38U | NR | NR |
| Chrysene | | NR | NR | NR | NR | NR | 0.17J* | 0.086J* | NR | NR |
| Fluoranthene | | NR | NR | NR | NR | NR | 0.33J* | 0.14J* | NR | NR |
| Indeno(1,2,3-cd)pyrene | | NR | NR | NR | NR | NR | <0.37U | <0.38U | NR | NR |
| Phenanthrene | | NR | NR | NR | NR | NR | 0.12J* | <0.38U | NR | NR |
| Pyrene | None | NR | NR | NR | NR | NR | 0.22J* | <0.38U | NR | NR |
| | 1 | | T | | Pesticides/PCBs (mg/ | | T | T | | |
| PCB-1254 | None | | NR | NR | NR | NR | NR | NR | <0.0092U | <0.0095U |
| gamma-Chlordane | None | NR | NR | NR | NR | NR | NR | NR | <0.0046U | <0.0047U |
| | |) ID |) TD |) ID | VOCs (mg/kg) | l vin | Lara | Lara | 0.0001 | 0.00677 |
| Acetone | None | NK | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|------------------------|--------------|----------------------------------|-----------------------|----------------------------------|---------------------------|--------------------|--------------------|---------------------|--------------------|----------------------------------|
| Aggregate Station | - | LL11ss-002 | LL11ss-003 | LL11ss-004 | LL11ss-004 | LL11ss-005 | LL11ss-006 | LL11ss-008 | LL11ss-009 | LL11ss-010 |
| Sample ID | | LL11ss-002 LL11ss-002-0001-SO | LL11ss-003-0001-SO | LL11ss-004 LL11ss-004-0001-FD | LL11ss-004-0001-SO | LL11ss-005-0001-SO | LL11ss-006-0001-SO | LL11ss-008-0001-SO | LL11ss-009-0001-SO | LL11ss-010 LL11ss-010-0001-FD |
| Date | - | 11/06/00 | 11/07/00 | 11/07/00 | 11/07/00 | 11/17/00 | 11/17/00 | 11/07/00 | 11/08/00 | 11/13/00 |
| Depth (ft) | - | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | Misc., TAL Metals, |
| Analyte | Criteria | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes | Explosives |
| Tilaryte | Criteria | Lapiosives | Lapiosives | Lapiosives | Metals (mg/kg) | Lapiosives | Lapiosives | Lapiosives | anarytes | LAPIOSIVES |
| Aluminum | 17700 | 9270 | 9700 | 8530 | 7840 | 13300 | 15900 | 9660 | 9260 | 12200 |
| Antimony | 0.96 | <0.26U | 0.51 | <0.28U | 0.54 | 0.3 | 0.85 | 0.35 | <0.27U | <0.29U |
| Arsenic | 15.4 | 13.2 | 14.3 | 14.2 | 20.2* | 24* | 11.4 | 17.9* | 5.3 | 8.4 |
| Barium | 88.4 | | 98.6* | 41.7 | 48.5 | 62.4 | 62.6 | 59.9 | 55.9 | 91.1* |
| Beryllium | | 0.66 | 0.58 | 0.46 | 0.45 | 0.78 | 0.6 | 0.57 | 0.61 | 0.63 |
| Cadmium | 0 | <0.17U | 0.35* | <0.18U | <0.19U | 0.23* | 0.17* | <0.18U | 0.22* | 0.32* |
| Calcium | 15800 | 19800* | 3200 | 1410 | 1940 | 1530 | 2620 | 2090 | 860 | 56100* |
| Chromium | 17.4 | 13.6 | 12.8 | 11.6 | 10.9 | 17.2 | 18.7* | 14 | 11.5 | 13.6 |
| Cobalt | 10.4 | 6.4 | 8.2 | 7.2 | 7.4 | 9.8 | 7.9 | 7.8 | 8.8 | 10.4 |
| Copper | | 17.7 | 35.5* | 23* | 20.7* | 20.5* | 11.1 | 17.1 | 15.2 | 12 |
| Cyanide | | 0.58* | <0.35U | <0.39U | <0.52U | <0.46U | <0.6U | 0.68* | <0.55U | <0.56U |
| Iron | | 23200* | 21700 | 20300 | 18300 | 26700* | 24900* | 22300 | 17600 | 16000 |
| Lead | | 15.2 | 63.4* | 34.4* | 34.6* | 5.6 | 5.3 | 58.1* | 6.4 | 20.5 |
| Magnesium | | 4570* | 2230 | 2290 | 2070 | 3090* | 2950 | 2150 | 1960 | 2560 |
| Manganese | | 576 | 800 | 313 | 501 | 286 | 481 | 569 | 843 | 1240 |
| Mercury | 0.036 | <0.04U | 0.08* | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | 0.04* | 0.06* |
| Nickel | 21.1 | 12.8 | 14.8 | 16.7 | 14.5 | 23.2* | 14.6 | 15.1 | 15.6 | 13.6 |
| Potassium | | 1050* | 948* | 1060* | 988* | 1840* | 1540* | 1300* | 1020* | 1290* |
| Selenium | 1.4 | <0.44U | <0.46U | <0.47U | <0.43U | <0.36U | 1.4* | <0.9U | <0.45U | <0.41U |
| Sodium | | 591* | 632* | 623* | 612* | 879* | 829* | 656* | 719* | 789* |
| Thallium | 0 | <0.18U | <0.19U | <0.19U | <0.17U | <0.14U | 0.22* | <0.18U | <0.18U | <0.19U |
| Vanadium | 31.1 61.8 | 14.7 | 18.9 92.7 * | 15 85.3 * | 14.2 66.4* | 21.2 59.9 | 31.4* 51.9 | 19.7 88 * | 17.4 59.4 | 21.9 68 * |
| Zinc | 01.8 | 40.9 | 92.1* | 92.3° | Anions (mg/kg) | 39.9 | 31.9 | 00" | 39.4 | 00" |
| Nitrate | None | <0.94U | 1.3* | <1U | <1.2U | <1.1U | <0.92U | <1.1U | <1.1U | 1.5* |
| Sulfide | | 154* | <29U | 81.5* | 52.1* | <29U | <30.2U | 35* | <27.8U | 45.4* |
| Surrice | rvone | 104 | 1270 | | xplosives and Propellants | | 30.20 | 33 | 27.00 | 10,1 |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | 1.1* | NR |
| Nitroguanidine | None | | | | | | | NR | <0.29U | NR |
| | Į. | | | | SVOCs (mg/kg) | | | | 1 | |
| Benz(a)anthracene | None | NR | NR | NR | | NR | NR | NR | <0.39U | NR |
| Benzo(a)pyrene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Benzo(b)fluoranthene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Benzo(ghi)perylene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Benzo(k)fluoranthene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Chrysene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Fluoranthene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Indeno(1,2,3-cd)pyrene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Phenanthrene | None | | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| Pyrene | None | NR | NR | NR | NR | NR | NR | NR | <0.39U | NR |
| DGD 1051 | | N.D. | N.T.D. | N.D. | Pesticides/PCBs (mg/ | |) ID | N.D. | 0.000077 | 0.000717 |
| PCB-1254 | None | | NR | NR | | NR | NR | NR | <0.0092U | <0.0095U |
| gamma-Chlordane | None | NK | NR | NR | NR WOG (/l -) | NR | NR | NR | <0.0046U | <0.0047U |
| A4 | ът | ND | ND | ND | VOCs (mg/kg) | ND | ND | ND | <0.00(11 | <0.00(11 |
| Acetone | None | NK | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|------------------------|------------|--------------------|----------------------------------|--------------------|----------------------------------|----------------------------------|--------------------|----------------------------------|--------------------|----------------------------------|
| Aggregate Station | - | LL11ss-011 | LL11ss-012 | LL11ss-013 | LL11ss-014 | LL11ss-015 | LL11ss-016 | LL11ss-017 | LL11ss-018 | LL11ss-019 |
| Sample ID | | LL11ss-011-0001-SO | LL11ss-012 LL11ss-012-0001-SO | LL11ss-013-0001-SO | LL11ss-014 LL11ss-014-0001-SO | LL11ss-015 LL11ss-015-0001-SO | LL11ss-016-0001-SO | LL11ss-017 LL11ss-017-0001-SO | LL11ss-018-0001-SO | LL11ss-019 LL11ss-019-0001-SO |
| Date | - | 11/17/00 | 11/17/00 | 11/08/00 | 11/06/00 | 11/08/00 | 11/13/00 | 11/08/00 | 11/09/00 | 11/10/00 |
| Depth (ft) | _ | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | Misc., TAL Metals, | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyte | Criteria | Explosives | analytes | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives |
| Analyte | Criteria | Explosives | analytes | Explosives | Metals (mg/kg) | Explosives | Explosives | Explosives | Lapiosives | Explosives |
| Aluminum | 17700 | 11400 | 7870 | 14500 | 7410 | 7900 | 8780 | 10700 | 13500 | 11900 |
| Antimony | | 0.5 | 0.34 | <0.24U | | 0.32 | 0.27 | <0.29U | 0.55 | 0.56 |
| Arsenic | 15.4 | | 24.2* | 12.1 | | 6.3 | 12.2 | 11.5 | 10.6 | 11.5 |
| Barium | 88.4 | | 36.1 | 107* | | 52.9 | 46.7 | 59.9 | 101* | 54 |
| Beryllium | | 0.47 | 0.49 | 1.1* | | 0.54 | 0.49 | 0.65 | 1* | 0.61 |
| Cadmium | | <0.15U | | 0.37* | | 1.3* | 0.32* | 0.25* | 0.16* | 0.2* |
| Calcium | 15800 | 167 | 1470 | 23700* | 64400* | 2030 | 1040 | 2140 | 27500* | 2010 |
| Chromium | 17.4 | | 11.4 | 11.4 | 12.9 | 12.4 | 11.4 | 16.1 | 14 | 14.5 |
| Cobalt | 10.4 | | 8 | 5 | | 9.2 | 8.7 | 11.2* | 6.9 | 10.8* |
| Copper | 17.7 | | 18.9* | 12.7 | 41.2* | 45.7* | 16.5 | 19.5* | 10.7 | 11.8 |
| Cyanide | | <0.53U | <0.42U | <0.53U | 1.6* | <0.52U | <0.55U | <0.52U | <0.53U | <0.56U |
| Iron | 23100 | 19600 | 20600 | 18100 | 13100 | 20400 | 16700 | 26800* | 19100 | 18400 |
| Lead | | 6.6 | 9.8 | 28.2* | 78.3* | 43* | 97.1* | 5.5 | 17.8 | 19.8 |
| Magnesium | | 1940 | 2410 | 4280* | 5390* | 2090 | 1720 | 3520* | 4800* | 2480 |
| Manganese | 1450 | 158 | 354 | 1500* | 642 | 544 | 507 | 390 | 1540* | 536 |
| Mercury | 0.036 | <0.04U | | 0.34* | | 0.05* | 0.04* | <0.04U | <0.04U | <0.04U |
| Nickel | | 12.7 | 18.5 | 11.2 | 13.3 | 17.6 | 11.8 | 26.8* | 10.1 | 13.4 |
| Potassium | | 949* | 1240* | 1560* | | 981* | 1060* | 1270* | 1110* | 1010* |
| Selenium | | 0.5 | | 0.89 | <0.87U | <0.44U | <0.32U | <0.49U | 0.45 | <0.47U |
| Sodium | | 604* | 770* | 858* | 663* | 734* | 647* | 888* | 178* | 840* |
| Thallium | | <0.15U | 0.15* | <0.16U | <0.17U | <0.18U | <0.18U | <0.2U | <0.18U | <0.19U |
| Vanadium | | 20.2 | 13.5 | 15.2 | 8.9 | 14.9 | 16.8 | 17.8 | 20.2 | 21.2 |
| Zinc | 61.8 | | 52.2 | 45.5 | 107* | 105* | 58.4 | 71.5* | 38.9 | 50.5 |
| | 1 | | | | Anions (mg/kg) | | | | | |
| Nitrate | None | 8.1* | 4.1* | <1U | <0.99U | <1.1U | <1U | <1.2U | <1.1U | <1.2U |
| Sulfide | None | <28.5U | <27.4U | 40.5* | 226* | 29.9* | 36.8* | 42.6* | 105* | 30.7* |
| | | | | E | xplosives and Propellants | (mg/kg) | | | | |
| Nitrocellulose | None | | 1.1* | NR | NR | NR | NR | NR | NR | NR |
| Nitroguanidine | None | NR | <0.28U | NR | | NR | NR | NR | NR | NR |
| | | | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | | | NR | | NR | NR | NR | NR | NR |
| Benzo(a)pyrene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Benzo(b)fluoranthene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Benzo(ghi)perylene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Benzo(k)fluoranthene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Chrysene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Fluoranthene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Indeno(1,2,3-cd)pyrene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Phenanthrene | None | | | NR | NR | NR | NR | NR | NR | NR |
| Pyrene | None | NR | <0.36U | NR | NR | NR | NR | NR | NR | NR |
| | | | | | Pesticides/PCBs (mg/ | | | | | |
| PCB-1254 | None | | | NR | NR | NR | NR | NR | NR | NR |
| gamma-Chlordane | None | NR | <0.00089U | NR | NR | NR | NR | NR | NR | NR |
| 1 | | ND | -0.00 <i>C</i> II | ND | VOCs (mg/kg) |) ID |) ID |) ID | ND | ND |
| Acetone | None | NK | <0.006U | NR | NR | NR | NR | NR | NR | NR |

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | NPA | NPA |
|----------------------|------------|--------------------|--------------------|--------------------|-------------------------------------|----------------------------------|--------------------|----------------------------------|--------------------|---------------------------------------|
| Aggregate Station | - | LL11ss-020 | LL11ss-021 | LL11ss-022 | LL11ss-023 | LL11ss-023 | LL11ss-024 | LL11ss-025 | LL11sb-002 | LL11sb-005 |
| Sample ID | _ | LL11ss-020-0001-SO | LL11ss-021-0001-SO | LL11ss-022-0001-SO | LL11ss-023-0002-FD | LL11ss-023 LL11ss-023-0001-SO | LL11ss-024-0001-SO | LL11ss-025 LL11ss-025-0001-SO | | |
| Date | - | 11/10/00 | 11/09/00 | 11/09/00 | 11/09/00 | 11/09/00 | 11/09/00 | 11/09/00 | 10/30/00 | 10/31/00 |
| Depth (ft) | F | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite |
| Analyte | Criteria | analytes | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes |
| | | | <u> </u> | <u> </u> | Metals (mg/kg) | <u> </u> | <u> </u> | <u> </u> | <u> </u> | , , , , , , , , , , , , , , , , , , , |
| Aluminum | 17700 | 7800 | 6880 | 11300 | 14900 | 7880 | 14100 | 9570 | 8870 | 13800 |
| Antimony | 0.96 | 0.4 | <0.28U | 0.41 | 0.73 | 0.36 | 0.61 | 0.54 | 0.32 | <0.25U |
| Arsenic | 15.4 | 13.9 | 10.3 | 30.2* | 14.8 | 11.1 | 16* | 12.6 | 12.1 | 9.4 |
| Barium | 88.4 | 39 | 44.6 | 45 | 128* | 33.9 | 79.8 | 52.5 | 48 | 69.8 |
| Beryllium | 0.88 | 0.5 | 0.39 | 0.53 | 0.78 | 0.44 | 0.68 | 0.54 | 0.35 | 0.45 |
| Cadmium | 0 | 0.4* | <0.14U | <0.16U | <0.15U | <0.14U | <0.16U | <0.16U | <0.14U | <0.14U |
| Calcium | 15800 | 1200 | 14900 | 999 | 1330 | 1200 | 2320 | 1400 | 3000 | 1580 |
| Chromium | 17.4 | 10.4 | 9.1 | 14.1 | 19.1* | 10.4 | 19.4* | 12.3 | 14.2 | 16.7 |
| Cobalt | 10.4 | 6.6 | 5.9 | 5.7 | 10.5* | 6.1 | 8.5 | 8 | 7.2 | 8.4 |
| Copper | 17.7 | 20.7* | 19.4* | 19.4* | 15.2 | 19.8* | 21.4* | 14.2 | 14.5 | 16.2 |
| Cyanide | 0 | <0.41U | <0.46U | <0.52U | <0.46U | <0.42U | <0.43U | <0.53U | <0.43U | <0.55U |
| Iron | 23100 | 17600 | 15100 | 20100 | 23300* | 18800 | 25900* | 18400 | 18800 | 22500 |
| Lead | 26.1 | 33.1* | 28.1* | 17.1 | 16.1 | 21.6 | 28* | 20.1 | 21.2 | 20.1 |
| Magnesium | 3030 | 1970 | 2800 | 2230 | 3250* | 2160 | 3410* | 2150 | 2290 | 2680 |
| Manganese | 1450 | 483 | 383 | 134 | 691 | 334 | 296 | 471 | 395 | 484 |
| Mercury | 0.036 | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U |
| Nickel | 21.1 | 17 | 12 | 13.5 | 22.1* | 15 | 21.2* | 14.4 | 15.5 | 18.2 |
| Potassium | 927 | 868 | 743 | 945* | 2130* | 1210* | 1730* | 939* | 1200* | 1980* |
| Selenium | 1.4 | <0.46U | <0.35U | <0.41U | <0.37U | <0.35U | <0.39U | 0.61 | <0.42U | <0.41U |
| Sodium | | 593* | <70.5U | <81.4U | 99.5 | <69.2U | 87.3 | <78.9U | 745* | 901* |
| Thallium | | <0.18U | <0.19U | <0.2U | <0.18U | <0.16U | <0.2U | <0.17U | <0.17U | <0.16U |
| Vanadium | | 13.2 | 12.7 | 21.6 | 24.9 | 13.9 | 24.4 | 17.3 | 16.2 | 24.6 |
| Zinc | 61.8 | 79.9* | 72.8* | 60.8 | 59.7 | 59.9 | 67.1* | 53.6 | 56.8 | 71.7* |
| NT' | NT. | ×11.1 | 2.2% | ×1.011 | Anions (mg/kg) | 2111 | -1 ATT | 2.1% | 21 1TT | 21.1TI |
| Nitrate | | <1U | 2.3* | <1.2U | <1.2U 36.5 * | <1U | <1.2U | 2.1* | <1.1U | <1.1U |
| Sulfide | None | 30.2* | 52.3* | 41.8* | 36.5* Explosives and Propellants | 56.8* | 43.1* | 46.9* | <28.5U | 31.1* |
| Nitrocellulose | None | 1* | NR | NR | NR | NR | NR | NR | NR | 0.89* |
| Nitroguanidine | | <0.28U | NR NR | NR NR | NR | NR | NR | NR | NR NR | <0.3U |
| Nitroguamume | None | <u> </u> | INK | INK | SVOCs (mg/kg) | INK | INK | INK | INK | <u> </u> |
| Benz(a)anthracene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | 0.071J* |
| Benzo(a)pyrene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| Benzo(b)fluoranthene | | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| Benzo(ghi)perylene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| Benzo(k)fluoranthene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| Chrysene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | 0.083J* |
| Fluoranthene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | 0.14J* |
| Indeno(1,2,3- | None | 0.070 | 1120 | 1,11 | 112 | 112 | 1.12 | 1,11 | 1.11 | 012 10 |
| cd)pyrene | 1,522 | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| Phenanthrene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| Pyrene | None | <0.37U | NR | NR | NR | NR | NR | NR | NR | <0.39U |
| | 1 | | | | Pesticides/PCBs (mg/ | (kg) | | | | |
| PCB-1254 | None | | NR | NR | NR | NR | NR | NR | NR | <0.0098U |
| gamma-Chlordane | None | <0.00094U | NR | NR | NR | NR | NR | NR | NR | <0.0049U |
| | | | | | VOCs (mg/kg) | T | | | | |
| Acetone | None | NR | <0.006U | NR | <0.006U | NR | <0.006U | NR | <0.006U | NR |

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|-----------------------------|------------|------------------------|--------------------|------------------------|---------------------------|--------------------|--------------------|--------------------|--------------------|------------------------|
| Station | - | LL11sb-006 | LL11sb-007 | LL11sb-010 | LL11sd-014 | LL11sd-015 | LL11sd-016 | LL11sd-020 | LL11sd-022 | LL11sd-029 |
| Sample ID | | LL11sb-006-0001-SO | LL11sb-007-0001-SO | LL11sb-010-0001-SO | | LL11sd-015-0001-SD | LL11sd-016-0001-SD | LL11sd-020-0001-SD | LL11sd-022-0001-SD | LL11sd-029-0001-FD |
| Date | - | 10/31/00 | 10/31/00 | 10/31/00 | 11/14/00 | 11/14/00 | 11/14/00 | 11/15/00 | 11/15/00 | 11/16/00 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite |
| Analyte | Criteria | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes |
| 1111111111111 | 01100110 | 211001100 | 23.19.10521105 | 20191001100 | Metals (mg/kg) | 23191051105 | 23191001100 | 2.191001, 00 | 2.191001100 | u11u1j vos |
| Aluminum | 17700 | 8570 | 11500 | 13400 | 12500 | 8640 | 11800 | 6170 | 10100 | 15400 |
| Antimony | | 0.41 | <0.26U | <0.27U | <0.28U | <0.29U | 0.77 | <0.25U | <0.25U | <0.22U |
| Arsenic | | 17.8* | 11.4 | 6.7 | 12.7 | 9.7 | 14.1 | 11.4 | 12.5 | 9.3 |
| Barium | | 45.1 | 51.6 | 64.6 | 98.1* | 48.5 | 74.5 | 37.7 | 46.7 | 86.3 |
| Beryllium | 0.88 | 0.39 | 0.42 | 0.36 | 0.68 | 0.52 | 0.71 | 0.41 | 0.53 | 0.84 |
| Cadmium | 0 | <0.15U | <0.13U | <0.15U | <0.2U | <0.22U | <0.2U | <0.19U | <0.16U | <0.17U |
| Calcium | 15800 | 13500 | 1140 | 802 | 1940 | 1790 | 18100* | 1780 | 1240 | 3510 |
| Chromium | | 11.9 | 13.9 | 13.1 | 14.6 | 12.3 | 17.2 | 10.5 | 15.3 | 20.9* |
| Cobalt | | 7.6 | 11* | 6.2 | 13.9* | 8.1 | 11.8* | 6.2 | 5.6 | 11.1* |
| Copper | 17.7 | 21* | 19.1* | 5.8 | 11.3 | 15 | 19.6* | 18.2* | 14.3 | 20.9* |
| Cyanide | 0 | <0.53U | <0.41U | <0.58U | <0.43U | <0.65U | <0.35U | <0.58U | <0.46U | <0.44U |
| Iron | | 21600 | 23400* | 14300 | 19500 | 18600 | 23900* | 17400 | 21100 | 26900* |
| Lead | | 20 | 15.3 | 13.3 | 16.2 | 18.3 | 20.3 | 27.2* | 10.8 | 14 |
| Magnesium | | 3520* | 2260 | 1580 | 2220 | 2050 | 6390* | 1740 | 2190 | 4440* |
| Manganese | | 373 | 664 | 237 | 1930* | 302 | 356 | 266 | 210 | 367 |
| Mercury | 0.036 | <0.04U | <0.04U | <0.04U | 0.05* | <0.04U | <0.04U | 0.08* | <0.04U | <0.04U |
| Nickel | 21.1 | 17.3 | 17.3 | 8.5 | 14.7 | 15.3 | 24.7* | 13.3 | 14.1 | 27.5* |
| Potassium | 927 | 1610* | 1560* | 1210* | 1120* | 971* 0.69 | 1870* | 934* 0.64 | 1040* 0.68 | 2530* |
| Selenium Sodium | 1.4 | <0.39U 922 * | <0.43U 905* | <0.46U 990 * | 0.59 807 * | 553* | <0.5U 736* | <92.9U | <81.5U | <0.43U 864 * |
| Thallium | 123 | <0.16U | <0.17U | <0.18U | <0.19U | <0.19U | <0.2U | <0.17U | 0.23* | <0.15U |
| Vanadium | 31.1 | 16.1 | 19.9 | 22.8 | 24.4 | 16.9 | 21.4 | 12.2 | 19.9 | 26.4 |
| Zinc | | 98.4* | 59.7 | 40.8 | 53 | 73.3* | 63.4* | 68.1* | 55.8 | 62.3* |
| Zinc | 01.0 | 70.4 | 37.1 | 70.0 | Anions (mg/kg) | 13.3 | 03.7 | 00.1 | 33.0 | 02.3 |
| Nitrate | None | <1U | 1.2* | <1.2U | 3* | <0.57U | 0.69* | <1.1U | 1.3* | <0.6U |
| Sulfide | | <27.3U | <29.6U | <31U | 79.7* | 58* | 68.2* | 48* | 51.5* | 37.4* |
| | | · - | | | Explosives and Propellant | | · | ı | | |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | NR | 0.97* |
| Nitroguanidine | None | | NR | | | | | | NR | <0.3U |
| | | | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Benzo(a)pyrene | None | | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Benzo(b)fluoranthene | None | | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Benzo(ghi)perylene | | NR | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Benzo(k)fluoranthene | | NR | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Chrysene | | NR | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Fluoranthene | | NR | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Indeno(1,2,3-cd)pyrene | | NR | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Phenanthrene | None | | NR NB | NR | NR | NR NB | NR | NR | NR | <0.4U |
| Pyrene | None | NK | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| DCD 1054 | None | ND | ND | ND | Pesticides/PCBs (mg | | ND | ND | ND | <0.0111 |
| PCB-1254 gamma-Chlordane | None None | | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | <0.01U <0.001U |
| gamma-Uniordane | None | NIC | NK | NIC | VOCs (mg/kg) | NK | NIC | NK | INK | <u></u> ~0.001∪ |
| Acetone | None | NR | NR | NR | NR | NR | NR | NR | NR | <0.006U |
| 1 tottoffe | THORE | 1111 | 1117 | 1111 | 1111 | 1110 | 1111 | 1110 | 1417 | -0.000 |

Table 4-6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|--------------------------------|------------|-------------------------|------------------------|--------------------|---------------------------|--------------------|--------------------|--------------------|------------------------|
| Station | | LL11sd-029 | LL11ss-026 | LL11ss-027 | LL11ss-028 | LL11ss-029 | LL11ss-029 | LL11ss-030 | LL11ss-031 |
| Sample ID | | LL11sd-029-0001-SD | LL11ss-026-0001-SO | LL11ss-027-0001-SO | LL11ss-028-0001-SO | LL11ss-029-0001-FD | LL11ss-029-0001-SO | LL11ss-030-0001-SO | LL11ss-031-0001-SO |
| Date | | 11/16/00 | 11/10/00 | 11/10/00 | 11/10/00 | 11/14/00 | 11/14/00 | 11/14/00 | 11/10/00 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyte | Criteria | analytes | Explosives | Explosives | Explosives | Explosives, VOCs | Explosives, VOCs | Explosives | Explosives |
| | | | | | Metals (mg/kg) | | | | |
| Aluminum | 17700 | 11000 | 9070 | 9340 | 14400 | 11600 | 11700 | 14400 | 10000 |
| Antimony | 0.96 | 0.43 | 0.34 | <0.27U | <0.3U | 0.63 | 0.39 | <0.29U | 0.36 |
| Arsenic | 15.4 | 18.2* | 6.6 | 18.8* | 9.5 | 11.3 | 11.5 | 11.5 | 17.9* |
| Barium | 88.4 | 61 | 62.4 | 28.4 | 55.2 | 71.9 | 71.5 | 58.3 | 87.7 |
| Beryllium | 0.88 | 0.64 | 0.48 | 0.53 | 0.52 | 0.64 | 0.63 | 0.53 | 0.74 |
| Cadmium | 0 | <0.16U | 0.2* | 0.19* | 0.23* | <0.2U | <0.2U | <0.19U | 0.38* |
| Calcium | 15800 | 5670 | 790 | 292 | 215 | 520 | 543 | 363 | 1370 |
| Chromium | 17.4 | 16.4 | 11.1 | 10.6 | 16.7 | 13.8 | 13.7 | 19* | 12.2 |
| Cobalt | 10.4 | 9.4 | 10.5* | 7.8 | 6.4 | 14* | 10.7* | 7 | 7.2 |
| Copper | 17.7 | 17.7 | 6.7 | 19.5* | 8.6 | 9.1 | 8.9 | 11.6 | 8.7 |
| Cyanide | 0 | <0.4U | <0.59U | <0.57U | <0.53U | <0.37U | <0.53U | 0.54* | <0.49U |
| Iron | 23100 | 22200 | 14800 | 19100 | 20800 | 20200 | 18500 | 27300* | 17400 |
| Lead | 26.1 | 15.6 | 16.2 | 13.7 | 18.5 | 22.6 | 23.2 | 15.4 | 22.4 |
| Magnesium | 3030 | 3510* | 1540 | 1750 | 2160 | 1850 | 1820 | 2800 | 1630 |
| Manganese | 1450 | 285 | 822 | 300 | 270 | 2080* | 1520* | 288 | 901 |
| Mercury | 0.036 | <0.04U | <0.04U | <0.04U | <0.04U | 0.05* | 0.05* | <0.04U | 0.05* |
| Nickel | 21.1 | 22* | 10.5 | 15.3 | 12.4 | 12.8 | 12.1 | 15.9 | 13.1 |
| Potassium | 927 | 1580* | 585 | 1020* | 1000* | 755 | 798 | 1430* | 752 |
| Selenium | 1.4 | 0.62 | <0.48U | <0.46U | <0.5U | 0.85 | 0.77 | 0.77 | <0.51U |
| Sodium | 123 | 746* | 684* | 637* | 934* | 704* | 775* | 647* | 706* |
| Thallium | | 0.16* | <0.19U | <0.18U | <0.2U | 0.24* | <0.21U | <0.19U | <0.2U |
| Vanadium | 31.1 | 20.1 | 19.7 | 14.3 | 28.4 | 25.5 | 24.6 | 26 | 19.9 |
| Zinc | 61.8 | 60.5 | 39.5 | 58.8 | 46.6 | 79.6* | 74.3* | 50.9 | 60.1 |
| NI't t | N | <0.5011 | <1.2U | | Anions (mg/kg) | <0.5211 | 0.67* | 0.64* | <1.211 |
| Nitrate | None | <0.58U | <1.20 42.6 * | <1.1U 52.7* | <1.2U 48.6 * | <0.53U | <31.8U | 85.5* | <1.3U 54.7 * |
| Sulfide | None | <26U | 42.0** | | s and Propellants (mg/kg) | 42.1* | <31.8U | 85.5* | 54./* |
| Nitrocellulose | None | 0.91* | NR | NR | NR | 1.2* | 1.2* | NR | NR |
| Nitroguanidine Nitroguanidine | None | <0.29U | NR NR | NR NR | NR NR | <0.33U | <0.32U | NR NR | NR NR |
| rvinoguamume | INUITE | N.47U | 1111 | 1 | SVOCs (mg/kg) | \0.JJU | \0.J2U | 111 | 1111 |
| Benz(a)anthracene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Benzo(a)pyrene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Benzo(b)fluoranthene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Benzo(ghi)perylene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Benzo(k)fluoranthene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Chrysene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Fluoranthene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Indeno(1,2,3-cd)pyrene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Phenanthrene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| Pyrene | None | <0.4U | NR | NR | NR | NR | NR | NR | NR |
| | 1,5110 | | <u> </u> | | cides/PCBs (mg/kg) | <u> </u> | 1 | • | |
| PCB-1254 | None | <0.01U | NR | NR | NR | NR | NR | NR | NR |
| gamma-Chlordane | | | NR | NR | NR | NR | NR | NR | NR |
| Samma Chioradic | 1 10110 | 0.0010 | 1 111 | 1111 | 1111 | 1111 | 1111 | 1111 | 1 111 |

Table 4–6. Analytes Detected in Phase I RI Surface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|---------------------|------------|-------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11sd-029 | LL11ss-026 | LL11ss-027 | LL11ss-028 | LL11ss-029 | LL11ss-029 | LL11ss-030 | LL11ss-031 |
| Sample ID | | LL11sd-029-0001-SD | LL11ss-026-0001-SO | LL11ss-027-0001-SO | LL11ss-028-0001-SO | LL11ss-029-0001-FD | LL11ss-029-0001-SO | LL11ss-030-0001-SO | LL11ss-031-0001-SO |
| Date | | 11/16/00 | 11/10/00 | 11/10/00 | 11/10/00 | 11/14/00 | 11/14/00 | 11/14/00 | 11/10/00 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed | Background | RVAAP Full-suite | Misc., TAL Metals, |
| Analyte | Criteria | analytes | Explosives | Explosives | Explosives | Explosives, VOCs | Explosives, VOCs | Explosives | Explosives |
| | | | | - | VOCs (mg/kg) | | | | |
| Acetone | None | <0.006U | NR | NR | NR | <0.006U | <0.006U | NR | NR |

FPA = Former production area.

ft = Feet.

ID = Identification.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.

NPA= Non-production Area.

NR = Not reported/not analyzed.

PCB = Polychlorinated biphenyl

RI = Remedial Investigation

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

TAL = Target analyte list.

U = Not detected.

VOC = Volatile organic compound.

VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|----------------------|---------------------|------------------|--------------------|--------------------|--------------------------------|--------------------|--------------------|--------------------|--------------------|-----------------------|
| Station | | LL11sb-001 | LL11sb-003 | LL11sb-004 | LL11sb-008 | LL11sb-008 | LL11sb-009 | LL11sb-011 | LL11sb-012 | LL11sb-012 |
| Station | | LL11sb-001-0002- | EL1130-003 | LL1130-004 | LLI13D-000 | LL1130-000 | LLII3D-007 | 221130-011 | LL1130-012 | EE1150-012 |
| Sample ID | | SO | LL11sb-003-0002-SO | LL11sb-004-0002-SO | LL11sb-008-0002-FD | LL11sb-008-0002-SO | LL11sb-009-0002-SO | LL11sb-011-0001-SO | LL11sb-012-0001-FD | LL11sb-012-0001-SO |
| Date | | 11/03/00 | 11/09/00 | 11/08/00 | 11/08/00 | 11/08/00 | 11/13/00 | 08/21/00 | 08/22/00 | 08/22/00 |
| Depth (ft) | | 18.0 - 20.0 | 20.0 - 40.0 | 10.0 - 12.0 | 12.0 - 13.0 | 12.0 - 13.0 | 4.0 - 6.0 | 4.0 - 6.0 | 8.0 - 10.0 | 8.0 - 10.0 |
| Parameters Analyzed | | Misc., TAL | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite |
| Analyte | Background Criteria | | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes | analytes |
| | | | | | Metals (mg/kg) | | | | | |
| Aluminum | 19500 | 8130 | 8650 | 6670 | 8300 | 7450 | 11600 | 10300 | 12000 | 11900 |
| Antimony | 0.96 | <0.3U | <0.29U | <0.27U | <0.26U | <0.26U | 0.74 | <0.21U | 0.58 | <0.2U |
| Arsenic | 19.8 | | 24.5* | 7.2 | 9.3 | 9.5 | 22.9* | 15.7 | 12.6 | 11.3 |
| Barium | | 35.2 | 64.7 | 26.9 | 42.5 | 30.2 | 68.8 | 54.4 | 58.5 | 58.8 |
| Beryllium | 0.88 | 0.44 | 0.61 | 0.41 | 0.47 | 0.45 | 0.72 | 0.72 | 0.84 | 0.61 |
| Cadmium Calcium | 35500 | <0.21U | <0.16U 993 | <0.18U 9050 | 0.22* 17900 | <0.17U 6990 | <0.16U 1130 | <0.14U 4460 | <0.14U 5990 | 0.16* 23500 |
| Chromium | 27.2 | | 12.4 | 11.5 | 15.8 | 13.5 | 16.2 | 14.4 | 15.8 | 16 |
| Cobalt | 23.2 | | 12.7 | 7.1 | 10.1 | 8.9 | 8 | 9.2 | 9.9 | 9.4 |
| Copper | 32.3 | | 20.2 | 17.4 | 18.9 | 21.4 | 17.2 | 18.5 | 20.6 | 18.3 |
| Cyanide | 0 | <0.34U | <0.55U | <0.32U | <0.37U | <0.48U | <0.3U | <0.6U | <0.42U | <0.5U |
| Iron | 35200 | 22700 | 22500 | 19000 | 21700 | 22900 | 24100 | 22700 | 21800 | 23200 |
| Lead | | 8.3 | 13.7 | 6.5 | 7.5 | 47.5* | 13.8 | 14.6 | 20.4* | 15.8 |
| Magnesium | 8790 | 4850 | 2370 | 4420 | 4960 | 4260 | 3360 | 3840 | 4040 | 5400 |
| Manganese | 3030 | 232 | 456 | 320 | 353 | 271 | 324 | 296 | 206 | 435 |
| Mercury | 0.044 | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U |
| Nickel | 60.7 | 19.7 | 19.2 | 17.5 | 23.3 | 20 | 21.2 | 23.3 | 21.8 | 21.5 |
| Potassium | 3350 | 1650 | 1020 | 1580 | 1510 | 1210 | 1130 | 1550 | 1870 | 1920 |
| Selenium | 1.5 | < 0.5 U | 0.39 | <0.46U | <0.44U | <0.44U | <0.4U | <1.8U | <1.7U | <1.7U |
| Sodium | 145 | 693* | <77.8U | 684* | 672* | 725* | 840* | 789* | 731* | 716* |
| Thallium | 0.91 | <0.2U | <0.19U | <0.18U | <0.18U | <0.17U | <0.2U | 0.2 | 0.23 | 0.14 |
| Vanadium | 37.6 | 13.8 | 14.8 | 11.7 | 14.1 | 12.8 | 18.7 | 17.3 | 21.7 | 21.3 |
| Zinc | 93.3 | 33./ | 49.5 | 58 | 53.8 | 56.5 | 56.9 | 54.8 | 57.7 | 58 |
| Nitrate | None | <1.3U | <1U | <1.1U | Anions (mg/kg) | <1.1U | <1.2U | <1.1U | <1.1U | <1.2U |
| Sulfate | None | <63.6U | <49.8U | <54.7U | <56.7U | <55.8U | <59U | <58.2U | <57.5U | <59.8U |
| Sulfide | None | 36.5* | 44.2* | 46.9* | <27.4U | <27U | 46.6* | <29.7U | <27U | <28.5U |
| Bullide | None | 30.3 | 77,2 | 10.7 | Miscellaneous (mg/kg) | | 10.0 | 12).10 | 1270 | 120.30 |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | | | | | | | | | NR | NR |
| | | | | | xplosives/Propellants (mg | | | | | |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | NR | NR | NR | NR | NR | NR | NR | <0.4U | <0.41U |
| Benzo(a)pyrene | None | NR | NR | NR | NR | NR | NR | NR | <0.4U | <0.41U |
| Benzo(b)fluoranthene | None | NR | NR | NR | NR | NR | NR | NR | <0.4U | <0.41U |
| Chrysene | None | NR | NR | NR | NR | NR | NR | NR | <0.4U | <0.41U |
| Fluoranthene | None | NR | NR | NR | NR | NR | NR | NR | <0.4U | <0.41U |
| Phenanthrene | None | NR | NR NB | NR NB | NR NR | NR NR | NR | NR NB | <0.4U | <0.41U |
| Pyrene | None | NR | NR | NR | NK Pesticides/PCBs (mg/kg | | NR | NR | <0.4U | <0.41U |
| PCB-1254 | None | NR | NR | NR | NR | NR | NR | NR | <0.01U | <0.01U |
| 1 CD-12JT | TYOHC | 1117 | INIX | 1117 | VOCs (mg/kg) | TAIX | INIX | INIX | | \0.01U |
| Acetone | None | NR | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |
| Carbon tetrachloride | None | NR | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |
| Toluene | None | NR | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |
| L | | | | | • | | | | | <u> </u> |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|----------------------|----------------------|--------------------|----------------------|--------------------|--|--------------------|--------------------|--------------------|--------------------|------------------|
| Station | - | LL11sb-017 | LL11sb-018 | LL11sb-019 | LL11sb-020 | LL11sb-021 | LL11sb-022 | LL11sb-023 | LL11sb-041 | LL11sb-042 |
| Sample ID | - | LL11sb-017-0001-SO | | LL11sb-019-0002-SO | | | LL11sb-022-0002-SO | | LL11sb-041-0001-SO | |
| Date | - | 08/22/00 | 08/22/00 | 08/21/00 | 08/21/00 | 08/21/00 | 08/22/00 | 08/22/00 | 03/13/01 | 03/13/01 |
| Depth (ft) | 1 | 4.0 - 6.0 | 4.0 - 6.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 |
| Parameters | 1 | 4.0 - 0.0 | 4.0 - 0.0 | 0.0 - 0.0 | 0.0 - 0.0 | 0.0 - 0.0 | 0.0 - 0.0 | 0.0 - 0.0 | 0.0 - 0.0 | 0.0 - 0.0 |
| Analyzed | | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite |
| Analyte | Background Criteria | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes | analytes |
| - Imaly te | Duengi vana Criteria | 2Aprosi ves | Emplosives | <u> </u> | Metals (mg/kg) | Emprosives | Emplosives | Emplosives | undiy tes | undly tes |
| Aluminum | 19500 | 6120 | 11800 | 13100 | 13700 | 13500 | 7330 | 9910 | 6020 | 11900 |
| Antimony | 0.96 | 0.24 | <0.2U | <0.2U | <0.2U | <0.21U | 0.33 | 0.21 | <0.29U | <0.27U |
| Arsenic | | | 16.5 | 11.8 | 11.4 | 15.3 | 25.8* | 10.7 | 11.2 | 20.6* |
| Barium | | 24.1 | 64.1 | 60 | 73.2 | 69.6 | 49.4 | 63.4 | 23.5 | 69.2 |
| Beryllium | | 0.48 | 0.86 | 0.74 | 0.73 | 0.75 | 0.59 | 0.64 | <0.38U | 0.74 |
| Cadmium | | <0.15U | <0.14U | 0.24* | 0.25* | 0.16* | <0.14U | <0.14U | <0.19U | <0.18U |
| Calcium | | 820 | 11300 | 27600 | 32600 | 16200 | 984 | 2930 | 844 | 7850 |
| Chromium | 27.2 | | 15.4 | 18.9 | 19.2 | 19 | 11.1 | 12.9 | 9.8 | 16 |
| Cobalt | | 7.1 | 8.7 | 10.6 | 13.3 | 10.6 | 7.4 | 7.8 | 6.3 | 7.9 |
| Copper | | | 18.7 | 20.6 | 20.8 | 20.8 | 21 | 14.6 | 21.1 | 16.3 |
| Cyanide | | 0.270 | <0.33U | <0.35U | <0.42U | <0.3U | <0.27U | <0.23U | <0.23U | <0.24U |
| Iron Lead | | 19200 13.7 | 23000 | 25500 15.2 | 24500 | 27300 16.2 | 24700 | 19200 | 20600 | 21300 |
| Magnesium | 19.1 8790 | 1660 | 19.2* 3810 | 6770 | 15.7 8270 | 6300 | 22.1* 2250 | 65.1* 2370 | 9.1 | 11.8 4630 |
| Manganese | | 346 | 527 | 359 | 545 | 313 | 330 | 507 | 145 | 578 |
| Mercury | | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.042U | <0.04U |
| Nickel | 60.7 | 14.4 | 22.7 | 27.5 | 28.6 | 27.2 | 18 | 16.5 | 16 | 22.1 |
| Potassium | | 690 | 1920 | 2860 | 3090 | 2970 | 900 | 1160 | 679 | 1480 |
| Selenium | | <1.9U | <1.7U | <0.33U | <1.6U | <1.7U | <0.31U | <1.7U | <0.47U | <0.44U |
| Sodium | | 611* | 884* | 817* | 805* | 837* | 619* | 623* | 712* | 923* |
| Thallium | | <0.15U | <0.13U | 0.17 | 0.17 | 0.21 | <0.13U | <0.14U | <0.2U | 0.19 |
| Vanadium | | 11.7 | 19 | 21.9 | 23 | 22.9 | 14 | 18.8 | 11.8 | 18.1 |
| Zinc | 93.3 | 54.5 | 60.2 | 62.4 | 57.7 | 59.9 | 52.7 | 54.6 | 53.3 | 55.3 |
| | | | | | Anions (mg/kg) | | | | | |
| Nitrate | None | <1.3U | <1.1U | <1.1U | <1.2U | <1.2U | <1.1U | <1.2U | 0.15* | 0.14* |
| Sulfate | None | <63.6U | <57.5U | <56.3U | <58.6U | <59.8U | <57.6U | <59.8U | 23* | 44.2* |
| Sulfide | None | <30.2U | <29.3U | <28.6U | <29U | <25.9U | <29.2U | 44.7* | <9.3U | <8.4U |
| | T T | | T | | Miscellaneous (mg/kg) | T | T | T | T | |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| NI'4 | None | ND | NR | | <i>plosives/Propellants (mg/</i> NR | (kg) NR | NR | NR | 1.04 | 1.64 |
| Nitrocellulose | None | NR | NK | NR | SVOCs (mg/kg) | NK | NK | NK | 1.9* | 1.6* |
| Benz(a)anthracene | None | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| Benzo(a)pyrene | | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| Benzo(b)fluoranthene | | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| Chrysene | None | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| Fluoranthene | None | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| Phenanthrene | | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| Pyrene | None | NR | NR | NR | NR | NR | NR | NR | <0.42U | <0.4U |
| <i>J</i> | | - | <u> </u> | | Pesticides/PCBs (mg/kg) | | <u> </u> | | | |
| PCB-1254 | None | NR | NR | NR | NR | NR | NR | NR | <0.011U | <0.01U |
| | • | | • | | VOCs (mg/kg) | • | • | • | • | • |
| Acetone | None | NR | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |
| Carbon tetrachloride | None | NR | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |
| Toluene | None | NR | NR | NR | NR | NR | NR | NR | <0.006U | <0.006U |
| · | · | | | | | | | | | |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|------------------------------|---------------------|-------------------------|--------------------|--------------------|---------------------------|--------------------|--------------------|--------------------|-------------------------|--------------------|
| Station | 1 | LL11sb-043 | LL11ss-001 | LL11ss-002 | LL11ss-003 | LL11ss-004 | LL11ss-007 | LL11ss-008 | LL11ss-009 | LL11ss-010 |
| Sample ID | 1 | LL11sb-043-0001-SO | | | LL11ss-003-0002-SO | LL11ss-004-0002-SO | | LL11ss-008-0002-SO | | |
| Date | 1 | 03/13/01 | 11/07/00 | 11/06/00 | 11/06/00 | 11/07/00 | 11/13/00 | 11/07/00 | 11/08/00 | 11/13/00 |
| Depth (ft) | 1 | 6.0 - 8.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 |
| Parameters | 1 | | | | | | | | | |
| Analyzed | | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite |
| Analyte | Background Criteria | analytes | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | analytes |
| | | • | • | <u> </u> | Metals (mg/kg) | - | - | - | - | |
| Aluminum | 19500 | | 11600 | 7190 | 9190 | 9540 | 11700 | 9810 | 8510 | 13200 |
| Antimony | | <0.29U | 0.69 | 0.28 | 0.51 | 0.92 | 0.37 | 0.37 | <0.26U | <0.29U |
| Arsenic | | 22.4* | 19.5 | 20* | 22.6* | 15.9 | 12.3 | 13.7 | 6.3 | 11.8 |
| Barium | | 42.3 | 53.8 | 47.4 | 65.9 | 52.3 | 75.5 | 58.2 | 42.9 | 76.7 |
| Beryllium | | 0.49 | 0.52 | 0.51 | 0.61 | 0.53 | 0.72 | 0.65 | 0.53 | 0.71 |
| Cadmium | | <0.2U | <0.18U | <0.18U | <0.2U | <0.19U | <0.16U | <0.17U | 0.2* | <0.16U |
| Claramina | 35500 27.2 | 1260 | 8060 13.6 | 1390 10.4 | 2260 13.7 | 1680 13.3 | 2600 | 2020 13.3 | 848 11.6 | 1720 15.4 |
| Chromium Cobalt | | | 7.5 | 8.8 | 10.9 | 9.9 | 9.4 | 7.8 | 8 | 10.2 |
| Copper | 32.3 | | 17.2 | 22.7 | 27.6 | 16.7 | 11 | 14.8 | 16.5 | 10.2 |
| Cyanide | | <0.23U | <0.4U | <0.45U | <0.43U | 0.79* | 0.29* | <0.46U | <0.37U | <0.52U |
| Iron | | 25000 | 20800 | 22000 | 23200 | 22600 | 19100 | 23400 | 17400 | 18900 |
| Lead | 19.1 | 12.7 | 20.1* | 13.4 | 32.9* | 16.8 | 16.4 | 18.2 | 8.1 | 18.8 |
| Magnesium | | 2550 | 3580 | 2160 | 2420 | 2480 | 2550 | 2150 | 1860 | 2090 |
| Manganese | | 231 | 382 | 325 | 395 | 413 | 895 | 594 | 737 | 935 |
| Mercury | | <0.043U | <0.04U | <0.04U | 0.04 | <0.04U | 0.04 | <0.04U | <0.04U | 0.05* |
| Nickel | 60.7 | 19.5 | 14.4 | 19.2 | 18.9 | 17.9 | 16 | 15.9 | 15.1 | 14 |
| Potassium | | 802 | 1320 | 814 | 990 | 1060 | 1320 | 1200 | 1070 | 1060 |
| Selenium | | <0.5U | <0.47U | <0.43U | <0.89U | <0.48U | 0.56 | <0.89U | <0.43U | 0.66 |
| Sodium | | 909* | 791* | 534* | 668* | 588* | 751* | 632* | 712* | 776* |
| Thallium | | <0.19U | <0.19U | <0.17U | <0.18U | <0.19U | <0.19U | <0.18U | 0.2 | <0.19U |
| Vanadium | | 15.1 | 20 | 12.3 | 16.6 | 17.2 | 21.3 | 20.2 | 16.5 | 26.4 |
| Zinc | 93.3 | 61.6 | 51.9 | 63.3 | 74.4 | 54 | 54.1 | 63.6 | 56 | 50.6 |
| NI'AA. | NT | A 154 | A± | z1 1II | Anions (mg/kg) | <1.011 | 1 44 | 21 111 | 21 111 | 1 1 4 |
| Nitrate | | 0.15* | 2* <59.4U | <1.1U | <1U <51U | <1.2U <59.7U | 1.4* <58.8U | <1.1U | <1.1U | 1.1* <50.9U |
| Sulfate Sulfide | None None | 22.8* <9.4U | <39.40 <27.3U | <56.7U <27.9U | 34.5* | 71.5* | 63.8* | <56.4U 35.2* | <53.3U 34.5 * | 54.7 * |
| Sumde | None | <u> </u> | ~27.30 | ~27.90 | Miscellaneous (mg/kg) | /1.5 | 03.0 | 33.2 | 34.3 | 34.7 |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| THI GRO | TVOICE | 1111 | 1111 | | plosives/Propellants (mg/ | | 1111 | 1111 | 1111 | 1111 |
| Nitrocellulose | None | 2* | NR | NR | NR | NR | NR | NR | NR | 1.3* |
| | | | • | | SVOCs (mg/kg) | • | • | • | | |
| Benz(a)anthracene | None | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Benzo(a)pyrene | None | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Benzo(b)fluoranthene | | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Chrysene | None | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Fluoranthene | None | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Phenanthrene | None | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| Pyrene | None | <0.43U | NR | NR | NR | NR | NR | NR | NR | <0.4U |
| DCD 1254 | NI | <0.01111 | ND | | Pesticides/PCBs (mg/kg) | | ND | ND | ND | 0.022* |
| PCB-1254 | None | <0.011U | NR | NR | NR VOCa (ma/ha) | NR | NR | NR | NR | 0.033* |
| Agatona | None | <0.006U | NR | NR | VOCs (mg/kg) NR | NR | NR | NR | NR | <0.006U |
| Acetone Carbon tetrachloride | None | <0.006U | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | <0.006U <0.006U |
| Toluene Carbon tetrachioride | None | <0.006U | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | <0.006U <0.006U |
| 1 OTUCITE | TNOHE | ~0.000U | INIX | INIX | INIX | INIX | INIX | INIX | INIX | \0.000U |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|----------------------|---------------------|-------------------------|--------------------|--------------------|---------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11ss-013 | LL11ss-014 | LL11ss-015 | LL11ss-016 | LL11ss-016 | LL11ss-017 | LL11ss-018 | LL11ss-019 | LL11ss-020 |
| Sample ID | | LL11ss-013-0002-SO | LL11ss-014-0002-SO | LL11ss-015-0002-SO | LL11ss-016-0002-FD | LL11ss-016-0002-SO | LL11ss-017-0002-SO | LL11ss-018-0002-SO | LL11ss-019-0002-SO | LL11ss-020-0002-SO |
| Date | | 11/08/00 | 11/07/00 | 11/08/00 | 11/13/00 | 11/13/00 | 11/08/00 | 11/09/00 | 11/10/00 | 11/10/00 |
| Depth (ft) | | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 |
| Parameters | | | | | | | | | | |
| Analyzed | | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | RVAAP Full-suite | RVAAP Full-suite | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyte | Background Criteria | analytes | Explosives | Explosives | analytes | analytes | Explosives | Explosives | Explosives | Explosives |
| · | | • | | • | Metals (mg/kg) | • | | | - | |
| Aluminum | 19500 | | 14000 | 9290 | 11400 | 11700 | 9480 | 14600 | 12700 | 9840 |
| Antimony | 0.96 | <0.26U | 0.49 | <0.26U | 0.47 | 0.42 | <0.26U | <0.3U | 0.44 | <0.29U |
| Arsenic | 19.8 | | 10.9 | 9.2 | 9.6 | 14.7 | 9.3 | 13.7 | 8.1 | 15.3 |
| Barium | | 39.5 | 80.7 | 63.4 | 41.9 | 39.4 | 42.7 | 48.5 | 58.1 | 44.7 |
| Beryllium | | 0.36 | 0.8 | 0.68 | 0.53 | 0.54 | 0.6 | 0.55 | 0.68 | 0.5 |
| Cadmium | | 0.24* | <0.18U | 0.29* | <0.15U | <0.14U | 0.17* | <0.16U | 0.17* | 0.27* |
| Calcium | | 891 | 10100 | 1600 | 850 | 710 | 1730 | 1840 | 597 | 1620 |
| Chromium | | 7.7 | 15.8 | 14.4 | 14.2 | 13.7 | 14.1 | 19 | 14.4 | 13.3 |
| Cobalt | 23.2 | | 8.2 | 10.2 | 6.9 | 6.9 | 9.6 | 6.4 | 11 | 7.9 |
| Copper | 32.3 | | 20.4 | 24.5 | 18.4 | 17.6 | 19.6 | 20 | 11 | 13.8 |
| Cyanide | 0 | <0.33U | <0.48U | <0.4U | <0.43U | <0.55U | <0.39U | <0.52U | <0.45U | <0.48U |
| Iron | 35200 | 14200 | 21400 | 23100 | 20800 | 20300 | 23900 | 26500 | 18800 | 21500 |
| Lead | | 4.8 | 23.8* | 14.1 | 19.7* | 22.2* | 5.4 | 12.9 | 15.9 | 25.9* |
| Magnesium | 8790 | 1210 | 3370 | 2430 | 2310 | 2270 | 2960 | 3040 | 2020 | 2210 |
| Manganese | | 245 | 788 | 332 | 292 | 303 | 338 | 178 | 638 | 337 |
| Mercury | 0.044 | <0.04U | 0.04 | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U |
| Nickel | 60.7 | 10.2 | 16.2 | 22.7 | 14.3 | 13.9 | 23 | 16.1 | 12.7 | 15 |
| Potassium | | 595 | 1390 | 1310 | 942 | 1030 | 1450 | 1570 | 975 | 704 |
| Selenium | 1.5 | <0.44U | <0.48U | <0.44U | 0.43 | 0.5 | <0.44U | <0.4U | <0.5U | <0.48U |
| Sodium | | 478* | 793* | 885* | 634* | 622* | 868* | 90.9 | 835* | 631* |
| Thallium | | <0.18U | <0.19U | <0.17U | <0.19U | <0.19U | 0.21 | <0.2U | <0.2U | 0.23 |
| Vanadium | 37.6 93.3 | | 23 65.3 | 15.9 | 20.9 51.4 | 20.9 46.5 | 16.1 65.2 | 24.3 | 21.6 44.1 | 17.1 52.2 |
| Zinc | 93.3 | 38.2 | 03.3 | 66.4 | | 40.3 | 03.2 | 47.6 | 44.1 | 32.2 |
| Nitrate | None | <0.94U | <1.2U | <1.1U | Anions (mg/kg) | 13.6* | 1.4* | <1.1U | <1.2U | <1.2U |
| Sulfate | None | <47.1U | <59.4U | <51.4U | <57.7U | <56.5U | <53.8U | <56.7U | <55.5U | <60.4U |
| Sulfide | None | <27.5U | 125* | 44.6* | <27.5U | <25.8U | 74* | 49.7 * | 52.9 * | 31.8* |
| Sumde | None | ~27.30 | 123 | 11.0 | Miscellaneous (mg/kg) | ~23.80 | / + | 47.1 | 32.9 | 31.0 |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TH GRE | TVOICE | 1110 | Titt | | plosives/Propellants (mg/ | | 1110 | 1110 | 1110 | 111 |
| Nitrocellulose | None | 0.87* | NR | NR | 1* | 0.98* | NR | NR | NR | NR |
| TitleGellalose | Tione | 0.07 | 1111 | 1112 | SVOCs (mg/kg) | 0.50 | 1112 | 1112 | 1111 | 1112 |
| Benz(a)anthracene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| Benzo(a)pyrene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| Benzo(b)fluoranthene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| Chrysene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| Fluoranthene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| Phenanthrene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| Pyrene | None | <0.37U | NR | NR | <0.39U | <0.39U | NR | NR | NR | NR |
| • | | | • | | Pesticides/PCBs (mg/kg) | | • | • | • | • |
| PCB-1254 | None | 0.042* | NR | NR | 0.45* | 0.79* | NR | NR | NR | NR |
| | | | | | VOCs (mg/kg) | | | | | |
| Acetone | None | <0.006U | NR | NR | <0.006U | <0.006U | NR | NR | NR | NR |
| Carbon tetrachloride | None | <0.006U | NR | NR | <0.006U | <0.006U | NR | NR | NR | NR |
| Toluene | None | <0.006U | NR | NR | <0.006U | <0.006U | NR | NR | NR | NR |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | NPA | NPA | NPA | NPA |
|----------------------------------|---------------------|--------------------|--------------------|--------------------|---------------------------|--------------------|-----------------------|--------------------|--------------------|--------------------|
| Station | | LL11ss-021 | LL11ss-022 | LL11ss-023 | LL11ss-024 | LL11ss-025 | LL11sb-002 | LL11sb-005 | LL11sb-006 | LL11sb-007 |
| Sample ID | | LL11ss-021-0002-SO | | | | | | | | |
| Date | 1 | 11/09/00 | 11/09/00 | 11/09/00 | 11/09/00 | 11/09/00 | 11/10/00 | 11/09/00 | 11/13/00 | 11/07/00 |
| Depth (ft) | | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 6.0 - 8.0 | 6.0 - 8.0 | 2.0 - 4.0 | 14.0 - 16.0 |
| Parameters | | | | | | | | | | |
| Analyzed | | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyte | Background Criteria | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives | Explosives |
| | 1 | | 1 | T | Metals (mg/kg) | | 1 | | T | |
| Aluminum | | 11100 | 9840 | 16400 | 15000 | 11900 | 8990 | 4720 | 6140 | 7980 |
| Antimony | | 0.4 | 0.33 | 0.5 | 0.33 | 0.39 | <0.27U | <0.29U | 0.31 | 0.4 |
| Arsenic | | 13.6 | 16.4 | 11.8 | 12.8 | 12 | 18.9 | 12 | 18.5 | 21.8* |
| Barium | 124 0.88 | | 46.3 | 88.8 0.81 | 65 0.68 | 38.9 | 39.2 | 20.6 <0.32U | 26.1 0.42 | 0.39 |
| Beryllium Cadmium | | | 0.62 <0.16U | <0.15U | <0.16U | 0.56 <0.15U | 0.55 0.24 * | <0.32U <0.16U | <0.15U | <0.18U |
| Calcium | | 1710 | 610 | 1650 | 2060 | 977 | 6710 | 143 | 7020 | 9150 |
| Chromium | | 15.1 | 13 | 19.8 | 2000 | 15.3 | 16.1 | 7.2 | 11.1 | 14.2 |
| Cobalt | | 7.5 | 8.5 | 9.5 | 9.6 | 14.7 | 8.8 | 4 | 6.9 | 9.1 |
| Copper | 32.3 | 16.2 | 18.3 | 14 | 17.4 | 17.5 | 19.4 | 13.7 | 18.7 | 19.5 |
| Cyanide | 0 | <0.45U | <0.47U | <0.46U | <0.35U | <0.46U | <0.5U | <0.47U | <0.31U | <0.35U |
| Iron | 35200 | 20100 | 20800 | 24200 | 23600 | 21300 | 22500 | 12200 | 18500 | 24400 |
| Lead | 19.1 | 13.5 | 15.3 | 14.4 | 23.1* | 12.6 | 12.5 | 8.9 | 10.8 | 10.3 |
| Magnesium | 8790 | 2750 | 2210 | 3070 | 3160 | 2480 | 3470 | 1040 | 3970 | 5750 |
| Manganese | | 257 | 228 | 763 | 188 | 493 | 297 | 186 | 269 | 366 |
| Mercury | 0.044 | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U |
| Nickel | | 15.4 | 14.8 | 18.6 | 18.5 | 18 | 22 | 8.6 | 15.4 | 21.3 |
| Potassium | 3350 | | 722 | 2210 | 1410 | 1470 | 1570 | 716 | 1210 | 1510 |
| Selenium | | <0.36U | 0.48 | <0.38U | 0.42 | <0.39U | <0.45U | <0.4U | <0.38U | <0.46U |
| Sodium | | <71.6U | <79.1U | 94.7 | 86.8 | <77.4U | 858* | <80.4U | 671* | 452* |
| Thallium Vanadium | 0.91 37.6 | <0.19U | <0.19U 18.4 | <0.19U 28.5 | <0.18U 29 | <0.17U 21.6 | <0.18U 15.3 | <0.2U 8.9 | <0.19U 11.2 | <0.18U 12.7 |
| Zinc | 93.3 | | 46 | 65.1 | 51.7 | 44.1 | 58.8 | 42.4 | 51.3 | 49.9 |
| Zilic | 75.5 | 31.0 | 10 | 03.1 | Anions (mg/kg) | TT.1 | 36.6 | 72.7 | 31.3 | 77.7 |
| Nitrate | None | 1.4* | <0.96U | <1.2U | <1.2U | <1.2U | 1.5* | <1.2U | <0.96U | <1.1U |
| Sulfate | None | <58.5U | <47.8U | <59.5U | <59.8U | <57.8U | <48.4U | <59.1U | <48.2U | <54.8U |
| Sulfide | None | 37.1* | 28.7* | 161* | 52.3* | 44.4* | 39.9* | 52.5* | 46.9* | 29.9* |
| | | | | | Miscellaneous (mg/kg) | | | | | |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | , , | | 1 | | plosives/Propellants (mg/ | | 1 | 1 | T | |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| D () 1 | 1 3 7 | N.I.D. | Lam | 1 3 m | SVOCs (mg/kg) | 310 | 1 vm | 310 | 1 3 ID | Lib |
| Benz(a)anthracene | None | NR | NR | NR | NR | NR NR | NR | NR | NR | NR |
| Benzo(a)pyrene | None | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR |
| Benzo(b)fluoranthene Chrysene | None None | NR | NR NR | NR | NR | NR | NR | NR | NR | NR NR |
| Fluoranthene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Phenanthrene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Pyrene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| 1 110110 | 1.5110 | | 1 | | Pesticides/PCBs (mg/kg) | I. | 1 2.20 | 1 | 1 2.20 | 1 2 1 2 2 |
| PCB-1254 | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| - | | | • | | VOCs (mg/kg) | • | | • | | |
| Acetone | None | NR | NR | NR | NR NR | NR | NR | NR | NR | NR |
| Carbon tetrachloride | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Toluene | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | | | | | | |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|----------------------|---------------------|--------------------|------------------|--------------------|--------------------------------|-------------------|--------------------|--------------------|--------------------|----------------|
| Station | - | LL11sb-010 | LL11sb-024 | LL11sb-025 | LL11sb-026 | LL11sb-027 | LL11sb-027 | LL11sb-028 | LL11sb-029 | LL11sb-030 |
| Sample ID | - | LL11sb-010-0002-SO | 1 | LL11sb-025-0001-SO | | | | | LL11sb-029-0001-SO | |
| Date | - | 11/07/00 | 11/16/00 | 11/17/00 | 11/16/00 | 11/16/00 | 11/16/00 | 11/16/00 | 11/16/00 | 11/17/00 |
| Depth (ft) | - | 14.0 - 16.0 | 7.0 - 9.0 | 5.0 - 7.0 | 4.0 - 6.0 | 6.0 - 8.0 | 6.0 - 8.0 | 4.0 - 6.0 | 6.0 - 8.0 | 8.0 - 10.0 |
| Parameters | - | 1410 1010 | 710 710 | 2.0 7.0 | 4.0 0.0 | 0.0 0.0 | 0.0 0.0 | 410 010 | 0.0 0.0 | 0.0 10.0 |
| Analyzed | | Misc., TAL Metals, | | | | | | | | |
| Analyte | Background Criteria | Explosives | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs |
| Tinaryte | Duckground Criteria | LAPIOSIVES | 5 1 0 CB, 1 0 CB | 5 1 0 CB, 1 0 CB | Metals (mg/kg) | B 1 0 CB, 1 0 CB | 5 1 0 CB, 1 0 CB | 5 1 0 CB, 1 0 CB | 51003,1003 | b v des, v des |
| Aluminum | 19500 | 5750 | NR | NR | NR | NR | NR | NR | NR | NR |
| Antimony | 0.96 | 0.35 | NR | NR | NR | NR | NR | NR | NR | NR |
| Arsenic | | 14.7 | NR | NR | NR | NR | NR | NR | NR | NR |
| Barium | | 17.5 | NR | NR | NR | NR | NR | NR | NR | NR |
| Beryllium | | | NR | NR | NR | NR | NR | NR | NR | NR |
| Cadmium | | 0.170 | NR | NR | NR | NR | NR | NR | NR | NR |
| Calcium | | 6920 | NR | NR | NR | NR | NR | NR | NR | NR |
| Chromium | | 9.4 | NR | NR | NR | NR | NR | NR | NR | NR |
| Cobalt | 23.2 | | NR | NR | NR | NR | NR | NR | NR | NR |
| Copper Cyanide | 32.3 | <0.3U | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR | NR NR |
| | 35200 | 16800 | NR NR | NR | NR NR | NR NR | NR | NR | NR NR | NR NR |
| Iron Lead | 19.1 | 8.1 | NR | NR | NR | NR | NR | NR | NR | NR |
| Magnesium | | 3850 | NR | NR | NR | NR | NR | NR | NR | NR |
| Manganese | | 346 | NR | NR | NR | NR | NR | NR | NR | NR |
| Mercury | | <0.04U | NR | NR | NR | NR | NR | NR | NR | NR |
| Nickel | | 14.8 | NR | NR | NR | NR | NR | NR | NR | NR |
| Potassium | 3350 | | NR | NR | NR | NR | NR | NR | NR | NR |
| Selenium | 1.5 | <0.41U | NR | NR | NR | NR | NR | NR | NR | NR |
| Sodium | | 456* | NR | NR | NR | NR | NR | NR | NR | NR |
| Thallium | | <0.16U | NR | NR | NR | NR | NR | NR | NR | NR |
| Vanadium | 37.6 | | NR | NR | NR | NR | NR | NR | NR | NR |
| Zinc | 93.3 | 41.3 | NR | NR | NR | NR | NR | NR | NR | NR |
| AT' | 3.7 | -0.07II |) ID | 3.10 | Anions (mg/kg) | 310 | N.D. |) ID | ND. | ND. |
| Nitrate | None | <0.97U | NR | NR | NR | NR | NR | NR | NR | NR |
| Sulfate | None | <48.5U <27.6U | NR NR | NR NR | NR NR | NR NR | NR NR | NR NB | NR NB | NR |
| Sulfide | None | <27.00 | NK | INK | Miscellaneous (mg/kg) | NK | INK | NR | NR | NR |
| TPH-DRO | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | | NR | NR | NR | NR | NR | NR |
| IIII GRO | TVOICE | 1110 | TVIC | | plosives/Propellants (mg/ | | 1410 | TVIC | TVIC | 1110 |
| Nitrocellulose | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| | | | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| Benzo(a)pyrene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| Benzo(b)fluoranthene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| Chrysene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| Fluoranthene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| Phenanthrene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| Pyrene | None | NR | <0.39U | <0.38U | <0.38U | <0.39U | <0.38U | <0.39U | <0.38U | <0.38U |
| DCD 1254 | No | NID | ND | | Pesticides/PCBs (mg/kg) | | LND | LND | ND | ND |
| PCB-1254 | None | NR | NR | NR | NR VOCs (mg/kg) | NR | NR | NR | NR | NR |
| Acetone | None | NR | <0.006U | <0.006U | <i>VOCs (mg/kg)</i> <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U |
| Carbon tetrachloride | None | NR NR | <0.006U | <0.006U | 0.001J* | <0.006U | <0.006U <0.006U | <0.006U <0.006U | <0.006U | <0.006U |
| Toluene | None | NR | <0.006U | <0.006U | <0.006U | 0.000U 0.002J* | <0.006U | <0.006U | <0.006U | <0.006U |
| Torucile | 110110 | 1111 | 0.000 | | 0.000 | V•VV#U | 1 0.0000 | 1 0.000 | 0.0000 | |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|--------------------------|---------------------|-------------------------|--------------------|--------------------|---------------------------|--------------------|--------------------|--------------------|---------------------|--------------------|
| Station | | LL11sb-035 | LL11sb-035 | LL11sb-037 | LL11sb-038 | LL11sb-039 | LL11sb-040 | LL11ss-026 | LL11ss-027 | LL11ss-028 |
| Sample ID | | LL11sb-035-0001-FD | LL11sb-035-0001-SO | LL11sb-037-0001-SO | LL11sb-038-0001-SO | LL11sb-039-0001-SO | LL11sb-040-0001-SO | LL11ss-026-0002-SO | LL11ss-027-0002-SO | LL11ss-028-0002-SO |
| Date | | 11/17/00 | 11/17/00 | 03/13/01 | 03/13/01 | 03/13/01 | 03/13/01 | 11/10/00 | 11/10/00 | 11/10/00 |
| Depth (ft) | | 4.0 - 6.0 | 4.0 - 6.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 | 6.0 - 8.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 |
| Parameters | | | | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, | | | |
| Analyzed | | | | Misc., Explosives, | Misc., Explosives, | Misc., Explosives, | Misc., Explosives, | | | |
| | | RVAAP Full-suite | RVAAP Full-suite | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyte | Background Criteria | analytes | analytes | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | SVOCs, VOCs | Explosives | Explosives | Explosives |
| | 10500 | 7 410 | T 400 | | Metals (mg/kg) | T 4.60 | L 0.000 | 14000 | T 50 50 | 12000 |
| Aluminum | 19500 | | 7490 | 6990 | 7170 | 7460 | 8690 | 14000 | 7250 | 13900 |
| Antimony | 0.96 | | 0.35 | <0.29U | <0.27U | <0.27U | <0.28U | 0.41 | <0.28U | <0.29U |
| Arsenic | 19.8 | | 11.5 | 15.4 | 9.8 | 22.3* | 23.4* | 10.6 | 18.4 | 20.1* |
| Barium | | | 28.9 | 27 | 25.3 | 28.8 | 34.7 | 49.6 | 19.7 | 48.3 |
| Beryllium | 0.88 | | 0.43 | <0.37U | <0.38U | 0.44 | 0.48 | 0.6 | 0.47 | 0.66 |
| Cadmium | | <0.15U 173 | <0.14U | <0.19U 297 | <0.19U 192 | <0.18U 206 | <0.19U | 0.26* | 0.18* 112 | 0.27* |
| Calcium Chromium | 27.2 | | 186 8.5 | 9.4 | 9.1 | 11.6 | 247 | 898 17 | 9.2 | 319 18.1 |
| Cobalt | | 7.2 | 6 | 5.3 | 5.8 | 8 | 8.6 | 6.2 | 8 | 8.5 |
| Copper | 32.3 | | 18.2 | 16.4 | 19.5 | 21.2 | 21.9 | 17.6 | 21.8 | 17.8 |
| Cyanide | | <0.59U | <0.54U | <0.25U | <0.18U | <0.24U | <0.22U | <0.43U | <0.45U | <0.54U |
| Iron | | 20100 | 15600 | 16200 | 17800 | 22100 | 24700 | 23000 | 20000 | 25800 |
| Lead | 19.1 | 7 | 5.3 | 8.8 | 11.5 | 11.3 | 10.9 | 17.6 | 13 | 13.3 |
| Magnesium | 8790 | 1680 | 1360 | 1350 | 1480 | 2160 | 2500 | 2440 | 1770 | 2680 |
| Manganese | | 248 | 265 | 250 | 245 | 277 | 323 | 306 | 300 | 229 |
| Mercury | | <0.04U | <0.04U | <0.042U | <0.041U | <0.04U | <0.04U | <0.04U | <0.04U | <0.04U |
| Nickel | 60.7 | 14 | 11.4 | 10.9 | 11.6 | 18.6 | 21.5 | 14.9 | 16.4 | 19.4 |
| Potassium | | 888 | 941 | 695 | 707 | 1110 | 1340 | 1400 | 680 | 1530 |
| Selenium | 1.5 | | 0.61 | <0.47U | 0.54 | <0.45U | <0.47U | <0.48U | <0.47U | <0.49U |
| Sodium | | 724* | 562* | 552* | 631* | 873* | 970* | 691* | 618* | 961* |
| Thallium | 0.91 | | 0.14 | <0.19U | <0.18U | <0.18U | <0.19U | <0.19U | <0.19U | <0.2U |
| Vanadium | 37.6 | | 13 | 13.5 | 13.3 | 13.2 | 14.8 | 26.6 | 10.9 | 25 |
| Zinc | 93.3 | | 50.8 | 47.4 | 47.7 | 57.5 | 63.9 | 44.5 | 57.7 | 55.2 |
| | | | | | Anions (mg/kg) | • | • | | | |
| Nitrate | None | <1.2U | <1.1U | <0.083U | <0.08U | 0.41* | 0.54* | <1.2U | <1.1U | <1.2U |
| Sulfate | None | <61.6U | <54.5U | 36.3* | 40.3* | 27.3* | 22.7* | <57.9U | <57.4U | <59.9U |
| Sulfide | None | <28.3U | <29.9U | <9.2U | <8.3U | <9.1U | <8.6U | 30.5* | 30.5* | <29.3U |
| | | | T | T | Miscellaneous (mg/kg) | T | T | | 1 | |
| | | NR | NR | <5.1U | <5U | <5U | <5U | NR | NR | NR |
| TPH-GRO | None | NR | NR | <0.063U | <0.062U | <0.06U | <0.061U | NR | NR | NR |
| 277 | T | | Γ | | plosives/Propellants (mg/ | | T | | 1 | Lim |
| Nitrocellulose | None | 1.1* | 0.87* | 2* | 1.8* | 2* | 1.8* | NR | NR | NR |
| D(.) 41 | N | <0.41II | <0.411 | <0.4111 | SVOCs (mg/kg) | <0.411 | <0.4IT | ND | ND | NID |
| Benz(a)anthracene | None | <0.41U | <0.4U | <0.41U | <0.4U | <0.4U | <0.4U | NR | NR | NR |
| Benzo(a)pyrene | None | <0.41U | <0.4U | <0.41U | <0.4U | <0.4U | <0.4U | NR | NR | NR |
| Benzo(b)fluoranthene | None None | <0.41U | <0.4U <0.4U | <0.41U | <0.4U | <0.4U <0.4U | <0.4U | NR NR | NR NR | NR NR |
| Chrysene Fluoranthene | | <0.41U | | <0.41U | <0.4U | | <0.4U | NR | NR NR | NR NR |
| Phenanthrene | None None | <0.41U <0.41U | <0.4U <0.4U | <0.41U <0.41U | <0.4U <0.4U | <0.4U <0.4U | <0.4U <0.4U | NR NR | NR NR | NR NR |
| Pyrene | None | <0.41U <0.41U | <0.4U | <0.41U <0.41U | <0.4U | <0.4U | <0.4U | NR NR | NR NR | NR NR |
| 1 yrene | TNOHE | \U. \\ 11U | | V.41U | Pesticides/PCBs (mg/kg) | | | INIX | TAIX | INIX |
| PCB-1254 | None | <0.01U | <0.01U | <0.01U | <0.01U | <0.01U | <0.01U | NR | NR | NR |
| 100 1201 | 1.0110 | 3.010 | 1 0.010 | 1 0.010 | VOCs (mg/kg) | 1 0.010 | 1 0.010 | 1 2 22 | 1 111 | 1 2 1 2 2 |
| Acetone | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | NR | NR | NR |
| Carbon tetrachloride | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | NR | NR | NR |
| Toluene | None | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | <0.006U | NR | NR | NR |
| | | 217 T T T | 1 | 1 | 1 | 1 | 1 | | | |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA |
|----------------------|---------------------|-------------------------------|-------------------------------|-------------------------------|-------------------------------|
| Station | 1 | LL11ss-029 | LL11ss-030 | LL11ss-030 | LL11ss-031 |
| Sample ID | 7 | LL11ss-029-0002-SO | LL11ss-030-0002-FD | LL11ss-030-0002-SO | LL11ss-031-0002-SO |
| Date | | 11/14/00 | 11/14/00 | 11/14/00 | 11/10/00 |
| Depth (ft) | 7 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 |
| Parameters Analyzed | 7 | | | | |
| Analyte | Background Criteria | Misc., TAL Metals, Explosives |
| | | | Metals (mg/kg) | | |
| Aluminum | 19500 | 13200 | 13700 | 19000 | 8560 |
| Antimony | | 0.56 | 0.92 | 0.49 | <0.29U |
| Arsenic | 19.8 | 19.3 | 15.7 | 13.2 | 21.7* |
| Barium | 124 | 60.6 | 60.9 | 84.2 | 56.6 |
| Beryllium | 0.88 | 0.6 | 0.66 | 0.89* | 0.64 |
| Cadmium | 0 | <0.18U | <0.19U | <0.19U | 0.26* |
| Calcium | 35500 | 548 | 642 | 872 | 2090 |
| Chromium | 27.2 | | 18.7 | 24.7 | 13.3 |
| Cobalt | 23.2 | | 7.7 | 9.8 | 8.4 |
| Copper | 32.3 | 19.8 | 17 | 19.9 | 16.5 |
| Cyanide | 0 | 1 | <0.52U | 1* | <0.56U |
| Iron | 35200 | | 26900 | 30700 | 22300 |
| Lead | 19.1 | 14.5 | 14.9 | 15.1 | 21.6* |
| Magnesium | | 2690 | 3400 | 4340 | 2190 |
| Manganese | 3030 | | 159 | 191 | 527 |
| Mercury | 0.044 | | <0.04U | <0.04U | <0.04U |
| Nickel | 60.7 | 17.4 | 21.3 | 25.8 | 17.5 |
| Potassium | 3350 | 1210 | 1370 | 2520 | 1110 |
| Selenium | 1.5 | 1 | <0.48U | 0.52 | <0.48U |
| Sodium | 145 | 706* | 517* | 793* | 734* |
| Thallium | 0.91 | 0.22 | 0.21 | 0.21 | <0.19U |
| Vanadium | | 27.3 | 23.3 | 31.5 | 17.3 |
| Zinc | 93.3 | 44.5 | 50 | 61.1 | 55 |
| Zine | 75.5 | | Anions (mg/kg) | 01.1 | 33 |
| Nitrate | None | <0.59U | <0.57U | <0.53U | <1.2U |
| Sulfate | None | <58.9U | <57.1U | <52.7U | <57.8U |
| Sulfide | None | 76.2* | 81.8* | 81.8* | 33.3* |
| 2511140 | 1.010 | | rellaneous (mg/kg) | 02.0 | |
| TPH-DRO | None | NR | NR | NR | NR |
| TPH-GRO | None | NR | NR | NR | NR |
| TITI GRE | 110110 | | es/Propellants (mg/kg) | Titt | 1112 |
| Nitrocellulose | None | NR | NR | NR | NR |
| Titaloccitatose | 110110 | | SVOCs (mg/kg) | Titt | Tite |
| Benz(a)anthracene | None | NR | NR | NR | NR |
| Benzo(a)pyrene | None | NR | NR | NR | NR |
| Benzo(b)fluoranthene | None | NR | NR | NR | NR |
| Chrysene | None | NR | NR NR | NR | NR |
| Fluoranthene | None | NR | NR | NR | NR |
| Phenanthrene | None | NR | NR NR | NR | NR |
| Pyrene | None | NR | NR NR | NR | NR |
| 1 yronc | TAORE | | rides/PCBs (mg/kg) | 1111 | IVIX |
| PCB-1254 | None | NR Testic | NR | NR | NR |
| 1 CD-1234 | TAULIC | 1117 | 1117 | 1117 | 1117 |

Table 4–7. Analytes Detected in Phase I RI Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA |
|----------------------|---------------------|-------------------------------|-------------------------------|-------------------------------|-------------------------------|
| Station | | LL11ss-029 | LL11ss-030 | LL11ss-030 | LL11ss-031 |
| Sample ID | | LL11ss-029-0002-SO | LL11ss-030-0002-FD | LL11ss-030-0002-SO | LL11ss-031-0002-SO |
| Date | | 11/14/00 | 11/14/00 | 11/14/00 | 11/10/00 |
| Depth (ft) | | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 | 1.0 - 3.0 |
| Parameters Analyzed | | | | | |
| Analyte | Background Criteria | Misc., TAL Metals, Explosives |
| | | 1 | VOCs (mg/kg) | | |
| Acetone | None | NR | NR | NR | NR |
| Carbon tetrachloride | None | NR | NR | NR | NR |
| Toluene | None | NR | NR | NR | NR |

FPA = Former Production Area.

ft = Feet.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram. PCB = Polychlorinated biphenyl.

NPA= Non-production area.

NR = Not reported/not analyzed.

RI = Remedial investigation.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

TAL = Target analyte list.

TPH-GRO/DRO = Total petroleum hydrocarbons-gasoline range organics/diesel range organics.

U = Not detected.

VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

Table 4–8. Analytes Detected in Phase I RI Surface Water Samples

| Aggregate | | East Ditch EU | East Ditch EU | West Ditch EU | West Ditch EU | West Ditch EU | West Ditch EU |
|---|---------------------|---------------------|-------------------|---------------------|---------------|---------------------|---------------|
| Station | 1 | LL11sw-012 | LL11sw-012 | LL11sw-014 | LL11sw-014 | LL11sw-015 | LL11sw-015 |
| Station | 1 | LL11sw-012- | LL11sw-012- | LL11sw-014- | LL11sw-014- | LL11sw-015- | LL11sw-015- |
| Sample ID | | 0001-FD | 0001-SW | 0001-SW | 0001b-SW | 0001-SW | 0001b-SW |
| Date | - | 11/17/00 | 11/17/00 | 11/20/00 | 02/07/01 | 11/20/00 | 02/07/01 |
| Parameters Analyzed | - | RVAAP Full-suite | RVAAP Full- | RVAAP Full-suite | 02/07/01 | RVAAP Full-suite | 02/07/01 |
| Parameters Analyzeu | - | analytes + Nitrate, | suite analytes + | analytes + Nitrate, | | analytes + Nitrate, | |
| | | Sulfate, and | Nitrate, Sulfate, | Sulfate, and | | Sulfate, and | TAL Metals, |
| Analyte | Background Criteria | Sulfide Sulfide | and Sulfide | Sulfide | Explosives | Sulfide | Explosives |
| Analyte | background Criteria | Sumae | | | Explosives | Sunde | Explosives |
| | | 0.00077 | Metals (mg/l | | 1 | | T 2.75 |
| Antimony | 0 | | <0.003U | <0.003U | NR | 0.0059* | NR |
| Barium | 0.0475 | 0.0252 | 0.0258 | 0.0416 | NR | 0.0235 | NR |
| Calcium | 41.4 | 75* | 76.8* | 85.4* | NR | 67.2* | NR |
| Iron | 2.56 | <0.05U | <0.05U | <0.05U | NR | 0.0587 | NR |
| Magnesium | 10.8 | 24.7* | 25.4* | 21* | NR | 15.5* | NR |
| Manganese | 0.391 | 0.0139 | 0.0149 | 0.0181 | NR | 0.15 | NR |
| Potassium | 3.17 | 1.88 | 1.92 | 1.67 | NR | 1.24 | NR |
| Sodium | 21.3 | 13.5 | 13.7 | 14.7 | NR | 9.05 | NR |
| Zinc | 0.042 | <0.01U | <0.01U | 0.0115 | NR | 0.0155 | NR |
| | | | Anions (mg/ | L) | | 1 | • |
| Sulfate | None | 46.7* | 46.2* | 50.6* | NR | 32* | NR |
| | | | SVOCs (mg/ | L) | | • | |
| Bis(2-ethylhexyl)phthalate | None | <0.01U | <0.01U | <0.01U | NR | 0.012* | NR |
| \ J J/I | 1 | | VOCs (mg/L | 2) | | ı | |
| Trichloroethene | None | 0.0006* | 0.002* | <0.0005U | NR | <0.0005U | NR |
| EU = Exposure unit. | | | ***** | | 1 | | 1 |
| ID = Identification. | | | | | | | |
| mg/L = Milligrams per liter | | | | | | | |
| NR = Not reported/not analyzed. | | | | | | | |
| RI = Remedial Investigation | | | | | | | |
| RVAAP = Ravenna Army Ammu | | | | | | | |
| SVOC = Semi-volatile organic co | ompound. | | | | | | |
| TAL = Target analyte list. | | | | | | | |
| U = Not detected. VOC = Volatile organic compour | nd | | | | | | |
| v OC – v otatile organic compour | iu. | | | | | | |

VOC = Volatile organic compound.

^{* =} Result exceeds background criteria or no background criteria was available.

< = Less than.

Table 4–9. Analytes Detected in Phase I RI Sediment Samples

| Aggregate | | East Ditch EU | East Ditch EU | Sewer Outfall EU | West Ditch EU | West Ditch EU |
|---------------------|---------------------|---------------------------|---------------------------|--------------------|--------------------|--------------------|
| Station | | LL11sd-017 | LL11sd-017 | LL11sd-021 | LL11sd-024 | LL11sd-030 |
| Sample ID | | LL11sd-017-0001-FD | LL11sd-017-0001-SD | LL11sd-021-0001-SD | LL11sd-024-0001-SD | LL11sd-030-0001-SD |
| Date | | 11/15/00 | 11/15/00 | 11/16/00 | 11/16/00 | 11/16/00 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters | | | | | | |
| Analyzed | | | | Misc., TAL Metals, | Misc., TAL Metals, | Misc., TAL Metals, |
| Analyte | Background Criteria | RVAAP Full-suite analytes | RVAAP Full-suite analytes | Explosives | Explosives | Explosives |
| | | | Metals (mg/ | kg) | | |
| Aluminum | 13900 | | 10700 | 4300 | 15400* | 21400* |
| Antimony | 0 | 0.42* | <0.4U | <0.34U | 0.9* | <0.88U |
| Arsenic | 19.5 | 11.9 | 21.1* | 12.8 | 19.5 | 11.9 |
| Barium | 123 | 89.3 | 90.6 | 48 | 111 | 159* |
| Beryllium | 0.38 | 0.61* | 0.69* | <0.51U | 1* | <1.3U |
| Cadmium | 0 | <0.26U | <0.27U | <0.26U | <0.24U | <0.65U |
| Calcium | 5510 | 2630 | 2730 | 2960 | 2520 | 6490* |
| Chromium | 18.1 | 14 | 15.5 | 7.1 | 21* | 26.2* |
| Cobalt | 9.1 | 9.5* | 9.5* | 5.3 | 14.1* | 11.5* |
| Copper | 27.6 | 15.3 | 15.9 | 13.4 | 21.4 | 32.7* |
| Iron | 28200 | 20700 | 23300 | 12100 | 27800 | 24900 |
| Lead | 27.4 | 26.8 | 39.8* | 19.3 | 19.7 | 47.6* |
| Magnesium | 2760 | 2110 | 2160 | 1340 | 3630* | 4260* |
| Manganese | 1950 | 1980* | 2020* | 1760 | 530 | 311 |
| Nickel | 17.7 | 13.4 | 14 | 11.1 | 27.7* | 31.1* |
| Potassium | 1950 | 1020 | 1110 | 521 | 2010* | 2240* |
| Selenium | 1.7 | 1.9* | <0.67U | <0.64U | 0.66 | <1.6U |
| Sodium | 112 | <129U | <133U | 396* | 1060* | 1450* |
| Thallium | 0.89 | <0.26U | <0.27U | <0.22U | <0.22U | <0.59U |
| Vanadium | 26.1 | 19.9 | 22.2 | 8.9 | 30.8* | 35.6* |
| Zinc | 532 | 63.4 | 64.3 | 47.2 | 79.5 | 202 |
| | | | Anions | | | |
| Sulfate | None | | 252* | 248* | <83.6U | 4220* |
| Sulfide | None | 137* | 210* | 540* | 43.4* | 431* |
| | | | Explosives (m | | | |
| Nitrocellulose | None | 1.9* | 1.8* | NR | NR | NR |
| | | | Pesticides/PCBs | | | |
| PCB-1254 | None | 0.037* | 0.056* | NR | NR | NR |
| | | | VOCs (mg/l | | | |
| Acetone | None | 0.028* | 0.03* | NR | NR | NR |
| EU = Exposure unit. | · | | | | | |

EU = Exposure unit. ft = Feet.

ID = Identification.

ID = Identification.

mg/kg = Milligrams per kilogram.

NR = Not reported/not analyzed.

PCB = Polychlorinated biphenyl.

RI = Remedial Investigation

RVAAP = Ravenna Army Ammunition Plant.

TAL = Target analyte list.

U = Not detected.

^{* =} Result exceeds background criteria or no background criteria was available.

< = Less than.

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Table 4-10. Human Health COPCs per the Phase I RI

| Surface Soil (0-1 ft bgs) | Subsurface Soil (>1 ft bgs) | Sediment | Surface Water | Groundwater |
|------------------------------|--------------------------------|----------------|-------------------|-------------------|
| Aluminum | Aluminum | Aluminum | Antimony | Arsenic |
| Arsenic | Arsenic | Arsenic | Manganese | Bis(2-ethylhexyl) |
| Manganese | Barium | Manganese | Bis(2-ethylhexyl) | phthalate |
| Vanadium | Thallium | Vanadium | phthalate | Trichloroethene |
| Benzo(a)pyrene | Aroclor-1254 | Benzo(a)pyrene | | |
| | Benzo(a)pyrene | | | |

Source: Report for the Remedial Investigation at Load Line 11 (AOC 44) (MKM 2005a).

Sediment and surface water samples collected during the investigation associated with infrastructure (i.e., sumps, basins, and sewers) are not included in the summary table.

bgs = Below ground surface.

ft = Feet.

RI = Remedial Investigation.

Table 4-11. Summary of Ecological Risk Calculation Results per the Phase I RI

| Terrestrial Plants and Soil Invertebrates | Benthic | Aquatic | Terrestrial Wildlife: | Terrestrial Wildlife: |
|---|----------------------|----------------------|---|---------------------------|
| | Invertebrates | Organisms | Insectivores/Herbivores | Carnivores and Piscivores |
| No COPCs retained | No COPCs retained | No COPCs retained | Arsenic Aroclor-1254 Lead Chromium Zinc | Aroclor-1254 |

Source: Report for the Remedial Investigation at Load Line 11 (AOC 44) (MKM 2005a).

Sediment and surface water samples collected during the investigation associated with infrastructure (i.e., sumps, basins, and sewers) and are not included in the summary.

COPC = Chemical of potential concern.

RI = Remedial investigation.

Table 4-12. Chemicals Detected at Concentrations above Screening Criteria in Previous Investigations

| Surface Soil | Subsurface Soil | Sediment | Surface Water |
|------------------------|-----------------|-----------|----------------------------|
| Aluminum | Aluminum | Aluminum | Antimony |
| Arsenic | Arsenic | Arsenic | Manganese |
| Chromium | Barium | Barium | Bis(2-ethylhexyl)phthalate |
| Cobalt | Cobalt | Chromium | Trichloroethene |
| Manganese | Thallium | Manganese | |
| Thallium | PCB-1254 | | |
| Benz(a)anthracene | Benzo(a)pyrene | | |
| Benzo(a)pyrene | | | |
| Benzo(b)fluoranthene | | | |
| Indeno(1,2,3-cd)pyrene | | | |

Note: This table was generated using data from the Report for the Remedial Investigation at Load Line 11 (AOC 44) (MKM 2005a).

PCB = Polychlorinated biphenyl.

Table 4-13. PBA08 RI Surface Soil Samples and Rationales

| | | | I | | | | |
|---------------------|---|--|--------|------------|------|---------------------|------|
| PBA08 RI Station | Targeted Area | Purpose | Metals | Explosives | VOCs | Pesticides/ PCBs | SVOC |
| LL11ss-070 | Adjacent to Earthen Barricade associated with Building AP-18 (Percussion Elements Storage) | Characterize an area not previously sampled. | Y | Y | N | PCBs | РАН |
| LL11ss-071 | Building AP-9 (Shipping Building) | Characterize an area not previously sampled. | Y | Y | N | PCBs | PAH |
| LL11ss-074 | Building AP-14 (Change House) | Delineate previously identified contamination. | Y | Y | N | PCBs | PAH |
| LL11ss-076 | LL11ss-022; Drainage ditch near Buildings | Delineate previously identified contamination. | Y | Y | N | PCBs | PAH |
| LL1188-070 | AP-1 and AP-3. | QA/QC | Y | Y | N | PCBs | PAH |
| | | | Y | Y | N | PCBs | PAH |
| LL11ss-077 | Drainage ditch exiting AOC in north-west portion of NPA | Characterize an area not previously sampled. | Y | Y | N | PCBs | PAH |
| LL11ss-078 | Drainage ditch in north-central portion of NPA across road from Building AP-10 (downgradient of LL11sb-006) | Delineate previously identified contamination. | Y | Y | N | PCBs | РАН |
| LL11ss-079 | Northern portion of NPA, upgradient of FPA | Characterize an area not previously sampled. Analyzed for RVAAP full-suite analytes. | Y | Y | Y | Y | Y |
| LL11ss-080 | Downgradient of IRA excavation area in north-east portion of NPA | Characterize an area not previously sampled. | Y | Y | N | PCBs | PAH |
| LL11ss-081 | Confluence of drainage ditches in northeastern portion of NPA | Characterize an area not previously sampled. | Y | Y | N | PCBs | PAH |

FPA = Former production area.

IRA = Interim removal action.

NPA = Non-production area.

PAH = Polycyclic aromatic hydrocarbon.

PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

PCB = Polychlorinated biphenyl.

QA = Quality assurance.

QC = Quality control.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

VOC = Volatile organic compound.

Table 4–14. Analytes Detected in PBA08 RI Discrete Surface Soil Samples

| | 1 | | | <u> </u> | | screte surface son samp | | | T | |
|-----------------------|-----------------------|---------------------|--------------------|---------------------------|--------------------|-------------------------|------------------|--------------------|--------------------|--------------------|
| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
| Station | | LL11sb-060 | LL11sb-061 | LL11sb-062 | LL11sb-062 | LL11sb-064 | LL11ss-071 | LL11ss-072 | LL11ss-073 | LL11ss-074 |
| Sample ID | | LL11sb-060-5551-SO | LL11sb-061-5555-SO | LL11sb-062-6189-FD | LL11sb-062-5559-SO | LL11sb-064-5569-SO | | LL11ss-072-5598-SO | LL11ss-073-5599-SO | LL11ss-074-5600-SO |
| Date | | 03/18/10 | 03/17/10 | 03/18/10 | 03/18/10 | 03/18/10 | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed a | | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | | | TAL Metals, |
| | n | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | GI · | CI. | Explosives, |
| A T4 - | Background | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, SVOCs | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Chromium | Chromium | Pesticides/PCBs, |
| Analyte | Criteria ^b | SVOCs | SVOCs | SVOCS | SVOCs | SVOCs | SVOCs | Speciation | Speciation | SVOCs |
| A 1 | 17700 | 1.4000 | 0150 | 11200 | Metals (mg/kg) | 789 | 11100 | ND | ND | 0240 |
| Aluminum | 17700 0.96 | 14800 0.099J | 8150 0.081J | 11300 | 11200 | | 11100 | NR NR | NR NR | 9340 |
| Antimony | | 13.3J | | 0.12J 10.2J | 0.1J 10.6J | 0.08J | 0.1J 11.2 | NR NR | NR NR | 0.12J |
| Arsenic | 15.4 | | 12.1J | | | 2.7J | 57.7 | | NR NR | 11.1 |
| Barium | 88.4 | 57.5J 0.65 | 45.5J 0.49 | 66.1J | 73.8J | 14.6J | 0.56 | NR NR | NR NR | 0.48 |
| Beryllium | 0.88 | 0.03 0.21J* | 0.49 0.068J* | 0.55 0.26J* | 0.58 0.24J* | 0.13 0.053J* | 0.30 0.13J* | NR NR | NR NR | 0.48 0.17J* |
| Calaine | 15800 | 2000 | 2300J | 2220 | 2030 | 1690J | 1450 | NR NR | NR NR | 2630 |
| Calcium Chromium | 17.4 | 18* | 12.1 | 13.5 | 13.5 | 3.5 | 14.9 | 10.1 | 17.1 | 12.9 |
| Chromium, hexavalent | 1 / .4 | NR | NR | NR | NR | NR | NR | 0.44J* | <1U | NR |
| Cobalt | 10.4 | 8.3J | 9.4 | 10.5J* | 13J* | 2.2 | 10.4 | NR | NR | 8.5 |
| Copper | 17.7 | 16.6 | 17.5 | 12.6 | 11 | 2.7 | 16.2J | NR NR | NR | 16.4J |
| Iron | 23100 | 30100* | 23900* | 20600 | 22300 | 11400 | 23500* | NR NR | NR | 21800 |
| Lead | 26.1 | 20.5J | 15.3J | 23.4J | 21.4J | 3.9J | 17 | NR NR | NR | 17.5 |
| Magnesium | 3030 | 2270 | 2880J | 2210 | 2140 | 121J | 2670 | NR NR | NR | 2600 |
| Manganese | 1450 | 257 | 263 | 887 | 1320 | 174 | 525 | NR NR | NR | 302 |
| Mercury | 0.036 | 0.077J* | <0.12U | 0.041J* | 0.035J | <0.11U | 0.031J | NR NR | NR | 0.018J |
| Nickel | 21.1 | 19 | 19.6 | 15.7 | 15.4 | 4.5 | 18.9 | NR NR | NR | 18 |
| Potassium | 927 | 852J | 1100* | 804J | 658J | 232 | 803 | NR NR | NR | 748 |
| Selenium | 1.4 | 1.1J | 0.72J | 0.94J | 0.9J | 0.25J | 0.83J | NR | NR NR | 0.84J |
| Silver | 0 | <0.04UJ | 0.023J* | <0.033UJ | <0.036UJ | 0.0046J* | <0.028UJ | NR | NR | <0.027UJ |
| Sodium | 123 | 44.5J | 63.6J | 35.9J | 34.8J | 21.9J | 34.5J | NR | NR | 36.2J |
| Thallium | 0 | 0.19J* | 0.15J* | 0.16J* | 0.17J* | <0.22U | 0.16J* | NR | NR | 0.15J* |
| Vanadium | 31.1 | 25 | 14.8 | 19.7 | 20.8 | 2.9 | 19.9 | NR | NR | 17.2 |
| Zinc | 61.8 | 128* | 52.6J | 55.7 | 49.9 | 20.9J | 52.8J | NR | NR | 54.9J |
| 2 | 01.0 | 120 | 52.00 | 5517 | Explosives (mg/kg) | 20.50 | 52.00 | 1,120 | 1.11 | 0.150 |
| HMX | None | <0.24UJ | <0.24U | <0.25U | <0.24U | <0.24U | <0.26U | NR | NR | <0.24U |
| PETN | None | <0.49U | 0.049J* | <0.5U | <0.48U | <0.48U | <0.52U | NR | | |
| Tetryl | None | <0.24U | 0.02J* | <0.25U | <0.44U | <0.24U | <0.26U | NR | NR | <0.44U |
| Tetryi | None | \0.2 4 0 | U.U2J · | \0.230 | SVOCs (mg/kg) | \0.2 4 0 | \0.200 | NIX | INIX | V0.240 |
| Acenaphthene | None | 0.033* | 0.081* | <0.009U | <0.0088U | <0.0074U | <0.0082U | NR | NR | <0.0078U |
| Acenaphthylene | None | 0.012* | <0.008U | <0.009U | <0.0088U | <0.0074U | <0.0082U | NR | NR | <0.0078U |
| Anthracene | None | 0.049* | 0.16* | <0.009U | <0.0088U | <0.0074U | <0.0082U | NR | NR | <0.0078U |
| Benz(a)anthracene | None | 0.37* | 0.35* | <0.009U | <0.0088U | 0.0089* | <0.0082U | NR | NR | <0.0078U |
| Benzo(a)pyrene | None | 0.45* | 0.31* | 0.0099* | 0.0091* | <0.0074U | <0.0082U | NR | NR | <0.0078U |
| Benzo(b)fluoranthene | None | 0.56* | 0.42* | <0.009U | <0.0088U | 0.014* | <0.0082U | NR NR | NR | 0.011* |
| Benzo(ghi)perylene | None | 0.32* | 0.23* | <0.009U | <0.0088U | 0.012* | <0.0082U | NR | NR | <0.0078U |
| Benzo(k)fluoranthene | None | 0.37* | 0.17* | <0.009U | <0.0088U | <0.0074U | <0.0082U | NR NR | NR | <0.0078U |
| Chrysene | None | 0.47* | 0.31* | <0.009U | <0.0088U | 0.014* | <0.0082U | NR NR | NR | <0.0078U |
| Dibenz(a,h)anthracene | None | 0.076* | 0.072* | <0.009U | <0.0088U | 0.014 | <0.0082U | NR NR | NR | <0.0078U |
| Fluoranthene | None | 1* | 0.9* | 0.017* | 0.014* | 0.022* | 0.00028 | NR | NR | 0.013* |
| Fluorene | None | 0.025* | 0.079* | <0.009U | <0.0088U | <0.0074U | <0.0082U | NR NR | NR | <0.0078U |
| Tuorene | INOILE | 0.023 | 0.079 | \0.0090 | \0.0000U | \0.00/ 4 0 | \0.0002U | INIX | INIX | \ 0.0076 U |

Table 4-14. Analytes Detected in PBA08 RI Discrete Surface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA | FPA |
|------------------------|-----------------------|--------------------|--------------------|--------------------|------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11sb-060 | LL11sb-061 | LL11sb-062 | LL11sb-062 | LL11sb-064 | LL11ss-071 | LL11ss-072 | LL11ss-073 | LL11ss-074 |
| Sample ID | | LL11sb-060-5551-SO | LL11sb-061-5555-SO | LL11sb-062-6189-FD | LL11sb-062-5559-SO | LL11sb-064-5569-SO | LL11ss-071-5597-SO | LL11ss-072-5598-SO | LL11ss-073-5599-SO | LL11ss-074-5600-SO |
| Date | | 03/18/10 | 03/17/10 | 03/18/10 | 03/18/10 | 03/18/10 | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed a | | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | | | TAL Metals, |
| | | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | | | Explosives, |
| | Background | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Chromium | Chromium | Pesticides/PCBs, |
| Analyte | Criteria ^b | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | Speciation | Speciation | SVOCs |
| | | | | | SVOCs (mg/kg), continu | ued | | | | |
| Indeno(1,2,3-cd)pyrene | None | 0.27* | 0.2* | <0.009U | <0.0088U | 0.021* | <0.0082U | NR | NR | <0.0078U |
| Naphthalene | None | 0.01* | 0.017* | <0.009U | 0.0089* | <0.0074U | <0.0082U | NR | NR | <0.0078U |
| Phenanthrene | None | 0.37* | 0.62* | 0.011* | <0.0088U | 0.012* | <0.0082U | NR | NR | <0.0078U |
| Pyrene | None | 0.83* | 0.69* | 0.014* | 0.012* | 0.013* | <0.0082U | NR | NR | 0.011* |
| | | · | | · | PCBs (mg/kg) | | · | | | |
| PCB-1254 | None | <0.046U | 0.42* | <0.045U | <0.043U | 0.02J* | <0.041U | NR | NR | 0.02J* |

Table 4-14. Analytes Detected in PBA08 RI Discrete Surface Soil Samples (continued)

| | 1 | 77D.4 | | | | Surface Son Samples (Co | | NTD (| T 3754 | NTD 4 |
|-----------------------|-------------------------------------|--------------------|---------------------------|---------------------------|--------------------------------|---------------------------|---------------------------|---------------------------|--------------------|---------------------------|
| Aggregate | | FPA | FPA | FPA | NPA | NPA | NPA | NPA | NPA | NPA |
| Station | | LL11ss-075 | LL11ss-076 | LL11ss-076 | LL11sb-063 | LL11sb-065 | LL11sb-066 | LL11sb-067 | LL11sb-068 | LL11sb-069 |
| Sample ID | | LL11ss-075-5601-SO | LL11ss-076-6183-FD | LL11ss-076-5602-SO | LL11sb-063-5563-SO | LL11sb-065-5573-SO | LL11sb-066-5577-SO | | LL11sb-068-5585-SO | LL11sb-069-5589-SO |
| Date | | 04/12/10 | 04/12/10 | 04/12/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/18/10 | 03/18/10 | 03/22/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed a | | | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | | TAL Metals, |
| | Do alsanoun d | Charamia | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | DVA AD Fall and | Explosives, |
| Amolysto | Background Criteria ^b | Chromium | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | RVAAP Full-suite | Pesticides/PCBs, SVOCs |
| Analyte | Criteria | Speciation | SVOCS | SVOCS | | SVOCS | SVUCS | SVUCS | analytes | SVOCS |
| Aluminum | 17700 | NR | 13500 | 13000 | <i>Metals (mg/kg)</i> 8570J | 12900J | 10200 | 8740 | 8920 | 6960 |
| | 0.96 | NR NR | 0.15J | 0.26J | 0.1J | 0.087J | <0.7UJ | 0.099J | 0.091J | <0.7UJ |
| Antimony Arsenic | 15.4 | NR | 18.7* | 26.4* | 9.3J | 10.3J | 9.6 | 10.1J | 12.3J | 7.6 |
| Barium | 88.4 | NR | 46.1 | 44.2 | 64.5 | 45.5 | 69.5J | 57.7J | 55.9J | 50.4J |
| Beryllium | 0.88 | NR | 0.5 | 0.45 | 0.65 | 0.41 | 0.48 | 0.49 | 0.5 | 0.51 |
| Cadmium | 0.88 | NR | 0.08J* | 0.43 0.071J* | 0.24J* | 0.41 0.073J* | 0.13J* | 0.14J* | 0.15J* | 0.31 |
| Calcium | 15800 | NR | 1120 | 1030 | 8480 | 446 | 180J | 2250J | 1260J | 1870J |
| Chromium | 17.4 | 15.4 | 16.6 | 15.4 | 11.3 | 14.4 | 11.9 | 11.8 | 11.6 | 9.3 |
| Chromium, hexavalent | 17.7 | 0.71J* | NR | NR | NR | NR | NR | NR | NR | NR |
| Cobalt | 10.4 | NR | 6.1 | 6 | 7.3 | 6.4 | 9.5 | 8.2 | 9.1 | 5.4 |
| Copper | 17.7 | NR | 18J* | 17.9J* | 14 | 10.7 | 6.3 | 11.9 | 14 | 10.4 |
| Iron | 23100 | NR | 31900* | 26200* | 17100 | 21200 | 19700 | 19700 | 20900 | 15100 |
| Lead | 26.1 | NR | 14.7 | 14.2 | 27.2* | 11.6 | 17 | 16.9J | 14.8J | 16.3 |
| Magnesium | 3030 | NR | 2430 | 2490 | 3060J* | 2210J | 1430 | 2380J | 2250J | 1570 |
| Manganese | 1450 | NR | 171 | 136 | 748 | 243 | 1100 | 543 | 608 | 247 |
| Mercury | 0.036 | NR | 0.034J | 0.039J* | 0.041J* | 0.043J* | 0.053J* | <0.12U | 0.023J | 0.054J* |
| Nickel | 21.1 | NR | 16.2 | 16.3 | 14.7 | 12.9 | 9 | 15.6 | 17.9 | 12.8 |
| Potassium | 927 | NR | 662 | 675 | 721 | 635 | 512 | 660 | 709 | 549 |
| Selenium | 1.4 | NR | 0.88J | 0.81J | 1.1J | 0.89J | 1 | 0.72J | 0.75J | 1 |
| Silver | 0 | NR | <0.024UJ | <0.024UJ | 0.025J* | <0.017UJ | 0.05J* | 0.024J* | <0.025UJ | 0.036J* |
| Sodium | 123 | NR | 33.2J | 36J | 74.3J | 30.4J | 30.3J | 39.4J | 39.2J | 33J |
| Thallium | 0 | NR | 0.16J* | 0.16J* | 0.13J* | 0.13J* | 0.18J* | 0.14J* | 0.16J* | 0.1J* |
| Vanadium | 31.1 | NR | 24.4 | 24 | 15 | 24.4 | 22.3 | 16.3 | 16.1 | 15.1 |
| Zinc | 61.8 | NR | 40.8J | 38.8J | 64.3* | 46.1 | 38.7 | 46.9J | 56J | 67.6* |
| | | | | | Explosives (mg/kg) | | | | | |
| HMX | None | NR | <0.25U | <0.24U | <0.26U | <0.26U | <0.25U | <0.24U | <0.25U | <0.24U |
| PETN | None | NR | <0.5U | <0.48U | <0.51U | <0.51U | <0.5U | 0.036J* | <0.5U | <0.48U |
| Tetryl | None | NR | <0.25U | <0.24U | <0.26U | <0.26U | <0.25U | <0.24U | <0.25U | <0.24U |
| | | | V-20- | * | SVOCs (mg/kg) | V.2.V.2 | V | * | V.=V- | V.2.12 |
| Acenaphthene | None | NR | <0.0083U | <0.0085U | <0.0098U | <0.0086U | <0.0093U | 0.065* | <0.061U | <0.0094U |
| Acenaphthylene | None | NR | <0.0083U | <0.0085U | <0.0098U | <0.0086U | <0.0093U | <0.008U | <0.061U | <0.0094U |
| Anthracene | None | NR | <0.0083U | <0.0085U | 0.011* | <0.0086U | <0.0093U | 0.13* | <0.061U | <0.0094U |
| Benz(a)anthracene | None | NR | 0.018* | 0.0096* | 0.048* | <0.0086U | <0.0093U | 0.32* | 0.015J* | 0.046* |
| Benzo(a)pyrene | None | NR | 0.022* | 0.011* | 0.049* | <0.0086U | <0.0093U | 0.28* | 0.014J* | 0.05* |
| Benzo(b)fluoranthene | None | NR | 0.025* | 0.016* | 0.085* | <0.0086U | <0.0093U | 0.41* | 0.022J* | 0.081* |
| Benzo(ghi)perylene | None | NR | 0.019* | <0.0085U | 0.041* | <0.0086U | <0.0093U | 0.21* | 0.014J* | 0.034* |
| Benzo(k)fluoranthene | None | NR | <0.0083U | <0.0085U | 0.036* | <0.0086U | <0.0093U | 0.13* | 0.0094J* | 0.037* |
| Chrysene | None | NR | 0.017* | 0.0097* | 0.054* | <0.0086U | <0.0093U | 0.28* | 0.016J* | 0.056* |
| Dibenz(a,h)anthracene | None | NR | <0.0083U | <0.0085U | <0.0098U | <0.0086U | <0.0093U | 0.069* | <0.061U | <0.0094U |
| Fluoranthene | None | NR | 0.03* | 0.016* | 0.1* | <0.0086U | 0.011* | 0.78* | 0.034J* | 0.12* |
| Fluorene | None | NR | <0.0083U | <0.0085U | <0.0098U | <0.0086U | <0.0093U | 0.049* | <0.061U | <0.0094U |
| <u> </u> | | | | | | | | | | |

Table 4-14. Analytes Detected in PBA08 RI Discrete Surface Soil Samples (continued)

| Aggregate | | FPA | FPA | FPA | NPA | NPA | NPA | NPA | NPA | NPA |
|----------------------------------|-----------------------|--------------------|--------------------|--------------------|------------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11ss-075 | LL11ss-076 | LL11ss-076 | LL11sb-063 | LL11sb-065 | LL11sb-066 | LL11sb-067 | LL11sb-068 | LL11sb-069 |
| Sample ID | | LL11ss-075-5601-SO | LL11ss-076-6183-FD | LL11ss-076-5602-SO | LL11sb-063-5563-SO | LL11sb-065-5573-SO | LL11sb-066-5577-SO | LL11sb-067-5581-SO | LL11sb-068-5585-SO | LL11sb-069-5589-SO |
| Date | | 04/12/10 | 04/12/10 | 04/12/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/18/10 | 03/18/10 | 03/22/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed ^a | | | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | | TAL Metals, |
| | | | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | | Explosives, |
| | Background | Chromium | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | RVAAP Full-suite | Pesticides/PCBs, |
| Analyte | Criteria ^b | Speciation | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | analytes | SVOCs |
| | | | | | SVOCs (mg/kg), continu | ued | | | | |
| Indeno(1,2,3-cd)pyrene | None | NR | 0.015* | <0.0085U | 0.028* | <0.0086U | <0.0093U | 0.19* | 0.011J* | 0.03* |
| Naphthalene | None | NR | <0.0083U | <0.0085U | 0.01* | <0.0086U | <0.0093U | 0.0094* | <0.061U | <0.0094U |
| Phenanthrene | None | NR | 0.015* | <0.0085U | 0.046* | <0.0086U | <0.0093U | 0.52* | 0.017J* | 0.058* |
| Pyrene | None | NR | 0.022* | 0.013* | 0.096* | <0.0086U | 0.01* | 0.59* | 0.025J* | 0.11* |
| | | | | | PCBs (mg/kg) | | | | | |
| PCB-1254 | None | NR | <0.041U | <0.042U | <0.049U | <0.043U | <0.046UJ | <0.039U | <0.041U | <0.046UJ |

Table 4-14. Analytes Detected in PBA08 RI Discrete Surface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA |
|----------------------------------|-----------------------|-------------------------|-------------------------|-------------------------|---------------------------|-------------------------|-------------------------|
| Station | | LL11ss-070 | LL11ss-077 | LL11ss-078 | LL11ss-079 | LL11ss-080 | LL11ss-081 |
| Sample ID | | LL11ss-070-5596-SO | LL11ss-077-5603-SO | LL11ss-078-5604-SO | LL11ss-079-5605-SO | LL11ss-080-5606-SO | LL11ss-081-5607-SO |
| Date | | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed ^a | | TAL Metals, Explosives, | TAL Metals, Explosives, | TAL Metals, Explosives, | | TAL Metals, Explosives, | TAL Metals, Explosives, |
| Analyte | Background Criteria b | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | RVAAP Full-suite analytes | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs |
| Aluminum | 17700 | 7900 | 12400 | 9350 | 8070 | 11100 | 11400 |
| Antimony | 0.96 | 0.74J | <0.62UJ | 0.16J | 0.084J | <0.72UJ | 0.12J |
| Arsenic | 15.4 | 40.4* | 10.2 | 10.8 | 9.7 | 10.4 | 6.5 |
| Barium | 88.4 | 65.8 | 57.6 | 60.1 | 26.8 | 85.6 | 128* |
| Beryllium | 0.88 | 0.59 | 0.36 | 0.55 | 0.36 | 0.76 | 0.69 |
| Cadmium | 0 | 0.17J* | 0.057J* | 0.92* | 0.092J* | 0.34* | 0.45* |
| Calcium | 15800 | 10100 | 859 | 2650 | <113UJ | 1770 | 2800 |
| Chromium | 17.4 | 12.1 | 14.5 | 13.4 | 9.5 | 12.8 | 15.4 |
| Chromium, hexavalent | | NR | NR | NR | NR | NR | NR |
| Cobalt | 10.4 | 12.7* | 8.4 | 8.4 | 5.9 | 9.2 | 7.6 |
| Copper | 17.7 | 19.6J* | 15.5J | 17.3J | 11.2J | 10.2J | 18.1J* |
| Iron | 23100 | 26300* | 20100 | 21700 | 16200 | 20600 | 16500 |
| Lead | 26.1 | 29.5* | 9.9 | 23.4 | 12.6 | 17.8 | 33.9* |
| Magnesium | 3030 | 2750 | 2230 | 2560 | 1430 | 2060 | 2650 |
| Manganese | 1450 | 1030 | 393 | 337 | 239 | 506 | 445 |
| Mercury | 0.036 | 0.044J* | <0.12U | 0.031J | 0.02J | 0.037J* | 0.04J* |
| Nickel | 21.1 | 26.2* | 16.7 | 19.4 | 11.2 | 16.8 | 20.4 |
| Potassium | 927 | 588 | 967* | 912 | 437 | 788 | 908 |
| Selenium | 1.4 | 0.91J | 0.78J | 0.94J | 0.7J | 0.88J | 1.2J |
| Silver | 0 | <0.029UJ | 0.032J* | 0.035J* | <0.021UJ | 0.053J* | 0.06J* |
| Sodium | 123 | 85.9J | 35.2J | 40.6J | 22.5J | 38.5J | 48.2J |
| Thallium | 0 | 0.15J* | 0.21J* | 0.16J* | 0.11J* | 0.16J* | 0.21J* |
| Vanadium | 31.1 | 14.4 | 16.5 | 17 | 13.7 | 21.4 | 19.1 |
| Zinc | 61.8 | 57.8J | 45.6J | 478J* | 40.4J | 78.6J* | 11 7 J* |
| HMX | None | <0.24U | <0.24U | <0.25U | <0.24U | <0.26U | 0.013J* |
| PETN | None | <0.48U | <0.48U | <0.5U | <0.48U | <0.51U | <0.5U |
| Tetryl | None | <0.24U | <0.24U | <0.25U | <0.24U | <0.26U | <0.25U |
| Acenaphthene | None | <0.0079U | <0.0083U | <0.0082U | <0.059U | <0.0096U | 0.012* |
| Acenaphthylene | None | <0.0079U | <0.0083U | <0.0082U | <0.059U | <0.0096U | <0.011U |
| Anthracene | None | <0.0079U | <0.0083U | 0.016* | <0.059U | <0.0096U | 0.018* |
| Benz(a)anthracene | None | 0.042* | <0.0083U | 0.056* | 0.013J* | <0.0096U | 0.089* |
| Benzo(a)pyrene | None | 0.053* | <0.0083U | 0.052* | 0.017J* | <0.0096U | 0.11* |
| Benzo(b)fluoranthene | None | 0.12* | <0.0083U | 0.068* | <0.059U | <0.0096U | 0.15* |
| Benzo(ghi)perylene | None | 0.047* | <0.0083U | 0.039* | 0.013J* | <0.0096U | 0.066* |
| Benzo(k)fluoranthene | None | 0.062* | <0.0083U | 0.036* | <0.059U | <0.0096U | 0.077* |
| Chrysene | None | 0.076* | <0.0083U | 0.059* | 0.018J* | <0.0096U | 0.11* |
| Dibenz(a,h)anthracene | None | <0.0079U | <0.0083U | <0.0082U | <0.059U | <0.0096U | <0.011U |
| Fluoranthene | None | 0.075* | <0.0083U | 0.14* | 0.028J* | 0.02* | 0.26* |
| Fluorene | None | <0.0079U | <0.0083U | <0.0082U | <0.059U | <0.0096U | <0.011U |

Table 4–14. Analytes Detected in PBA08 RI Discrete Surface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA |
|----------------------------------|-----------------------|-------------------------|-------------------------|-------------------------|---------------------------|-------------------------|-------------------------|
| Station | | LL11ss-070 | LL11ss-077 | LL11ss-078 | LL11ss-079 | LL11ss-080 | LL11ss-081 |
| Sample ID | | LL11ss-070-5596-SO | LL11ss-077-5603-SO | LL11ss-078-5604-SO | LL11ss-079-5605-SO | LL11ss-080-5606-SO | LL11ss-081-5607-SO |
| Date | | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 | 04/12/10 |
| Depth (ft) | | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 | 0.0 - 1.0 |
| Parameters Analyzed ^a | | TAL Metals, Explosives, | TAL Metals, Explosives, | TAL Metals, Explosives, | | TAL Metals, Explosives, | TAL Metals, Explosives, |
| Analyte | Background Criteria b | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | RVAAP Full-suite analytes | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs |
| SVOCs (mg/kg), continued | | | | | | | |
| Indeno(1,2,3-cd)pyrene | None | 0.037* | <0.0083U | 0.033* | <0.059U | <0.0096U | 0.067* |
| Naphthalene | None | <0.0079U | <0.0083U | <0.0082U | <0.059U | <0.0096U | <0.011U |
| Phenanthrene | None | 0.025* | <0.0083U | 0.069* | 0.012J* | <0.0096U | 0.11* |
| Pyrene | None | 0.061* | <0.0083U | 0.1* | 0.023J* | 0.016* | 0.19* |
| $PCBs\ (mg/kg)$ | | | | | | | |
| PCB-1254 | None | <0.039U | <0.041U | 0.035J* | <0.039U | <0.047U | <0.056U |

^aOnly detected SRCs are presented in the table.

FPA = Former production area. HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocane.

ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.

NPA = Non-production area.

NR = Not reported/not analyzed.

PBA08 RI = Performance-Based Acquisition 2008 Remedial Investigation.

PCB = Polychlorinated biphenyl.

PETN = Pentaerythritol tetranitrate.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

TAL = Target analyte list.

U = Non-detectable concentration.

^b Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b). ft = Feet.

^{* =} Result exceeds background criteria or no background criteria was available.

Table 4–15. Chromium Speciation Samples under PBA08 RI

| PBA08 RI Location | Rationale for Sample Selection |
|----------------------|--|
| LL11ss-072 | Discrete sample collected to assess chromium speciation. Previous chromium result represents elevated chromium concentration (LL11sb-004 at 20.7 mg/kg) |
| LL11ss-073 | Discrete sample collected to assess chromium speciation. Previous chromium result represents near background chromium concentration (LL11sd-025 at 17.5 mg/kg) |
| LL11ss-075 | Discrete sample collected to assess chromium speciation. Previous chromium result represents elevated chromium concentration (LL11cs-011 at 28.4 mg/kg) |

mg/kg = Milligrams per kilogram.
PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

Table 4–16. Subsurface Soil Rationale and Analyses

| PBA08 RI Location | Comments/Rationale | Sample Type | Depth (ft bgs) | Analyses Performed Metals | Explosives | VOCs | Pesticides/ PCBs | svoc |
|----------------------|---|-----------------------|-------------------|---------------------------------|------------|------|---------------------|------|
| | | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| T T 1 1 1 0 60 | Confirm presence of contamination in | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL11sb-060 | previously sampled area (LL11sb-001) | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | | NA | 7-13 | Y | Y | N | N | N |
| | | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| 11111001 | Further investigation of previously | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL11sb-061 | identified contamination (LL11ss-003 and LL11ss-005, Building AP-8) | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | LL11ss-003, Building AP-8) | NA | 7-13 | N | N | N | N | N |
| | D.F | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| | Delineate vertical and lateral extent of | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| 1111-1-062 | previously identified contamination (LL11sb-004 and LL11sb-017 | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | (LLT180-004 and LLT180-017 | NA | 7-13 | N | N | N | N | N |
| | QA/QC | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| | QA/QC | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| LL11sb-085 | Geotechnical. Water encountered at 6.5 ft | Discrete | 2-3.9 | N | N | N | N | N |
| LL1180-083 | bgs. South of Building AP-11 | Discrete | 4-5.5 | N | N | N | N | N |
| | Confirm presence of contamination in | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| | previously sampled area (LL11ss-024; | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL11sb-063 | LL11cs-018, LL11cs-019); Geoprobe | Discrete | 4-5 | Y | Y | N | PCBs | PAH |
| | could not access location safely over steep ditch slopes; hand auger refusal at 5 ft bgs | NS | 7-13 | N | N | N | N | N |
| | Confirm presence of contamination in previously sampled area (LL11ss-012 and LL11ss-013; Building AP-11 QA/QC | Discrete | 0-1 | Y | Y | N | PCBs | РАН |
| LL11sb-064 | QA/QC | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| | | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | QA/QC | Discrete ^b | 7-13 | Y | Y | N | PCBs | PAH |
| | QA/QC | Discrete | 7-13 | Y | Y | N | PCBs | PAH |
| | | Discrete | 7-13 | Y | Y | N | PCBs | PAH |
| | | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| LL11sb-065 | Confirm presence of contamination in | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL1180-003 | previously sampled area (LL11ss-029) | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | | Discrete ^a | 7-13 | Y | Y | N | PCBs | PAH |

Table 4-16. Subsurface Soil Rationale and Analyses (continued)

| PBA08 RI | | Sample | Depth ft | Analyses Performed | | | Pesticides/ | |
|--------------|--|----------|----------|-----------------------|------------|------|-------------|------|
| Location | Comments/Rationale | Туре | (bgs) | Metals | Explosives | VOCs | PCBs | SVOC |
| | Delineate vertical extent of previously identified contamination (LL11ss- | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| I I 11ah 066 | | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL11sb-066 | 028) | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | 028) | NA | 7-13 | N | N | N | N | N |
| | | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| | Confirm presence of contamination in | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL11sb-067 | previously sampled area (LL11cs- | Discrete | 4-7 | Y | Y | N | PCBs | PAH |
| | 045) | NA | 7-13 | N | N | N | N | N |
| | QA/QC | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| | QA/QC | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| | Delineate previously identified | Discrete | 0-1 | Y | Y | Y | Y | Y |
| | contamination (LL11cs-050, LL11sb- | Discrete | 1-4 | Y | Y | Y | Y | Y |
| LL11sb-068 | 040, and LL11sb-039). Analyzed for | Discrete | 4-7 | Y | Y | Y | Y | Y |
| LL1180-008 | RVAAP full-suite analytes | NA | 7-13 | N | N | N | N | N |
| | QA/QC. Analyzed for RVAAP full- | Discrete | 4-7 | Y | Y | Y | Y | Y |
| | suite analytes | Discrete | 4-7 | Y | Y | Y | Y | Y |
| | Confirm presence of contamination in | Discrete | 0-1 | Y | Y | N | PCBs | PAH |
| | previously sampled area (LL11ss-031, | Discrete | 1-4 | Y | Y | N | PCBs | PAH |
| LL11sb-069 | drainage ditch); Geoprobe could not | Discrete | 4-5 | Y | Y | N | PCBs | PAH |
| a C 1 1 | access location safely over steep ditch slopes; hand auger refusal at 5 ft bgs | NS | 7-13 | N | N | N | N | N |

^a Sample analyzed by the laboratory based on exceedance of preliminary screening criteria of the 4–7 ft sample interval.

bgs = Below ground surface.

ft = Feet.

NA = Sample not analyzed by the laboratory based on preliminary screening criteria results of the 4–7 ft sample interval.

NS = Not sampled due to refusal.

PAH = Polycyclic aromatic hydrocarbon.

PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

PCB = Polychlorinated biphenyl.

QA = Quality assurance.

QC = Quality control.

RVAAP = Ravenna Army and Ammunition Plant.

SVOC = Semi-volatile organic compound.

VOC = Volatile organic compound.

^b One sample (10%) from 7–13 ft was submitted for laboratory analysis to characterize subsurface soil to 13 ft bgs.

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Table 4-17. Analytes Detected in PBA08 RI Subsurface Soil Samples

| Aggregate | | FPA |
|----------------------------------|-----------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|
| Station | | LL11sb-060 | LL11sb-060 | LL11sb-061 | LL11sb-061 | LL11sb-062 | LL11sb-062 | LL11sb-064 | LL11sb-064 | LL11sb-064 |
| Sample ID | | LL11sb-060-5552-SO | LL11sb-060-5553-SO | LL11sb-061-5556-SO | LL11sb-061-5557-SO | LL11sb-062-5560-SO | LL11sb-062-5561-SO | LL11sb-064-5570-SO | LL11sb-064-5571-SO | LL11sb-064-6188-FD |
| Date | | 03/18/10 | 03/18/10 | 03/17/10 | 03/17/10 | 03/18/10 | 03/18/10 | 03/18/10 | 03/18/10 | 03/18/10 |
| Depth (ft) | | 1.0 - 4.0 | 4.0 - 7.0 | 1.0 - 4.0 | 4.0 - 7.0 | 1.0 - 4.0 | 4.0 - 7.0 | 1.0 - 4.0 | 4.0 - 7.0 | 7.0 - 13.0 |
| Parameters Analyzed ^a | | TAL Metals, |
| _ | | Explosives, |
| | Background | Pesticides/PCBs, |
| Analyte | Criteria ^b | SVOCs |
| | | | | | Metals (mg/kg) | | | | | |
| Aluminum | 19500 | 13100 | 6110 | 6950 | 6650 | 16200 | 6550 | 5340 | 5130 | 6110 |
| Antimony | 0.96 | 0.084J | 0.093J | 0.084J | 0.079J | <0.63R | 0.077J | 0.083J | 0.077J | <0.62UJ |
| Arsenic | 19.8 | 15.9J | 16.8J | 15.1J | 15.4 | 11.7J | 18.2J | 12.5J | 13.9 | 11.5J |
| Barium | 124 | 103J | 56.4 | 31.9J | 30 | 65.9J | 30.8 | 28.8J | 25.1 | 26.4J |
| Beryllium | 0.88 | 0.84 | 0.28J | 0.39 | 0.34 | 0.69 | 0.34J | 0.32 | 0.26 | 0.3 |
| Cadmium | 0 | 0.074J* | 0.077J* | 0.044J* | 0.053J* | <0.065UJ | 0.061J* | 0.068J* | 0.064J* | <0.054UJ |
| Calcium | 35500 | 1600 | 1140 | 1870J | 7710 | 1070 | 3970 | 6320J | 6760 | 5080J |
| Chromium | 27.2 | 16.6 | 9 | 10.7 | 10.6 | 19 | 10.3 | 8.5 | 8.4 | 9.5 |
| Cobalt | 23.2 | 27.1J* | 8 | 7.7 | 8.5 | 11.5J | 8.6 | 7.1 | 7.1 | 7.4 |
| Copper | 32.3 | 22.1 | 23.9J | 18.6 | 18.5 | 20.4 | 18J | 16.8 | 18.4 | 16.7 |
| Iron | 35200 | 29000 | 21600 | 22400 | 22600 | 27000 | 22600 | 19800 | 19500 | 18600 |
| Lead | 19.1 | 15.8J | 15.3 | 14.9J | 12.2 | 16.5J | 11.4 | 10.4J | 10 | 8.8J |
| Magnesium | 8790 | 3260 | 2140J | 2600J | 3820 | 3900 | 3040J | 3010J | 3440 | 3740J |
| Manganese | 3030 | 788 | 593 | 260 | 355 | 256 | 361 | 382 | 325 | 279 |
| Mercury | 0.044 | <0.12U | <0.12U | <0.12U | <0.12U | 0.026J | <0.12U | <0.11U | <0.12U | <0.12U |
| Nickel | 60.7 | 30.4 | 17.7J | 18 | 19.1 | 28.8 | 19.2J | 16.7 | 16.8 | 17 |
| Potassium | 3350 | 1050J | 574 | 848 | 953 | 1290J | 807 | 895 | 815 | 1000 |
| Selenium | 1.5 | 1.4J | 0.73J | 0.65J | 0.72 | 0.84J | 0.64J | 0.5J | 0.59J | 0.81J |
| Silver | 0 | <0.035UJ | 0.019J* | 0.019J* | 0.021J* | <0.018UJ | 0.017J* | 0.017J* | 0.019J* | <0.018UJ |
| Sodium | 145 | 63.2J | 34.3J | 64.2J | 56.6J | 48.5J | 41.7J | 46.5J | 34.8J | 61.3J |
| Thallium | 0.91 | 0.19J | 0.12J | 0.14J | 0.13J | 0.18J | 0.12J | 0.11J | 0.11J | 0.11J |
| Vanadium | 37.6 | 22.2 | 10.4 | 12.3 | 11.9 | 22.3 | 11.4 | 10.2 | 9.5 | 10.4 |
| Zinc | 93.3 | 62.5 | 56 | 52.3J | 55.3 | 57.6 | 53 | 55.2J | 51 | 47.7J |
| | | | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Benzo(a)pyrene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Benzo(b)fluoranthene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Benzo(ghi)perylene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Bis(2-ethylhexyl)phthalate | None | NR |
| Chrysene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Di-n-butyl phthalate | None | NR |
| Fluoranthene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Phenanthrene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |
| Pyrene | None | <0.0081U | <0.0079U | <0.0077U | <0.0077U | <0.0084U | <0.0077U | <0.0077U | <0.008U | <0.0083U |

Table 4-17. Analytes Detected in PBA08 RI Subsurface Soil Samples (continued)

| Aggregate | | FPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|----------------------------------|-----------------------|--------------------|--------------------|--------------------|--------------------|--------------------|------------------|--------------------|--------------------|--------------------|
| Station | - - | LL11sb-064 | LL11sb-063 | LL11sb-063 | LL11sb-065 | LL11sb-065 | LL11sb-065 | LL11sb-066 | LL11sb-066 | LL11sb-067 |
| Sample ID | | LL11sb-064-5572-SO | LL11sb-063-5564-SO | LL11sb-063-5565-SO | LL11sb-065-5574-SO | LL11sb-065-5575-SO | | LL11sb-066-5578-SO | LL11sb-066-5579-SO | LL11sb-067-6186-FD |
| Date | - - | 03/18/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/22/10 | 03/18/10 |
| Depth (ft) | - - | 7.0 - 13.0 | 1.0 - 4.0 | 4.0 - 5.0 | 1.0 - 4.0 | 4.0 - 7.0 | 7.0 - 13.0 | 1.0 - 4.0 | 4.0 - 7.0 | 1.0 - 4.0 |
| Parameters Analyzed ^a | - - | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, | TAL Metals, |
| | 1 | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, | Explosives, |
| | Background | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, | Pesticides/PCBs, |
| Analyte | Criteria ^b | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs | SVOCs |
| | | | | | Metals (mg/kg) | | | | | |
| Aluminum | 19500 | 4760 | 8900J | 8210J | 10500 | 7280 | 6880 | 9320 | 6720 | 11700 |
| Antimony | 0.96 | <0.63UJ | <0.61UJ | <0.63UJ | 0.092J | 0.087J | 0.078J | 0.086J | 0.077J | 0.12J |
| Arsenic | 19.8 | 11.5J | 12J | 15.5J | 15.5 | 24.7J* | 14.1 | 16.6 | 15.1J | 15.5J |
| Barium | 124 | 23.9 | 45.2 | 34.9 | 88.9J | 32.4 | 25.4 | 70J | 39.2 | 48.3J |
| Beryllium | 0.88 | 0.25 | 0.45 | 0.43 | 0.63 | 0.46 | 0.41 | 0.53 | 0.41 | 0.5 |
| Cadmium | 0 | 0.065J* | 0.088J* | <0.049UJ | 0.13J* | 0.051J* | 0.039J* | 0.054J* | 0.059J* | 0.082J* |
| Calcium | 35500 | 5850J | 20400 | 6000 | 1340J | 1140 | 4420 | 790J | 1190 | 1720J |
| Chromium | 27.2 | 7.4 | 13.6 | 12.4 | 14.7 | 11.5 | 10.5 | 13.8 | 11.4 | 15 |
| Cobalt | 23.2 | 6.4 | 9.5 | 9.4 | 16.8 | 9.1 | 8.7 | 9.4 | 9.4 | 5.2 |
| Copper | 32.3 | 16.7 | 17 | 17.7 | 21.2 | 24.5 | 15.5J | 19.7 | 18.5 | 18.4 |
| Iron | 35200 | 17300 | 23500 | 25000 | 25800 | 26100 | 20600 | 24600 | 21200 | 27800 |
| Lead | 19.1 | 9.7J | 11.4 | 10.8 | 14.3 | 12.5 | 10.2J | 13.7 | 10.9 | 8.6J |
| Magnesium | 8790 | 3460J | 4710J | 4250J | 3210 | 2540J | 3530J | 2670 | 2540J | 2220J |
| Manganese | 3030 | 282 | 408 | 375 | 867 | 243 | 367 | 388 | 478 | 132 |
| Mercury | 0.044 | <0.13U | <0.12U | <0.13U | <0.12U | 0.025J | <0.12U | 0.045J* | <0.12U | 0.031J |
| Nickel | 60.7 | 14.8 | 22.3 | 21.3 | 33.2 | 22.1 | 19.6J | 23.3 | 24.3 | 13.6 |
| Potassium | 3350 | 748 | 1250 | 1110 | 1080 | 838 | 1370 | 901 | 801 | 585 |
| Selenium | 1.5 | 0.55J | 0.8J | 1.1J | 1.5 | 0.84J | 1 | 1.4 | 0.8J | 0.91J |
| Silver | 0 | 0.017J* | <0.015UJ | <0.02UJ | <0.013UJ | 0.014J* | <0.021UJ | <0.015UJ | 0.011J* | <0.01UJ |
| Sodium | 145 | 53J | 62.2J | 65.8J | 51.2J | 40.3J | 52.5J | 39.7J | 41.6J | 38.9J |
| Thallium | 0.91 | 0.1J | 0.15J | 0.11J | 0.19J | 0.13J | 0.11J | 0.14J | 0.13J | 0.26 |
| Vanadium | 37.6 | 8.7 | 15.3 | 12.7 | 19.8 | 14.1 | 12.1 | 15.3 | 12.1 | 23 |
| Zinc | 93.3 | 46.9J | 60.2 | 54.7 | 51.9 | 54.3 | 52.2 | 54.8 | 54.9 | 39.6J |
| | 1 | T | | | SVOCs (mg/kg) | | | | | |
| Benz(a)anthracene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | <0.0077U | <0.0079U | <0.0079U | <0.008U |
| Benzo(a)pyrene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | <0.0077U | <0.0079U | <0.0079U | <0.008U |
| Benzo(b)fluoranthene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | 0.022* | <0.0079U | <0.0079U | <0.008U |
| Benzo(ghi)perylene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | <0.0077U | <0.0079U | <0.0079U | <0.008U |
| Bis(2-ethylhexyl)phthalate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Chrysene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | <0.0077U | <0.0079U | <0.0079U | <0.008U |
| Di-n-butyl phthalate | None | NR | NR | NR | NR | NR | NR | NR | NR | NR |
| Fluoranthene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | <0.0077U | <0.0079U | <0.0079U | 0.0098* |
| Phenanthrene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | 0.01* | <0.0079U | <0.0079U | <0.008U |
| Pyrene | None | <0.0084U | <0.0082U | <0.0085U | <0.008U | <0.0078U | 0.0094* | <0.0079U | <0.0079U | <0.008U |

Table 4–17. Analytes Detected in PBA08 RI Subsurface Soil Samples (continued)

| Aggregate | | NPA | NPA | NPA | NPA | NPA | NPA | NPA |
|---|-----------------------------------|-------------------------|----------------------------|-----------------------------------|--------------------|--------------------|-------------------------|-------------------------|
| Station | ╡ | LL11sb-067 | LL11sb-067 | LL11sb-068 | LL11sb-068 | LL11sb-068 | LL11sb-069 | LL11sb-069 |
| Sample ID | | LL11sb-067-5582-SO | LL11sb-067-5583-SO | LL11sb-068-5586-SO | LL11sb-068-6187-FD | LL11sb-068-5587-SO | LL11sb-069-5590-SO | LL11sb-069-5591-SO |
| Date | | 03/18/10 | 03/18/10 | 03/18/10 | 03/18/10 | 03/18/10 | 03/22/10 | 03/22/10 |
| Depth (ft) | | 1.0 - 4.0 | 4.0 - 7.0 | 1.0 - 4.0 | 4.0 - 7.0 | 4.0 - 7.0 | 1.0 - 4.0 | 4.0 - 5.0 |
| Parameters Analyzed ^a | | TAL Metals, Explosives, | TAL Metals, Explosives, | RVAAP Full-suite | RVAAP Full-suite | RVAAP Full-suite | TAL Metals, Explosives, | TAL Metals, Explosives, |
| Analyte | Background Criteria b | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs | analytes | analytes | analytes | Pesticides/PCBs, SVOCs | Pesticides/PCBs, SVOCs |
| · | | / | / | Metals (mg/kg) | v I | v | , | , |
| Aluminum | 19500 | 11700 | 6730 | 12700 | 7580 | 6670 | 6840 | 6750 |
| Antimony | 0.96 | 0.081J | 0.082J | 0.11J | 0.076J | 0.082J | <0.6UJ | <0.6UJ |
| Arsenic | 19.8 | 12.5J | 14.4 | 9.7J | 9.5J | 9 | 13.8 | 13.9 |
| Barium | 124 | 46.7J | 38.4 | 48.3J | 26J | 23.9 | 21.6J | 48.3J |
| Beryllium | 0.88 | 0.5 | 0.35 | 0.42 | 0.29 | 0.25 | 0.35 | 0.36 |
| Cadmium | 0 | 0.066J* | 0.051J* | 0.063J* | <0.036UJ | 0.037J* | 0.058J* | 0.06J* |
| Calcium | 35500 | 915J | 205J | 1230J | 459J | 426 | 1650J | 8520J |
| Chromium | 27.2 | 14.8 | 9.6 | 14.9 | 8.5 | 8.1 | 10.9 | 13 |
| Cobalt | 23.2 | 10.6 | 7.5 | 7.4 | 4.5 | 4.6 | 7 | 9.8 |
| Copper | 32.3 | 19.1 | 21.5 | 12.2 | 12.1 | 10.9 | 18.6 | 16.6 |
| Iron | 35200 | 24600 | 22500 | 22200 | 14700 | 14500 | 22000 | 21700 |
| Lead | 19.1 | 11.6J | 10.7 | 12.9J | 8J | 8.2 | 10.2 | 10.2 |
| Magnesium | 8790 | 2890J | 2010 | 2770J | 1440J | 1390 | 2640 | 4130 |
| Manganese | 3030 | 352 | 286 | 375 | 194 | 189 | 325 | 516 |
| Mercury | 0.044 | <0.12U | <0.12U | 0.036J | <0.12U | <0.12U | <0.12U | 0.068J* |
| Nickel | 60.7 | 21.4 | 16.7 | 15.9 | 10.5 | 9.8 | 19.9 | 21.1 |
| Potassium | 3350 | 891 | 581 | 670 | 528 | 449 | 1070 | 1200 |
| Selenium | 1.5 | 0.87J | 0.66 | 0.65J | 0.5J | 0.56J | 0.98 | 0.96 |
| Silver | 0 | 0.0076J* | 0.0059J* | 0.019J* | <0.0063UJ | 0.01J* | <0.024UJ | <0.022UJ |
| Sodium | 145 | 44.6J | 31.4J | 41.6J | 28.4J | 27.4J | 47.6J | 52.4J |
| Thallium | 0.91 | 0.15J | 0.12J | 0.16J | 0.11J | 0.097J | 0.1J | 0.11J |
| Vanadium | 37.6 | 18.4 | 11.4 | 24.7 | 12.6 | 12.2 | 11.9 | 11.5 |
| Zinc | 93.3 | 52.2J | 55.8 | 42.9J | 36J | 32.5 | 56.3 | 53 |
| | | | | SVOCs (mg/kg) | | | | |
| Benz(a)anthracene | None | <0.008U | <0.0079U | 0.011J* | <0.059U | <0.059U | <0.008U | <0.008U |
| Benzo(a)pyrene | None | <0.008U | <0.0079U | 0.011J* | <0.059U | <0.059U | <0.008U | <0.008U |
| Benzo(b)fluoranthene | None | <0.008U | <0.0079U | 0.016J* | <0.059U | <0.059U | <0.008U | 0.0085* |
| Benzo(ghi)perylene | None | <0.008U | <0.0079U | <0.061U | <0.059U | <0.059U | <0.008U | 0.01* |
| Bis(2-ethylhexyl)phthalate | None | NR | NR | 0.024J* | <0.39U | <0.39UJ | NR | NR |
| Chrysene | None | <0.008U | <0.0079U | 0.01J* | <0.059U | <0.059U | <0.008U | 0.0082* |
| Di-n-butyl phthalate | None | NR | NR | 0.019J* | 0.019J* | <0.39UJ | NR | NR |
| Fluoranthene | None | <0.008U | <0.0079U | 0.025J* | <0.059U | <0.059U | <0.008U | <0.008U |
| Phenanthrene | None | <0.008U | <0.0079U | 0.013J* | <0.059U | <0.059U | <0.008U | <0.008U |
| Pyrene | None | <0.008U | <0.0079U | 0.019J* | <0.059U | <0.059U | <0.008U | <0.008U |
| ^a Only detected site-related contami | nants are presented in the table. | | PBA08 RI = Performance-bas | ed Acquisition 2008 Remedial Inve | stigation. | | | |

PCB = Polychlorinated biphenyl.

PETN = Pentaerythritol tetranitrate.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

TAL = Target Analyte List.

^b Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

ID = Identification.

FPA = Former production area.

ft = Feet.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram. NPA = Non-production area.

NR = Not reported/not analyzed.

PAH = Polycyclic aromatic hydrocarbon.

U= Not detected.

UJ = Not detected, reporting limit estimated.

VOC = Volatile organic compound.

^{* =} Result exceeds background criteria or no background criteria was available.

< = Less than.

Table 4–18. PBA08 RI Surface Water and Sediment Samples and Rationales

| | | | | | | Analy | ses Perfo | ormed | |
|---|---|--|-----------------------|--------------------|--------|------------|-----------|---------------------|------|
| PBA08 RI Location | Targeted Area | Comments/Rationale | Sample | Depth (ft | Metals | Explosives | VOCs | Pesticides/ PCBs | SVOC |
| LL11sd-082 | Targeted Area | Assess potential off-AOC migration of | Type Composite | bgs) 0–0.5 | V | V | N | PCBs | PAH |
| LL11sw-082 | Sand Creek near Newton Falls Road | contamination; sample collected in Sand Creek north of the AOC. | Discrete | N/A | Y | Y | Y | Y | Y |
| LL11sd-083 | | Confirm presence of contamination in | Composite | 0-0.5 | Y | Y | Y | Y | Y |
| LL11sw-083 | LL11sd-017 (drainage ditch within FPA) | previously sampled area; sediment sample analyzed for RVAAP full-suite analytes. | Discrete | N/A | Y | Y | Y | Y | Y |
| LL11sd-084 | LL11sd-030 (drainage ditch west of FPA) | Delineate lateral extent of previously identified | Composite | 0-0.5 | Y | Y | N | PCBs | PAH |
| LL11sw-084 | LLTISG-050 (drainage ditch west of FPA) | contamination. | Discrete | N/A | Y | Y | Y | Y | Y |
| LL11sd-096 | LL11sd-030 and LL11sd-084 (drainage ditch west of FPA) | Recollect sediment near LLsd-084 for RVAAP full-suite analytes in August 2012. No surface water present. | Composite | 0-0.5 | Y | Y | Y | Y | Y |
| OC = Area of concern. gs = Below ground surfac = Feet. /A = Not applicable. AH = Polycyclic aromatic BA08 RI = Performance- CB = Polychlorinated bip VAAP = Ravenna Army VOC= Semi-volatile orga OC = Volatile organic co | c hydrocarbon. based Acquisition 2008 Remedial Investigation. henyl. Ammunition Plant. unic compound. | | | | | | | | |

Table 4–19. Analytes Detected in PBA08 RI Surface Water Samples

| Aggregate | | East Ditch EU | Off AOC | West Ditch EU |
|----------------------------------|-----------------------|---------------------------|----------------------------------|---------------------------|
| Sample location | | LL11sw-083 | LL11sw-082 | LL11sw-084 |
| Sample ID | | LL11sw-083-5609-SW | LL11sw-082-5608-SW | LL11sw-084-5610-SW |
| Date | | 02/25/10 | 02/24/10 | 02/25/10 |
| Parameters Analyzed ^a | Background | | | |
| Analyte | Criteria ^b | RVAAP Full-suite analytes | RVAAP Full-suite analytes | RVAAP Full-suite analytes |
| - | | Metals (mg/L) | | |
| Aluminum | 3.37 | 0.674 | 0.58 | 2.72 |
| Antimony | 0 | <0.005U | <0.005U | 0.00036J* |
| Arsenic | 0.0032 | 0.00058J | 0.00066J | 0.0043J* |
| Barium | 0.0475 | 0.0136 | 0.0179 | 0.0576* |
| Beryllium | 0 | <0.001U | <0.001U | 0.00011J* |
| Cadmium | 0 | <0.002U | 0.00004J* | 0.00024J* |
| Calcium | 41.4 | 33.1 | 18.6 | 61.9* |
| Chromium | 0 | 0.00078J* | 0.00079J* | 0.0033J* |
| Cobalt | 0 | 0.00017J* | 0.00026J* | 0.0019J* |
| Copper | 0.0079 | 0.0014J | 0.0024J | 0.0062 |
| Iron | 2.56 | 0.637 | 0.864 | 5.83* |
| Lead | 0 | 0.00044J* | 0.00043J* | 0.0062* |
| Magnesium | 10.8 | 5.59 | 4.5 | 12.8* |
| Manganese | 0.391 | 0.0161 | 0.21 | 0.559* |
| Nickel | 0 | 0.00091J* | 0.0012J* | 0.0039J* |
| Potassium | 3.17 | 1.21 | 1.3 | 1.92 |
| Selenium | 0 | 0.00026J* | <0.005U | 0.00073J* |
| Sodium | 21.3 | 1.86 | 4.79 | 6.26 |
| Vanadium | 0 | 0.0011J* | 0.00086J* | 0.0051J* |
| | | SVOCs (mg/L) | | |
| Benz(a)anthracene | | <0.01U | <0.01U | 0.00035J* |
| Benzo(a)pyrene | | <0.01U | <0.01U | 0.00033J* |
| Benzo(b)fluoranthene | | <0.01UJ | <0.01U | 0.00041J* |
| Benzo(ghi)perylene | | <0.01U | <0.01U | 0.00024J* |
| Benzo(k)fluoranthene | | <0.01U | <0.01U | 0.00024J* |
| Chrysene | | <0.01U | <0.01U | 0.00035J* |
| Fluoranthene | | <0.01U | <0.01U | 0.00073J* |
| Indeno(1,2,3-cd)pyrene | | <0.01U | <0.01U | 0.00021J* |
| Phenanthrene | | <0.01U | <0.01U | 0.00022J* |
| Pyrene | | <0.01U | <0.01U | 0.00057J* |
| | 1 | Pesticides (mg/L) | | |
| beta-BHC | None | 0.00001J* | <0.00005U | 9.6E-6J* |
| gamma-Chlordane | None | 0.00002J* | <0.00005U | <0.00005U |

^a Only detected SRCs are presented in the table.

^b Background concentrations are published in the *Phase II Remedial Investigation Report for Winklepeck Burning Grounds* (USACE 2001b).

AOC = Area of concern BHC = Hexachlorocyclohexane.

EU = Exposure unit. ID = Identification.

J = Estimated value less than reporting limits.

mg/L = Milligrams per liter.
PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

RVAAP = Ravenna Army Ammunition Plant.

SVOC = Semi-volatile organic compound.

^{* =} Result exceeds background criteria or no background criteria was available.

Table 4–20. Analytes Detected in PBA08 RI Sediment Samples

| Aggregate | | East Ditch EU | Off AOC | West Ditch EU | West Ditch EU |
|----------------------------------|----------------------------------|---------------------------|-------------------------------------|-------------------------------------|---------------------------|
| Sample location | | LL11sd-083 | LL11sd-082 | LL11sd-084 | LL11sd-096 |
| Sample ID | | LL11sd-083-5594-SD | LL11sd-082-5593-SD | LL11sd-084-5595-SD | LL11sd-096-5874-SD |
| Date | | 02/25/10 | 02/24/10 | 02/25/10 | 08/09/12 |
| Depth (ft) | | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 0.5 |
| Parameters Analyzed ^a | | | | | |
| Analyte | Background Criteria ^b | RVAAP Full-suite analytes | TAL Metals, Explosives, SVOCs, VOCs | TAL Metals, Explosives, SVOCs, VOCs | RVAAP Full-suite analytes |
| | | | Metals (mg/kg) | | |
| Aluminum | 13900 | 15100* | 4070 | 10100 | 15000* |
| Antimony | 0 | 0.13J* | <0.65R | <0.78R | 0.14J* |
| Arsenic | 19.5 | 19.7* | 9.3 | 7.9 | 10 |
| Barium | 123 | 98.2 | 16.7 | 64 | 84 |
| Beryllium | 0.38 | 1* | 0.22 | 0.61* | 0.91* |
| Cadmium | 0 | <0.13UJ | <0.056UJ | <0.42U | 0.51* |
| Calcium | 5510 | 3190J | 1490J | 3100J | 3800 |
| Chromium | 18.1 | 18.6* | 5.7 | 12.2 | 17 |
| Cobalt | 9.1 | 17.4* | 4.3 | 7.1 | 9.4* |
| Copper | 27.6 | 31.4* | 15.8 | 14.6 | 18 |
| Iron | 28200 | 39300* | 13500 | 17100 | 23000 |
| Lead | 27.4 | 25.7 | 8.4 | 22 | 27 |
| Magnesium | 2760 | 3660* | 1540 | 1820 | 2700 |
| Manganese | 1950 | 665 | 317 | 164 | 270 |
| Mercury | 0.059 | <0.15U | <0.13U | 0.049J | 0.1J* |
| Nickel | 17.7 | 32.6* | 10.5 | 15.6 | 21* |
| Potassium | 1950 | 1230J | 390J | 685J | 1100 |
| Selenium | 1.7 | 3.2* | 0.48J | 1.4 | 1.7 |
| Silver | 0 | <0.022UJ | <0.027UJ | <0.062UJ | 0.072J* |
| Sodium | 112 | 68.1J | 33.2J | 36J | 35J |
| Thallium | 0.89 | 0.14J | <0.26U | 0.13J | 0.21J |
| Vanadium | 26.1 | 21.4 | 6.5 | 17.8 | 25 |
| Zinc | 532 | 77.6 | 27.5 | 80.9 | 120 |
| Nitrocellulose | None | <7.5U | Propellants (mg/kg) NR | NR | 1.4J* |
| Nitrocentitiose | None | <7.30 | SVOCs (mg/kg) | INK | 1.45** |
| 2-Methylnaphthalene | None | 0.013J* | 0.0097J* | <0.51U | <0.0045U |
| Acenaphthene | None | <0.075U | <0.065U | <0.078U | 0.013* |
| Anthracene | None | 0.025J* | <0.065U | <0.078U | 0.013 |
| Benz(a)anthracene | None | 0.11* | 0.014J* | 0.019J* | 0.077* |
| Benzenemethanol | None | <0.5U | <0.43U | <0.51U | 0.054J* |
| Benzo(a)pyrene | None | 0.098* | 0.012J* | 0.021J* | 0.09* |
| Benzo(b)fluoranthene | None | 0.13* | 0.017J* | 0.038J* | 0.14* |
| Benzo(ghi)perylene | None | 0.064J* | 0.011J* | 0.017J* | 0.068* |
| Benzo(k)fluoranthene | None | 0.056J* | 0.0089J* | 0.011J* | 0.066* |
| Chrysene | None | 0.096* | 0.014J* | 0.023J* | 0.099* |
| Dibenz(a,h)anthracene | None | 0.017J* | <0.065U | <0.078U | 0.016* |
| Fluoranthene | None | 0.23* | 0.025J* | 0.043J* | 0.21* |
| Fluorene | None | 0.012J* | <0.065U | <0.078U | <0.0045U |
| Indeno(1,2,3-cd)pyrene | None | 0.056J* | <0.065U | 0.015J* | 0.061* |
| Naphthalene | None | 0.01J* | <0.065U | <0.078U | <0.0045U |
| Phenanthrene | None | 0.1* | 0.017J* | 0.016J* | 0.098* |
| Pyrene | None | 0.18* | 0.019J* | 0.033J* | 0.16* |

Table 4–20. Analytes Detected in PBA08RI Sediment Samples (continued)

| Aggregate | | East Ditch EU | Off AOC | West Ditch EU | West Ditch EU | |
|----------------------------------|----------------------------------|---------------------------|-------------------------------------|-------------------------------------|---------------------------|--|
| Sample location | | LL11sd-083 | LL11sd-082 | LL11sd-084 | LL11sd-096 | |
| Sample ID | | LL11sd-083-5594-SD | LL11sd-082-5593-SD | LL11sd-084-5595-SD | LL11sd-096-5874-SD | |
| Date | | 02/25/10 | 02/24/10 | 02/25/10 | 08/09/12 | |
| Depth (ft) | | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 0.5 | 0.0 - 0.5 | |
| Parameters Analyzed ^a | | | | | | |
| Analyte | Background Criteria ^b | RVAAP Full-suite analytes | TAL Metals, Explosives, SVOCs, VOCs | TAL Metals, Explosives, SVOCs, VOCs | RVAAP Full-suite analytes | |
| | | | VOCs (mg/kg) | | | |
| Carbon disulfide | | <0.0075U | 0.00063J* | <0.0078UJ | <0.00068UJ | |
| Toluene | | <0.0075U | 0.0004J* | <0.0078UJ | <0.00068UJ | |

EU = Exposure unit.

ft = Feet.ID = Identification.

J = Estimated value less than reporting limits.

mg/kg = Milligrams per kilogram.

NR = Not reported/not analyzed.
PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.
RVAAP = Ravenna Army Ammunition Plant.

SVOC= Semi-volatile organic compound.

TAL = Target analyte list. U = Not detected.

UJ = Not detected, reporting limit estimated.

VOC = Volatile organic compound.

* = Result exceeds background criteria or no background criteria was available.

< = Less than

^a Only detected site-related contaminants are presented in the table.
^b Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b). AOC = Area of concern.

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Table 4–21. Changes from the PBA08 SAP

| | | Date | | | |
|------------|---------------------------|-----------|---|--|--|
| Location | Affected Sample | Sampled | Change/Rationale | | |
| | LL11SB-060-5551-SO | 3/18/2010 | | | |
| LL11sb-060 | LL11SB-060-5552-SO | 3/18/2010 | Relocated to the southwest to the bottom of the | | |
| LLI180-000 | LL11SB-060-5553-SO | 3/18/2010 | ditch | | |
| | LL11SB-060-5554-SO | 3/18/2010 | | | |
| | LL11SB-063-5567-SO | | Geotechnical boring renamed LL11sb-085 due to | | |
| LL11sb-063 | LL11SB-063-5568-SO | 3/18/2010 | the duplicate usage of LL11sb-063 in Work Plan. | | |
| LL1180-003 | LL11SB-085-5567-SO | 3/18/2010 | (Note: SDG AOC 190549 indicates geotechnical | | |
| | LL11SB-085-5568-SO | | sample was collected at LL11sb-063) | | |
| | | | Sample could not be hand augered because location | | |
| LL11sb-063 | LL11SB-063-5563-SO | 3/22/2010 | was in asphalt, so the 0-1 ft bgs sample was | | |
| | | | collected by the Geoprobe | | |
| LL11sb-067 | LL11SB-067-6186-FD | 3/18/2010 | A twin boring had to be drilled to collect volume | | |
| LL1180-007 | LL11SB-067-6190-QA | 3/18/2010 | for QA/QC samples | | |
| | LL11SB-068-5586-SO | 3/18/2010 | A 4-2-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1- | | |
| LL11sb-068 | LL11SB-068-5587-SO | 3/18/2010 | A twin boring had to be drilled to collect adequate | | |
| | LL11SB-068-6187-FD | 3/18/2010 | volume for samples collected from the 1-4 ft bgs and 4-7 ft bgs | | |
| | LL11SB-068-6191-QA | 3/18/2010 | and 4-7 it ogs | | |
| | LL11SB-069-5589-SO | 3/22/2010 | | | |
| LL11sb-069 | LL11SB-069-5590-SO | 3/22/2010 | Relocated to the southwest to the bottom of a ditch | | |
| | LL11SB-069-5591-SO | 3/22/2010 | | | |
| | LL11SS-076-5602-SO | 4/12/2010 | | | |
| LL11ss-076 | LL11SS-076-6183-FD | 4/12/2010 | Relocated to the southwest to the bottom of a ditch | | |
| | LL11SS-076-6182-QA | 4/12/2010 | | | |
| LL11ss-078 | LL11SS-078-5604-SO | 4/12/2010 | Relocated to the northwest to the bottom of a ditch | | |
| LL11ss-080 | LL11SS-080-5606-SO | 4/12/2010 | Relocated to the north to the bottom of a ditch | | |
| LL11ss-081 | LL11SS-081-5607-SO | 4/12/2010 | Relocated to the east to the bottom of a ditch | | |
| LL11sd-082 | LL11SD-082-5593-SD | 2/24/2010 | Relocated downstream so samples could be | | |
| LL11sw-082 | LL11SW-082-5608-SW | 2/24/2010 | collected safely | | |
| LL11sd-083 | LL11SD-083-5594-SD | 2/25/2010 | | | |
| LL11sw-083 | LL11SW-083-5609-SW | 2/25/2010 | | | |
| LL11sb-085 | See note above for LL11sl | | • | | |
| LLsd-096 | LL11sd-096-5874-SD | 8/9/2012 | Recollected near LL11sd-084 for RVAAP full-suite analytes | | |

bgs = Below ground surface.

Ft = Feet.

PBA08 SAP = Performance-Based Acquisition 2008 Supplemental Investigation Sampling and Analysis Plan Addendum No. 1.

N/A = Not applicable. QA/QC = Quality assurance/quality control.

Table 4-22. RVAAP Background Concentrations

| | Surface Soil | Subsurface soil | Sediment | Surface Water | | r-Unconsolidated ng/L) | Croundwater | -Bedrock (mg/L) |
|-----------|--------------|-----------------|----------|---------------|----------|---------------------------|-------------|-----------------|
| Chemical | (mg/kg) | (mg/kg) | (mg/kg) | (mg/L) | Filtered | Unfiltered | Filtered | Unfiltered |
| Aluminum | 17700 | 19500 | 13900 | 3.37 | NA | 48 | NA | 9.41 |
| Antimony | 0.96 | 0.96 | 0 | 0 | 0 | 0.0043 | 0 | 0 |
| Arsenic | 15.4 | 19.8 | 19.5 | 0.0032 | 0.0117 | 0.215 | 0 | 0.0191 |
| Barium | 88.4 | 124 | 123 | 0.0475 | 0.0821 | 0.327 | 0.256 | 0.241 |
| Beryllium | 0.88 | 0.88 | 0.38 | 0 | 0 | 0 | 0 | 0 |
| Cadmium | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calcium | 15800 | 35500 | 5510 | 41.4 | 115 | 194 | 53.1 | 48.2 |
| Chromium | 17.4 | 27.2 | 18.1 | 0 | 0.0073 | 0.0852 | 0 | 0.0195 |
| Cobalt | 10.4 | 23.2 | 9.1 | 0 | 0 | 0.0463 | 0 | 0 |
| Copper | 17.7 | 32.3 | 27.6 | 0.0079 | 0 | 0.289 | 0 | 0.017 |
| Cyanide | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Iron | 23100 | 35200 | 28200 | 2.56 | 0.279 | 195 | 1.43 | 21.5 |
| Lead | 26.1 | 19.1 | 27.4 | 0 | 0 | 0.183 | 0 | 0.023 |
| Magnesium | 3030 | 8790 | 2760 | 10.8 | 43.3 | 58.4 | 15 | 13.7 |
| Manganese | 1450 | 3030 | 1950 | 0.391 | 1.02 | 2.86 | 1.34 | 1.26 |
| Mercury | 0.036 | 0.044 | 0.059 | 0 | 0 | 0.00025 | 0 | 0 |
| Nickel | 21.1 | 60.7 | 17.7 | 0 | 0 | 0.117 | 0.0834 | 0.0853 |
| Potassium | 927 | 3350 | 1950 | 3.17 | 2.89 | 7.48 | 5.77 | 6.06 |
| Selenium | 1.4 | 1.5 | 1.7 | 0 | 0 | 0.0057 | 0 | 0 |
| Silver | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sodium | 123 | 145 | 112 | 21.3 | 45.7 | 44.7 | 51.4 | 49.7 |
| Thallium | 0 | 0.91 | 0.89 | 0 | 0 | 0.0024 | 0 | 0 |
| Vanadium | 31.1 | 37.6 | 26.1 | 0 | 0 | 0.0981 | 0 | 0.0155 |
| Zinc | 61.8 | 93.3 | 532 | 0.042 | 0.0609 | 0.888 | 0.0523 | 0.193 |

Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

mg/kg = Milligrams per kilogram.

mg/L = Milligrams per liter.

NA = Not available. Aluminum results were rejected in validation. RVAAP = Ravenna Army Ammunition Plant.

Table 4-23. Recommended Dietary Allowances/Reference Daily Intake Values

| Essential Human Nutrient | USDA RDA/RDI ^a Value |
|---------------------------------|---------------------------------|
| Calcium | 1000 mg/d |
| Chloride ^b | 3400 mg/d |
| Iodine | 150 μg/d |
| Iron | 8 mg/d |
| Magnesium | 400 mg/d |
| Potassium ^b | 4700 mg/d |
| Phosphorous | 700 mg/d |
| Sodium ^b | 2300 mg/d |

^a Dietary reference intakes vary by gender and age, values present are for life stage group: Males 19-30 years.

mg/d = Milligram per day.

RDA= Recommended dietary allowance.

RDI= Reference daily intake.

 $\mu g/d = Micrograms per day.$

USDA = U.S. Department of Agriculture.

Source= Values were obtained from http://fnic.nal.usda.gov charts.

^b Adequate intake value.

Table 4–24. SRC Screening Summary for FPA Surface Soil

| Analyte | CAS Number | Freq. of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? (yes/no) | SRC Justification |
|----------------|---------------|--------------------|------------------------------|------------------------------|------------------------------|--|------------------|--------------------|
| | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 47/47 | 789 | 23800 | 11400 | 17700 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 35/47 | 0.08 | 0.63 | 0.306 | 0.96 | No | Below background |
| Arsenic | 7440-38-2 | 47/47 | 2.7 | 30.2 | 13.2 | 15.4 | Yes | Exceeds background |
| Barium | 7440-39-3 | 47/ 47 | 14.6 | 107 | 61 | 88.4 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 47/47 | 0.13 | 1.1 | 0.587 | 0.88 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 26/47 | 0.053 | 1.3 | 0.215 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 47/47 | 830 | 64400 | 6500 | 15800 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 47/ 47 | 3.5 | 28.4 | 14.6 | 17.4 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 47/ 47 | 2.2 | 33.8 | 8.54 | 10.4 | Yes | Exceeds background |
| Copper | 7440-50-8 | 47/ 47 | 2.7 | 45.7 | 18 | 17.7 | Yes | Exceeds background |
| Cyanide | 57-12-5 | 3/34 | 0.58 | 1.6 | 0.301 | 0 | Yes | Exceeds background |
| Iron | 7439-89-6 | 47/47 | 11400 | 38300 | 21600 | 23100 | No | Essential Nutrient |
| Lead | 7439-92-1 | 47/ 47 | 3.9 | 102 | 29.6 | 26.1 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 47/47 | 121 | 5390 | 2790 | 3030 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 47/ 47 | 63.9 | 1540 | 511 | 1450 | Yes | Exceeds background |
| Mercury | 7439-97-6 | 17/ 47 | 0.018 | 0.34 | 0.0397 | 0.036 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 47/ 47 | 4.5 | 33.2 | 16.4 | 21.1 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 47/47 | 232 | 2910 | 1260 | 927 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 19/ 47 | 0.25 | 1.3 | 0.48 | 1.4 | No | Below background |
| Silver | 7440-22-4 | 2/47 | 0.0046 | 0.023 | 0.181 | 0 | No | <5% Detected |
| Sodium | 7440-23-5 | 42/47 | 21.9 | 943 | 471 | 123 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 16/ 47 | 0.13 | 0.26 | 0.124 | 0 | Yes | Exceeds background |
| Vanadium | 7440-62-2 | 47/47 | 2.9 | 38.8 | 20.1 | 31.1 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 47/47 | 20.9 | 465 | 77.3 | 61.8 | Yes | Exceeds background |
| | | | | Anions | | | | |
| Nitrate | 14797-55-8 | 7/34 | 0.92 | 3.2 | 0.834 | None | Yes | Exceeds background |
| Sulfide | 18496-25-8 | 22/ 34 | 29.9 | 244 | 49.3 | None | Yes | Exceeds background |
| | | | | losives and Prop | | | | |
| Nitrocellulose | 9004-70-0 | 4/4 | 0.89 | 1.1 | 0.985 | None | Yes | Detected organic |
| Nitroguanidine | 556-88-7 | 1/4 | 0.077 | 0.077 | 0.132 | None | Yes | Detected organic |

Table 4-24. SRC Screening Summary for FPA Surface Soil (continued)

| Analyte Number Detect (mg/kg) (mg/kg) | 0.216 None 0.101 None 0.114 None 0.107 None | SRC? (yes/no) Yes Yes | SRC Justification Detected organic Detected organic | | | | | | | |
|---|---|--------------------------------|---|--|--|--|--|--|--|--|
| Explosives and Propellants (continued PETN 78-11-5 1/7 0.049 0.049 | ued) None 0.216 None 0.101 None 0.114 None 0.107 None | Yes Yes | Detected organic Detected organic | | | | | | | |
| PETN 78-11-5 1/7 0.049 0.049 Tetryl 479-45-8 1/41 0.02 0.02 SVOCs Acenaphthene 83-32-9 2/14 0.033 0.081 Acenaphthylene 208-96-8 1/14 0.012 0.012 Anthracene 120-12-7 2/14 0.049 0.16 | 0.216 None 0.101 None 0.114 None 0.107 None | Yes | Detected organic | | | | | | | |
| Tetryl 479-45-8 1/41 0.02 0.02 SVOCs Acenaphthene 83-32-9 2/14 0.033 0.081 Acenaphthylene 208-96-8 1/14 0.012 0.012 Anthracene 120-12-7 2/14 0.049 0.16 | 0.101 None 0.114 None 0.107 None | Yes | Detected organic | | | | | | | |
| SVOCs Acenaphthene 83-32-9 2/14 0.033 0.081 Acenaphthylene 208-96-8 1/14 0.012 0.012 Anthracene 120-12-7 2/14 0.049 0.16 | 0.114 None 0.107 None | | | | | | | | | |
| Acenaphthene 83-32-9 2/14 0.033 0.081 Acenaphthylene 208-96-8 1/14 0.012 0.012 Anthracene 120-12-7 2/14 0.049 0.16 | 0.107 None | Yes | | | | | | | | |
| Acenaphthylene 208-96-8 1/14 0.012 0.012 Anthracene 120-12-7 2/14 0.049 0.16 | 0.107 None | Yes | | | | | | | | |
| Anthracene 120-12-7 2/14 0.049 0.16 | | | Detected organic | | | | | | | |
| | 0.101 | Yes | Detected organic | | | | | | | |
| D () 41 | 0.121 None | Yes | Detected organic | | | | | | | |
| Benz(a)anthracene 56-55-3 7/14 0.0089 0.37 | 0.147 None | Yes | Detected organic | | | | | | | |
| Benzo(a)pyrene 50-32-8 7/14 0.0091 0.45 | 0.151 None | Yes | Detected organic | | | | | | | |
| Benzo(b)fluoranthene 205-99-2 6/14 0.011 0.56 | 0.188 None | Yes | Detected organic | | | | | | | |
| Benzo(ghi)perylene 191-24-2 4/14 0.012 0.32 | 0.145 None | Yes | Detected organic | | | | | | | |
| Benzo(k)fluoranthene 207-08-9 2/14 0.17 0.37 | 0.145 None | Yes | Detected organic | | | | | | | |
| Chrysene 218-01-9 8/14 0.0097 0.47 | 0.153 None | Yes | Detected organic | | | | | | | |
| Dibenz(a,h)anthracene 53-70-3 3/14 0.022 0.076 | 0.118 None | Yes | Detected organic | | | | | | | |
| Fluoranthene 206-44-0 11/14 0.0093 1 | 0.277 None | Yes | Detected organic | | | | | | | |
| Fluorene 86-73-7 2/14 0.025 0.079 | 0.114 None | Yes | Detected organic | | | | | | | |
| Indeno(1,2,3-cd)pyrene 193-39-5 4/14 0.021 0.27 | 0.142 None | Yes | Detected organic | | | | | | | |
| Naphthalene 91-20-3 3/14 0.0089 0.017 | 0.108 None | Yes | Detected organic | | | | | | | |
| Phenanthrene 85-01-8 5/14 0.012 0.62 | 0.172 None | Yes | Detected organic | | | | | | | |
| Pyrene 129-00-0 8/14 0.011 0.83 | 0.255 None | Yes | Detected organic | | | | | | | |
| Pesticides/PCBs | | | | | | | | | | |
| PCB-1254 11097-69-1 4/12 0.02 0.42 | 0.0562 None | Yes | Detected organic | | | | | | | |
| VOCs | | | | | | | | | | |
| Acetone 67-64-1 1/8 0.011 0.011 | 0.0112 None | Yes | Detected organic | | | | | | | |

^a Background concentrations are published in the *Phase II Remedial Investigation Report* for Winklepeck Burning Grounds (USACE 2001b). Site-related contaminant screening tables include all available and appropriate data as presented in Section 4.5.4.

CAS = Chemical Abstract Service.

FPA = Former production area.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl. PETN = Pentaerythritol tetranitrate.

SRC = Site-related contaminant.

SVOC= Semi-volatile organic compound

VOC = Volatile organic compound.

Table 4–25. SRC Screening Summary for NPA Surface Soil

| Analyte | CAS Number | Freq. of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? (yes/no) | SRC Justification |
|----------------|---------------|--------------------|------------------------------|------------------------------|------------------------------|--|------------------|--------------------|
| · | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 30/30 | 6170 | 14400 | 10300 | 17700 | No | Below background |
| Antimony | 7440-36-0 | 15/30 | 0.084 | 0.77 | 0.246 | 0.96 | No | Below background |
| Arsenic | 7440-38-2 | 29/ 30 | 6.5 | 40.4 | 12 | 15.4 | Yes | Exceeds background |
| Barium | 7440-39-3 | 30/ 30 | 26.8 | 128 | 61.3 | 88.4 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 30/30 | 0.35 | 0.76 | 0.528 | 0.88 | No | Below background |
| Cadmium | 7440-43-9 | 17/30 | 0.057 | 0.92 | 0.191 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 29/30 | 180 | 18100 | 2970 | 15800 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 30/ 30 | 9.3 | 19 | 13.3 | 17.4 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 30/ 30 | 5.4 | 13.9 | 8.46 | 10.4 | Yes | Exceeds background |
| Copper | 7440-50-8 | 30/ 30 | 5.8 | 21 | 13.5 | 17.7 | Yes | Exceeds background |
| Cyanide | 57-12-5 | 1/ 17 | 0.54 | 0.54 | 0.269 | 0 | Yes | Exceeds background |
| Iron | 7439-89-6 | 30/30 | 14300 | 27300 | 19900 | 23100 | No | Essential Nutrient |
| Lead | 7439-92-1 | 29/ 30 | 9.9 | 33.9 | 18.1 | 26.1 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 30/30 | 1430 | 6390 | 2340 | 3030 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 30/ 30 | 210 | 1930 | 550 | 1450 | Yes | Exceeds background |
| Mercury | 7439-97-6 | 14/30 | 0.02 | 0.08 | 0.034 | 0.036 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 30/ 30 | 8.5 | 26.2 | 15.5 | 21.1 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 30/30 | 437 | 1980 | 991 | 927 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 20/30 | 0.59 | 1.2 | 0.626 | 1.4 | No | Below background |
| Silver | 7440-22-4 | 8/ 30 | 0.024 | 0.06 | 0.139 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 28/ 30 | 22.5 | 990 | 432 | 123 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 14/30 | 0.1 | 0.23 | 0.124 | 0 | Yes | Exceeds background |
| Vanadium | 7440-62-2 | 30/30 | 12.2 | 28.4 | 19.3 | 31.1 | No | Below background |
| Zinc | 7440-66-6 | 30/ 30 | 38.7 | 478 | 78.2 | 61.8 | Yes | Exceeds background |
| | | T | | Anions | T | | | |
| Nitrate | 14797-55-8 | 6/ 17 | 0.64 | 3 | 0.778 | None | Yes | Exceeds background |
| Sulfide | 18496-25-8 | 11/ 17 | 31.1 | 85.5 | 41.6 | None | Yes | Exceeds background |
| | | T | | losives and Prop | | | | T |
| HMX | 2691-41-0 | 1/29 | 0.013 | 0.013 | 0.12 | None | Yes | Detected organic |
| Nitrocellulose | 9004-70-0 | 3/5 | 0.89 | 1.2 | 1.8 | None | Yes | Detected organic |

Table 4–25. SRC Screening Summary for NPA Soil (continued)

| | | | Minimum | Maximum | | Background | | | | | | | |
|------------------------|--|----------|---------|---------|----------------|-----------------------|----------|-------------------|--|--|--|--|--|
| | CAS | Freq. of | Detect | Detect | Average Result | Criteria ^a | SRC? | | | | | | |
| Analyte | Number | Detect | (mg/kg) | (mg/kg) | (mg/kg) | (mg/kg) | (yes/no) | SRC Justification | | | | | |
| | Explosives and Propellants (continued) | | | | | | | | | | | | |
| PETN | 78-11-5 | 1/12 | 0.036 | 0.036 | 0.23 | None | Yes | Detected organic | | | | | |
| SVOCs | | | | | | | | | | | | | |
| Acenaphthene | 83-32-9 | 2/ 14 | 0.012 | 0.065 | 0.0405 | None | Yes | Detected organic | | | | | |
| Anthracene | 120-12-7 | 4/ 14 | 0.011 | 0.13 | 0.0469 | None | Yes | Detected organic | | | | | |
| Benz(a)anthracene | 56-55-3 | 9/ 14 | 0.013 | 0.32 | 0.0656 | None | Yes | Detected organic | | | | | |
| Benzo(a)pyrene | 50-32-8 | 8/ 14 | 0.014 | 0.28 | 0.0741 | None | Yes | Detected organic | | | | | |
| Benzo(b)fluoranthene | 205-99-2 | 7/ 14 | 0.022 | 0.41 | 0.0985 | None | Yes | Detected organic | | | | | |
| Benzo(ghi)perylene | 191-24-2 | 8/ 14 | 0.013 | 0.21 | 0.0626 | None | Yes | Detected organic | | | | | |
| Benzo(k)fluoranthene | 207-08-9 | 7/ 14 | 0.0094 | 0.13 | 0.0593 | None | Yes | Detected organic | | | | | |
| Chrysene | 218-01-9 | 9/ 14 | 0.016 | 0.28 | 0.0693 | None | Yes | Detected organic | | | | | |
| Dibenz(a,h)anthracene | 53-70-3 | 1/ 14 | 0.069 | 0.069 | 0.0404 | None | Yes | Detected organic | | | | | |
| Fluoranthene | 206-44-0 | 11/ 14 | 0.011 | 0.78 | 0.137 | None | Yes | Detected organic | | | | | |
| Fluorene | 86-73-7 | 1/ 14 | 0.049 | 0.049 | 0.0389 | None | Yes | Detected organic | | | | | |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 7/ 14 | 0.011 | 0.19 | 0.0599 | None | Yes | Detected organic | | | | | |
| Naphthalene | 91-20-3 | 2/ 14 | 0.0094 | 0.01 | 0.0365 | None | Yes | Detected organic | | | | | |
| Phenanthrene | 85-01-8 | 8/ 14 | 0.012 | 0.52 | 0.0907 | None | Yes | Detected organic | | | | | |
| Pyrene | 129-00-0 | 10/ 14 | 0.01 | 0.59 | 0.116 | None | Yes | Detected organic | | | | | |
| | Pesticides/PCBs | | | | | | | | | | | | |
| PCB-1254 | 11097-69-1 | 1/ 14 | 0.035 | 0.035 | 0.0206 | None | Yes | Detected organic | | | | | |

^a Background concentrations are published in the *Phase II Remedial Investigation Report* for Winklepeck Burning Grounds (USACE 2001b). SRC screening tables include all available and appropriate data as presented in Section 4.5.4.

CAS = Chemical Abstract Service.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocane.

mg/kg = Milligrams per kilogram.

NPA = Non-production area.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

SVOC = Semi-volatile organic compound.

Table 4–26. SRC Screening Summary for FPA Subsurface Soil

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification |
|----------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|--------------------|
| | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 30/ 30 | 6170 | 14400 | 10300 | 17700 | No | Below background |
| Antimony | 7440-36-0 | 15/30 | 0.084 | 0.77 | 0.246 | 0.96 | No | Below background |
| Arsenic | 7440-38-2 | 29/30 | 6.5 | 40.4 | 12 | 15.4 | Yes | Exceeds background |
| Barium | 7440-39-3 | 30/30 | 26.8 | 128 | 61.3 | 88.4 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 30/30 | 0.35 | 0.76 | 0.528 | 0.88 | No | Below background |
| Cadmium | 7440-43-9 | 17/ 30 | 0.057 | 0.92 | 0.191 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 29/ 30 | 180 | 18100 | 2970 | 15800 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 30/ 30 | 9.3 | 19 | 13.3 | 17.4 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 30/ 30 | 5.4 | 13.9 | 8.46 | 10.4 | Yes | Exceeds background |
| Copper | 7440-50-8 | 30/ 30 | 5.8 | 21 | 13.5 | 17.7 | Yes | Exceeds background |
| Cyanide | 57-12-5 | 1/ 17 | 0.54 | 0.54 | 0.269 | 0 | Yes | Exceeds background |
| Iron | 7439-89-6 | 30/30 | 14300 | 27300 | 19900 | 23100 | No | Essential Nutrient |
| Lead | 7439-92-1 | 29/ 30 | 9.9 | 33.9 | 18.1 | 26.1 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 30/ 30 | 1430 | 6390 | 2340 | 3030 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 30/ 30 | 210 | 1930 | 550 | 1450 | Yes | Exceeds background |
| Mercury | 7439-97-6 | 14/ 30 | 0.02 | 0.08 | 0.034 | 0.036 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 30/30 | 8.5 | 26.2 | 15.5 | 21.1 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 30/30 | 437 | 1980 | 991 | 927 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 20/30 | 0.59 | 1.2 | 0.626 | 1.4 | No | Below background |
| Silver | 7440-22-4 | 8/ 30 | 0.024 | 0.06 | 0.139 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 28/30 | 22.5 | 990 | 432 | 123 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 14/30 | 0.1 | 0.23 | 0.124 | 0 | Yes | Exceeds background |
| Vanadium | 7440-62-2 | 30/30 | 12.2 | 28.4 | 19.3 | 31.1 | No | Below background |
| Zinc | 7440-66-6 | 30/ 30 | 38.7 | 478 | 78.2 | 61.8 | Yes | Exceeds background |
| | | | | Anions | | | | |
| Nitrate | 14797-55-8 | 6/ 17 | 0.64 | 3 | 0.778 | None | Yes | Exceeds background |
| Sulfide | 18496-25-8 | 11/ 17 | 31.1 | 85.5 | 41.6 | None | Yes | Exceeds background |
| | | | | losives/Prope | | | | |
| HMX | 2691-41-0 | 1/ 29 | 0.013 | 0.013 | 0.12 | None | Yes | Detected organic |
| Nitrocellulose | 9004-70-0 | 3/5 | 0.89 | 1.2 | 1.8 | None | Yes | Detected organic |
| PETN | 78-11-5 | 1/ 12 | 0.036 | 0.036 | 0.23 | None | Yes | Detected organic |

Table 4–26. SRC Screening Summary for FPA Subsurface Soil (continued)

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification | | | |
|------------------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|-------------------|--|--|--|
| Analyte | Number | Detect | (mg/kg) | Metals | (IIIg/Kg) | (mg/kg) | y es/110 | SKC Justification | | | |
| Acenaphthene | 83-32-9 | 2/14 | 0.012 | 0.065 | 0.0405 | None | Yes | Detected organic | | | |
| Anthracene | 120-12-7 | 4/14 | 0.012 | 0.13 | 0.0469 | None | Yes | Detected organic | | | |
| Benz(a)anthracene | 56-55-3 | 9/14 | 0.011 | 0.13 | 0.0455 | None | Yes | Detected organic | | | |
| Benzo(a)pyrene | 50-33-8 | 8/14 | 0.013 | 0.32 | 0.0030 | None | Yes | Detected organic | | | |
| Benzo(b)fluoranthene | 205-99-2 | 7/14 | 0.022 | 0.41 | 0.0985 | None | Yes | Detected organic | | | |
| Benzo(ghi)perylene | 191-24-2 | 8/ 14 | 0.013 | 0.21 | 0.0626 | None | Yes | Detected organic | | | |
| Benzo(k)fluoranthene | 207-08-9 | 7/ 14 | 0.0094 | 0.13 | 0.0593 | None | Yes | Detected organic | | | |
| Chrysene | 218-01-9 | 9/ 14 | 0.016 | 0.28 | 0.0693 | None | Yes | Detected organic | | | |
| Dibenz(a,h)anthracene | 53-70-3 | 1/ 14 | 0.069 | 0.069 | 0.0404 | None | Yes | Detected organic | | | |
| Fluoranthene | 206-44-0 | 11/ 14 | 0.011 | 0.78 | 0.137 | None | Yes | Detected organic | | | |
| Fluorene | 86-73-7 | 1/ 14 | 0.049 | 0.049 | 0.0389 | None | Yes | Detected organic | | | |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 7/ 14 | 0.011 | 0.19 | 0.0599 | None | Yes | Detected organic | | | |
| Naphthalene | 91-20-3 | 2/ 14 | 0.0094 | 0.01 | 0.0365 | None | Yes | Detected organic | | | |
| Phenanthrene | 85-01-8 | 8/ 14 | 0.012 | 0.52 | 0.0907 | None | Yes | Detected organic | | | |
| Pyrene | 129-00-0 | 10/ 14 | 0.01 | 0.59 | 0.116 | None | Yes | Detected organic | | | |
| | Pesticide/PCB | | | | | | | | | | |
| PCB-1254 | 11097-69-1 | 1/ 14 | 0.035 | 0.035 | 0.0206 | None | Yes | Detected organic | | | |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

CAS = Chemical Abstract Service.

FPA = Former production area.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocane.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

PETN = Pentaerythritol tetranitrate.

SRC = Site-related contaminant.

SRC screening tables include all available and appropriate data as presented in Section 4.5.4.

Table 4–27. SRC Screening Summary for NPA Subsurface Soil

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification | | |
|------------------------------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|--------------------|--|--|
| | | | Λ | 1etals | | | | | | |
| Aluminum | 7429-90-5 | 45/ 45 | 4720 | 24500 | 9790 | 19500 | Yes | Exceeds background | | |
| Antimony | 7440-36-0 | 15/45 | 0.077 | 0.56 | 0.18 | 0.96 | No | Below background | | |
| Arsenic | 7440-38-2 | 39/ 45 | 6 | 44.1 | 14.2 | 19.8 | Yes | Exceeds background | | |
| Barium | 7440-39-3 | 45/ 45 | 19.7 | 4190 | 139 | 124 | Yes | Exceeds background | | |
| Beryllium | 7440-41-7 | 39/ 45 | 0.25 | 3.2 | 0.531 | 0.88 | Yes | Exceeds background | | |
| Cadmium | 7440-43-9 | 21/45 | 0.037 | 0.38 | 0.119 | 0 | Yes | Exceeds background | | |
| Calcium | 7440-70-2 | 45/45 | 112 | 20400 | 2000 | 35500 | No | Essential Nutrient | | |
| Chromium | 7440-47-3 | 45/45 | 5.9 | 24.7 | 12.8 | 27.2 | No | Below background | | |
| Cobalt | 7440-48-4 | 45/ 45 | 3.3 | 24.7 | 8.42 | 23.2 | Yes | Exceeds background | | |
| Copper | 7440-50-8 | 45/ 45 | 6 | 62.3 | 18.6 | 32.3 | Yes | Exceeds background | | |
| Cyanide | 57-12-5 | 7/ 25 | 0.0097 | 1 | 0.163 | 0 | Yes | Exceeds background | | |
| Iron | 7439-89-6 | 45/45 | 9560 | 54800 | 21800 | 35200 | No | Essential Nutrient | | |
| Lead | 7439-92-1 | 40/ 45 | 5.3 | 160 | 14.2 | 19.1 | Yes | Exceeds background | | |
| Magnesium | 7439-95-4 | 45/45 | 786 | 4710 | 2440 | 8790 | No | Essential Nutrient | | |
| Manganese | 7439-96-5 | 45/45 | 103 | 1930 | 411 | 3030 | No | Below background | | |
| Mercury | 7439-97-6 | 6/45 | 0.025 | 0.068 | 0.0317 | 0.044 | Yes | Exceeds background | | |
| Nickel | 7440-02-0 | 45/45 | 8.6 | 33.5 | 17.6 | 60.7 | No | Below background | | |
| Potassium | 7440-09-7 | 45/45 | 449 | 2600 | 1100 | 3350 | No | Essential Nutrient | | |
| Selenium | 7782-49-2 | 23/45 | 0.52 | 1.5 | 0.543 | 1.5 | No | Below background | | |
| Silver | 7440-22-4 | 6/45 | 0.0059 | 0.019 | 0.159 | 0 | Yes | Exceeds background | | |
| Sodium | 7440-23-5 | 35/45 | 27.4 | 1530 | 396 | 145 | No | Essential Nutrient | | |
| Thallium | 7440-28-0 | 19/ 45 | 0.097 | 2.5 | 0.169 | 0.91 | Yes | Exceeds background | | |
| Vanadium | 7440-62-2 | 45/45 | 8.9 | 31.5 | 16.9 | 37.6 | No | Below background | | |
| Zinc | 7440-66-6 | 45/ 45 | 29 | 390 | 64.9 | 93.3 | Yes | Exceeds background | | |
| | | | A | nions | | | | | | |
| Nitrate | 14797-55-8 | 14/ 25 | 0.4 | 2.9 | 0.662 | None | Yes | Exceeds background | | |
| Sulfate | 14808-79-8 | 15/ 25 | 13.6 | 44.9 | 28.1 | None | Yes | Exceeds background | | |
| Sulfide | 18496-25-8 | 9/ 25 | 12.2 | 81.8 | 19.8 | None | Yes | Exceeds background | | |
| Miscellaneous | | | | | | | | | | |
| TPH-Diesel Range Organics | NS791 | 1/ 15 | 34 | 34 | 5.42 | None | Yes | Exceeds background | | |
| TPH-Gasoline Range Organics | NS834 | 2/ 14 | 0.026 | 0.054 | 0.0319 | None | Yes | Exceeds background | | |
| Explosives | | | | | | | | | | |
| Nitrocellulose | 9004-70-0 | 15/ 17 | 0.77 | 2 | 1.4 | None | Yes | Detected organic | | |

Table 4–27. SRC Screening Summary for NPA Subsurface Soil (continued)

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification | | |
|----------------------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|-------------------|--|--|
| | | | S | SVOCs | | | | | | |
| Benz(a)anthracene | 56-55-3 | 1/36 | 0.011 | 0.011 | 0.128 | None | No | <5% Detected | | |
| Benzo(a)pyrene | 50-32-8 | 2/36 | 0.011 | 0.12 | 0.126 | None | Yes | Detected organic | | |
| Benzo(b)fluoranthene | 205-99-2 | 3/36 | 0.0085 | 0.022 | 0.129 | None | Yes | Detected organic | | |
| Benzo(ghi)perylene | 191-24-2 | 1/36 | 0.01 | 0.01 | 0.129 | None | No | <5% Detected | | |
| Bis(2-ethylhexyl)phthalate | 117-81-7 | 1/25 | 0.024 | 0.024 | 0.19 | None | No | <5% Detected | | |
| Chrysene | 218-01-9 | 2/36 | 0.0082 | 0.01 | 0.128 | None | Yes | Detected organic | | |
| Di-n-butyl phthalate | 84-74-2 | 1/25 | 0.019 | 0.019 | 0.19 | None | No | <5% Detected | | |
| Fluoranthene | 206-44-0 | 2/36 | 0.025 | 0.16 | 0.128 | None | Yes | Detected organic | | |
| Phenanthrene | 85-01-8 | 2/36 | 0.01 | 0.013 | 0.129 | None | Yes | Detected organic | | |
| Pyrene | 129-00-0 | 2/36 | 0.0094 | 0.019 | 0.129 | None | Yes | Detected organic | | |
| VOCs | | | | | | | | | | |
| Carbon tetrachloride | 56-23-5 | 1/26 | 0.001 | 0.001 | 0.00298 | None | No | <5% Detected | | |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

SRC screening tables include all available and appropriate data as presented in Section 4.5.4.

CAS = Chemical Abstract Service.

mg/kg = Milligrams per kilogram.

NPA = Non-production area.

SRC = Site-related contaminant.

SVOC = Semi-volatile organic compound.

TPH = Total petroleum hydrocarbon.

VOC = Volatile organic compound.

Table 4–28. SRC Screening Summary for East Ditch Sediment

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification |
|-----------------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|--------------------|
| | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 2/2 | 10600 | 15100 | 12900 | 13900 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 1/2 | 0.13 | 0.13 | 0.155 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 1/2 | 19.7 | 19.7 | 10.5 | 19.5 | Yes | Exceeds background |
| Barium | 7440-39-3 | 2/2 | 85.2 | 98.2 | 91.7 | 123 | No | Below background |
| Beryllium | 7440-41-7 | 2/2 | 0.61 | 1 | 0.805 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 1/2 | 0.47 | 0.47 | 0.268 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 2/2 | 2390 | 3190 | 2790 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 2/2 | 16.5 | 18.6 | 17.6 | 18.1 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 2/2 | 8.2 | 17.4 | 12.8 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 2/2 | 19.7 | 31.4 | 25.6 | 27.6 | Yes | Exceeds background |
| Iron | 7439-89-6 | 2/2 | 19200 | 39300 | 29300 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 1/2 | 25.7 | 25.7 | 15.3 | 27.4 | No | Below background |
| Magnesium | 7439-95-4 | 2/2 | 2140 | 3660 | 2900 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 2/2 | 665 | 754 | 710 | 1950 | No | Below background |
| Nickel | 7440-02-0 | 2/2 | 15.2 | 32.6 | 23.9 | 17.7 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 2/2 | 856 | 1230 | 1040 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 2/2 | 2 | 3.2 | 2.6 | 1.7 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 2/2 | 68.1 | 817 | 443 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 1/2 | 0.14 | 0.14 | 0.13 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 2/2 | 21.4 | 24.2 | 22.8 | 26.1 | No | Below background |
| Zinc | 7440-66-6 | 2/2 | 77.6 | 234 | 156 | 532 | No | Below background |
| | • | | • | SVOCs | 1 | | • | |
| 2-Methylnaphthalene | 91-57-6 | 1/1 | 0.013 | 0.013 | 0.013 | None | Yes | Detected organic |
| Anthracene | 120-12-7 | 1/1 | 0.025 | 0.025 | 0.025 | None | Yes | Detected organic |
| Benz(a)anthracene | 56-55-3 | 1/1 | 0.11 | 0.11 | 0.11 | None | Yes | Detected organic |
| Benzo(a)pyrene | 50-32-8 | 1/1 | 0.098 | 0.098 | 0.098 | None | Yes | Detected organic |
| Benzo(b)fluoranthene | 205-99-2 | 1/1 | 0.13 | 0.13 | 0.13 | None | Yes | Detected organic |
| Benzo(ghi)perylene | 191-24-2 | 1/1 | 0.064 | 0.064 | 0.064 | None | Yes | Detected organic |
| Benzo(k)fluoranthene | 207-08-9 | 1/1 | 0.056 | 0.056 | 0.056 | None | Yes | Detected organic |
| Chrysene | 218-01-9 | 1/1 | 0.096 | 0.096 | 0.096 | None | Yes | Detected organic |
| Dibenz(a,h)anthracene | 53-70-3 | 1/1 | 0.017 | 0.017 | 0.017 | None | Yes | Detected organic |
| Fluoranthene | 206-44-0 | 1/1 | 0.23 | 0.23 | 0.23 | None | Yes | Detected organic |

Table 4–28. SRC Screening Summary for East Ditch Sediment (continued)

| | | | Minimum | Maximum | Average | Background | | |
|------------------------|----------|---------|---------|---------|---------|-----------------------|--------|-------------------|
| | CAS | Freq of | Detect | Detect | Result | Criteria ^a | SRC? | |
| Analyte | Number | Detect | (mg/kg) | (mg/kg) | (mg/kg) | (mg/kg) | yes/no | SRC Justification |
| Fluorene | 86-73-7 | 1/1 | 0.012 | 0.012 | 0.012 | None | Yes | Detected organic |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/1 | 0.056 | 0.056 | 0.056 | None | Yes | Detected organic |
| Naphthalene | 91-20-3 | 1/1 | 0.01 | 0.01 | 0.01 | None | Yes | Detected organic |
| Phenanthrene | 85-01-8 | 1/1 | 0.1 | 0.1 | 0.1 | None | Yes | Detected organic |
| Pyrene | 129-00-0 | 1/1 | 0.18 | 0.18 | 0.18 | None | Yes | Detected organic |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

mg/kg = Milligrams per kilogram.
SRC = Site-related contaminant.
SVOC= Semi-volatile organic compound.

SRC screening tables include all available and appropriate data as presented in Section 4.5.4.

CAS = Chemical Abstract Service.

Table 4–29. SRC Screening Summary for West Ditch Sediment

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification |
|-------------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|--------------------|
| | | | | Metals | | | | |
| Aluminum | 7429-90-5 | 4/4 | 10100 | 17700 | 14600 | 13900 | Yes | Exceeds background |
| Antimony | 7440-36-0 | 3/3 | 0.14 | 0.9 | 0.483 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 4/4 | 7.9 | 19.5 | 12.3 | 19.5 | No | Below background |
| Barium | 7440-39-3 | 4/4 | 64 | 111 | 86.3 | 123 | No | Below background |
| Beryllium | 7440-41-7 | 4/4 | 0.61 | 1 | 0.835 | 0.38 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 1/4 | 0.51 | 0.51 | 0.231 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 4/4 | 2170 | 3800 | 2900 | 5510 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 4/4 | 12.2 | 21 | 17.7 | 18.1 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 4/4 | 7.1 | 14.1 | 9.63 | 9.1 | Yes | Exceeds background |
| Copper | 7440-50-8 | 4/4 | 14.6 | 26.3 | 20.1 | 27.6 | No | Below background |
| Iron | 7439-89-6 | 4/4 | 17100 | 30000 | 24500 | 28200 | No | Essential Nutrient |
| Lead | 7439-92-1 | 4/4 | 14.7 | 27 | 20.9 | 27.4 | No | Below background |
| Magnesium | 7439-95-4 | 4/4 | 1820 | 3630 | 2810 | 2760 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 4/4 | 164 | 530 | 289 | 1950 | No | Below background |
| Mercury | 7439-97-6 | 2/4 | 0.049 | 0.1 | 0.05 | 0.059 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 4/4 | 15.6 | 27.7 | 21 | 17.7 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 4/4 | 685 | 2010 | 1440 | 1950 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 4/4 | 0.5 | 1.7 | 1.07 | 1.7 | No | Below background |
| Silver | 7440-22-4 | 1/4 | 0.072 | 0.072 | 0.155 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 4/4 | 35 | 1060 | 517 | 112 | No | Essential Nutrient |
| Thallium | 7440-28-0 | 3/4 | 0.13 | 0.23 | 0.17 | 0.89 | No | Below background |
| Vanadium | 7440-62-2 | 4/4 | 17.8 | 30.8 | 24.9 | 26.1 | Yes | Exceeds background |
| Zinc | 7440-66-6 | 4/4 | 65.3 | 120 | 86.4 | 532 | No | Below background |
| | | • | | Anions | | | • | |
| Sulfide | 18496-25-8 | 1/1 | 43.4 | 43.4 | 43.4 | None | Yes | Exceeds background |
| | | • | E | Explosives | | | • | |
| Nitrocellulose | 9004-70-0 | 1/1 | 1.4 | 1.4 | 1.4 | None | Yes | Detected organic |
| | • | - | 1 | SVOCs | <u>'</u> | | • | |
| Acenaphthene | 83-32-9 | 1/2 | 0.013 | 0.013 | 0.026 | None | Yes | Detected organic |
| Anthracene | 120-12-7 | 1/2 | 0.02 | 0.02 | 0.0295 | None | Yes | Detected organic |
| Benz(a)anthracene | 56-55-3 | 2/2 | 0.019 | 0.077 | 0.048 | None | Yes | Detected organic |
| Benzenemethanol | 100-51-6 | 1/2 | 0.054 | 0.054 | 0.155 | None | Yes | Detected organic |

Table 4–29. SRC Screening Summary for West Ditch Sediment (continued)

| Analyte | CAS Number | Freq of Detect | Minimum Detect (mg/kg) | Maximum Detect (mg/kg) | Average Result (mg/kg) | Background Criteria ^a (mg/kg) | SRC? yes/no | SRC Justification |
|------------------------|---------------|-------------------|------------------------------|------------------------------|------------------------------|--|----------------|-------------------|
| Benzo(a)pyrene | 50-32-8 | 2/2 | 0.021 | 0.09 | 0.0555 | None | Yes | Detected organic |
| Benzo(b)fluoranthene | 205-99-2 | 2/2 | 0.038 | 0.14 | 0.089 | None | Yes | Detected organic |
| Benzo(ghi)perylene | 191-24-2 | 2/2 | 0.017 | 0.068 | 0.0425 | None | Yes | Detected organic |
| Benzo(k)fluoranthene | 207-08-9 | 2/2 | 0.011 | 0.066 | 0.0385 | None | Yes | Detected organic |
| Chrysene | 218-01-9 | 2/2 | 0.023 | 0.099 | 0.061 | None | Yes | Detected organic |
| Dibenz(a,h)anthracene | 53-70-3 | 1/2 | 0.016 | 0.016 | 0.0275 | None | Yes | Detected organic |
| Fluoranthene | 206-44-0 | 2/2 | 0.043 | 0.21 | 0.127 | None | Yes | Detected organic |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 2/2 | 0.015 | 0.061 | 0.038 | None | Yes | Detected organic |
| Phenanthrene | 85-01-8 | 2/2 | 0.016 | 0.098 | 0.057 | None | Yes | Detected organic |
| Pyrene | 129-00-0 | 2/2 | 0.033 | 0.16 | 0.0965 | None | Yes | Detected organic |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

SVOC= Semi-volatile organic compound.

SRC screening tables include all available and appropriate data as presented in Section 4.5.4.

CAS = Chemical Abstract Service.

mg/kg = Milligrams per kilogram. SRC = Site-related contaminant.

Table 4–30. SRC Screening Summary for Sewer Outfall Sediment

| | CAS | Freq of | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? | | | | | | |
|---|------------|---------|-------------------|-------------------|-------------------|-------------------------------------|--------|--------------------|--|--|--|--|--|
| Analyte | Number | Detect | (mg/kg) | (mg/kg) | (mg/kg) | (mg/kg) | yes/no | SRC Justification | | | | | |
| | | | | Metals | | | | | | | | | |
| Aluminum 7429-90-5 1/1 4300 4300 4300 13900 No Below background | | | | | | | | | | | | | |
| Arsenic | 7440-38-2 | 1/1 | 12.8 | 12.8 | 12.8 | 19.5 | No | Below background | | | | | |
| Barium | 7440-39-3 | 1/1 | 48 | 48 | 48 | 123 | No | Below background | | | | | |
| Calcium | 7440-70-2 | 1/1 | 2960 | 2960 | 2960 | 5510 | No | Essential Nutrient | | | | | |
| Chromium | 7440-47-3 | 1/1 | 7.1 | 7.1 | 7.1 | 18.1 | No | Below background | | | | | |
| Cobalt | 7440-48-4 | 1/1 | 5.3 | 5.3 | 5.3 | 9.1 | No | Below background | | | | | |
| Copper | 7440-50-8 | 1/1 | 13.4 | 13.4 | 13.4 | 27.6 | No | Below background | | | | | |
| Iron | 7439-89-6 | 1/1 | 12100 | 12100 | 12100 | 28200 | No | Essential Nutrient | | | | | |
| Lead | 7439-92-1 | 1/1 | 19.3 | 19.3 | 19.3 | 27.4 | No | Below background | | | | | |
| Magnesium | 7439-95-4 | 1/1 | 1340 | 1340 | 1340 | 2760 | No | Essential Nutrient | | | | | |
| Manganese | 7439-96-5 | 1/1 | 1760 | 1760 | 1760 | 1950 | No | Below background | | | | | |
| Nickel | 7440-02-0 | 1/1 | 11.1 | 11.1 | 11.1 | 17.7 | No | Below background | | | | | |
| Potassium | 7440-09-7 | 1/1 | 521 | 521 | 521 | 1950 | No | Essential Nutrient | | | | | |
| Sodium | 7440-23-5 | 1/1 | 396 | 396 | 396 | 112 | No | Essential Nutrient | | | | | |
| Vanadium | 7440-62-2 | 1/1 | 8.9 | 8.9 | 8.9 | 26.1 | No | Below background | | | | | |
| Zinc | 7440-66-6 | 1/1 | 47.2 | 47.2 | 47.2 | 532 | No | Below background | | | | | |
| | | | | Anions | | | | - | | | | | |
| Sulfate | 14808-79-8 | 1/1 | 248 | 248 | 248 | None | Yes | Exceeds background | | | | | |
| Sulfide | 18496-25-8 | 1/1 | 540 | 540 | 540 | None | Yes | Exceeds background | | | | | |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

SRC screening tables include all available and appropriate data as presented in Section 4.5.4.

CAS = Chemical Abstract Service.

mg/kg = Milligrams per kilogram.

SRC = Site-related contaminant.

Table 4–31. SRC Screening for East Ditch Surface Water

| | CAS | Freq of | Minimum | Maximum | Average | Background | SRC? | |
|-----------------|-----------|---------|----------|----------------|-----------|-----------------------|----------|--------------------|
| Analyte | Number | Detect | Detect | Detect | Result | Criteria ^a | (yes/no) | SRC Justification |
| | | | | Metals (m | g/L) | | | |
| Aluminum | 7429-90-5 | 1/1 | 0.674 | 0.674 | 0.674 | 3.37 | No | Below background |
| Arsenic | 7440-38-2 | 1/1 | 0.00058 | 0.00058 | 0.00058 | 0.0032 | No | Below background |
| Barium | 7440-39-3 | 1/1 | 0.0136 | 0.0136 | 0.0136 | 0.0475 | No | Below background |
| Calcium | 7440-70-2 | 1/1 | 33.1 | 33.1 | 33.1 | 41.4 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 1/1 | 0.00078 | 0.00078 | 0.00078 | 0 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 1/1 | 0.00017 | 0.00017 | 0.00017 | 0 | Yes | Exceeds background |
| Copper | 7440-50-8 | 1/1 | 0.0014 | 0.0014 | 0.0014 | 0.0079 | No | Below background |
| Iron | 7439-89-6 | 1/1 | 0.637 | 0.637 | 0.637 | 2.56 | No | Essential Nutrient |
| Lead | 7439-92-1 | 1/1 | 0.00044 | 0.00044 | 0.00044 | 0 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 1/1 | 5.59 | 5.59 | 5.59 | 10.8 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 1/1 | 0.0161 | 0.0161 | 0.0161 | 0.391 | No | Below background |
| Nickel | 7440-02-0 | 1/1 | 0.00091 | 0.00091 | 0.00091 | 0 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 1/1 | 1.21 | 1.21 | 1.21 | 3.17 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 1/1 | 0.00026 | 0.00026 | 0.00026 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 1/1 | 1.86 | 1.86 | 1.86 | 21.3 | No | Essential Nutrient |
| Vanadium | 7440-62-2 | 1/1 | 0.0011 | 0.0011 | 0.0011 | 0 | Yes | Exceeds background |
| | | | | Pesticides/PCI | Bs (mg/L) | | | |
| beta-BHC | 319-85-7 | 1/1 | 0.000013 | 0.000013 | 0.000013 | None | Yes | Detected organic |
| gamma-Chlordane | 5103-74-2 | 1/1 | 0.000015 | 0.000015 | 0.000015 | None | Yes | Detected organic |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

BHC = Hexachlorocyclohexane.

CAS = Chemical Abstract Service.

mg/L = Milligrams per liter.

PCB = Polychlorinated biphenyl.

SRC = Site-related contaminant.

Table 4–32. SRC Screening for West Ditch Surface Water

| | CAS | Freq of | Minimum | Maximum | Average | Background | SRC? | |
|------------------------|-----------|---------|---------|------------|---------|-----------------------|----------|--------------------|
| Analyte | Number | Detect | Detect | Detect | Result | Criteria ^a | (yes/no) | SRC Justification |
| | | | | Metals (mg | ·/L) | | | |
| Aluminum | 7429-90-5 | 1/ 1 | 2.72 | 2.72 | 2.72 | 3.37 | No | Below background |
| Antimony | 7440-36-0 | 1/1 | 0.00036 | 0.00036 | 0.00036 | 0 | Yes | Exceeds background |
| Arsenic | 7440-38-2 | 1/1 | 0.0043 | 0.0043 | 0.0043 | 0.0032 | Yes | Exceeds background |
| Barium | 7440-39-3 | 1/1 | 0.0576 | 0.0576 | 0.0576 | 0.0475 | Yes | Exceeds background |
| Beryllium | 7440-41-7 | 1/1 | 0.00011 | 0.00011 | 0.00011 | 0 | Yes | Exceeds background |
| Cadmium | 7440-43-9 | 1/1 | 0.00024 | 0.00024 | 0.00024 | 0 | Yes | Exceeds background |
| Calcium | 7440-70-2 | 1/ 1 | 61.9 | 61.9 | 61.9 | 41.4 | No | Essential Nutrient |
| Chromium | 7440-47-3 | 1/1 | 0.0033 | 0.0033 | 0.0033 | 0 | Yes | Exceeds background |
| Cobalt | 7440-48-4 | 1/1 | 0.0019 | 0.0019 | 0.0019 | 0 | Yes | Exceeds background |
| Copper | 7440-50-8 | 1/ 1 | 0.0062 | 0.0062 | 0.0062 | 0.0079 | No | Below background |
| Iron | 7439-89-6 | 1/ 1 | 5.83 | 5.83 | 5.83 | 2.56 | No | Essential Nutrient |
| Lead | 7439-92-1 | 1/1 | 0.0062 | 0.0062 | 0.0062 | 0 | Yes | Exceeds background |
| Magnesium | 7439-95-4 | 1/1 | 12.8 | 12.8 | 12.8 | 10.8 | No | Essential Nutrient |
| Manganese | 7439-96-5 | 1/1 | 0.559 | 0.559 | 0.559 | 0.391 | Yes | Exceeds background |
| Nickel | 7440-02-0 | 1/1 | 0.0039 | 0.0039 | 0.0039 | 0 | Yes | Exceeds background |
| Potassium | 7440-09-7 | 1/1 | 1.92 | 1.92 | 1.92 | 3.17 | No | Essential Nutrient |
| Selenium | 7782-49-2 | 1/1 | 0.00073 | 0.00073 | 0.00073 | 0 | Yes | Exceeds background |
| Sodium | 7440-23-5 | 1/1 | 6.26 | 6.26 | 6.26 | 21.3 | No | Essential Nutrient |
| Vanadium | 7440-62-2 | 1/1 | 0.0051 | 0.0051 | 0.0051 | 0 | Yes | Exceeds background |
| | | | | SVOCs (mg | g/L) | | | |
| Benz(a)anthracene | 56-55-3 | 1/1 | 0.00035 | 0.00035 | 0.00035 | None | Yes | Detected organic |
| Benzo(a)pyrene | 50-32-8 | 1/1 | 0.00033 | 0.00033 | 0.00033 | None | Yes | Detected organic |
| Benzo(b)fluoranthene | 205-99-2 | 1/1 | 0.00041 | 0.00041 | 0.00041 | None | Yes | Detected organic |
| Benzo(ghi)perylene | 191-24-2 | 1/1 | 0.00024 | 0.00024 | 0.00024 | None | Yes | Detected organic |
| Benzo(k)fluoranthene | 207-08-9 | 1/1 | 0.00024 | 0.00024 | 0.00024 | None | Yes | Detected organic |
| Chrysene | 218-01-9 | 1/1 | 0.00035 | 0.00035 | 0.00035 | None | Yes | Detected organic |
| Fluoranthene | 206-44-0 | 1/1 | 0.00073 | 0.00073 | 0.00073 | None | Yes | Detected organic |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | 1/1 | 0.00021 | 0.00021 | 0.00021 | None | Yes | Detected organic |
| Phenanthrene | 85-01-8 | 1/1 | 0.00022 | 0.00022 | 0.00022 | None | Yes | Detected organic |
| Pyrene | 129-00-0 | 1/1 | 0.00057 | 0.00057 | 0.00057 | None | Yes | Detected organic |

Table 4–32. SRC Screening for West Ditch Surface Water (continued)

| Analyte | CAS Number | Freq of Detect | Minimum Detect | Maximum Detect | Average Result | Background Criteria ^a | SRC? (yes/no) | SRC Justification |
|----------|---------------|-------------------|-------------------|-------------------|-------------------|-------------------------------------|------------------|-------------------|
| | | | | Pesticides/PCBs | (mg/L) | | | |
| beta-BHC | 319-85-7 | 1/1 | 0.0000096 | 0.0000096 | 0.0000096 | None | Yes | Detected organic |

^a Background concentrations are published in the Phase II Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b).

BHC = Hexachlorocyclohexane.

CAS = Chemical Abstract Service.

mg/L = Milligrams per liter.

PCB = Polychlorinated biphenyl.

SRC = Site-related Contaminant.

SVOC= Semi-volatile organic compound.

Table 4–33. Data Summary and Designated Use for RI

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|--------------------|------|----------|------------|------------------|-----------|-----|-----|------|-----|---|
| | | | | Surface and Subs | urface So | il | | | | |
| RV-264A | D | 01/31/96 | 0-0.5 | 1997 RRSE | 1 | - | | | | Used for initial evaluation of site. |
| RV-441 | D | 10/23/98 | 0-0.5 | 1998 RRSE | | | | | | Used for initial evaluation of site. |
| RV-442 | D | 10/23/98 | 0-0.5 | 1998 RRSE | | | | | | Used for initial evaluation of site. |
| RV-443 | D | 10/23/98 | 0-0.5 | 1998 RRSE | | | | | | Used for initial evaluation of site. |
| RV-444 | D | 10/23/98 | 0-0.5 | 1998 RRSE | | | | | | Used for initial evaluation of site. |
| RV-445 | D | 10/23/98 | 0-0.5 | 1998 RRSE | | | | | | Used for initial evaluation of site. |
| RV-446 | D | 10/23/98 | 0-0.5 | 1998 RRSE | | | | | | Used for initial evaluation of site. |
| LL11cs-001-0001-SO | D | 01/24/01 | 5–6 | IRA | | X | X | X | | Confirmation sample – Sump excavation. |
| LL11cs-002-0001-FD | D | 01/26/01 | 5–6 | IRA | X | | | | | Field duplicate. |
| LL11cs-002-0001-SO | D | 01/26/01 | 5–6 | IRA | | X | X | X | | Confirmation sample – Sump excavation. |
| LL11cs-003-0001-SO | D | 01/26/01 | 5–6 | IRA | | X | X | X | | Confirmation sample – Sump excavation. |
| LL11cs-004-0001-SO | D | 02/21/01 | 5–6 | IRA | | X | X | X | | Confirmation sample – Sump excavation. |
| LL11cs-005-0001-SO | D | 02/21/01 | 5–6 | IRA | | X | X | X | | Confirmation sample – Sump excavation. |
| LL11cs-006-0001-SD | D | 03/21/01 | 0–1 | IRA | | X | X | X | X | Confirmation sample – Ditch southeast of Building AP-3. |
| LL11cs-007-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | Confirmation sample – Ditch southeast of Building AP-3. |
| LL11cs-008-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | Confirmation sample – Ditch southeast of Building AP-3. |
| LL11cs-009-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | Confirmation sample – Ditch southeast of Building AP-3. |
| LL11cs-010-0001-SD | D | 03/21/01 | 0–1 | IRA | | X | X | X | X | Confirmation sample – Ditch southeast of Building AP-3. |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|--------------------|------|---------------|------------|----------------|-------|-----|-----|------|-----|---|
| LL11cs-011-0001-FD | D | 03/23/01 | 0–1 | IRA | X | | | | | Field duplicate. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-011-0001-SD | D | 03/23/01 | 0–1 | IRA | | X | X | X | X | east of Building AP-4. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-012-0001-SO | D | 03/23/01 | 1–2 | IRA | | X | X | X | | east of Building AP-4. |
| | _ | 00/00/01 | | | | | | | | Confirmation sample – Ditch |
| LL11cs-013-0001-SO | D | 03/23/01 | 1–2 | IRA | | X | X | X | | east of Building AP-4. |
| 1111 014 0001 00 | - | 02/22/01 | 1.0 | TD 4 | | *** | 37 | *** | | Confirmation sample – Ditch |
| LL11cs-014-0001-SO | D | 03/23/01 | 1–2 | IRA | | X | X | X | | east of Building AP-4. |
| 1111 016 0001 CD | ъ | 02/22/01 | 0 1 | TD 4 | | 37 | 37 | 37 | 37 | Confirmation sample – Ditch |
| LL11cs-016-0001-SD | D | 03/23/01 | 0-1 | IRA | X | X | X | X | X | north of Building AP-14. |
| LL11cs-017-0001-FD | D | 03/23/01 | 1–2 | IRA | X | | | | | Field duplicate. |
| LL11cs-017-0001-SO | D | 03/23/01 | 1–2 | IRA | | X | X | X | | Confirmation sample – Ditch |
| LL11cs-01/-0001-SO | ע | 03/23/01 | 1-2 | IKA | | Λ | Λ | Λ | | north of Building AP-14. Confirmation sample – Ditch |
| LL11cs-018-0001-SO | D | 03/23/01 | 1–2 | IRA | | X | X | X | | north of Building AP-14. |
| LL11CS-018-0001-SO | ע | 03/23/01 | 1-2 | IKA | | Λ | Λ | Λ | | Confirmation sample – Ditch |
| LL11cs-019-0001-SO | D | 03/23/01 | 1–2 | IRA | | X | X | X | | north of Building AP-14. |
| EE11cs-019-0001-30 | D | 03/23/01 | 1-2 | IICA | | Λ | Λ | Λ | | Confirmation sample – Ditch |
| LL11cs-020-0001-SO | D | 03/23/01 | 0–1 | IRA | | X | X | X | X | north of Building AP-14. |
| EE1163 020 0001 50 | Ъ | 03/23/01 | 0 1 | IIU I | | 71 | 71 | 21 | 71 | Confirmation sample – Ditch |
| | | | | | | | | | | east of Load Line 11 entrance |
| LL11cs-021-0001-SD | D | 03/21/01 | 0–1 | IRA | | X | X | X | X | road. |
| | | | | | | | | | | Confirmation sample – Ditch |
| | | | | | | | | | | east of Load Line 11 entrance |
| LL11cs-022-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | road. |
| | | | | | | | | | | Confirmation sample – Ditch |
| | | | | | | | | | | east of Load Line 11 entrance |
| LL11cs-023-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | road. |
| | | | | | | | | | | Confirmation sample – Ditch |
| | | | | | | | | | | east of Load Line 11 entrance |
| LL11cs-024-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | road. |
| | | | | | | | | | | Confirmation sample – Ditch |
| | | | | | | | | | | east of Load Line 11 entrance |
| LL11cs-025-0001-SD | D | 03/21/01 | 0–1 | IRA | | X | X | X | X | road. |
| | | 0.0 (0.1 (0.1 | | | | | | | | Confirmation sample – Ditch |
| LL11cs-026-0001-SD | D | 03/21/01 | 1–2 | IRA | | X | X | X | | west of Building AP-8. |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|--------------------|------|----------|------------|----------------|-------------|-----|-------|-------|-----|---|
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-027-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | west of Building AP-8. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-028-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | west of Building AP-8. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-029-0001-SO | D | 03/21/01 | 1–2 | IRA | | X | X | X | | west of Building AP-8. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-030-0001-SD | D | 03/21/01 | 0–1 | IRA | | X | X | X | X | west of Building AP-8. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-032-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-033-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| | | | | | | | | | | Confirmation sample – Ditch |
| LL11cs-034-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| 1111 025 0001 00 | - | 02/20/01 | 1.0 | TD 4 | | *** | *** | 7.7 | | Confirmation sample – Ditch |
| LL11cs-035-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| 1111 026 0001 00 | - | 02/20/01 | 1.0 | TD 4 | | *** | *** | 7.7 | | Confirmation sample – Ditch |
| LL11cs-036-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| 1111 027 0001 00 | ъ | 02/20/01 | 1 2 | TD 4 | | 37 | 37 | 37 | | Confirmation sample – Ditch |
| LL11cs-037-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| 1111 029 0001 90 | Ъ | 02/20/01 | 1.2 | ID A | | v | v | V | | Confirmation sample – Ditch |
| LL11cs-038-0001-SO | D | 03/20/01 | 1–2 | IRA | | X | X | X | | north of Building AP-19. |
| LL11cs-039-0001-SD | D | 03/20/01 | 0–1 | IRA | | v | X | X | X | Confirmation sample – Ditch |
| LL11cs-040-0001-FD | D | 03/20/01 | 3–7 | IRA | X | X | Λ | Λ | | north of Building AP-19. Field duplicate. |
| LL11CS-040-0001-FD | D | 03/22/01 | 3-7 | IKA | Λ | | | | | Confirmation sample – Hot |
| LL11cs-040-0001-SO | D | 03/22/01 | 3–7 | IRA | | X | X | X | | spot excavation. |
| LL11cs-040-0001-SO | ש | 03/22/01 | 3-7 | IKA | | Λ | Λ | Λ | | Confirmation sample – Hot |
| LL11cs-041-0001-SO | D | 03/22/01 | 3–7 | IRA | | X | X | X | | spot excavation. |
| LL11cs-041-0001-SO | ע | 03/22/01 | 3-7 | IKA | | Λ | Λ | Λ | 1 | Confirmation sample – Hot |
| LL11cs-042-0001-SO | D | 03/22/01 | 3–7 | IRA | | X | X | X | | spot excavation. |
| LL11cs-042-0001-SO | ע | 03/22/01 | 3-7 | IKA | | Λ | Λ | Λ | 1 | Confirmation sample – Hot |
| LL11cs-043-0001-SO | D | 03/22/01 | 3–7 | IRA | | X | X | X | | spot excavation. |
| LL1105-043-0001-30 | ע | 03/44/01 | 3-1 | IIVA | | Λ | Λ | Λ | | Confirmation sample – Hot |
| LL11cs-044-0001-SO | D | 03/22/01 | 7–8 | IRA | | X | X | X | | spot excavation. |
| LL11cs-045-0001-50 | D | 03/22/01 | 7–8 | IRA | X | | | | | Field duplicate. |
| LL11cs-045-0001-FD | D | 03/22/01 | 7–8 | IRA | | X | X | X | | Confirmation sample – Hot |
| LL1105-043-0001-30 | ע | 03/44/01 | 7-0 | шл | L | Λ | Λ | Λ | | Commination sample – Hot |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|---------------------|------|----------|------------|----------------|----|-----|-----|------|-----|----------------------------|
| | | | | | | | | | | spot excavation. |
| | | | | | | | | | | Confirmation sample – Test |
| LL11cs-046-0001-SO | D | 03/22/01 | 3–4 | IRA | | X | X | X | | trenches. |
| | | | | | | | | | | Confirmation sample – Test |
| LL11cs-047-0001-SO | D | 03/22/01 | 3–4 | IRA | | X | X | X | | trenches. |
| | | | | | | | | | | Confirmation sample – Test |
| LL11cs-048-0001-SO | D | 03/22/01 | 3–4 | IRA | | X | X | X | | trenches. |
| | | | | | | | | | | Confirmation sample – Test |
| LL11cs-049-0001-SO | D | 03/22/01 | 3–4 | IRA | | X | X | X | | trenches. |
| | _ | 00/00/01 | | | | | | | | Confirmation sample – Test |
| LL11cs-050-0001-SO | D | 03/22/01 | 3–4 | IRA | | X | X | X | | trenches. |
| LL11sb-001-0001-FD | D | 10/31/00 | 0-1 | Phase I RI | X | | | | | Field duplicate. |
| LL11sb-001-0001-SO | D | 10/31/00 | 0–1 | Phase I RI | | X | X | X | X | |
| TT 11 1 001 0000 00 | | 11/02/00 | 10.20 | DI IDI | | 37 | 77 | | | Sample below the exposure |
| LL11sb-001-0002-SO | D | 11/03/00 | 18–20 | Phase I RI | | X | X | | | depth of all receptors. |
| LL11sb-002-0001-SO | D | 10/30/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sb-002-0002-SO | D | 11/10/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-003-0001-SO | D | 10/30/00 | 0–1 | Phase I RI | | X | X | X | X | |
| TT 11 1 000 0000 00 | | 11/00/00 | 20.40 | DI IDI | | 37 | 77 | | | Sample below the exposure |
| LL11sb-003-0002-SO | D | 11/09/00 | 20–40 | Phase I RI | | X | X | | | depth of all receptors. |
| LL11sb-004-0001-SO | D | 10/31/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sb-004-0002-SO | D | 11/08/00 | 10–12 | Phase I RI | | X | X | X | | |
| LL11sb-005-0001-SO | D | 10/31/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sb-005-0002-SO | D | 11/09/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-006-0001-SO | D | 10/31/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sb-006-0002-SO | D | 11/13/00 | 2–4 | Phase I RI | | X | X | X | | |
| LL11sb-007-0001-SO | D | 10/31/00 | 0–1 | Phase I RI | | X | X | X | X | |
| | _ | 1110=100 | | | | | | | | Sample below the exposure |
| LL11sb-007-0002-SO | D | 11/07/00 | 14–16 | Phase I RI | | X | X | | | depth of all receptors. |
| LL11sb-008-0001-SO | D | 10/31/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sb-008-0002-FD | D | 11/08/00 | 12–13 | Phase I RI | X | | | | | Field duplicate. |
| LL11sb-008-0002-SO | D | 11/08/00 | 12–13 | Phase I RI | | X | X | X | | |
| LL11sb-009-0001-SO | D | 10/30/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sb-009-0002-SO | D | 11/13/00 | 4–6 | Phase I RI | | X | X | X | | |
| LL11sb-010-0001-SO | D | 10/31/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sb-010-0002-SO | D | 11/07/00 | 14–16 | Phase I RI | | X | X | | | Sample below the exposure |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|---------------------|------|----------|------------|----------------|----|-----|-----|------|-----|------------------------------|
| | | | | | | | | | | depth of all receptors. |
| LL11sb-011-0001-SO | D | 08/21/00 | 4–6 | Phase I RI | | X | X | X | | |
| LL11sb-012-0001-FD | D | 08/22/00 | 8–10 | Phase I RI | X | | | | | Field duplicate. |
| LL11sb-012-0001-SO | D | 08/22/00 | 8–10 | Phase I RI | | X | X | X | | |
| LL11sb-013-0001-SO | D | 08/22/00 | 8–10 | Phase I RI | | X | X | X | | |
| LL11sb-014-0001-SO | D | 08/22/00 | 11–13 | Phase I RI | | X | X | X | | |
| LL11sb-015-0001-SO | D | 08/22/00 | 10–12 | Phase I RI | | X | X | X | | |
| LL11sb-016-0001-SO | D | 08/22/00 | 8.5–11.5 | Phase I RI | | X | X | X | | |
| LL11sb-017-0001-SO | D | 08/22/00 | 4–6 | Phase I RI | | X | X | X | | |
| LL11sb-018-0001-SO | D | 08/22/00 | 4–6 | Phase I RI | | X | X | X | | |
| LL11sb-019-0001-FD | D | 08/21/00 | 0–1 | Phase I RI | X | | | | | Field duplicate. |
| LL11sb-019-0001-SO | D | 08/21/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sb-019-0002-SO | D | 08/21/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-020-0001-SO | D | 08/21/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sb-020-0002-SO | D | 08/21/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-021-0001-SO | D | 08/21/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sb-021-0002-SO | D | 08/21/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-022-0001-SO | D | 08/22/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sb-022-0002-SO | D | 08/22/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-023-0001-SO | D | 08/22/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sb-023-0002-SO | D | 08/22/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-024-0001-SO | D | 11/16/00 | 7–9 | Phase I RI | | X | X | X | | |
| LL11sb-025-0001-SO | D | 11/17/00 | 5–7 | Phase I RI | | X | X | X | | |
| LL11sb-026-0001-SO | D | 11/16/00 | 4–6 | Phase I RI | | X | X | X | | |
| LL11sb-027-0001-FD | D | 11/16/00 | 6–8 | Phase I RI | X | | | | | Field duplicate. |
| LL11sb-027-0001-SO | D | 11/16/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-028-0001-SO | D | 11/16/00 | 4–6 | Phase I RI | | X | X | X | | |
| LL11sb-029-0001-SO | D | 11/16/00 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-030-0001-SO | D | 11/17/00 | 8-10 | Phase I RI | | X | X | X | | |
| LL11sb-031-0001-SO | D | 11/16/00 | 4–6 | Phase I RI | | | | | | Excavated hot spot location. |
| LL11sb-031A-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | | | | | Excavated hot spot location. |
| LL11sb-032-0001-SO | D | 11/17/00 | 4–6 | Phase I RI | | | | | | Excavated hot spot location. |
| LL11sb-033-0001-SO | D | 11/17/00 | 6–8 | Phase I RI | | | | | | Excavated hot spot location. |
| LL11sb-034-0001-SO | D | 11/17/00 | 2–4 | Phase I RI | | | | | | Excavated hot spot location. |
| LL11sb-035-0001-FD | D | 11/17/00 | 4–6 | Phase I RI | X | | | | | Field duplicate. |
| LL11sb-035-0001-SO | D | 11/17/00 | 4–6 | Phase I RI | | X | X | X | | |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|--------------------|------|----------|------------|----------------|----|-----|-----|------|-----|------------------------------|
| LL11sb-036-0001-SO | D | 11/17/00 | 4–6 | Phase I RI | | | | | | Excavated hot spot location. |
| LL11sb-037-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-038-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-039-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-040-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-041-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-042-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-043-0001-SO | D | 03/13/01 | 6–8 | Phase I RI | | X | X | X | | |
| LL11sb-060-5551-SO | D | 03/18/10 | 0–1 | PBA08 RI | | X | X | X | X | |
| LL11sb-060-5552-SO | D | 03/18/10 | 1–4 | PBA08 RI | | X | X | X | | |
| LL11sb-060-5553-SO | D | 03/18/10 | 4–7 | PBA08 RI | | X | X | X | | |
| LL11sb-061-5555-SO | D | 03/17/10 | 0–1 | PBA08 RI | | X | X | X | X | |
| LL11sb-061-5556-SO | D | 03/17/10 | 1–4 | PBA08 RI | | X | X | X | | |
| LL11sb-061-5557-SO | D | 03/17/10 | 4–7 | PBA08 RI | | X | X | X | | |
| LL11sb-062-5559-SO | D | 03/18/10 | 0–1 | PBA08 RI | | X | X | X | X | |
| LL11sb-062-5560-SO | D | 03/18/10 | 1–4 | PBA08 RI | | X | X | X | | |
| LL11sb-062-5561-SO | D | 03/18/10 | 4–7 | PBA08 RI | | X | X | X | | |
| LL11sb-062-6189-FD | D | 03/18/10 | 0-1 | PBA08 RI | X | | | | | Field duplicate. |
| LL11sb-063-5563-SO | D | 03/22/10 | 0-1 | PBA08 RI | | X | X | X | X | |
| LL11sb-063-5564-SO | D | 03/22/10 | 1–4 | PBA08 RI | | X | X | X | | |
| LL11sb-063-5565-SO | D | 03/22/10 | 4–5 | PBA08 RI | | X | X | X | | |
| LL11sb-064-5569-SO | D | 03/18/10 | 0-1 | PBA08 RI | | X | X | X | X | |
| LL11sb-064-5570-SO | D | 03/18/10 | 1–4 | PBA08 RI | | X | X | X | | |
| LL11sb-064-5571-SO | D | 03/18/10 | 4–7 | PBA08 RI | | X | X | X | | |
| | | | | Sedime | nt | | | | | |
| | | | | IRA | | | | | | Confirmation sample – Ditch |
| LL11cs-015-0001-SD | D | 03/23/01 | 0–1 | | | X | X | X | X | east of Building AP-4. |
| LL11cs-031-0001-FD | D | 03/20/01 | 0–1 | IRA | X | | | | | Field duplicate. |
| | | | | IRA | | | | | | Confirmation sample – Ditch |
| LL11cs-031-0001-SD | D | 03/20/01 | 0–1 | | | X | X | X | X | north of Building AP-19. |
| | | | | Phase I RI | | | | | | Excavated former sump |
| LL11sd-002-0001-SD | D | 08/02/00 | 3–3 | | | | | | | location. |
| | | | | Phase I RI | | | | | | Excavated former sump |
| LL11sd-003-0001-SD | D | 08/02/00 | 3–3 | | | | | | | location. |
| | | | | Phase I RI | | | | | | Excavated former sump |
| LL11sd-005-0001-SD | D | 10/09/00 | 7–7 | | | | | | | location. |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|---------------------------------------|------|-----------|------------|-------------------|-------|-----|-----|------|-----|--------------------------------|
| LL11sd-007-0001-SD | D | 07/31/00 | 10.6-10.6 | Phase I RI | | | | | | Sample from sewer system. |
| LL11sd-008-0001-SD | D | 08/01/00 | 12-12 | Phase I RI | | | | | | Sample from sewer system. |
| LL11sd-009-0001-SD | D | 08/01/00 | 16–16 | Phase I RI | | | | | | Sample from sewer system. |
| LL11sd-010-0001-SD | D | 07/31/00 | 8–8 | Phase I RI | | | | | | Sample from sewer system. |
| LL11sd-011-0001-SD | D | 08/01/00 | 12-12 | Phase I RI | | | | | | Sample from sewer system. |
| LL11sd-017-0001-FD | D | 11/15/00 | 0–1 | Phase I RI | X | | | - | | Field duplicate. |
| | | | | Phase I RI | | | | | | Sample not used for HHRA |
| | | | | | | | | | | and ERA because more |
| | | | | | | | | | | recent sample was taken at |
| LL11sd-017-0001-SD | D | 11/15/00 | 0-1 | | | X | | | | the same location. |
| LL11sd-018-0001-SD | D | 11/15/00 | 0-0.5 | Phase I RI | | | | | | Excavated ditch location. |
| LL11sd-021-0001-SD | D | 11/16/00 | 0-1 | Phase I RI | | X | X | X | X | |
| LL11sd-024-0001-SD | D | 11/16/00 | 0–1 | Phase I RI | | X | X | X | X | |
| LL11sd-027-0001-SD | D | 11/15/00 | 0-0.5 | Phase I RI | | | | | | Excavated ditch location. |
| LL11sd-028-0001-SD | D | 11/15/00 | 0-0.5 | Phase I RI | | | | | | Excavated ditch location. |
| | | | | Phase I RI | | | | | | Sample not used for HHRA |
| | | | | | | | | | | and ERA because more |
| 1111 1020 0001 GD | _ | 11/16/00 | 0.1 | | | *** | | | | recent sample was taken at |
| LL11sd-030-0001-SD | D | 11/16/00 | 0–1 | | | X | | | | the same location. |
| 1111 1000 5500 GD | _ | 00/04/10 | 0 0 7 | DD 4 00 D4 | | *** | | | | Sample represents drainage |
| LL11sd-082-5593-SD | D | 02/24/10 | 0-0.5 | PBA08 RI | | X | | | | off AOC. |
| LL11sd-083-5594-SD | D | 02/25/10 | 0-0.5 | PBA08 RI | | X | X | X | X | |
| LL11sd-084-5595-SD | D | 02/25/10 | 0-0.5 | PBA08 RI | | X | X | X | X | |
| LL11sd-096-5874-SD | D | 08/09/12 | 0-0.5 | PBA08 RI | | X | X | X | X | |
| DV 447 | _ | 10/22/00 | 0.05 | 1000 DD GE | | | | | | Used for initial evaluation of |
| RV-447 | D | 10/23/98 | 0-0.5 | 1998 RRSE | 7 | | | | | site. |
| T T T T T T T T T T T T T T T T T T T | Б | 0.6/20/00 | 1111 | Surface W | | 1 | 1 | 1 | 1 | |
| LL11sw-001-0001-SW | D | 06/28/00 | 1.1–1.1 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-001-0002-SW | D | 07/31/00 | 3–3 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-002-0001-SW | D | 06/28/00 | 0.5-0.5 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-002-0002-SW | D | 07/31/00 | 3–3 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-003-0001-SW | D | 06/28/00 | 3–3 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-003-0002-SW | D | 07/31/00 | 3–3 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-004-0001-SW | D | 06/29/00 | 3.4–3.4 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-004-0002-SW | D | 07/31/00 | 3.4–3.4 | Phase I RI | | | | | | Excluded excavated sump. |
| LL11sw-005-0001-SW | D | 10/09/00 | 3–3 | Phase I RI | | | | | | Excluded excavated sump. |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|-----------------------|------|----------|------------|----------------|----|-----|-----|------|-----|------------------------------|
| | | | | Phase I RI | | | | | | Excluded excavated ditch |
| LL11sw-006-0001-SW | D | 06/28/00 | 1.2-1.2 | | | | | | | location. |
| | | | | Phase I RI | | | | | | Excluded excavated ditch |
| LL11sw-006-0002-SW | D | 07/31/00 | 1.2-1.2 | | | | | | | location. |
| LL11sw-007-0001-SW | D | 06/29/00 | 5.4-5.4 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-007-0002-SW | D | 07/31/00 | 8–8 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-008-0001-FD | D | 06/29/00 | 7.7–7.7 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-008-0001-SW | D | 06/29/00 | 7.7–7.7 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-008-0002-SW | D | 08/01/00 | 7–7 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-009-0001-SW | D | 06/29/00 | 10.5–10.5 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-009-0002-SW | D | 08/01/00 | 10–10 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-010-0001-SW | D | 06/29/00 | 5–5 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-010-0002-SW | D | 07/31/00 | 5–5 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-011-0001-FD | D | 06/28/00 | 7.8–7.8 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-011-0001-SW | D | 06/28/00 | 7.8–7.8 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-011-0002-FD | D | 07/12/00 | 7.8–7.8 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-011-0002-SW | D | 08/01/00 | 7.8–7.8 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-011-0002b-SW | D | 07/12/00 | 7.8–7.8 | Phase I RI | | | | | | Excluded sewer location. |
| LL11sw-012-0001-FD | D | 11/17/00 | 0.5-0.5 | Phase I RI | X | | | | | Field duplicate. |
| | | | | Phase I RI | | | | | | Sample used for nature and |
| | | | | | | | | | | extent evaluation only. More |
| | | | | | | | | | | recent sample at LL11sd-083 |
| LL11sw-012-0001-SW | D | 11/17/00 | 0.5-0.5 | | | X | | | | used for screening. |
| | | | | Phase I RI | | | | | | Excluded excavated ditch |
| LL11sw-013-0001-SW | D | 11/20/00 | 0-0.1 | | | | | | | location. |
| | | | | Phase I RI | | | | | | Excluded excavated ditch |
| LL11sw-013-0001b-SW | D | 02/07/01 | 0-0.1 | | | | | | | location. |
| | | | | Phase I RI | | | | | | Sample used for Nature and |
| | | | | | | | | | | Extent evaluation only. More |
| | _ | 11/20/00 | | | | | | | | recent sample at LL11sd-084 |
| LL11sw-014-0001-SW | D | 11/20/00 | 0-0.1 | | | X | | | | used for screening. |
| | | | | Phase I RI | | | | | | Sample used for nature and |
| | | | | | | | | | | extent evaluation only. More |
| 1111 014 00011 077 | D | 00/07/01 | 0.01 | | | 37 | | | | recent sample at LL11sd-084 |
| LL11sw-014-0001b-SW | D | 02/07/01 | 0-0.1 | DI I DI | | X | | | | used for screening. |
| I I 11 av 015 0001 SW | D | 11/20/00 | 0.01 | Phase I RI | | v | | | | Sample used for nature and |
| LL11sw-015-0001-SW | ע | 11/20/00 | 0-0.1 | | | X | | | | extent evaluation only. More |

Table 4–33. Data Summary and Designated Use for RI (continued)

| Sample ID | Type | Date | Depth (ft) | Sampling Event | QC | N&E | F&T | HHRA | ERA | Comments |
|---------------------|------|----------|------------|----------------|----|-----|-----|------|-----|------------------------------|
| | | | | | | | | | | recent sample at LL11sd-084 |
| | | | | | | | | | | used for screening. |
| | | | | Phase I RI | | | | | | Sample used for nature and |
| | | | | | | | | | | extent evaluation only. More |
| | | | | | | | | | | recent sample at LL11sd-084 |
| LL11sw-015-0001b-SW | D | 02/07/01 | 0-0.1 | | | X | | | | used for screening. |
| | | | | | | | | | | Sample represents drainage |
| | | | | | | | | | | off AOC used for nature and |
| LL11sw-082-5608-SW | D | 02/24/10 | | PBA08 RI | | X | | | | extent evaluation. |
| LL11sw-083-5609-SW | D | 02/25/10 | | PBA08 RI | | X | X | X | X | |
| LL11sw-084-5610-SW | D | 02/25/10 | | PBA08 RI | | X | X | X | X | |

AOC = Area of concern

D = Discrete.

ERA = Ecological risk assessment.

F&T = Fate and transport.

ft = Feet.

HHRA = Human health risk assessment.

ID = Identification.

IRA = Interim Removal Action.

N&E = Nature and extent.

PBA08 RI = Performance-based Acquisition 2008 Remedial Investigation.

QC = Quality control.

RI = Remedial investigation.

RRSE = Relative Risk Site Evaluation.

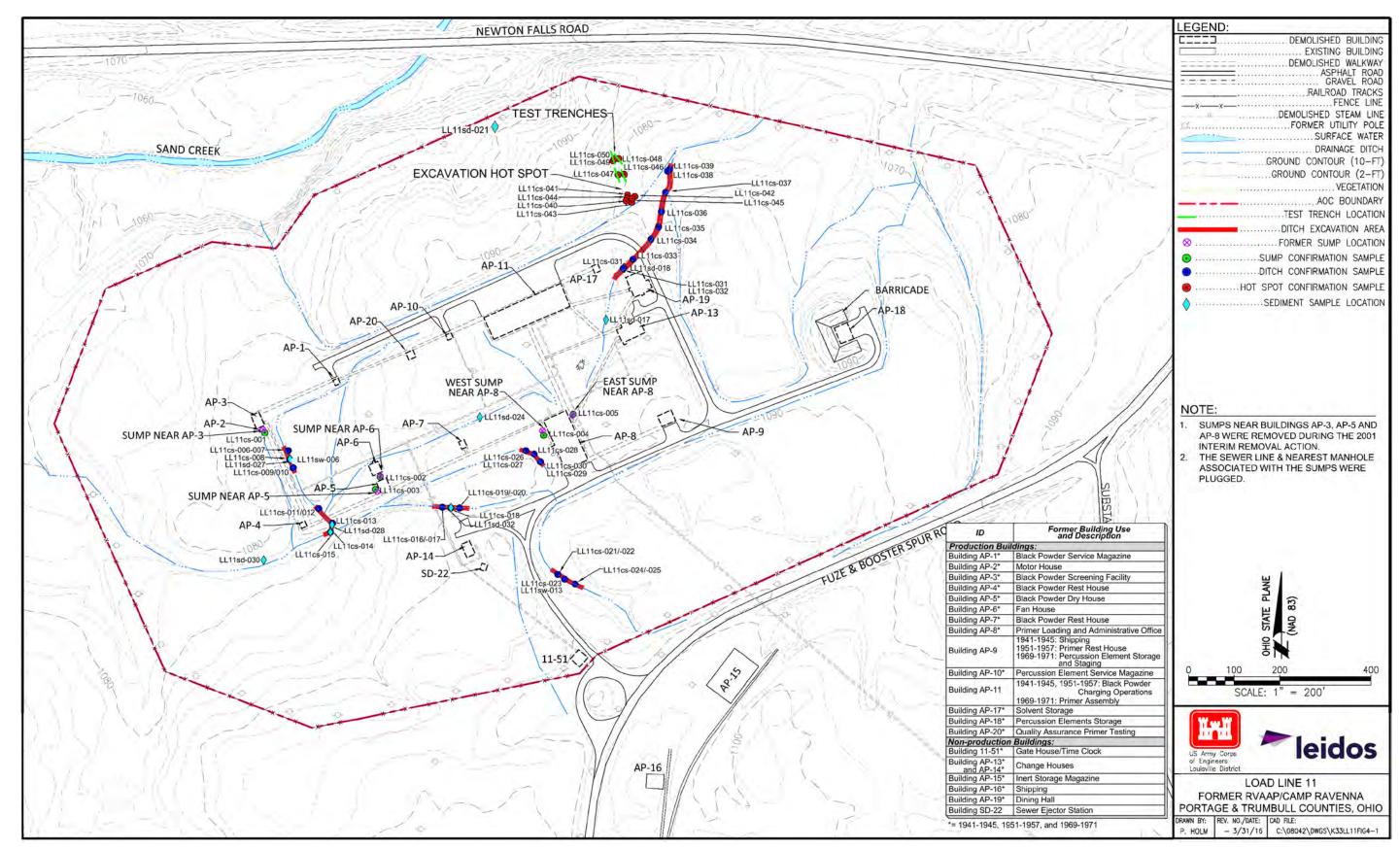


Figure 4-1. IRA Sample Locations at Load Line 11

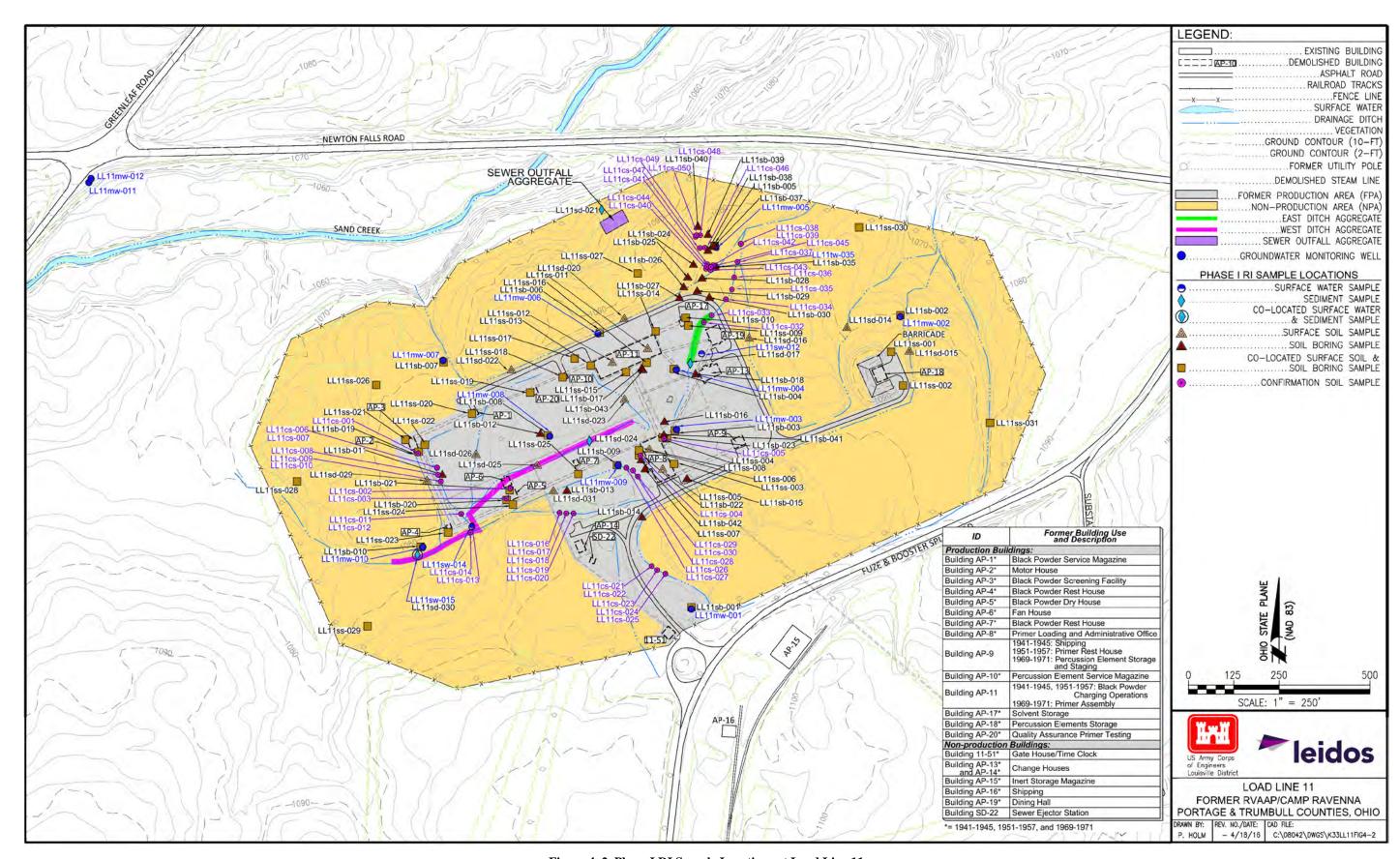


Figure 4–2. Phase I RI Sample Locations at Load Line 11

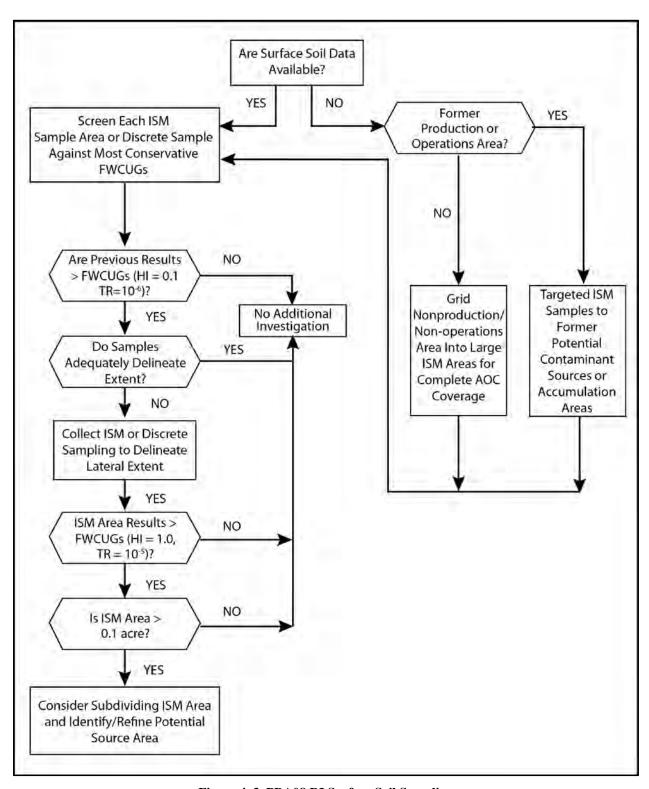


Figure 4-3. PBA08 RI Surface Soil Sampling

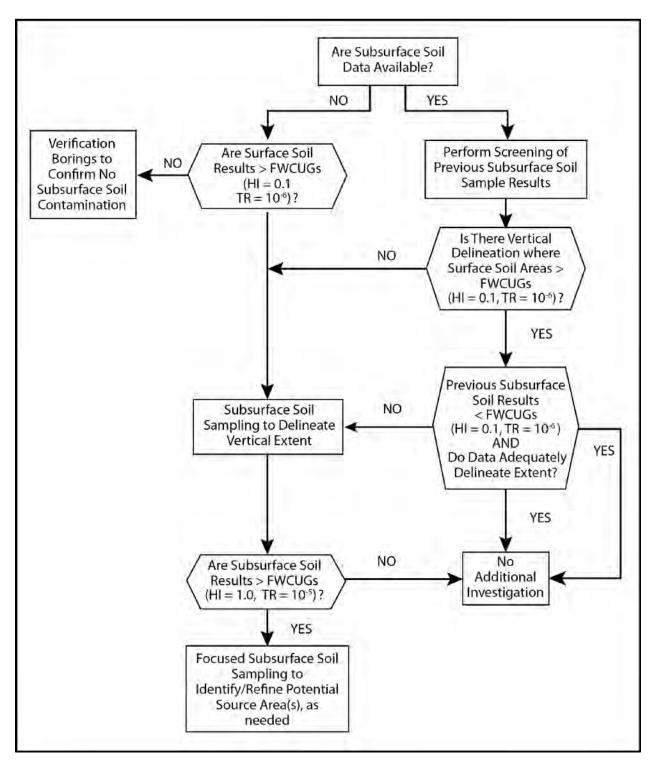


Figure 4-4. PBA08 RI Subsurface Soil Sampling

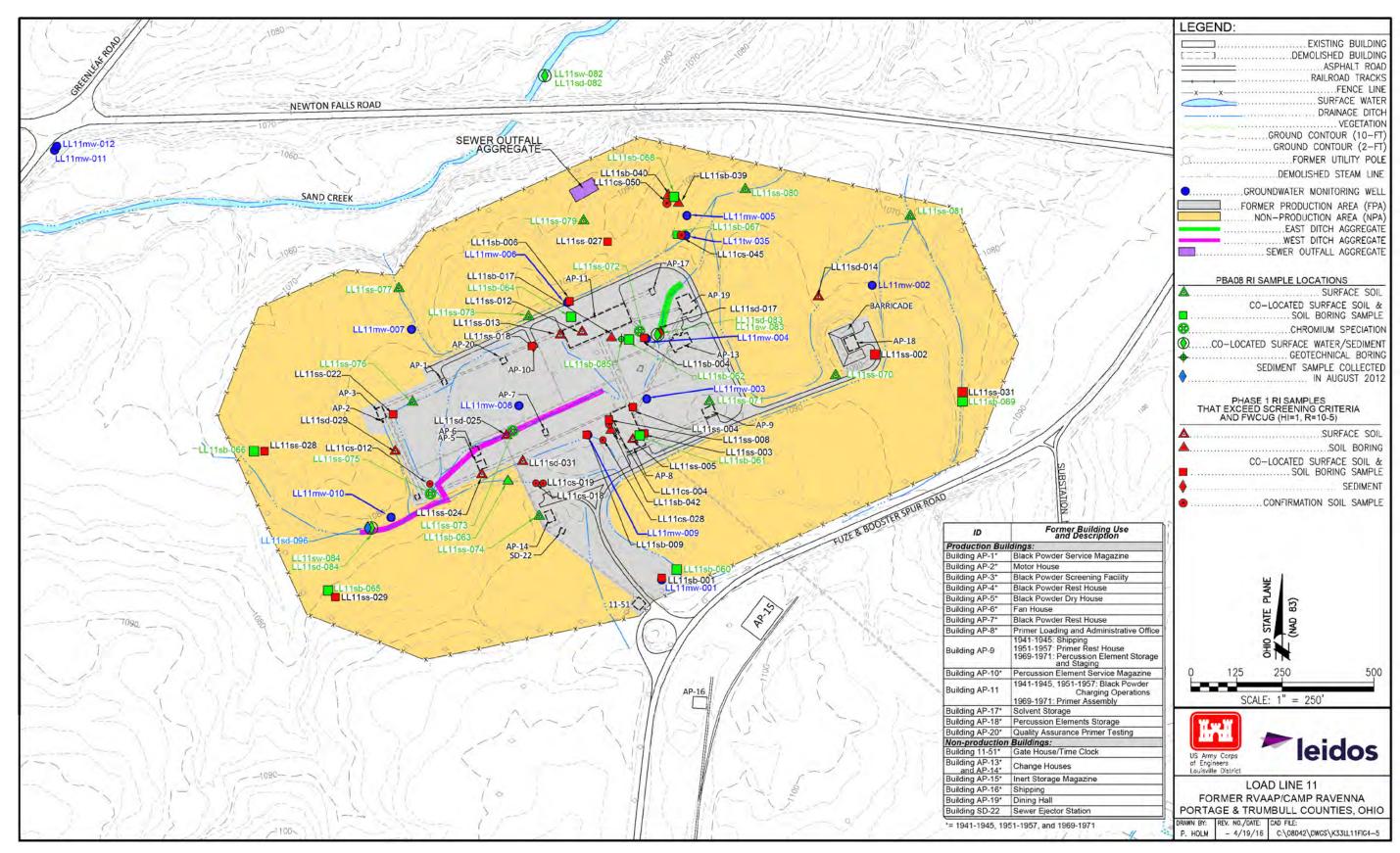


Figure 4–5. 2010 PBA08 RI Sample Locations

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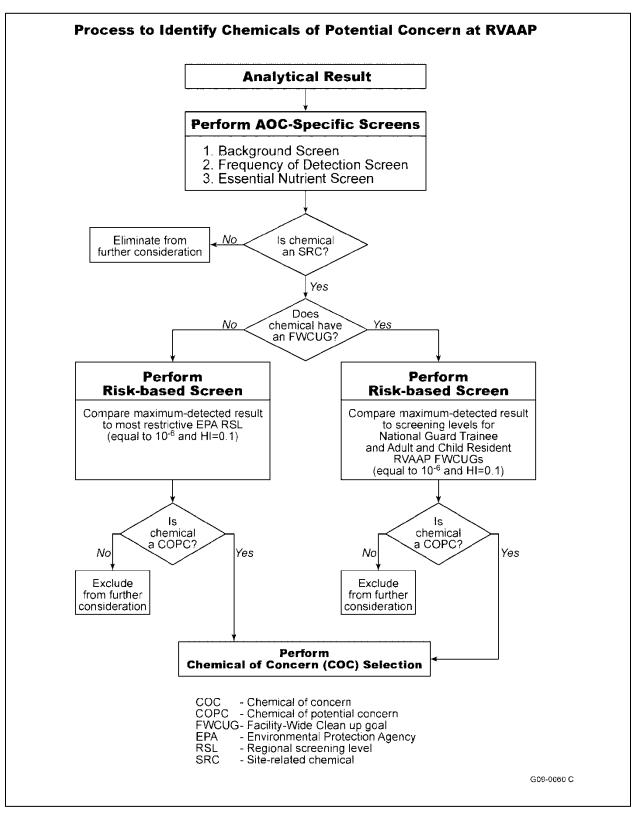


Figure 4-6. Process to Identify RVAAP COPCs in the HHRA (USACE 2010a)

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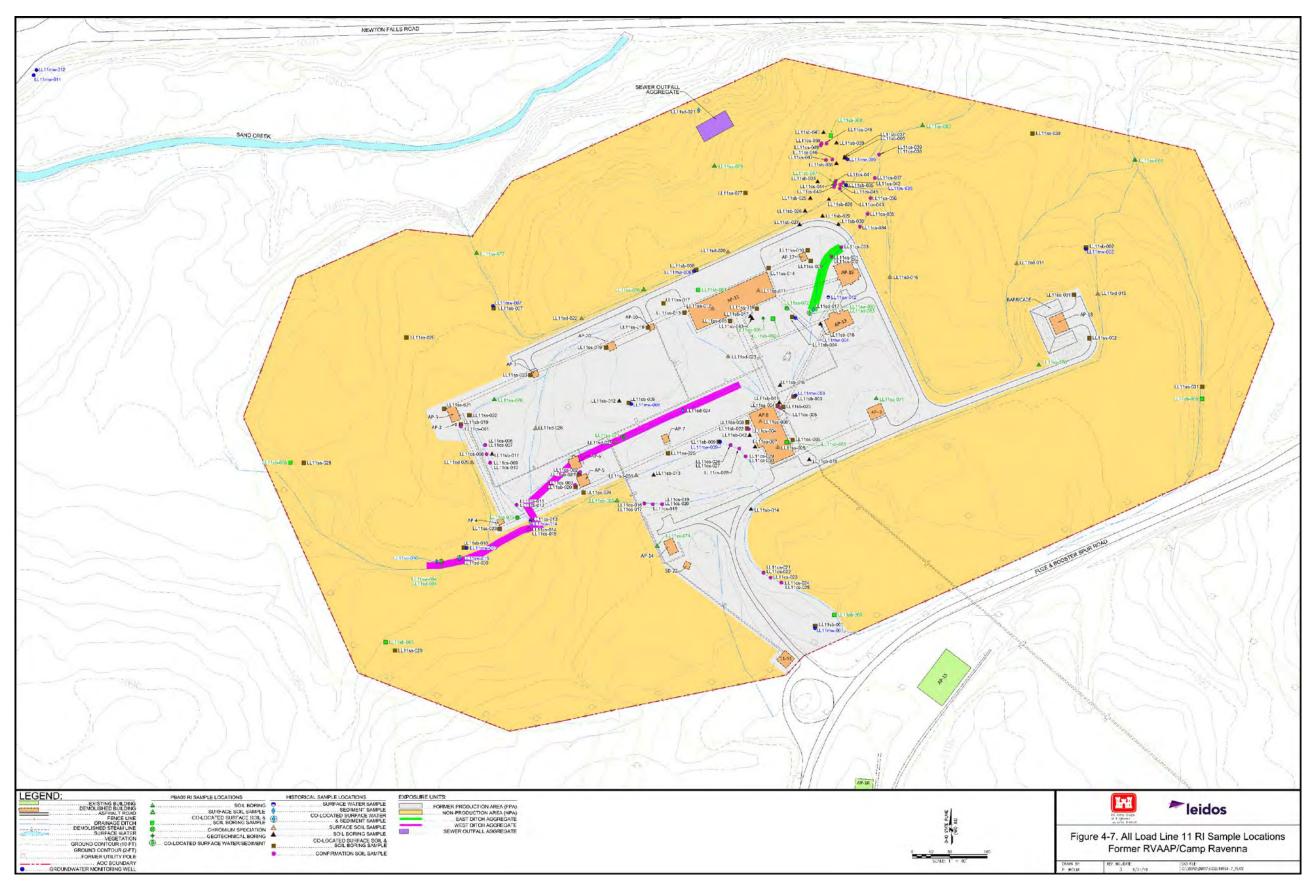


Figure 4–7. All Load Line 11 RI Sample Locations

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This section evaluates the nature and extent of contamination at Load Line 11. This evaluation includes two types of chemicals: 1) SRCs identified as being previously used during operational activities or that potentially were associated with operations, and 2) SRCs that do not appear to have been used during historical operations but were analyzed during investigations. The evaluation discusses the nature and extent of SRCs in environmental media at Load Line 11, focusing on chemicals previously used during operational activities and using analytical data results obtained during the 2000 Phase I RI, 2001 IRA, and 2010 and 2012 PBA08 RIs.

To support the evaluation of nature and extent of contamination, SRC concentrations were compared to SLs corresponding to the lowest FWCUG for the Resident Receptor (Adult and Child) and National Guard Trainee at a target HQ of 0.1 or TR of 1E-06, as presented in the FWCUG Report. The following figures illustrate the concentration and distribution of SRCs that exceed SLs.

- Figure 5-1 EUs and Sample Locations.
- Figure 5-2 Detected Concentrations of Explosives and Propellants in Soil at the FPA.
- Figure 5-3 Detected Concentrations of Explosives and Propellants in Soil at the NPA.
- Figure 5-4 Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) for Arsenic, Chromium, Lead, Mercury, Aluminum, Barium, Cobalt, Cyanide, Manganese, and Thallium in Soil at the FPA.
- Figure 5-5 Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) for Arsenic, Chromium, Lead, Mercury, Aluminum, Barium, Cobalt, Cyanide, Manganese, and Thallium in Soil at the NPA.
- Figure 5-6 PAH Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) in Soil at the FPA and NPA.
- Figure 5-7 Detected Concentrations of VOCs, Pesticides, and PCBs in Soil at the FPA and NPA.
- Figure 5-8 Detected Concentrations of Explosives and Propellants in Surface Water and Sediment.
- Figure 5-9 Exceedances of FWCUG (HQ of 0.1, TR 1E-06) for Arsenic, Chromium, Lead, Mercury, Aluminum, and Cobalt in Surface Water and Sediment.
- Figure 5-10 PAH Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) in Surface Water and Sediment.
- Figure 5-11 Detected Concentrations of VOCs, Pesticides, and PCBs in Surface Water and Sediment.

As discussed in Section 4.0, data from all eligible samples were combined and screened to identify SRCs representing current conditions at Load Line 11. All validated Load Line 11 data from the RIs (2000 Phase I RI and 2010 and 2012 PBA08 RIs) and 2001 IRA are included in Appendix D. Complete analytical data packages from the PBA08 RIs are also included in Appendix D.

The following sections discuss the source removal activities that took place under the IRA (MKM 2004a) as well as a current assessment of contaminant nature and extent for each medium and class of analyte.

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5.1 PREVIOUS CONTAMINANT SOURCE REMOVAL (INTERIM REMOVAL ACTION)

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As discussed in Section 4.2, contaminant source removal was conducted in 2001 under an IRA (MKM 2004a) as an early response action to mitigate migration of contaminants identified during the Phase I RI. This IRA is summarized in the IRA Report (MKM 2004a) and Phase I RI Report (MKM 2005a). The contaminant source removal consisted of the actions listed below.

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- Sump excavation and sewer line grouting Five sumps were excavated, removed, and disposed. One confirmation sample was collected from each sump location to confirm removal.
- Sewer line grouting Sewer manholes and sewer lines were filled with bentonite cement to
 prevent water from infiltrating back into the sumps during excavation and removal
 operations.
- Drainage ditch excavation A total of 230 yd³ of contaminated sediment was removed from six ditch lines within Load Line 11. These drainage ditches exhibited elevated concentrations of metals, VOCs, SVOCs, pesticides, and/or PCBs. Numerous samples were collected to confirm that the contaminant removal was complete.
- Hot spot excavation One location, a 30 by 30 by 8 ft hot spot area, located in an open field north of Building AP-17, was excavated due to petroleum contamination. Upon removing the contaminated soil, six samples were collected to confirm that the contaminated soil was removed.
- Test Trench Excavation Test trenches were excavated to assess high readings indicated on magnetometers used by UXO technicians immediately north of the hot spot excavation. Upon excavating test trenches, cable, scrap metal, bolts, and magnetite-containing rock were found. No UXO was identified. Five samples were collected to confirm there was no contamination present at this location.

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The information and results of these IRA activities were used in the current nature and extent evaluation of Load Line 11.

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5.2 SURFACE SOIL DISCRETE SAMPLE RESULTS FOR CHROMIUM SPECIATION

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During the 2010 PBA08 RI, surface soil samples were collected from three discrete sample locations and analyzed for hexavalent chromium and total chromium. One sample was collected from an area previously identified during the Phase I RI as having an elevated total chromium concentration (LL11ss-072), another sample was collected from an area previously identified during the IRA as having an elevated total chromium concentration (LL11ss-075), and an additional sample was collected from an area previously identified as having a total chromium concentration near background (LL11ss-073). This sampling attempted to determine the contribution of hexavalent

chromium to total chromium over a range of concentrations in soil at Load Line 11 for use in the HHRA (Section 7.2).

Chromium speciation results are shown in Table 5-1. All three samples were below the facility-wide background concentration of 17.4 mg/kg. The range of hexavalent chromium concentrations was 0.44J to less than 1 mg/kg and did not appear to be correlated to total chromium values. A detailed assessment of the speciation results respective to the HHRA is presented in Section 7.2.4.1.

5.3 CONTAMINANT NATURE AND EXTENT IN SURFACE SOIL

Tables 4-24 and 4-25 present the SRC screening results for surface soil at Load Line 11 that was still present after the source removals. The following subsections discuss the concentration and distribution of the discrete surface soil results.

5.3.1 Explosives and Propellants

The FPA and NPA were evaluated for explosives during both RIs and the IRA, comprising the Phase II RI data set. Specifically, the soil samples around the buildings were analyzed during the Phase I RI, the confirmation samples were evaluated following the IRA, and data gaps were assessed during the 2010 PBA08 RI. Figures 5-2 and 5-3 present locations that had detectable concentrations of explosives and propellants at the FPA and NPA.

Former Production Area

Two propellants (nitrocellulose and nitroguanidine) were identified as SRCs and as potentially related to previous AOC operations. Nitrocellulose was detected at concentrations ranging from 0.89 to 1.1 mg/kg in four of four samples with the maximum detection at location LL11ss-009. Nitroguanidine was detected only at LL11ss-001 at a concentration of 0.077J mg/kg. The maximum concentrations of nitrocellulose and nitroguanidine were below their respective SLs (19,000,000 and 630 mg/kg) and were not considered COPCs.

In addition to the explosive contaminants identified as being used historically, two explosives [pentaerythritol tetranitrate (PETN) and tetryl] were detected in the surface soil at the FPA. The explosives PETN and tetryl were detected only in PBA08 RI sample LL11sb-061 at concentrations of 0.049J and 0.02J mg/kg, respectively. LL11sb-061 is located immediately adjacent to former Building AP-8.

Non-Production Area

One explosive (HMX) and one propellant (nitrocellulose) were identified as SRCs and as potentially related to previous AOC operations. HMX was detected only in PBA08 RI sample LL11ss-081 (0.013J mg/kg), which is within a ditch line adjacent to the northeast fence line of Load Line 11. Nitrocellulose was detected in three of five samples analyzed, ranging from 0.89 (LL11sb-005) to 1.2 mg/kg (LL11ss-029). Nitrocellulose was not detected in either of the two PBA08 RI samples

analyzed. The maximum concentrations of HMX and nitrocellulose were below their respective SLs (359 and 19,000,000 mg/kg) and were not considered COPCs.

In addition to the explosive contaminants identified as being used historically, one explosive (PETN) was identified as a SRC in the surface soil at the NPA. PETN was detected only in PBA08 RI sample LL11sb-067 (0.036J mg/kg), located at the downstream portion of the northeast ditch line that was excavated during the IRA in 2001.

5.3.2 Inorganic Chemicals

Former Production Area

Arsenic, chromium, lead, and mercury were identified as potential SRCs and as potentially related to the previous site use. All four of these inorganic chemicals exceeded their respective background concentrations in the surface soil in the FPA, as summarized in the list below.

• Arsenic was detected above the background concentration of 15.4 mg/kg in 11 samples, with a maximum concentration of 30.2 mg/kg observed at the Phase I RI sample location LL11ss-022 that evaluated former Building AP-3.

Chromium was detected above the background concentration of 17.4 mg/kg in 10 samples, with a maximum concentration of 28.4 mg/kg observed at the IRA sample location LL11cs-011. Confirmation sample LL11cs-011 was collected to evaluate the effectiveness of the drainage ditch excavation near former Building AP-4.

• Lead was detected above the background concentration of 26.1 mg/kg in 21 samples, with a maximum concentration of 102 mg/kg observed at the IRA sample LL11cs-006. Confirmation sample LL11cs-006 was collected to evaluate the effectiveness of the drainage ditch excavation along the western side of the FPA.

 Mercury was detected above the background concentration of 0.036 mg/kg in 13 samples, with a maximum concentration of 0.34 mg/kg observed at the Phase I RI sample location LL11ss-013 that evaluated former Building AP-11.

Figure 5-4 presents the locations with concentrations that exceeded SLs and background for these four chemicals. Of the four chemicals, only arsenic was detected at concentrations above the SL and was considered a COPC.

Although not identified as previously used during historical operations, 12 other inorganic chemicals and 2 anions were identified as SRCs from the RVAAP screening process, as presented in Table 4-24. Of these constituents, four (aluminum, cobalt, cyanide, and manganese) exceeded their respective SLs and background concentrations. Figure 5-4 also presents exceedances of the SL and background for these four chemicals in the surface soil. Observations regarding other individual inorganic SRCs that exceeded their respective SLs in FPA surface soil are presented below:

• Aluminum was detected above the background concentration of 17,700 mg/kg in three samples, with a maximum concentration of 23,800 mg/kg observed at the IRA sample

- location LL11cs-025. Confirmation sample LL11cs-025 was collected to evaluate the effectiveness of the drainage ditch excavation on the south end of the AOC, just east of the entrance road.
- Cobalt was detected above the background concentration of 10.4 mg/kg in six samples, with a maximum concentration of 33.8 mg/kg observed at the Phase I RI sample location LL11sd-025 that evaluated the drainage ditch northwest of Building AP-7.
- Cyanide exceeded the SL of 0.27 mg/kg in three samples, with a maximum concentration of 1.6 mg/kg observed at the Phase I RI sample location LL11ss-014 that evaluated former Building AP-11.
- Manganese was detected above the background concentration of 1,450 mg/kg in two samples, with a maximum concentration of 1,540 mg/kg observed at the Phase I RI sample location LL11ss-018 that evaluated former Building AP-10.

Non-Production Area

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All four inorganic chemicals potentially related to previous site use exceeded their respective background concentrations in the surface soil in the NPA, as summarized in the list below.

- Arsenic was detected above the background concentration of 15.4 mg/kg in five samples, with a maximum concentration of 40.4 mg/kg observed at the 2010 PBA08 RI sample location LL11ss-070 that evaluated the earthen barricade associated with former Building AP-18.
- Chromium was detected above the background concentration of 17.4 mg/kg only in the Phase I RI sample location LL11ss-030, at a concentration of 19 mg/kg just inside the northern perimeter fence.
- Lead was detected above the background concentration of 26.1 mg/kg in four samples, with a
 maximum concentration of 33.9 mg/kg observed at the 2010 PBA08 RI sample location
 LL11ss-081 that evaluated the confluence of drainage ditches in the northeastern portion of
 the NPA.
- Mercury was detected above the background concentration of 0.036 mg/kg in 11 samples, with a maximum concentration of 0.08 mg/kg observed at the Phase I RI sample location LL11sd-020 that evaluated the drainage ditch along northern road, directly across from former Building AP-11.

Figure 5-5 presents the locations with concentrations that exceeded SLs and background for these four chemicals. Of the four chemicals, only arsenic was detected at concentrations above the SL and was considered a COPC.

Although not identified as previously used during historical operations, 10 other inorganic chemicals and 2 anions were identified as SRCs from the RVAAP screening process, as presented in Table 4-25. Of these constituents, three (cobalt, cyanide, and manganese) exceeded their respective SLs and background concentrations. Figure 5-5 also presents exceedances of the SL and background for these three chemicals in the surface soil.

Load Line 11

Observations regarding other individual inorganic SRCs that exceeded their respective SLs in NPA surface soil are presented below:

- Cobalt was detected above the background concentration of 10.4 mg/kg in seven samples, with a maximum concentration of 13.9 mg/kg observed at the Phase I RI sample location LL11sd-014 that evaluated the drainage ditch north of the former Building AP-18.
- Cyanide exceeded the SL of 0.27 mg/kg only in the Phase I RI sample location LLss-030, with a concentration of 0.54 mg/kg just inside the northern perimeter fence.
- Manganese was detected above the background concentration of 1,450 mg/kg in two samples, with a maximum concentration of 1,930 mg/kg observed at the Phase I RI sample location LL11sd-014 that evaluated the drainage ditch north of the former Building AP-18.

5.3.3 Semi-volatile Organic Compounds

Former Production Area

SVOCs do not have background concentrations for comparison to chemical results; consequently, a large number of SVOCs were identified as SRCs. A total of 16 SVOC SRCs were identified as a result of the data screening. Of these 16 SVOC SRCs, 5 were PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] that exceeded the SLs and were identified as COPCs. Figure 5-6 presents PAH exceedances of the SL in the FPA surface soil samples.

Although SVOCs were widely distributed in the FPA surface soil, the majority of detections and the highest concentrations were observed in the 2010 PBA08 RI sample locations LL11sb-060 and LL11sb-061 and IRA sample location LL11cs-020. With exception of benzo(a)pyrene at the these three sample locations, all PAH concentrations in samples collected in 2000, 2001, and 2010 were less than the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1.

Benzo(a)pyrene had a maximum concentration in surface soil of 0.45 mg/kg at sample location LL11sb-060. This location evaluated the ditch line immediately east of the entry road into the main gate of Load Line 11. The source of the PAHs observed could be runoff from the adjacent Fuze and Booster Spur Road. PBA08 RI sample location LL11sb-061 is located on the east side of former Building AP-8. IRA Confirmation sample LL11cs-020 evaluated the effectiveness of the excavation efforts of the drainage ditch removal on the south end of the AOC, just northwest of the entrance road. The surface soil samples that evaluated former Building AP-17 did not exceed the SLs.

Non-Production Area

Fifteen SVOCs were identified as SRCs in NPA surface soil, all of which were PAHs (Table 4-25). Of these 15 SVOC SRCs, 4 were PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and dibenz(a,h)anthracene] that exceeded the SLs and were identified as COPCs. Figure 5-6 presents PAH exceedances of the SL in the NPA surface soil samples.

 With exception of benzo(a)pyrene at the 2010 PBA08 RI sample location LL1sb-067, all PAH concentrations in samples collected in 2000, 2001, and 2010 were less than the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. Benzo(a)pyrene had a maximum concentration in NPA surface soil of 0.28 mg/kg at sample location LL11sb-067. This sample location is located at the east ditch line which was excavated in 2001 under the IRA. In PBA08 RI sample L11sb-068, located immediately downgradient of LL11sb-067, 10 PAHs were detected at concentrations approximately an order of magnitude less than those observed at LL11sb-067.

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5.3.4 Volatile Organic Compounds, Pesticides, and Polychlorinated Biphenyls

Former Production Area

VOCs and PCBs were identified as potential contaminants from previous site use. Figure 5-7 presents detected concentrations of VOCs, pesticides, and PCBs in FPA surface soil. One VOC (acetone) and one PCB (PCB-1254) were identified as SRCs in FPA surface soil. PCB-1254 was detected in 4 of 12 samples: 1 at the Phase I RI sample location LL11ss-020 (0.11 mg/kg) and 3 at the 2010 PBA08 RI sample locations. The maximum concentration of PCB-1254 detected was at 0.42 mg/kg at LL11sb-061, which exceeded the SL of 0.12 mg/kg. Sample location LL11sb-061 is located immediately adjacent to the east side of former Building AP-8. PCB-1254 also was detected at the 2010 PBA08 RI sample locations LL11sb-064 and LL11ss-074 at a concentration of 0.02J mg/kg. Acetone was detected in 1 of 10 samples (LL11cs-006 at 0.008 mg/kg) below the SL (6,100 mg/kg). No pesticides were detected in surface soil in the FPA.

Non-Production Area

One PCB (PCB-1254) was identified as an SRC in NPA surface soil. PCB-1254 was detected only in 2010 PBA08 RI sample location LL11ss-078 at a concentration of 0.035J mg/kg. LL11ss-078 is located in a ditch line alongside the access road north of former Building AP-10. No pesticides or VOCs were detected in NPA surface soil.

5.4 CONTAMINANT NATURE AND EXTENT IN SUBSURFACE SOIL

As discussed in Section 4.0, data from subsurface soil samples were screened to identify SRCs representing subsurface conditions at Load Line 11. The SRC screening data set included Phase I RI, IRA, and 2010 PBA08 RI samples. Samples excluded from the SRC screening data set are identified in Table 4-33 (e.g., samples impacted by excavations during the IRA). Additionally, Table 4-16 presents the classification of subsurface soil samples into the FPA and NPA. Subsurface soil samples collected during the PBA08 RI were analyzed for TAL metals, explosives, PAHs, and PCBs, with 10% submitted for RVAAP full-suite analysis (TAL metals, explosives, propellants, SVOCs, VOCs, PCBs, pesticides) to meet QA requirements. Table 4-26 and 4-27 present the results of the SRC screening for subsurface soil samples at the FPA and NPA, respectively.

5.4.1 Explosives and Propellants

Former Production Area

The propellant nitrocellulose was identified as an SRC in FPA subsurface soil. Nitrocellulose was detected in all eight samples analyzed, ranging in concentration from 0.87 (1-3 ft bgs at LL11ss-013) to 2 mg/kg (6-8 ft bgs at LL11sb-043). The location of the maximum detection is associated with former Building AP-11. The concentrations detected were all below the SL of 19,000,000 mg/kg.

No explosives were detected in the subsurface soil at the FPA.

Non-Production Area

Nitrocellulose was identified as an SRC in NPA subsurface soil. The propellant was detected in 15 of 17 samples analyzed, at concentrations comparable to those observed in FPA subsurface soil. A maximum nitrocellulose detection of 2 mg/kg was observed at the 2001 sample locations LL11sb-037 and LL11sb-039 within the 6-8 ft bgs interval, both of which were located within the ditch line excavated under an IRA.

No explosives were detected in the subsurface soil at the NPA.

5.4.2 Inorganic Chemicals

Former Production Area

Arsenic, chromium, lead, and mercury were identified as potential SRCs and as potentially related to the previous site use. Three of these four inorganic chemicals exceeded their respective background concentrations in the subsurface soil in the FPA, as summarized in the list below.

- Arsenic was detected above the background concentration of 19.8 mg/kg in six samples, with a maximum concentration of 25.8 mg/kg observed at the Phase I RI sample location LL11sb-022 from the 6-8 ft bgs interval that evaluated the sump near former Building AP-8.
- Chromium was not detected above its respective background concentration.

• Lead was detected above the background concentration of 19.1 mg/kg in 16 samples, with a maximum concentration of 65.1 mg/kg observed at the Phase I RI sample location LL11sb-023 from the 6-8 ft interval that evaluated the sump near former Building AP-8.

Mercury was detected above the background concentration of 0.044 mg/kg in one sample, with a maximum concentration of 0.05 mg/kg observed at 2000 Phase I RI sample location LL11ss-010 from the 1-3 ft bgs interval that evaluated former Building AP-17.

Figure 5-4 presents the locations with concentrations that exceeded SLs and background for these three chemicals. Of the three chemicals where concentrations were detected above background, only arsenic was detected at a concentration above the SL; therefore, arsenic is considered to be a COPC.

 Although not identified as previously used during historical operations, five other inorganic chemicals and three anions were identified as SRCs from the RVAAP screening process, as presented in Table 4-26. Of these constituents, one (cobalt) exceeded its respective SL and background concentration. Cobalt was detected above the background concentration of 23.2 mg/kg in one sample, with a maximum concentration of 27.1J mg/kg observed at the 2010 PBA08 RI sample location LL11sb-060 from the 1-4 ft bgs interval.

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5.4.3 Non-Production Area

Three of the four inorganic chemicals potentially related to previous site use exceeded their respective background concentrations in the subsurface soil in the NPA, as summarized in the list below.

- Arsenic was detected above the background concentration of 19.8 mg/kg in seven samples, with a maximum concentration of 44.1 mg/kg observed at sample LL11cs-045 from the 7-8 ft bgs interval that evaluated the effectiveness of the hotspot excavation efforts during the IRA.
- Chromium was not detected above its respective background concentration.
- Lead was detected above the background concentration of 19.1 mg/kg in three samples, with a maximum concentration of 160 mg/kg observed at sample LL11cs-050 from the 3-4 ft bgs interval that evaluated the effectiveness of the hotspot excavation efforts during the IRA.
- Mercury was detected above the background concentration of 0.044 mg/kg in four samples, with a maximum concentration of 0.068 mg/kg observed at 2010 PBA08 RI sample location LL11sb-069.

Figure 5-5 presents the locations with concentrations that exceeded SLs and background for these four chemicals. Of the three chemicals where concentrations were detected above background, only arsenic was detected at a concentration above the SL and was considered a COPC.

Ten other inorganic chemicals and three anions were identified as SRCs from the RVAAP screening process, as presented in Table 4-27. Of these constituents, five (aluminum, barium, cobalt, cyanide, and thallium) exceeded their respective SLs and background concentrations. Figure 5-5 also presents exceedances of the SL and background for these five chemicals in subsurface soil. Observations regarding other individual inorganic SRCs that exceeded their respective SLs in NPA subsurface soil are presented below:

Aluminum was detected above the background concentration of 19,500 mg/kg in one sample, with a concentration of 24,500 mg/kg observed at sample LL11cs-013 from the 1-2 ft bgs interval that evaluated the effectiveness of the excavation efforts in the drainage ditch east of former Building AP-4 during the IRA.

• Barium was detected above the background concentration of 124 mg/kg in two samples, with a maximum concentration of 4,190 mg/kg observed at sample LL11cs-050 from the 3-4 ft bgs interval that evaluated the effectiveness of the hotspot excavation efforts during the IRA.

• Cobalt was detected above the background concentration of 23.2 mg/kg in one sample, with a concentration of 24.7 mg/kg observed at sample LL11cs-036 that evaluated the effectiveness

- of the excavation efforts in the drainage ditch north of former Buildings AP-19 and AP-17 during the IRA.
 - Cyanide exceeded the SL of 0.27 mg/kg, with a maximum concentration of 1 mg/kg at the Phase I RI sample location LL11ss-030 from the 1-3 ft bgs interval. This is the same location as the MDC in NPA surface soil.
 - Thallium was detected above the background concentration of 0.91 mg/kg in one sample, with a maximum concentration of 2.5 mg/kg observed at sample LL11cs-050 from the 3-4 ft bgs interval that evaluated the effectiveness of the hotspot excavation efforts during the IRA.

5.4.4 Semi-volatile Organic Compounds

Former Production Area

No SVOCs were identified as SRCs in FPA subsurface soil. As shown in Table 4-26, SVOCs were detected in 1 of the 22 samples in the SRC screening data set. The seven PAHs shown in Table 4-26 were only identified at the IRA sample location LL11cs-019, from the 1-2 ft bgs interval; the concentrations of these PAHs were below laboratory detection limits. LL11cs-019 is located in a drainage ditch downstream of the former Building AP-8.

Non-Production Area

Six SVOCs, all of which were PAHs, were identified as SRCs in NPA subsurface soil. Of these six SVOC SRCs, one PAH [benzo(a)pyrene] exceeded the SL and was identified as a COPC. Benzo(a)pyrene was detected at a concentration of 0.12J mg/kg at the IRA sample location LL11cs-043 from the 3-7 ft bgs interval. The concentration detected was below the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1.

5.4.5 Volatile Organic Compounds, Pesticides, and Polychlorinated Biphenyls

Former Production Area

One PCB (PCB-1254) was identified as an SRC in the FPA subsurface soil and was detected in 4 of 19 samples in the SRC screening data set. The maximum concentration of PCB-1254 (0.79 mg/kg) was detected at the Phase I RI sample location LL11ss-016 from the 1-3 ft bgs interval, which exceeded the SL (0.12 mg/kg). LL11ss-016 is located in the interior portion of the FPA and immediately adjacent to the former Building AP-11.

No pesticides or VOCs were detected in the subsurface soil at the FPA.

Non-Production Area

No PCBs or pesticides were detected in the subsurface soil at the NPA. One VOC (carbon tetrachloride) was detected in 1 out of 26 samples. Based on its low frequency of detection and low concentration (0.001J mg/kg), carbon tetrachloride was not considered an SRC.

5.5 GEOTECHNICAL SUBSURFACE SOIL SAMPLES

 Two samples were collected from one soil boring completed at Load Line 11 for the purposes of obtaining geotechnical parameters to support vadose zone soil leaching and groundwater transport modeling. Samples were collected from the 2-3.9 and 4-5.5 ft bgs intervals at LL1sb-085. Groundwater was encountered at 6.5 ft bgs. Table 5-2 summarizes the results of the geotechnical characteristics of Load Line 11 soil. Laboratory analytical data package results are presented in Appendix D.

5.6 CONTAMINANT NATURE AND EXTENT IN SEDIMENT

Two Load Line 11 sediment samples were collected under the 2010 PBA08 RI, one representative sample each for the East Ditch and West Ditch aggregates. The West Ditch PBA08 RI sediment sample (LL11sd-084) was analyzed for TAL metals, explosives, SVOCs, and VOCs, and the East Ditch sample (LL11sd-083) was analyzed for RVAAP full-suite analytes (i.e., TAL metals, explosives, propellants, SVOCs, VOCs, PCBs, and pesticides). An additional sample (LL11sd-096) was taken under the 2012 PBA08 RI to provide full-suite analytical results in the West Ditch. Two additional samples were used to characterize the West Ditch: LL11sd-024 from the Phase I RI and LL11cs-015 from the IRA. One additional sample was used to characterize the East Ditch: LL11cs-031 from the IRA. Sediment sample LL11sd-017 from the East Ditch and LL11sd-030 from the West Ditch collected during the Phase I RI were not used for determining SRCs because samples from the more recent 2010 PBA08 RI were taken from the same locations.

The sample representative of the Sewer Outfall aggregate (LL11sd-021) was collected under the Phase I RI activities and was not resampled during the 2010 PBA08 RI. Conditions at this location were not considered to have changed, as the former sanitary sewer outfall is outside of Load Line 11 areas that would have been disturbed under IRA activities. Portions of the sanitary sewer system were plugged under the IRA activities, which would have prevented release of sediment to the overflow outfall.

Tables 4-28 through 4-30 present the results of the SRC screening for sediment for the East Ditch, West Ditch, and Sewer Outfall sediment, respectively. Complete copies of the laboratory analytical packages are presented in Appendix D. Figures 5-8 through 5-11 present the location and concentrations of selected SRCs identified in sediment at Load Line 11.

Additionally, under the 2010 PBA08 RI, one sediment sample was collected off-AOC at Sand Creek to assess conditions at the receiving water body north of Load Line 11. The data from this sample (LL11sd-082) were reviewed to assess off-AOC impacts from Load Line 11.

5.6.1 Explosives and Propellants

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East Ditch

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No explosives or propellants were identified as SRCs in sediment at the East Ditch. The propellant nitrocellulose was detected in the Phase I RI sample LL11sd-017 at a concentration of 1.8 mg/kg; however, nitrocellulose was not detected in the East Ditch 2010 PBA08 RI sample LL11sd-083. Sample LL11sd-017 was not utilized in the SRC screening as it was co-located with and superseded by 2010 PBA08 RI sample LL11sd-083.

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West Ditch

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The propellant nitrocellulose was detected at the 2012 PBA08 RI sample LL11sd-096 at a concentration of 1.4J mg/kg, which is below the SL of 19,000,000 mg/kg. No explosives were detected in the sediment at the West Ditch.

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Sewer Outfall

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No explosives or propellants were detected in the sediment at the Sewer Outfall.

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Off-AOC

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No explosives or propellants were detected in sediment at the off-AOC sample location LL11sd-082.

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5.6.2 Inorganic Chemicals

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East Ditch

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Arsenic, chromium, lead, and mercury were identified as potential inorganic contaminants from previous site use. Mercury was not detected in the sediment sample collected at the East Ditch; lead was detected at a concentration below its respective background concentration. Arsenic and chromium were detected at concentrations (19.7 and 18.6 mg/kg) that were slightly above their respective background concentrations (19.5 and 18.1 mg/kg). Arsenic exceeded the SL of 0.425 mg/kg. Chromium exceeded the SL for hexavalent chromium (1.64 mg/kg), but was detected at a concentration below the SL for trivalent chromium (8,147 mg/kg).

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- Eight other inorganic chemicals were identified as SRCs from the RVAAP screening process, as presented in Table 4-28. Of these constituents, aluminum and cobalt were detected at concentrations of 15,100 and 17.4 mg/kg, which were above their respective background concentrations (13,900 and
- 40 9.1 mg/kg) and exceeded their respective SLs (3,496 and 2.3 mg/kg).

West Ditch

Of the four inorganic contaminants (arsenic, chromium, lead, and mercury) potentially related to previous site use, arsenic and lead were not detected above their respective background concentrations. Mercury was detected above the background concentration (0.059 mg/kg) and below the SL (2.27 mg/kg), with a maximum concentration of 0.1 mg/kg observed at 2012 PBA08 sample location LL11sd-096. Chromium was detected above the background concentration (18.1 mg/kg) and exceeded the SL for hexavalent chromium (1.64 mg/kg) but not the SL for trivalent chromium (8,147 mg/kg), with a maximum concentration of 21 mg/kg observed at the Phase I RI sample location LL11sd-024.

Eight other inorganic chemicals were identified as SRCs from the RVAAP screening process, as presented in Table 4-29. Of these constituents, aluminum and cobalt were detected at concentrations that exceeded their respective SLs (3,496 and 2.3 mg/kg) and were above their background concentrations (13,900 and 9.1 mg/kg), with maximum concentrations of 17,700 mg/kg for aluminum at sample LL11cs-015 and 14.1 mg/kg for cobalt at sample LL11sd-024.

Sewer Outfall

No inorganic chemicals were identified as SRCs in Sewer Outfall sediment, as none were detected above their respective background concentrations.

Off-AOC

No inorganic chemicals were detected above background concentrations; therefore, no inorganic chemicals were identified as SRCs in off-AOC sample location LL11sd-082.

5.6.3 Semi-volatile Organic Compounds

East Ditch

Fifteen SVOCs were identified as SRCs in the East Ditch sediment. All of the SRCs in sediment were detected at the 2010 PBA08 RI sample location LL11sd-083 and had not previously been detected at this location in the Phase I RI sample LL11sd-017. One PAH [benzo(a)pyrene] exceeded the SL and was identified as a COPC. Benzo(a)pyrene was detected at a concentration of 0.098 mg/kg, which is below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1.

West Ditch

Fourteen SVOCs were identified as SRCs in the West Ditch sediment. One PAH [benzo(a)pyrene] exceeded the SL and was identified as a COPC. Benzo(a)pyrene was detected at a maximum concentration of 0.09 mg/kg at the 2012 PBA08 RI sample location LL11sd-096, which exceeded the SL of 0.022 mg/kg. The detected concentration was below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1.

| 1 | Sewer Outfall |
|----------|---|
| 2 3 | SVOC analysis was not performed on the former sanitary system overflow outfall sediment sample. |
| 4 | |
| 5 | Off-AOC |
| 6 | |
| 7 | Ten SVOCs were detected in sediment at the off-AOC sample location LL11sd-082. None of them |
| 8 | exceeded their SLs. |
| 9 | |
| 10 | 5.6.4 Volatile Organic Compounds, Pesticides, and Polychlorinated Biphenyls |
| 11 | |
| 12 | East Ditch |
| 13 | |
| 14 | No VOCs, pesticides, or PCBs were detected in the sediment at the East Ditch. |
| 15 | |
| 16 | West Ditch |
| 17 | No Words and the support of the section of the section of the West Disch |
| 18 19 | No VOCs, pesticides, or PCBs were detected in the sediment at the West Ditch. |
| 20 | Sewer Outfall |
| 21 | Sewer Outlan |
| 22 | Analysis for VOCs, pesticides, or PCBs was not performed on the former sanitary system overflow |
| 23 | outfall sediment sample. |
| 24 | |
| 25 | Off-AOC |
| 26 | |
| 27 | Two VOCs (carbon disulfide and toluene) were identified at concentrations below laboratory |
| 28 | reporting limits and below SLs at the off-AOC sediment sample location LL11sd-082. Pesticide and |
| 29 | PCB analysis was not performed on this sample. |
| 30 | |
| 31 | 5.7 CONTAMINANT NATURE AND EXTENT IN SURFACE WATER |
| 32 | |
| 33 | Two Load Line 11 surface water samples were collected under the 2010 PBA08 RI, one |
| 34 | representative sample each for the East Ditch and West Ditch aggregates. Both PBA08 RI surface |
| 35 | water samples were analyzed for RVAAP full-suite analytes. No surface water was collected at the |
| 36 | Sewer Outfall under the Phase I RI, as this location is predominantly dry. Tables 4-31 and 4-33 |
| 37 | present the results of the SRC screening for surface water at the East Ditch and West Ditch, |
| 38 | respectively. Only the PBA08 RI samples were utilized in the SRC screening, as the Phase I RI |
| 39 | surface water samples were not considered to represent current conditions. Complete copies of the |

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Additionally, under the 2010 PBA08 RI one surface water sample was collected off-AOC at Sand

laboratory analytical packages are presented in Appendix D. Figures 5-8 through 5-11 present the

Creek to assess conditions downstream of Load Line 11. The data from this sample (LL11sw-082)

location and concentrations of selected SRCs identified surface water at Load Line 11.

were reviewed to assess off-AOC and downstream impacts from Load Line 11. The data tables for 1 2 this sample are included in Appendix D. 3 4 **Explosives and Propellants** 5.7.1 5 6 No propellants or explosives were detected in the surface water at the East Ditch and West Ditch. In addition, no explosives or propellants were detected in the off-AOC surface water sample location 7 8 LL11sd-082. 9 10 5.7.2 **Inorganic Chemicals** 11 12 **East Ditch** 13 Arsenic, chromium, lead, and mercury were identified as potential inorganic contaminants from 14 15 previous site use. Mercury was not detected in the surface water sample collected at the East Ditch; 16 arsenic, chromium, and lead were detected at concentrations below their respective SLs. 17 18 The four other inorganic chemicals detected above background concentrations were not observed at 19 concentrations that exceeded their respective SLs. 20 21 **West Ditch** 22 23 Of the four inorganic contaminants (arsenic, chromium, lead, and mercury) potentially related to 24 previous site use, mercury was not detected in the surface water sample collected at the West Ditch 25 and arsenic was detected at a concentration (0.0043J mg/L) above the background concentration 26 (0.0032 mg/L) and exceeded the SL (0.0011 mg/L). In addition, chromium and lead were detected at 27 concentrations of 0.0033 and 0.0062 mg/L, which are above their respective background 28 concentrations and below their SLs (0.0245 and 0.015 mg/L, respectively). 29 30 Nine other inorganic chemicals were identified as SRCs from the RVAAP screening process, as 31 presented in Table 4-30. Of these constituents, cobalt was detected at a concentration of 0.0019J 32 mg/L, which exceeded its respective SL (0.0006 mg/L) and background concentration (0 mg/L). 33 34 Off-AOC 35 36 Six inorganic chemicals were detected above their respective background concentrations. All the 37 detections were at concentrations below their respective SLs. 38 39 5.7.3 **Semi-volatile Organic Compounds**

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East Ditch

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No SVOCs were detected in the East Ditch surface water sample.

| Wast | Ditch |
|------|-------|
| | |

Ten SVOCs were identified as SRCs in West Ditch surface water, all of which were PAHs. Of these 10 SVOCs, 4 [benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] exceeded the SLs and were identified as COPCs. The four PAH concentrations detected were all above the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1.

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Off-AOC

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No SVOCs were detected in surface water at the off-AOC sample location LL11sw-082.

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5.7.4 Volatile Organic Compounds, Pesticides, and Polychlorinated Biphenyls

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East Ditch

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Two pesticides were identified as SRCs in East Ditch surface water. The pesticides beta-BHC and gamma-chlordane were detected at PBA08 RI sample LL11sw-083 at low estimated concentrations of 0.000013J and 0.000015J mg/L, respectively. No pesticides were detected in the Phase I RI sample LL11sw-012 collected at this location in 2000.

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No VOCs or PCBs were detected in the East Ditch surface water sample.

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West Ditch

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One pesticide (beta-BHC) was identified as an SRC in West Ditch surface water. Beta-BHC was detected at LL11sw-084 at a low estimated concentration below laboratory detection limits (0.0000096J mg/L), but was not detected at the downstream, off-AOC location LL11sw-082.

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No VOCs or PCBs were detected in the West Ditch surface water sample.

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Off-AOC

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No VOCs, pesticides, or PCBs were detected in surface water at the off-AOC sample location LL11sw-082.

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5.8 SUMMARY OF CONTAMINANT NATURE AND EXTENT

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In 2001, a contaminant source removal was conducted under an IRA (MKM 2004a) as an early response action to mitigate migration of contaminants identified during the Phase I RI. The contaminant source removal consisted of removing sumps, contaminated media in drainage ditches, petroleum contamination in a hot spot area, and evaluated the potential for UXO in an area to the north of the hot spot area. Data from the Phase I RI, IRA, and 2010 and 2012 PBA08 RIs that represent the current status of Load Line 11 effectively characterized the nature and extent of the contamination.

To support the evaluation of nature and extent of contamination, SRC concentrations were compared to SLs corresponding to the lowest FWCUG for the Resident Receptor (Adult and Child) and National Guard Trainee at a target HQ of 0.1, TR of 1E-06, as presented in the FWCUG Report. Based on the information provided earlier in this section and the summary below, it can be concluded that:

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- The previous IRA removed contamination associated with Load Line 11,
- The vertical and horizontal extent of existing contamination is defined, and
- No further sampling is needed to evaluate Load Line 11.

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5.8.1 Surface Soil

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5.8.1.1 Former Production Area

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Sites where explosives were identified as potential contaminants from previous use were thoroughly evaluated, including around former process buildings and across the AOC as a whole. Two propellants (nitrocellulose and nitroguanidine) were identified as SRCs and as potentially related to previous AOC operations. The maximum concentrations of the propellants were below their respective SLs and were not considered COPCs in the FPA surface soil. Arsenic, chromium, lead, and mercury were identified as potential SRCs and as potentially related to the previous site use. All four of these inorganic chemicals exceeded their respective background concentrations in FPA surface soil; however, only arsenic was detected at a concentration above the SL and was considered to COPC. The maximum concentration of arsenic was observed at the Phase I RI sample location LL11ss-022 that evaluated former Building AP-3. PAH SRC detections were widespread and observed in 11 of 14 surface soil locations in the SRC screening data set. With the exception of benzo(a)pyrene at three sample locations, all PAH concentrations in samples collected in 2000, 2001, and 2010 were less than the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. One VOC (acetone) and PCB-1254 were identified as SRCs in FPA surface soil. PCB-1254 was detected in 4 of 12 SRC screening samples in Phase I RI sample LL11ss-020 (0.11 mg/kg) and in three of the PBA08 RI samples, with the maximum detection (0.42 mg/kg) observed at LL11sb-061 located immediately adjacent to the east side of former Building AP-8. The VOC acetone was detected in only one Phase I RI sample. No pesticides were detected in FPA surface soil.

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5.8.1.2 Non-Production Area

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One explosive (HMX) and one propellant (nitrocellulose) were identified as SRCs and as potentially related to previous AOC operations. HMX was detected only in 2010 PBA08 RI sample LL11ss-081, which is within a ditch line adjacent to the northeast fence line of Load Line 11. Nitrocellulose was detected in three of five samples analyzed with the maximum detection observed in Phase I RI sample location LL11ss-029. The maximum concentrations of HMX and nitrocellulose were below their respective SLs and were not considered COPCs. All four inorganic chemicals (arsenic, chromium, lead, and mercury) potentially related to previous site use exceeded their respective background concentrations in the surface soil in the NPA. Of the four chemicals, only arsenic was detected at a concentration above the SL and was considered a COPC. The maximum concentration of arsenic was

observed at the 2010 PBA08 RI sample location LL11ss-070 that evaluated the earthen barricade associated with former Building AP-18. PAH SRC detections were widespread throughout the AOC and were observed in 11 of 14 surface soil locations in the screening data set. Of the 15 SVOC SRCs, 4 were PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and dibenz(a,h)anthracene] that exceeded the SLs and were identified as COPCs. With the exception of benzo(a)pyrene at the 2010 PBA08 RI sample location LL1sb-067, all PAH concentrations in samples collected in 2000, 2001, and 2010 were less than the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. PBA08 RI sample location LL11sb-067 had the highest frequency and magnitude of PAH detections and is located at the east ditch line which was excavated in 2001 under the IRA. PCB-1254 was identified as an SRC in NPA surface soil and was detected only in PBA08 RI sample LL11ss-078, located in a ditch line alongside the access road north of former Building AP-10. No pesticides or VOCs were detected in NPA surface soil.

5.8.2 Subsurface Soil

5.8.2.1 Former Production Area

The propellant nitrocellulose was identified as an SRC and was detected in all eight samples analyzed, with the maximum detection associated with former Building AP-11. The highest nitrocellulose detections occurred from the 5-6 and 6-8 ft bgs intervals. The concentrations detected were all below the SL. No explosives were detected in the FPA subsurface soil. Arsenic was the only inorganic chemical of the four potentially related to previous site use that exceeded the SLs and the FWCUG at a TR of 1E-05, HQ of 1. PAHs were detected in 1 out of the 22 samples collected. Based on the low frequency of detection, PAHs were not considered SRCs for the FPA subsurface soil. PCB-1254 was identified as an SRC for the FPA and was detected in 4 of 19 samples in the SRC screening data set. The maximum PCB-1254 detection occurred in the 1-3 ft bgs interval in the Phase I RI sample LL11ss-016, located in the interior portion of the FPA and immediately adjacent to the former Building AP-11. No explosives, pesticides, or VOCs were detected in the FPA subsurface soil.

5.8.2.2 Non-Production Area

The propellant nitrocellulose was identified as an SRC in NPA subsurface soil, with the highest detections noted at the downstream portion of the northeast ditch line excavated under the IRA. Three of the four inorganic chemicals potentially related to previous site use exceeded their respective background concentrations in the NPA subsurface soil; however, only arsenic was detected at a concentration above the SL and was considered a COPC. Six SVOCs, all PAHs, were identified as SRCs in subsurface soil at the NPA. Of these six SVOC SRCs, one PAH [benzo(a)pyrene] exceeded the SL and was identified as a COPC. The concentration detected was below the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. A single VOC was detected in the 26 samples collected. Based on the low frequency of detection, PAHs were not considered SRCs for the NPA subsurface soil. TPH-DRO and TPH-GRO were identified as SRCs in subsurface soil at the NPA and were detected in confirmatory samples collected following the excavations at the northeast ditch line. No explosives, PCBs, or pesticides were detected in the NPA subsurface soil.

5.8.3 Sediment

Arsenic was the only inorganic chemical from previous site use that exceeded the SL and is considered a COPC. Chromium exceeded the SL for hexavalent chromium, but was detected at a concentration below the SL for trivalent chromium. Fifteen SVOCs were identified as SRCs in sediment at the East Ditch; all were PAHs. All of the SRCs in sediment were detected in the 2010 PBA08 RI sample LL11sd-083, but had not been previously detected at this location in Phase I RI sample LL11sd-017. One PAH [benzo(a)pyrene] exceeded the SL and was identified as a COPC. The benzo(a)pyrene concentration was detected at a concentration below the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. No explosives, propellants, VOCs, pesticides, or PCBs were detected in sediment at the East Ditch.

The propellant nitrocellulose was detected at the West Ditch in the 2012 PBA08 RI sample LL11sd-096 at a concentration below the SL. Of the four inorganic contaminants potentially related to previous site use, arsenic and lead were not detected above their respective background concentrations. Mercury was detected above the background concentration and below the SL. Chromium was detected above the background concentration and exceeded the SL for hexavalent chromium but not the SL for trivalent chromium. Fourteen SVOCs were identified as SRCs in sediment at the West Ditch, all of which were PAHs. All maximum PAH SRC detections occurred at 2012 PBA08 RI sample LL11sd-096. One PAH [benzo(a)pyrene] exceeded the SL and was identified as a COPC. The detected concentration was below the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. No explosives, PCBs, pesticides, or VOCs were detected in sediment at the West Ditch.

The only SRCs identified at the Sewer Outfall aggregate at Load Line 11 were the inorganic chemicals sulfate and sulfide, neither of which has an established background concentration for comparison. No explosives were detected at this overflow outfall from the former sanitary sewer system. Inorganic chemicals were not detected above background concentrations at the off-AOC location LL11sd-082 located on Sand Creek. No explosives, propellants, or SVOCs were detected. Two VOCs (carbon disulfide and toluene) were present at low, estimated concentrations below laboratory detection limits.

5.8.4 Surface Water

Of the four inorganic chemicals identified as potential inorganic contaminants from previous site use, arsenic, chromium, and lead were detected at concentrations below their respective SLs. Two pesticides (beta-BHC and gamma-chlordane) were identified as SRCs at low, estimated concentrations. No explosives, propellants, SVOCs, VOCs, or PCBs were detected in the East Ditch surface water.

Three of the four inorganic chemicals potentially related to previous site use were detected in the West Ditch surface water. Arsenic was detected at a concentration above the background concentration and exceeded the SL. Chromium and lead were detected at concentrations above their respective background concentrations and below the SLs. Ten SVOCs (all PAHs) were identified as

- 1 SRCs for the West Ditch aggregate. Four PAHs [benzo(a)anthracene, benzo(a)pyrene,
- benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] exceeded the SLs and were identified as COPCs.
- 3 The concentrations of the four PAHs detected were all above the Resident Receptor (Adult and Child)
- 4 FWCUGs at a TR of 1E-05, HQ of 1. One pesticide (beta-BHC) was identified as an SRC and was
- 5 detected at a low, estimated concentration below laboratory detection limits. No explosives,
 - propellants, VOCs, or PCBs were detected in the West Ditch surface water.

Six inorganic chemicals were the only analytes detected above their background screening criteria at the off-AOC sample location LL11sw-082. All the detections were at concentrations below their respective SLs. No explosives, propellants, SVOCs, VOCS, pesticides, or PCBs were observed at this Sand Creek location.

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Table 5–1. Chromium Speciation Results

| Sample Location | Hexavalent Chromium Concentration (mg/kg) | Total Chromium Concentration ^a (mg/kg) | Percent Hexavalent Chromium (%) |
|--------------------|---|---|---------------------------------------|
| LL11ss-072 | 0.44J | 10.1 | 4.4 |
| LL11ss-073 | <1U | 17.1 | NA |
| LL11ss-075 | 0.71J | 15.4 | 4.6 |

^aBackground screening value for total chromium = 17.4 mg/kg. No background concentration is available for hexavalent chromium.

Table 5-2. Summary of Geotechnical Parameters

| Sample ID: | | |
|------------------------|--------------------------------------|--|
| Parameters | L11sb-085-5567 | L11sb-085-5568 |
| Depth | 2-3.9 ft bgs | 4-5.5 ft bgs |
| Porosity | 33.4 % | 38.2 % |
| Density | 1.83 g/cm^3 | 1.67 g/cm^3 |
| Moisture content | 15.4 % | 23.9 % |
| Total organic carbon | 720J mg/kg | 1,500 mg/kg |
| | 6.1% gravel, 27.9% sand, 49.0% silt, | 0.0% gravel, 2.4% sand, 85.4% silt, 12.2 |
| Size fraction analysis | 16.9% clay | % clay |
| Permeability (K) | 1.7E-07 cm/sec | 1.7E-06 cm/sec |

bgs = Below ground surface.

cm/sec = Centimeters per second.

g/cm³ = Grams per cubic centimeter.

ID = Identification.

mg/kg = milligrams per kilogram.

J = Estimated value is less than reporting limits.

NA = Not applicable; hexavalent chromium not detected in sample

U= Non-detectable concentration.

ft = Feet.

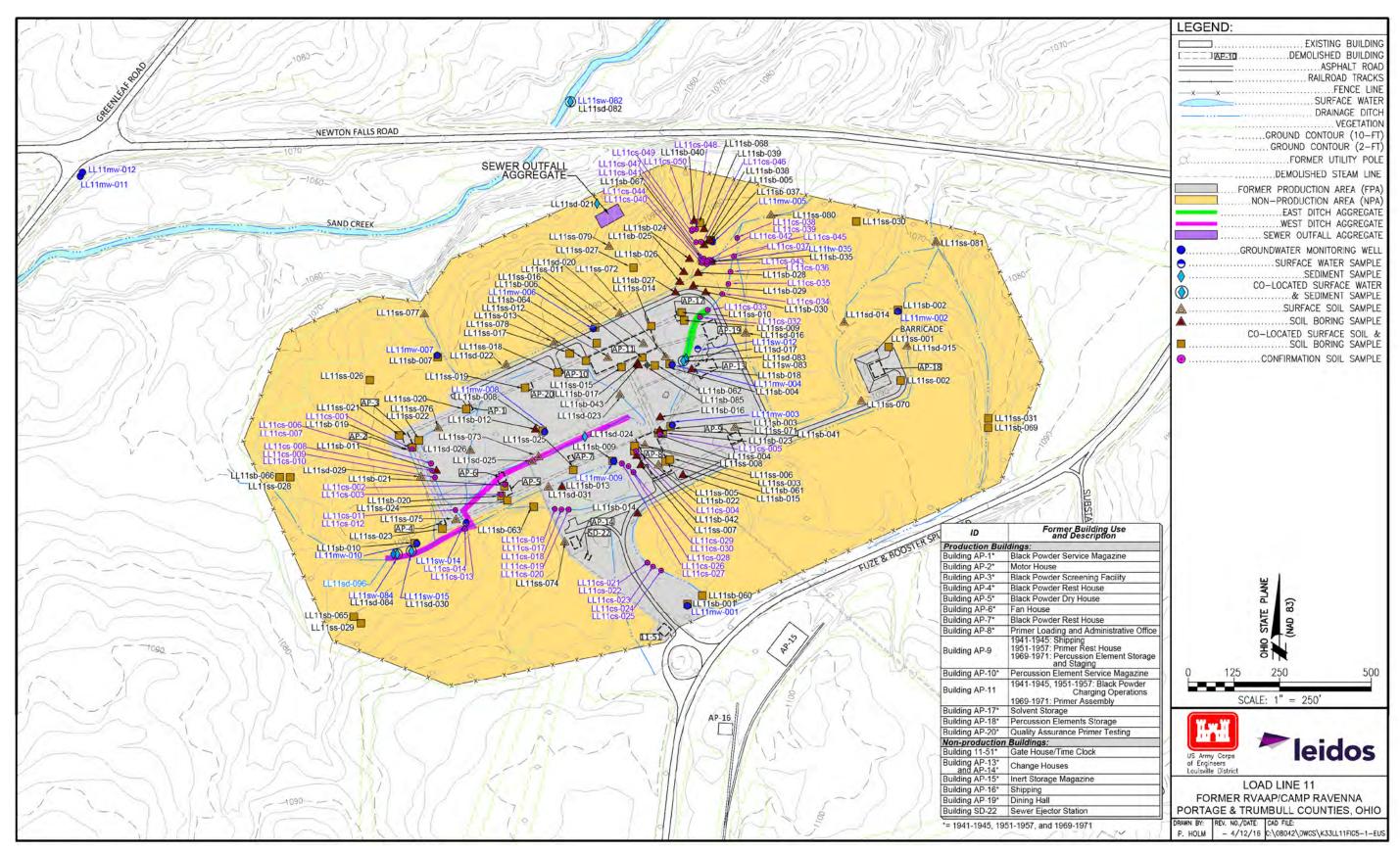


Figure 5-1. EUs and Sample Locations

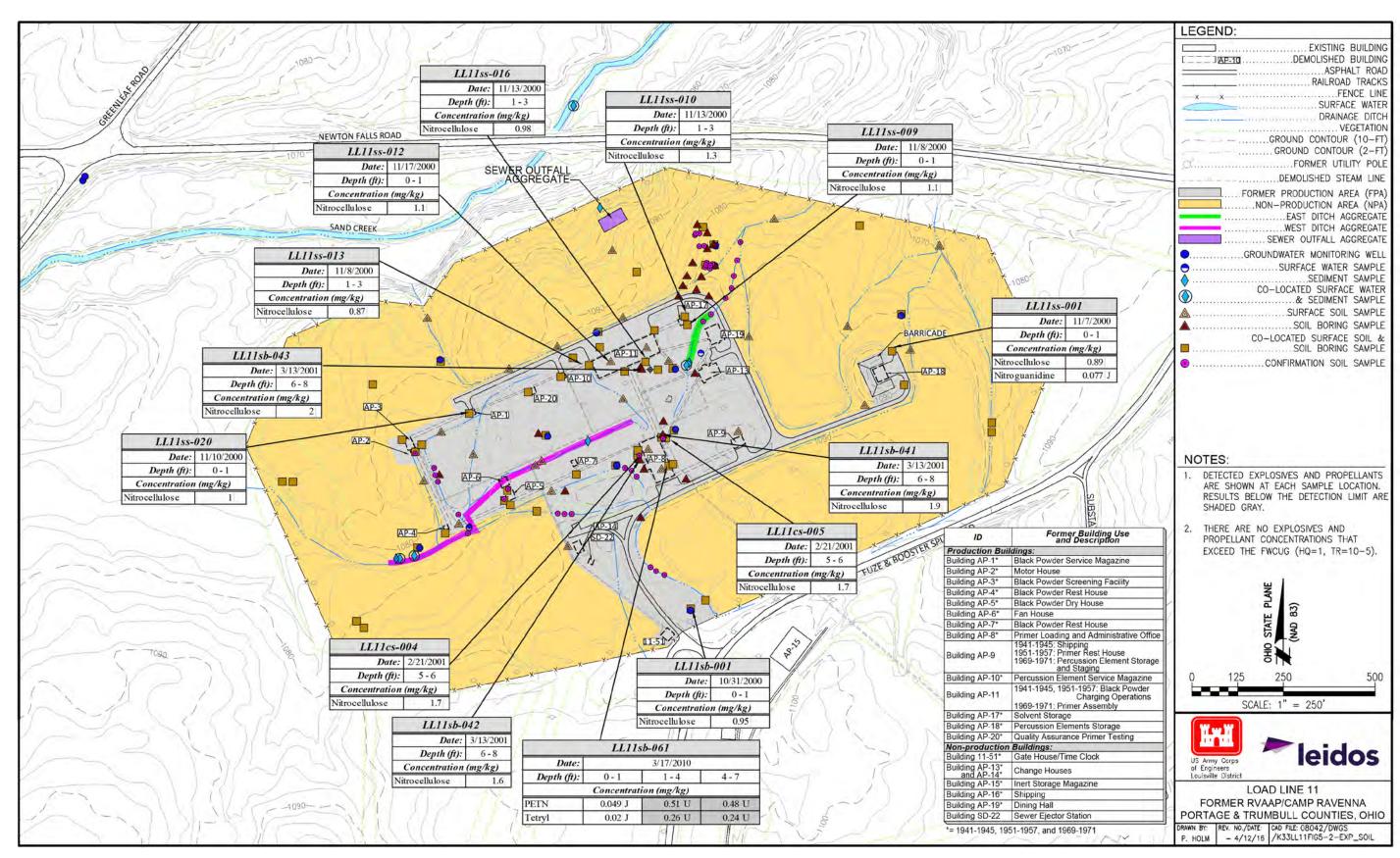


Figure 5–2. Detected Concentrations of Explosives and Propellants in Soil at the FPA

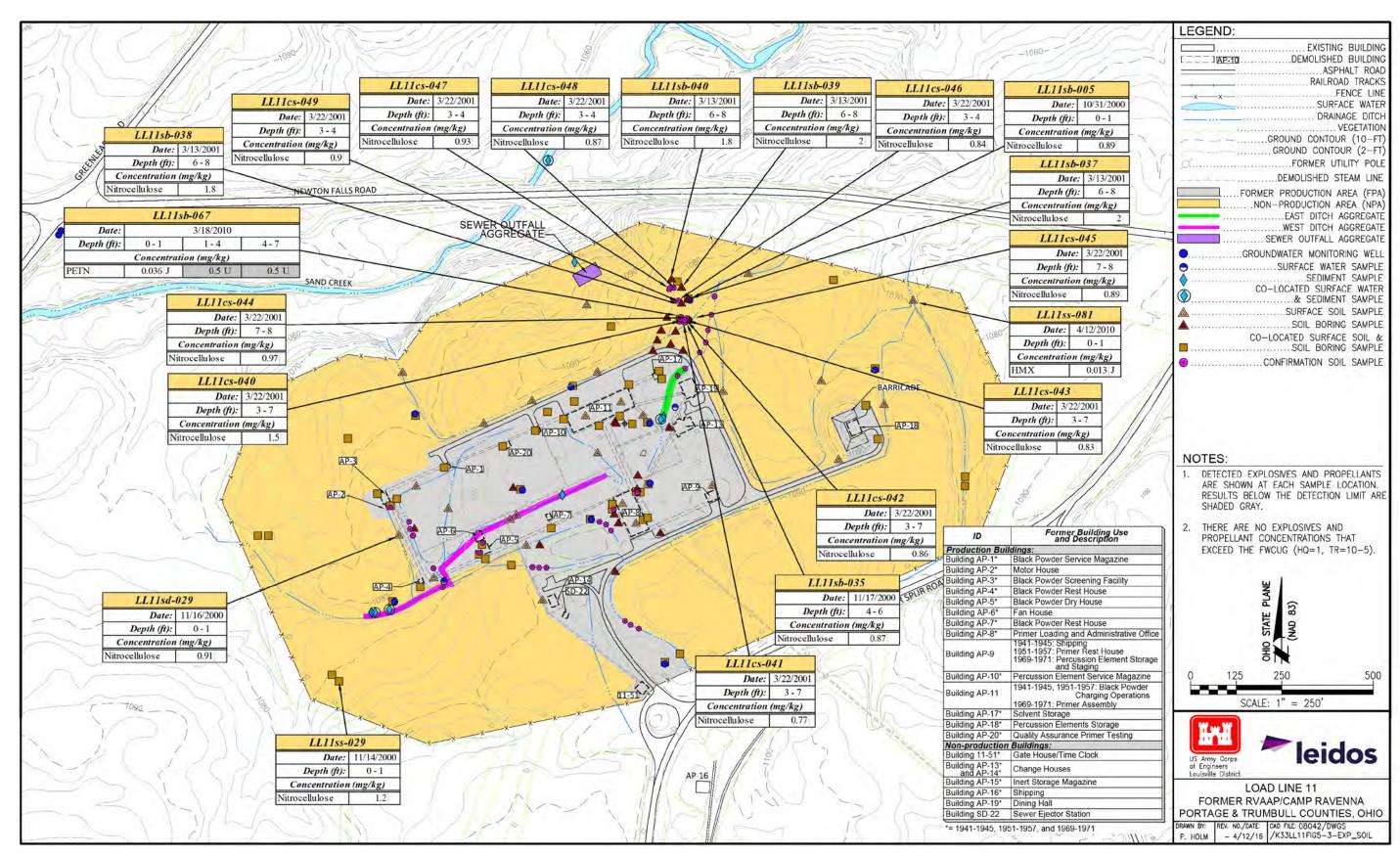


Figure 5-3. Detected Concentrations of Explosives and Propellants in Soil at the NPA

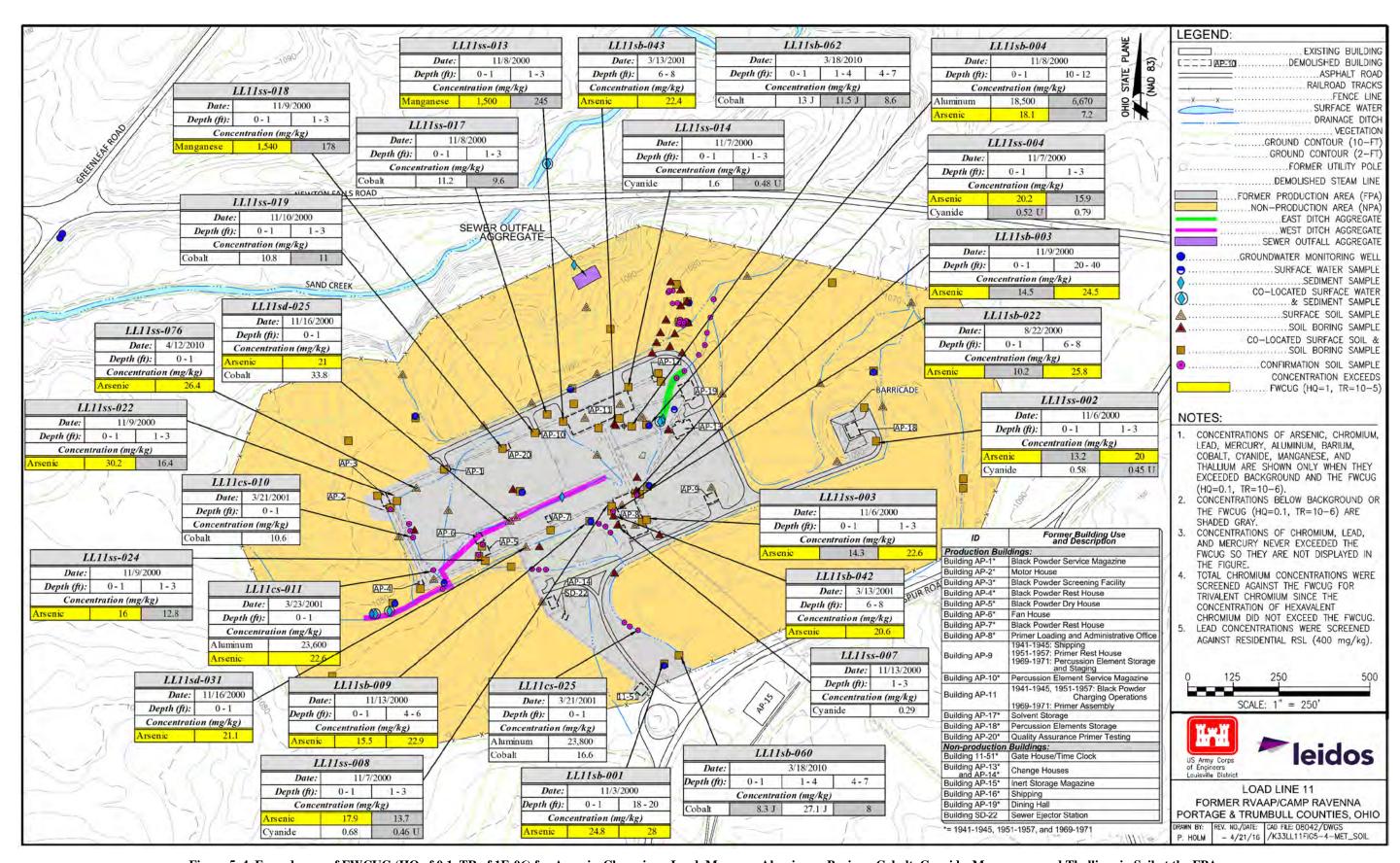


Figure 5-4. Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) for Arsenic, Chromium, Lead, Mercury, Aluminum, Barium, Cobalt, Cyanide, Manganese, and Thallium in Soil at the FPA

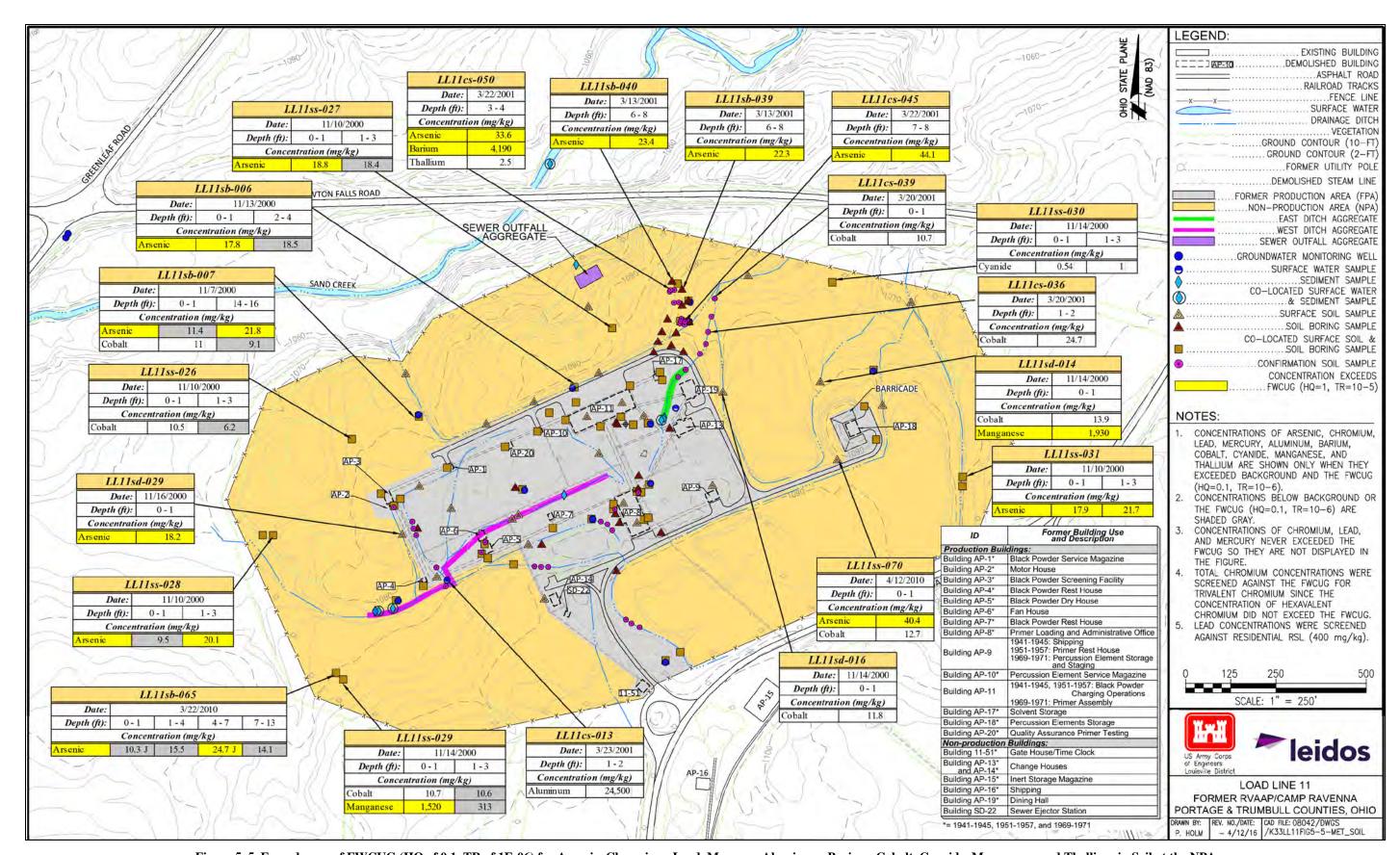


Figure 5–5. Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) for Arsenic, Chromium, Lead, Mercury, Aluminum, Barium, Cobalt, Cyanide, Manganese, and Thallium in Soil at the NPA

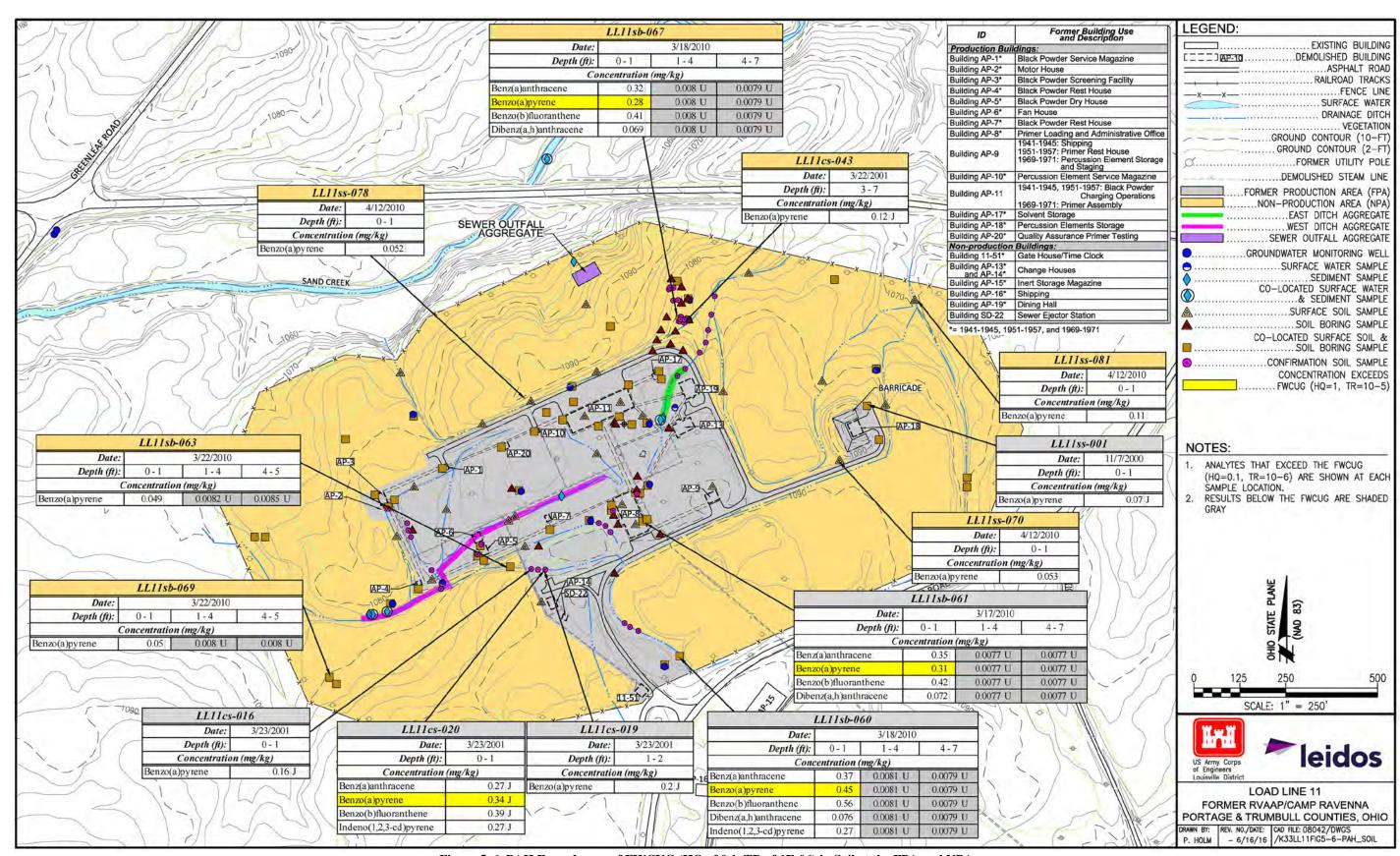


Figure 5-6. PAH Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) in Soil at the FPA and NPA

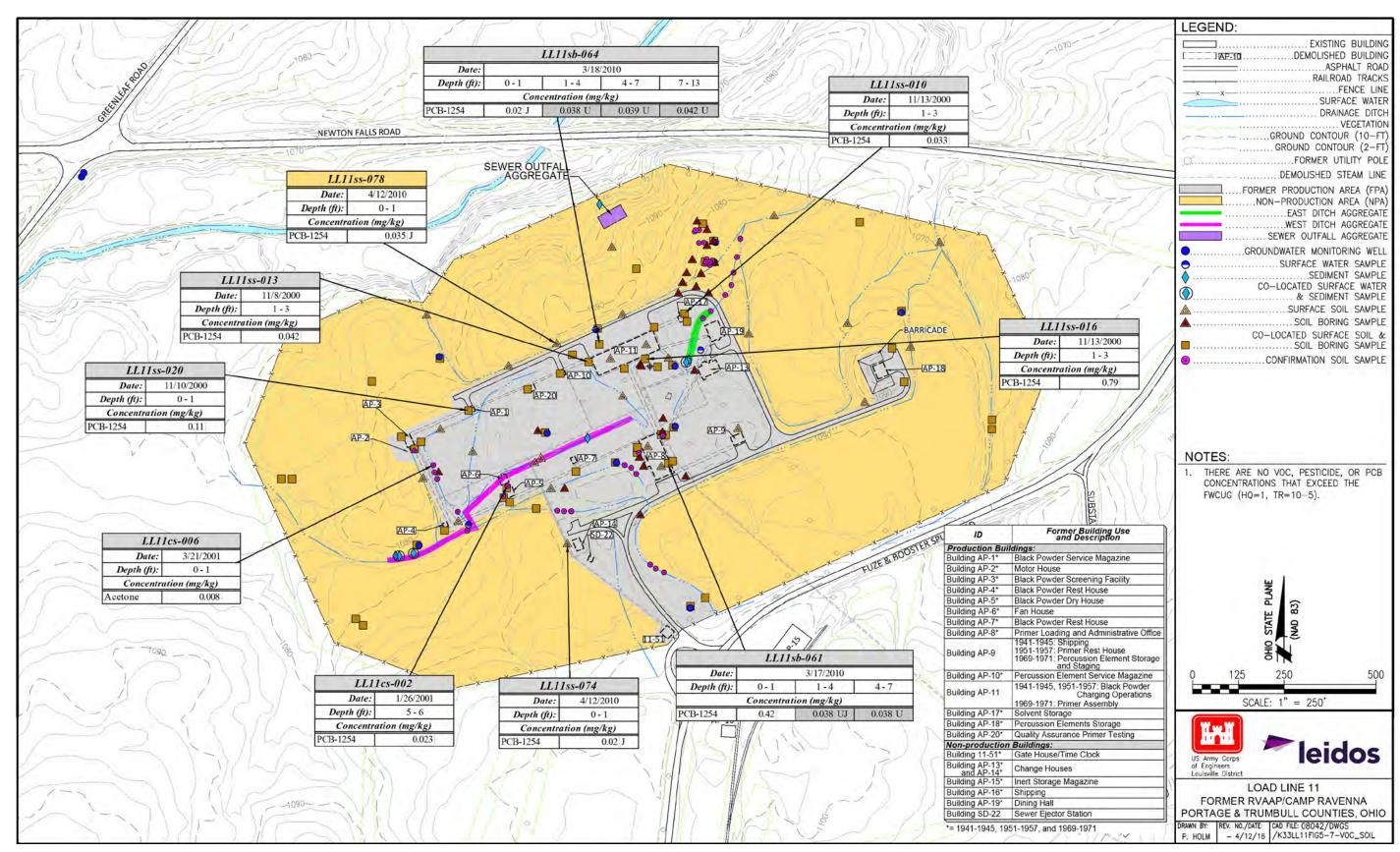


Figure 5-7. Detected Concentrations of VOCs, Pesticides, and PCBs in Soil at the FPA and NPA

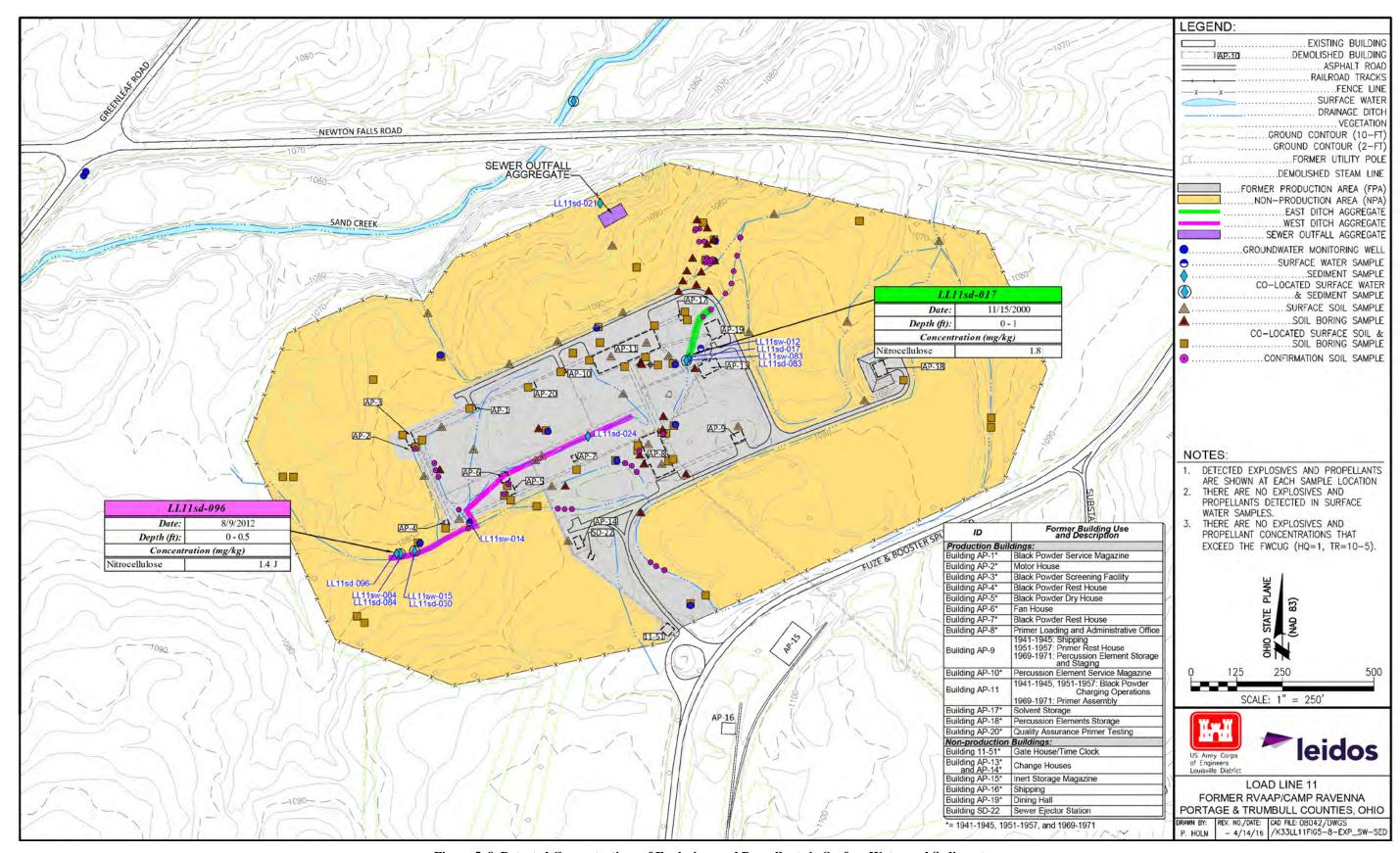


Figure 5–8. Detected Concentrations of Explosives and Propellants in Surface Water and Sediment

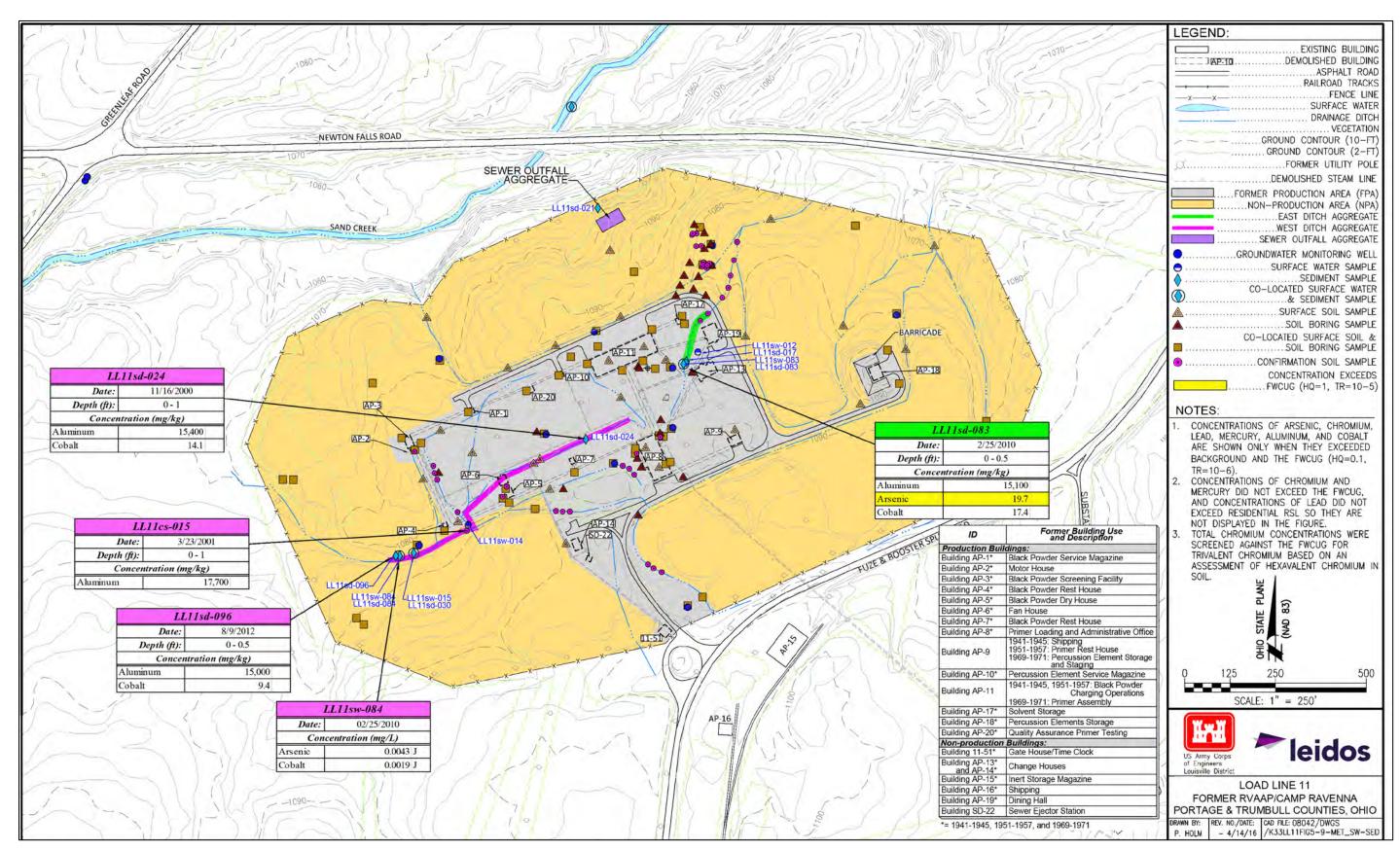


Figure 5-9. Exceedances of FWCUG (HQ of 0.1, TR 1E-06) for Arsenic, Chromium, Lead, Mercury, Aluminum, and Cobalt in Surface Water and Sediment

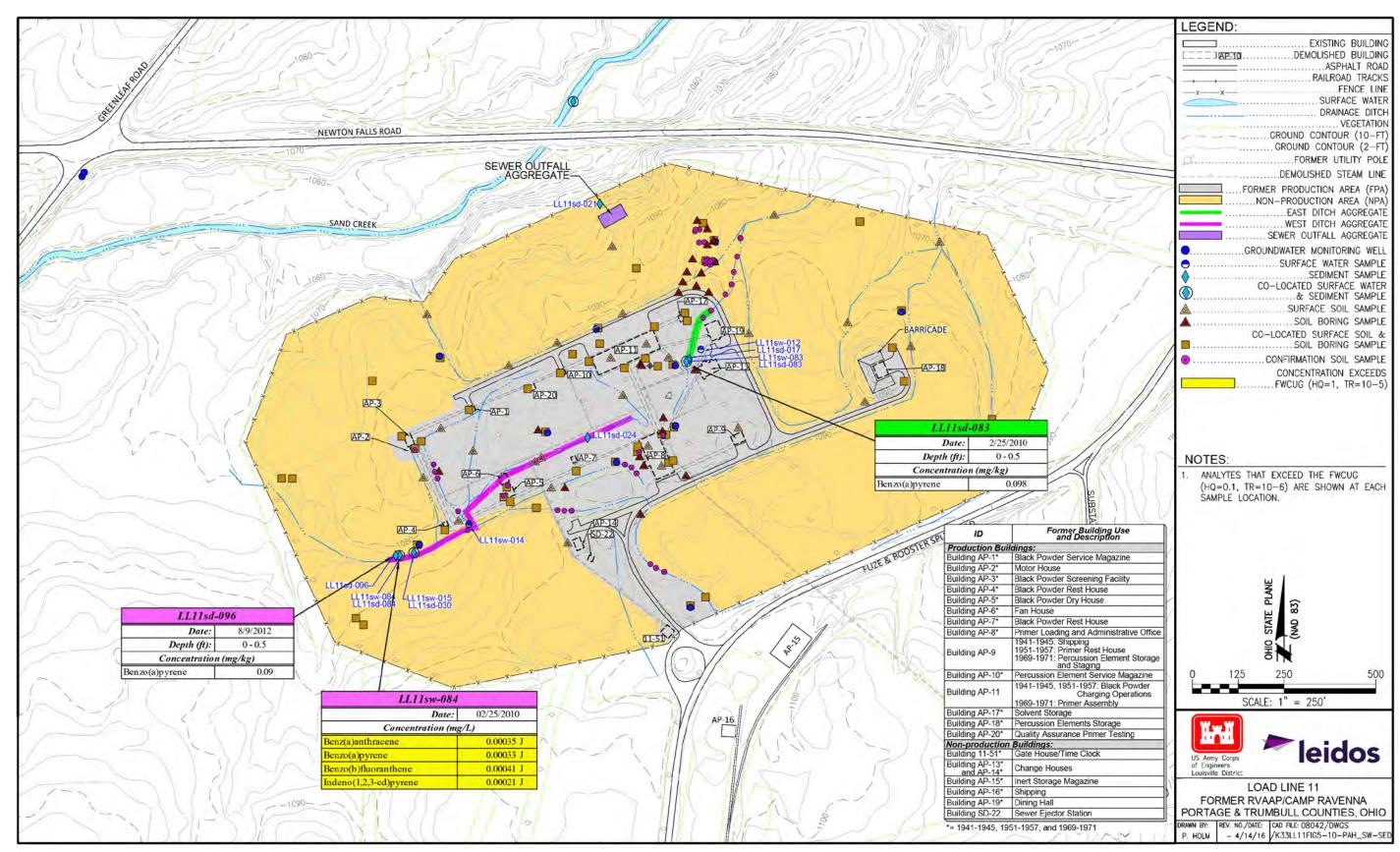


Figure 5-10. PAH Exceedances of FWCUG (HQ of 0.1, TR of 1E-06) in Surface Water and Sediment

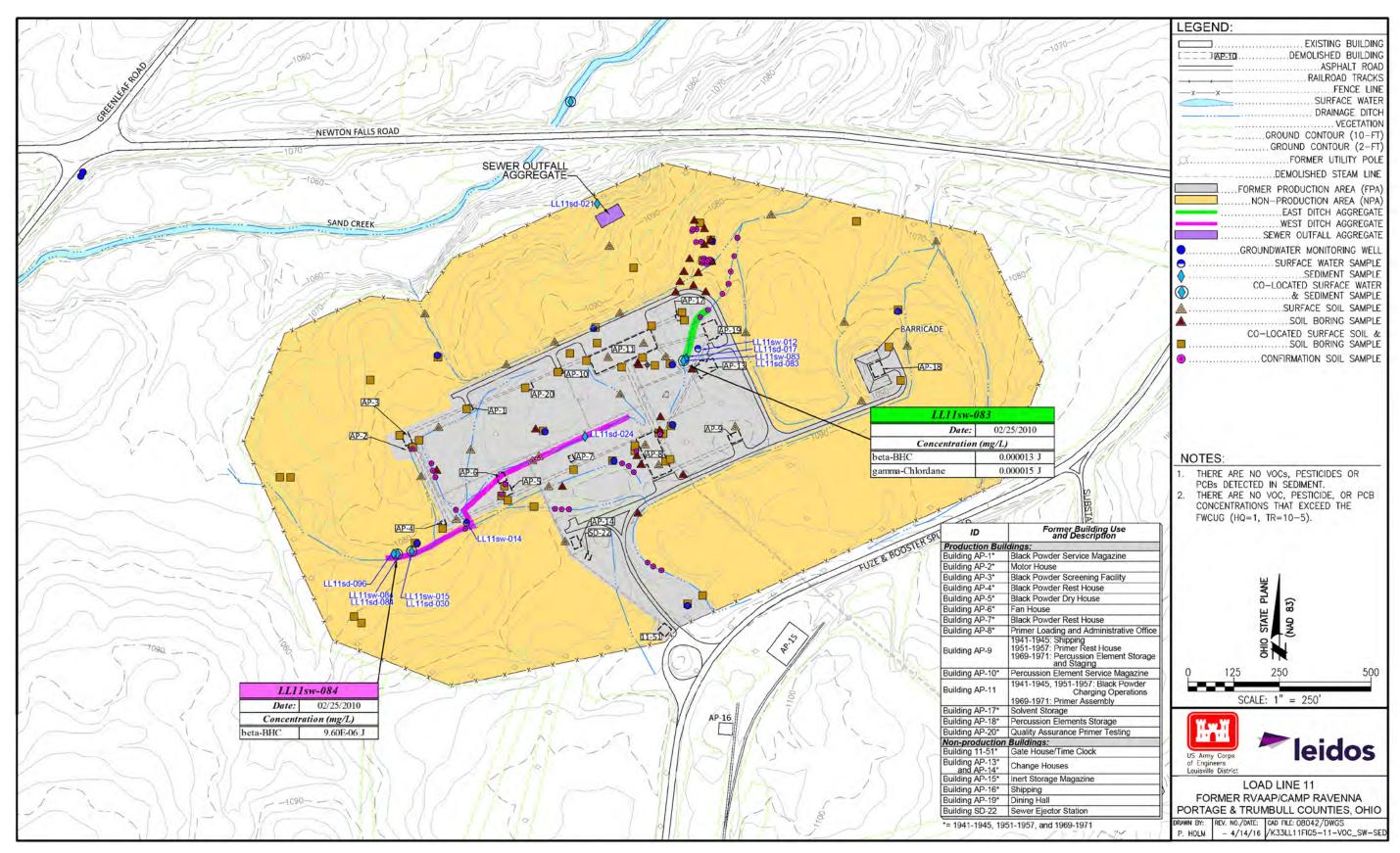


Figure 5–11. Detected Concentrations of VOCs, Pesticides, and PCBs in Surface Water and Sediment

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6.0 CONTAMINANT FATE AND TRANSPORT

Contaminant fate and transport modeling assesses the potential for SRCs to leach from surface soil, subsurface soil, and sediment sources at Load Line 11 and impact groundwater beneath the existing sources and downgradient receptor locations. As indicated in Section 4.2, sources identified at Load Line 11 have been removed as part of the IRA (MKM 2004a). Five sedimentation sumps, 230 yd³ of contaminated sediment within drainage ditches, and petroleum-contaminated soil within a 30 by 30 by 8 ft hotspot area have been removed from the site.

Modeling results were included in the decision-making process to determine whether performing remedial actions may be necessary to protect groundwater resources. Surface water exposure pathways are evaluated in the HHRA and ERA presented in Sections 7.2 and 7.3, respectively. A summary of the principles of contaminant fate and transport are presented in this section along with the results of the modeling.

Section 6.1 describes physical and chemical properties of SRCs found in soil and sediment at the AOC. Section 6.2 presents a conceptual model for contaminant fate and transport that considers AOC topography, hydrogeology, contaminant sources, and release mechanisms. Section 6.3 presents a soil screening analysis, and Section 6.4 presents a sediment screening analysis to identify SRCs with the potential to migrate from soil and sediment to groundwater as initial CMCOPCs. Section 6.5 describes fate and transport modeling of final CMCOPCs and presents CMCOCs. Section 6.6 provides a list of the remaining final CMCOPCs, a qualitative assessment of the results, and considerations of the limitations and assumptions. Section 6.7 summarizes the conclusions of the fate and transport analysis.

6.1 PHYSICAL AND CHEMICAL PROPERTIES OF SITE-RELATED CONTAMINANTS

The 1978 Installation Assessment identified the major contaminants of the former RVAAP as TNT, composition B (a combination of TNT and RDX), sulfates, nitrates, lead styphnate, and lead azide. Additional site-specific contaminants include black powder, TNT, RDX, HMX, nitroglycerine, nitrocellulose, nitroguanidine, and heavy metals (lead, chromium, mercury, and arsenic) from munitions assembly activities; VOCs from former Building AP-17 that was utilized for solvent storage and lacquer sealing materials used on finished primers in former Building AP-11; and PCBs from on-site transformers.

This evaluation of contaminant fate and transport evaluates not only those chemicals identified as potential contaminants from previous use but also includes an evaluation of chemicals that were evaluated as part of the overall RI.

The comprehensive list of surface and subsurface soil (including 20 inorganic chemicals and 22 organic chemicals in the FPA and 19 inorganic chemicals and 21 organic chemicals in the NPA) and sediment SRCs (including 10 inorganic chemicals and 15 organic chemicals in the East Ditch

aggregate, and 11 inorganic chemicals and 10 organic chemicals in the West Ditch aggregate) is detailed in Section 4.0 and are summarized below:

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- Inorganic SRCs in surface and subsurface soil: aluminum, arsenic, barium, beryllium, cadmium, chromium, cobalt, copper, cyanide, lead, manganese, mercury, nickel, nitrate, silver, sulfate, sulfide, thallium, vanadium, and zinc.
- Inorganic SRCs in sediment: aluminum, antimony, arsenic, beryllium, cadmium, chromium, cobalt, copper, mercury, nickel, selenium, silver, and vanadium.
- Organic SRCs in surface and subsurface soil: 2-methylnaphthalene, 4-nitrotoluene, acenaphthene, anthracene, benz(a)anthracene, benzenemethanol, benzo(a)pyrene, benzo(b)fluoranthene, benzo(ghi)perylene, benzo(k)fluoranthene, beta-BHC, bis(2ethylhexyl)phthalate, carbazole, chrysene, di-n-butyl phthalate, dibenz(a,h)anthracene, dibenzofuran. fluoranthene, fluorene, HMX, indeno(1,2,3-cd)pyrene, naphthalene, nitrocellulose, PCB-1254, phenanthrene, phenol, pyrene, and tetryl.
- Organic SRCs in sediment: 2-methylnaphthalene, acenaphthene, anthracene, benz(a)anthracene, benzenemethanol, benzo(a)pyrene, benzo(b)fluoranthene, benzo(ghi)perylene, benzo(k)fluoranthene, chrysene, dibenz(a,h)anthracene, fluoranthene, fluorene, indeno(1,2,3-cd)pyrene, naphthalene, nitrocellulose, phenanthrene, and pyrene.

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Chemicals released into the environment are susceptible to several degradation pathways, including hydrolysis, oxidation, reduction, isomerization, photolysis, photo-oxidation, biotransformation, and biodegradation. Transformed products resulting from these processes may behave differently than their parent chemical in the environment.

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The migration of chemicals is governed by their physical and chemical properties and the surface and subsurface media through which chemicals are transferred. In general, chemicals and structures with similar physical and chemical characteristics will show similar patterns of transformation, transport, or attenuation in the environment. Solubility, vapor pressure data, chemical partitioning coefficients, degradation rates, and Henry's Law Constant (HLC) provide information that can be used to evaluate contaminant mobility in the environment. Partitioning coefficients are used to assess relative affinities of chemicals for solution or solid phase adsorption. However, the synergistic effects of multiple migrating chemicals and complexity of soil/water interactions, including pH and oxidation-reduction potential, grain size, and clay mineral variability, are typically unknown.

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The physical properties of the chemicals defined as SRCs in surface and subsurface soil are summarized in Appendix E, Tables E-1 and E-2. These properties are used to assess the anticipated behavior of each chemical under environmental conditions. The physical properties of the chemicals defined as SRCs detected in soil and sediment are summarized in Sections 6.1.1 through 6.1.5.

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6.1.1 Chemical Factors Affecting Fate and Transport

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The water solubility of a chemical is a measure of the saturated concentration of the chemical in water at a given temperature and pressure. The tendency for a chemical to be transported by groundwater is directly related to its solubility and inversely related to its tendencies to adsorb to soil and volatilize

from water (OGE 1988). Chemicals with high water solubilities tend to desorb from soil, are less likely to volatilize from water, and are susceptible to biodegradation. The water solubility of a chemical varies with temperature, pH, and the presence of other dissolved chemicals (including organic carbon and humic acids).

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The octanol-water partition coefficient (K_{ow}) can be used to estimate the tendency for a chemical to partition between environmental phases of different polarity. The K_{ow} is a laboratory-determined ratio of the concentration of a chemical in the n-octanol phase of a two-phase system to the concentration in the water phase. Chemicals with log K_{ow} values less than one are highly hydrophilic, while chemicals with log K_{ow} values greater than four will partition to soil particles (Lyman et al. 1990).

The water/organic carbon partition coefficient (K_{oc}) is a measure of the tendency of an organic chemical to partition between water and organic carbon in soil. The K_{oc} is defined as the ratio of the absorbed chemical per unit weight of organic carbon to the aqueous solute concentration.

This coefficient can be used to estimate the degree to which an organic chemical will adsorb to soil and thus not migrate with groundwater. The higher the K_{oc} value, the greater is the tendency of the chemical to partition into soil (OGE 1988). The soil/water partitioning coefficient (K_d) is calculated by multiplying the K_{oc} value by the fraction of organic carbon in the soil.

Vapor pressure is a measure of the pressure at which a chemical and its vapor are in equilibrium. This value can be used to determine the extent to which a chemical would travel in air, as well as the rate of volatilization from soil and solution (OGE 1988). In general, chemicals with vapor pressures lower than 10^{-7} mm mercury will not be present in the atmosphere or air spaces in soil in significant amounts, while chemicals with vapor pressures higher than 10^{-2} mm mercury will exist primarily in the air (Dragun 1988).

The HLC value for a chemical is a measure of the ratio of the chemical's vapor pressure to its aqueous solubility. The HLC value can be used to make general predictions about a chemical's tendency to volatilize from water. Chemicals with HLC values less than 10^{-7} atm-m³/mol will generally volatilize slowly, while chemicals with a HLC greater than 10^{-3} atm-m³/mol will volatilize rapidly (Lyman et al. 1990).

6.1.2 Biodegradation

Organic chemicals with differing chemical structures will biodegrade at different rates. Primary biodegradation consists of any biologically induced structural change in an organic chemical. Complete biodegradation is the biologically mediated degradation of an organic chemical into carbon dioxide, water, oxygen, and other metabolic inorganic products (Dragun 1988).

Load Line 11

The first order biodegradation rate of an organic chemical is proportional to the concentration:

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-dC/dt = kC (Equation 6-1)

4 Where:

C = concentration

t = time

 $k = biodegradation rate constant = ln 2 / t_{1/2}$

 $t_{1/2}$ = biodegradation half-life

The biodegradation half-life is the time necessary for half of the chemical to degrade. The biodegradation rate of an organic chemical generally depends on the presence and population size of soil microorganisms that are capable of degrading the chemical.

6.1.3 Inorganic Chemicals

Inorganic chemicals detected in soil and sediment samples are associated with the aqueous phase and leachable metal ions on soil particles. The transport of this material from unsaturated soil to the underlying water table is controlled by the physical processes of precipitation percolation, chemical interaction with the soil, and downward transport of metal ions by continued percolation. The chemistry of inorganic chemical interactions with percolating precipitation and varying soil conditions is complex and includes numerous chemical transformations that may result in altered oxidation states, including ion exchange, adsorption, precipitation, or complexation. The chemical reactions, which are affected by environmental conditions (i.e., pH, oxidation/reduction conditions, type and amount of organic matter, clay content, and the presence of hydrous oxides), may act to enhance or reduce the mobility and toxicity of metal ions. In general, these reactions are reversible and add to the variability commonly observed in distributions of inorganic chemicals in soil.

The chemical form of an inorganic chemical determines its solubility and mobility in the environment; however, chemical speciation is complex and difficult to delineate in routine laboratory analysis. Inorganic chemicals in soil are commonly found in several forms, including dissolved concentrations in soil pore water, metal ions occupying exchange sites on inorganic soil constituents (adsorbed to inorganic soil constituents), metal ions associated with insoluble organic matter, precipitated inorganic chemicals as pure or mixed solids, and metal ions present in the structure of primary or secondary minerals.

The dissolved (aqueous) fraction and its equilibrium sorbed fraction are important when considering the migration potential of inorganic chemicals through soil. Of the inorganic chemicals that are likely to form, chlorides, nitrates, and nitrites are commonly the most soluble. Sulfate, carbonate, and hydroxides generally have low to moderate solubility. Soluble chemicals are transported in aqueous forms subject to attenuation, whereas less soluble chemicals remain as a precipitate and limit the overall dissolution of metal ions. The solubility of the metal ions is also regulated by ambient chemical conditions, including pH and oxidation/reduction.

 The attenuation of metal ions in the environment can be estimated numerically using the retardation factor (R), dispersion in higher flow systems (high conductivity environments), and diffusion in low conductivity environments. R defines the extent to which the velocity of the contaminant is slowed, which is largely derived from the K_d . R is calculated using the following equation:

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$$R = 1 + (K_d \rho_b)/\theta_w$$
 (Equation 6-2)

Where:

 ρ_b = the soil bulk dry density (g/cm³)

 $\theta_{\rm w}$ = soil moisture content (dimensionless)

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Metal ion concentrations in the environment do not attenuate by natural or biological degradation because of low volatility and solubility of the ions. Inorganic chemicals may be biotransformed or bioconcentrated through microbial activity.

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6.1.4 Organic Chemicals

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Organic chemicals, such as SVOCs or VOCs, may be transformed or degraded in the environment by processes including hydrolysis, oxidation/reduction, photolysis, volatilization, biodegradation, or biotransformation. The half-life of organic chemicals in transport media can vary from minutes to years, depending on environmental conditions and chemical structures. Some types of organic chemicals are very stable, and degradation rates can be very slow. Organic degradation may either enhance (by producing more toxic byproducts) or reduce (reducing concentrations) the toxicity of a chemical in the environment.

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6.1.5 Explosives – Related Chemicals

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Several explosive compounds were detected in soil and sediment at Load Line 11. Nitrocellulose is an aliphatic nitrate ester that will gelatinize with nitroglycerin when mixed together. Nitrocellulose occurs as a fibrous solid that can act as a sorbent that will dissolve in water under highly basic conditions with high temperatures. Nitrocellulose can undergo denitrification as a degradation pathway. Degradation of nitrocellulose to non-reactive nitrocellulose has been observed under methanogenic and fungus-mediated reducing conditions (USACE 2006). Tetryl may undergo biotransformation, hydrolysis, and photodegradation to produce several different transformation products, such as n-methylpicramide and 4-amino-n-methyl-n,2,6-trinitroaniline (USACE 1994). Nitroguanidine is a mobile nitramine chemical that can be biotransformed in aquatic environments and in soil under aerobic and anaerobic conditions. The rate of degradation of nitroguanidine in soil is dependent on the amount of organic carbon and microbial adaptation to nitroguanidine. Biotransformation products of nitroguanidine include cyanamide, melamine, and guanidine (USACE 2006). PETN undergoes biodegradation forming trinitrate and then dinitrate, which is then further degraded. PETN has low volatility and low solubility in water, and therefore has low bioavailability for most organisms. Neutron radiation degrades PETN, producing carbon dioxide and some pentaerythritol dinitrate and trinitrate. Like other nitrate esters, the primary degradation mechanism is the loss of nitrogen dioxide. Limited information exists regarding biotransformation or biodegradation of HMX. The biotransformation, primarily by anaerobic degradation (ERDC 2007), of

HMX occurs at a slow rate in the environment (USACHPPM 2001). HMX is primarily broken down 2 by photolysis and has a photolytic rate constant of 0.15 days (USEPA 1988). Breakdown products of 3 HMX include nitrate, nitrite, and formaldehyde (USACHPPM 2001). Appendix E, Figure E-1 4 (ATSDR 1997) shows the biotransformation pathway for HMX.

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6.2 CONCEPTUAL MODEL FOR FATE AND TRANSPORT

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The conceptual site model (CSM), which defines the framework for fate and transport modeling, describes conditions at Load Line 11, including the contaminant sources, surficial and subsurface hydrogeologic conditions, contaminant migration and pathways, and contaminant release mechanisms.

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AOC conditions described in Sections 2.0 through 5.0 include contaminant source information, the surrounding geologic and hydrologic conditions, and the magnitude of SRCs and their current spatial distribution. Information from Section 3.0 and the nature and extent evaluation in Section 5.0 were used to develop the CSM for fate and transport modeling by identifying SRCs and migration pathways. The CSM is based on information and data collected for historical investigations, this RI Report, and informed assumptions about the AOC. Assumptions contained in the CSM are reiterated throughout this section. The better the information and the greater the accuracy of the assumptions, the more accurately the CSM describes the AOC; therefore, the more reliable the fate and transport modeling predictions can be. The salient elements of the CSM that apply to fate and transport modeling are summarized in the following sections.

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6.2.1 **Contaminant Sources**

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No primary contaminant sources are located on the AOC. Secondary sources (i.e., contaminated media including sediment) identified in previous investigations at the AOC are further evaluated in this report.

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6.2.2 **Hydrogeology**

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A description of regional and AOC-specific geology and hydrology are provided in Sections 3.3.3 and 3.4.2, respectively, and are summarized below.

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The topography at Load Line 11 ranges from approximately 1,070 ft amsl near the northern fence line to 1,100 ft amsl near the southern fence line. Surface water drainage associated with heavy rainfall events follows the topography and drains into the east ditch that flows north and the west ditch that flows west northwest. Both ditches flow into Sand Creek (Figure 3-1).

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 Soil beneath the AOC consists of shallow clay to sand-rich silt tills with interbedded sands scattered throughout, as observed in subsurface borings installed during the PBA08 RI (Appendix A).

43 44 • Twelve groundwater monitoring wells were installed at the AOC to an average depth of 19 ft bgs. Eleven of these monitoring wells monitor groundwater in the unconsolidated zone to an

- average depth of 19 ft bgs. The other well (LL11mw-012) was screened in the bedrock consisting of Sharon Shale to a depth of 115 ft bgs.
- The potentiometric surface shows the groundwater flow direction is to the north with an average hydraulic gradient of 0.017 ft/ft (Figure 3-1).
- Water level elevations at the AOC range from 1,068.40 to 1,091.73 ft amsl with the highest elevation at well LL11mw-001 (Figure 3-1). Potentiometric data indicate the groundwater table occurs within the unconsolidated soil throughout the AOC.

6.2.3 Contaminant Release Mechanisms and Migration Pathways

Based on the information presented above, the following contaminant release mechanisms and migration pathways have been identified at the AOC:

- Contaminant leaching from soil to the water table (vertical migration) and lateral transport to the downgradient receptor (i.e., Sand Creek north of Load Line 11);
- Contaminated sediment transported to potential downstream receptors;
- Contaminated surface water migrating to potential downstream receptors; and
- Contaminated sediment from the wetlands as a secondary source of leaching to the water table (vertical migration) and lateral transport to potential downgradient receptors.

The first of these pathways, which considers a primary groundwater transport pathway, is treated explicitly in this fate and transport section. Sediment and surface water exposure pathways are evaluated in the HHRA and ERA presented in Sections 7.2 and 7.3, respectively. The fourth pathway listed above, which considers a secondary groundwater transport pathway, was evaluated using the sediment screening analysis presented in Section 6.4.

One of the principal migration pathways at the AOC is percolation through the unsaturated soil to the water table (i.e., vertical leaching of contaminants from soil into groundwater). However, because of the very heterogeneous nature of the unconsolidated glacial material, groundwater flow patterns within the unconsolidated soil are difficult to predict. Precipitation that does not leave the AOC as surface runoff percolates into the subsurface. Some of the percolating water leaves this environment via evapotranspiration after little or no vertical migration. The remainder of the water percolates into the water table. As discussed in Section 6.2.4, the rate of percolation is controlled by soil cover, ground slope, saturated conductivity of the soil, and meteorological conditions. Figure 6-1 illustrates the contaminant migration conceptual model.

Once the contaminant leachate percolates through the soil and reaches the water table, it migrates with the local groundwater and discharges at the downgradient receptors. Groundwater flow likely occurs along preferential pathways (e.g., sand seams, channel deposits, or other stratigraphic discontinuities) having higher permeabilities. For inorganic chemicals, lateral migration through groundwater will be very limited due to their high retardation by the bedrock material (USACE 2003b).

Additional factors that affect the leaching rate include a chemical's solubility, sorption capacity (expressed by the K_d), and the amount of percolation. Insoluble chemicals will precipitate out of the solution in the subsurface or remain in insoluble forms with little leaching.

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Another factor that affects whether a chemical will reach the water table through percolation of precipitation is the chemical's rate of decay. Most organic compounds decay at characteristic rates proportional to the chemical's half-life. For a given percolation rate, those chemicals with long half-lives have a greater potential for contaminating groundwater than those with shorter half-lives. For this analysis, the rate of decay/half-life was not considered.

Contaminant releases through gaseous emissions and airborne particulates are not significant at Load Line 11. The AOC is vegetated, located in a humid temperate climate, and soil moisture is typically high, which prevents dust borne contaminant migration. Therefore, there is likely little to no gaseous emission, and contaminant levels in the air pathway are minor to non-existent.

6.2.4 Water Budget

The potential for contaminant transport begins with precipitation. Percolation is the driving mechanism for soil contaminants leaching to groundwater. The actual amount of rainwater available for flow and percolation to groundwater is highly variable and depends upon soil type and climatic conditions. A water balance calculation can be used as a tool to quantitatively account for all components of the hydrologic cycle. The quantified elements of the water balance are used for inputs to the soil leaching and groundwater transport models discussed later. The components of a simple, steady-state water balance model include precipitation, evapotranspiration, surface runoff, and groundwater recharge or percolation.

These terms are defined as follows:

$$P = ET + Sr + q$$
 (Equation 6-3)

33 Rainwater available for flow = Sr + q = P - ET

(Equation 6-4)

Page 6-8

34 Where:

or

P = precipitation

Sr = surface runoff

ET = evapotranspiration

q = groundwater recharge or percolation

It is expected that runoff loss also occurs through evaporation. The remaining water, after runoff and evaporation, is available for percolation which includes loss to the atmosphere by evapotranspiration. The water balance estimations were developed using the Hydrologic Evaluation of Landfill Performance (HELP) model (USEPA 1994). See Appendix E, Table E-3 for parameters used in the HELP model to develop the water budget estimates used in the evaluation. Calculations using

precipitation and temperature data for a 100-year period were generated synthetically using coefficients for Cleveland, Ohio (e.g., the nearest weather station to Camp Ravenna with HELP model coefficients).

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The annual average water balance estimates indicate an evapotranspiration rate of 28% (10.3 inches) of total precipitation (37 inches). The remaining 72% (26.7 inches) of rainwater is available for surface water runoff and percolation to groundwater. Of the 72% (26.7 inches) of water available for runoff or percolation, groundwater recharge (percolation) accounts for 13% (3.6 inches), and surface runoff (along downgradient topography to nearest surface water bodies) accounts for the remaining 87% (23.1 inches).

6.3 SOIL SCREENING ANALYSIS

Soil screening analyses are screening evaluations performed to identify SRCs with the potential to leach to groundwater as CMCOPCs. This section describes the soil screening analysis approach and presents the limitations and assumptions.

6.3.1 Analysis Approach

The five steps for the soil leachability analysis are illustrated in Figure 6-2 and are described below.

The first step of the soil screening analysis is developing SRCs, as presented in Section 4.0. A summary of SRCs identified for soil and sediment is presented in Section 6.1.

The second step of the soil screening process (Figure 6-2) involves comparing the maximum concentrations of the SRCs with MCL-based generic soil screening levels (GSSLs). GSSLs were developed for Superfund sites for contaminant migration to groundwater (USEPA 1996b, USEPA 2015). The GSSL is defined as the concentration of a chemical in soil that represents a level of contamination below which there is no concern for impacts to groundwater under CERCLA, provided conditions associated with USEPA risk-based soil screening levels (SSLs) are met. Generally, if chemical concentrations in soil fall below the GSSL and there are no groundwater receptors of concern or anticipated exposures, then no further study or action is warranted for that chemical. If the GSSL for a chemical is not available, the USEPA risk-based SSL for groundwater migration, dated June 2015 (USEPA 2015), will be obtained from the USEPA RSL website and used. If neither the GSSL nor the USEPA risk-based SSL for a chemical is available, then no further evaluation of the chemical is performed and it is eliminated from the list of the initial CMCOPCs. However, some chemicals have been assigned surrogates by risk assessors if the chemical without an SSL is similar to another chemical with an SSL. Surrogates used for this analysis include acenaphthene for acenaphthylene and pyrene for benzo(ghi)perylene and phenanthrene.

The initial CMCOPC screen, as presented in Appendix E, Table E-4, eliminated 7 inorganic SRCs and 14 organic SRCs from the FPA and 6 inorganic SRCs and 11 organic SRCs from the NPA from further consideration. Inorganic anion SRCs nitrate, sulfate, and sulfide from FPA and NPA, and organic SRCs TPH-DRO and TPH-GRO from NPA were eliminated from further consideration as

there are no SSLs for these SRCs. A total of 10 inorganic and 8 organic SRCs from the FPA and 10 inorganic and 8 organic SRCs from the NPA were carried forward to the third step of the soil CMCOPC screening process.

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The third step of the soil screening process (Figure 6-2) involves comparing the maximum chemical concentrations with the site-specific soil screening levels (SSSLs). The SSSL is defined as the GSSL (or the USEPA risk-based SSL for groundwater protection if a GSSL is not available) multiplied by the AOC-specific dilution attenuation factor (DAF). Direct partitioning is used to derive GSSLs, assuming groundwater is in contact with the chemicals in soil and the groundwater concentration is equal to the leachate concentration. However, as leachate moves through soil, chemical concentrations are attenuated by adsorption and degradation. When the leachate reaches the water table, dilution by groundwater further reduces leachate concentrations. This concentration reduction can be expressed by a DAF. DAFs can vary based on AOC-specific characteristics (e.g., hydrogeologic properties, contaminated source area, and depth to contamination). As described in the *Soil Screening Guidance: Technical Background Document* (USEPA 1996b), chemical dilution in groundwater is estimated at each AOC from an AOC-specific DAF. The DAF, which is defined as the ratio of soil leachate concentration to a receptor point concentration, is minimally equal to one. Dilution in groundwater is derived from a simple mixing zone equation (Equation 6-5) and relies upon estimating the mixing zone depth (Equation 6-6).

DAF =
$$1 + \frac{(K \times i \times d)}{(q \times L)}$$
 (Equation 6-5)

Where:

DAF = dilution attenuation factor

K = aquifer hydraulic conductivity (m/yr)

i = horizontal hydraulic gradient (m/m)

q = percolation rate (m/yr)

L = source length parallel to groundwater flow (m)

d = mixing zone depth (m) (which is defined below)

30 and

32
$$d = \sqrt{0.0112 \times L^2} + d_a \times \left[1 - \exp\left(\frac{-L \times I}{K \times i \times d_a}\right) \right]$$
 (Equation 6-6)

33 Where:

 $d_a = aquifer thickness (m)$

 $d \leq d_a$

As stated above, if the aquifer thickness is less than the calculated mixing zone depth, then the aquifer thickness is used for "d" in the DAF calculation. The DAF calculation for the AOC is presented in Appendix E, Table E-5. It should be noted that the purpose of this screen is not to identify the chemicals that may pose risk at downgradient locations, but to target those chemicals that pose the greatest problem if they migrate from the AOC.

Based on this screening and an AOC-specific DAF of 1.04 for the FPA and 1.03 for the NPA, none of the chemicals were eliminated from further consideration. Therefore, all the remaining SRCs from the previous step also exceeded their published or calculated GSSL multiplied by the respective DAF and were identified as initial CMCOPCs based on leaching to groundwater. The SRCs identified as initial CMCOPCs are presented in Appendix E, Table E-6.

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The fourth step of the soil screening process (Figure 6-2) involves eliminating initial CMCOPCs identified in the SSSL evaluation which require more than 1,000 years to leach through the unsaturated zone before reaching the water table from further consideration. A period of 1,000 years was conservatively selected to evaluate eventual migration of the contaminant front to the water table despite uncertainties in vadose zone hydraulic parameters and groundwater recharge over time. Additionally, USACE suggests a screening value of 1,000 years be used due to the high uncertainty associated with predicting conditions beyond that time frame (USACE 2003b). Therefore, the initial CMCOPCs at the selected sources were screened against a travel time of greater than 1,000 years. The travel time in this screen is the time required for a CMCOPC to migrate vertically from the base of the soil interval detected above the background concentration to the water table. This distance is the leaching zone, which is evaluated in Appendix E, Table E-7, which may vary across the AOC based on the varying depths of soil sample concentrations above the facility-wide background concentrations and the elevation of the water table. The estimated travel time for each initial CMCOPC to reach the water table is determined using the following equations:

$$T = \frac{Lz \times R}{V_p}$$
 (Equation 6-7)

Where:

T = leachate travel time (year)

Lz = thickness of attenuation zone (ft)

R = retardation factor (dimensionless) (Equation 6-2)

V_p = porewater velocity (ft/year)

29 and

$$V_p = \frac{q}{\theta_w}$$
 (Equation 6-8)

32 Where:

q = percolation rate (ft/year)

 $\theta_{\rm w}$ = fraction of total porosity that is filled by water

If the travel time for a chemical from a source area exceeded 1,000 years, then the chemical was eliminated from the list of initial CMCOPCs. A total of 11 inorganic and organic SRCs at the FPA and 12 inorganic and organic SRCs at the NPA were eliminated from further consideration based on their travel times exceeding 1,000 years. Initial CMCOPCs with travel times less than 1,000 years were retained for further evaluation (Appendix E, Table E-7). The constituents selected for further evaluation with Seasonal Soil Compartment (SESOIL) modeling are listed in Table 6-1.

In the fifth step (Figure 6-2), the initial CMCOPCs were further evaluated using fate and transport models provided in Section 6.5.

6.3.2 Limitations and Assumptions of Soil Screening Analysis

It is important to recognize that acceptable soil concentrations for individual chemicals are highly AOC-specific. The GSSLs used in this screening represent USEPA's protection of groundwater SSLs that were developed based on a number of default assumptions chosen to be protective of human health for most AOC conditions (USEPA 1996b). These GSSLs are expected to be more conservative than SSSLs based on AOC conditions. The conservative assumptions included in this analysis are: (1) no adsorption in the unsaturated zone or in the aquifer, (2) no biological or chemical degradation in the soil or aquifer, and (3) contamination is uniformly distributed throughout the source. However, the GSSL does not incorporate the existence of contamination already present within the aquifer.

6.4 SEDIMENT SCREENING ANALYSIS

Sediment SRCs were developed in Section 4.0 and are presented in Section 6.1. The purpose of this screening analysis is to identify CMCOPCs based on contaminant migration from sediment to groundwater. Any identified CMCOPCs are modeled with Analytical Transient 1-, 2-, 3-Dimensional (AT123D) model to a downgradient receptor if present. The four steps for the sediment screening analysis are illustrated in Figure 6-3.

Sediment SRCs were screened by developing leachate concentrations assuming equilibrium between sediment and groundwater. The predicted leachate concentrations were diluted based on a sample-specific DAF calculated by dividing the calculated leachate concentrations by the co-located surface water concentrations. The DAF was calculated for each chemical that was detected in the sediment and surface water at the same sample location. The calculated DAF was then used to calculate the maximum groundwater concentration, considering dilution for sediment SRCs and assuming that the sediment is in direct contact with groundwater. The lowest DAF calculated for the sample area was used for sediment SRCs that did not have a sample-specific DAF. The DAFs calculated for each chemical are shown in Table 6-2 for the two aggregates (East Ditch and West Ditch).

This sediment screening analysis assumed that the sediment concentration and the recharging groundwater concentration were in equilibrium and a sample-specific DAF for each chemical was used for dilution in the aquifer. Based on this screening analysis (see Table 6-2), benz(a)anthracene and naphthalene at the East Ditch aggregate, and antimony, benz(a)anthracene, and benzo(b)fluoranthene at the West Ditch aggregate exceeded the risk-based screening criteria and were carried forward for further evaluation with the AT123D model. All other sediment SRCs were eliminated.

6.5 FATE AND TRANSPORT MODELING

Contaminant fate and transport modeling represents the fifth step in the fate and transport screening and evaluation process (Figure 6-2). SESOIL modeling was performed for chemicals identified as

initial CMCOPCs from the soil screening analysis presented in Section 6.3 and summarized in Table 6-1. SESOIL modeling was performed to predict chemical concentrations in the leachate immediately beneath the selected source areas and just above the water table. If the predicted maximum leachate concentration of an initial CMCOPC was higher than the facility-wide background concentration and the lowest risk-based screening value (i.e., Resident Receptor Adult FWCUG, MCL, or RSL), the CMCOPC was further evaluated using the AT123D model to predict future maximum concentrations in groundwater beneath the source, as well as at downgradient receptor locations, if applicable. The downgradient receptor location (if required) is the closest surface water body feature downgradient of the source areas that is connected to the groundwater. The predicted concentrations of CMCOPCs in groundwater beneath the source were compared to available groundwater monitoring results for the AOC to validate modeling results and provide WOE for identifying or eliminating CMCOCs.

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6.5.1 Modeling Approach

Contaminant transport includes the movement of water and dissolved material from the source areas to groundwater. This occurs as rainwater infiltrates the surface and percolates through the area of contamination, its surrounding soil, and into the saturated zone. The downward movement of water, driven by gravitational potential, capillary pressure, and other components of total fluid potential mobilizes the contaminants and carries them through the soil into the mixing zone with the water table. Lateral transport within the shallow bedrock is controlled by the groundwater gradient. Vertical transport (evaluated with the SESOIL model) through the overburden to the water table and horizontal transport (evaluated with the AT123D model) through the unconsolidated sediment to downgradient receptor locations are illustrated in Figure 6-1.

The output of the contaminant fate and transport modeling is presented as the expected maximum concentration of modeled contaminants at the selected receptor locations. For SESOIL, the receptor location is the groundwater table beneath the source area. For this analysis, six discrete sampling locations for the FPA and four discrete sampling locations for the NPA were considered sources of contamination based on the results of the soil screening analysis. A separate SESOIL analysis was performed for each initial CMCOPC listed in Table 6-1 and is presented in Figure 6-4.

The predicted maximum leachate concentration just above the water table, observed in the SESOIL results, was compared against its applicable RVAAP facility-wide background concentration, as well as RVAAP FWCUGs for the Resident Receptor Adult, MCLs, and RSLs. If the predicted maximum leachate concentration of an initial CMCOPC was higher than the facility-wide background concentration, and the lowest risk-based screening value (Resident Receptor Adult FWCUG, MCL, or RSL), the CMCOPC was further evaluated using the AT123D model to predict future maximum concentrations in groundwater beneath the source, as well as at downgradient receptor locations, if applicable.

If a predicted maximum leachate concentration was lower than the screening criteria, the chemical was no longer considered a CMCOPC.

 For chemicals identified as CMCOPCs, maximum concentrations predicted by AT123D in groundwater directly below the source areas and at the downgradient receptor locations were compared to the applicable RVAAP facility-wide background concentrations, as well as RVAAP FWCUGs for the Resident Receptor Adult, MCLs, and RSLs. If the predicted maximum concentration of a CMCOPC was higher than its facility-wide background concentration, and the lowest risk-based screening value (Resident Receptor Adult FWCUG, MCL, or RSL), the chemical was retained as a CMCOC. If the predicted maximum concentration of a CMCOPC in groundwater directly below the source areas and at the downgradient receptor location was lower than the screening criteria, the chemical was not considered a CMCOC.

CMCOCs identified by modeling results were evaluated with respect to WOE for retaining or eliminating CMCOCs from further consideration as a basis for potential soil or sediment remedial actions. Lines of evidence included validating modeling results using available AOC-specific groundwater monitoring data. Modeled timelines for potential leaching and lateral transport were evaluated with respect to estimated times for contaminant releases during RVAAP operations to determine if peak leaching concentrations would likely have occurred in the past. Some CMCOCs present at or below RVAAP soil or sediment background concentrations may have predicted leachate or groundwater concentrations exceeding risk-based criteria due to conservative model assumptions; therefore, these were also identified and considered in the evaluation. Additionally, identified CMCOCs were compared to COCs identified in the HHRA to determine if they had an associated risk related to direct exposure to soil or sediment or if CMCOCs and COCs were co-located and may be addressed simultaneously under a potential remedial action.

6.5.2 Model Applications

The SESOIL model (GSC 1998) used for leachate modeling, when applicable, estimates pollutant concentrations in the soil profile following introduction via direct application and/or interaction with transport media. AT123D (DOE 1992) is an analytical groundwater pollutant fate and transport model. It computes the spatial-temporal concentration distribution of waste in the aquifer system and predicts the transient spread of a contaminant plume through a groundwater aquifer. The application of both of these models is discussed in the following subsections.

6.5.2.1 SESOIL Modeling

The SESOIL model defines the soil column as compartments extending from the ground surface through the unsaturated zone and to the upper level of the saturated soil zone or top of bedrock. Processes simulated in SESOIL are categorized in three cycles: hydrologic, sedimentation, and pollutant. Each cycle is a separate submodule in the SESOIL code. The hydrologic cycle includes rainfall, surface runoff, percolation, soil-water content, evapotranspiration, and groundwater recharge. The sediment washload cycle includes erosion and sediment transport.

The pollutant cycle includes convective transport, volatilization, adsorption/desorption, and degradation/decay. A chemical in SESOIL can partition in up to four phases (i.e., liquid, adsorbed,

air, and pure). Data requirements for SESOIL are not extensive and utilize a minimum of AOC-specific soil and chemical parameters and monthly or seasonal meteorological values as input.

The SESOIL model output includes pollutant concentrations at various soil depths and pollutant loss from the unsaturated soil zone in terms of surface runoff, percolation to groundwater, volatilization, and degradation. The mathematical representations in SESOIL generally consider the rate at which the modeled processes occur, the interaction of different processes with each other, and the initial conditions of the waste area and surrounding subsurface matrix material.

The input data for SESOIL can be grouped into four types: climatic, chemical, soil, and application data. There are 61 separate parameters contained in these four data groups. Wherever possible, AOC-specific parameter values were used for modeling. However, certain parameters were not available for the source areas and were estimated based on pertinent scientific literature, geochemical investigations, and checks for consistency between model results and historical data. Conservative estimates were used when a range of values existed or parameter values were not available.

6.5.2.2 Climate Data

The climatic data file of SESOIL consists of an array of mean monthly temperature, mean monthly cloud cover fraction, average monthly relative humidity, average monthly reflectivity of the earth's surface (i.e., shortwave albedo), average daily evapotranspiration, monthly precipitation, mean number of storm events per month, mean duration of rainfall, and mean length of rainy season. The climatic data are presented in Appendix E, Table E-9. The data set was taken from the Youngstown National Weather Service Office weather station at the Youngstown-Warren Regional Airport in Vienna, Ohio, as it was determined to be most appropriate in corresponding to the latitude and longitude at Camp Ravenna.

Climate data from the Youngstown weather station did not have all of the necessary climatic parameters for the HELP model simulation. Accordingly, the water balance evaluation presented in Section 6.2.4 was based on the nearest available weather station data with all necessary coefficients stored within the HELP model (Cleveland, Ohio). Inputs for the SESOIL model (Youngstown station) and HELP model (Cleveland station) produced virtually the same recharge rate (i.e., 9.40 cm/yr for Cleveland and 9.42 cm/yr for Youngstown) for each location. Therefore, the use of the two different weather station data sets did not impact modeling results.

6.5.2.3 Chemical Data

The pollutant fate cycle of SESOIL focuses on various chemical transport and transformation processes that may occur in the soil zone. These processes include volatilization/diffusion, adsorption/desorption, cation exchange, biodegradation and hydrolysis, and metal complexation. The chemical-specific parameters used for SESOIL are presented in Appendix E, Table E-10. The distribution coefficients (K_d s) for inorganic chemicals and the K_{oc} values for organic chemicals were obtained from the chemical-specific parameter table associated with the USEPA Risk Based Generic Screening Tables (USEPA 2015). The K_d s for organic chemicals were estimated from organic,

carbon-based K_{oc} using the relationship $K_d = (f_{oc})(K_{oc})$, where $f_{oc} =$ mass fraction of the organic carbon soil content obtained from AOC-specific measurements. In general, biodegradation rates are not applicable for inorganic CMCOPCs and biodegradation was not considered for the organic chemicals in this evaluation.

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6.5.2.4 Soil Data

The soil data file of SESOIL contains input parameters describing the physical characteristics of the subsurface soil and is presented in Table 6-3. These parameters include soil bulk density, intrinsic permeability, soil disconnectedness index, soil porosity, organic carbon content, and cation exchange capacity. AOC-specific data were used from geotechnical samples collected at the AOC during the PBA08 RI (Table 5-4). There is, however, no measurement method for the soil disconnectedness index or a measured value of the Freundlich exponent. The soil disconnectedness index is a parameter that relates soil permeability to the moisture content, and the Freundlich exponent relates to concentration of a solute on the surface of an adsorbent to the concentration of the solute in the liquid with which it is in contact. Thus, SESOIL default values were used for these two parameters.

An average intrinsic permeability for the vadose zone, representing the unconsolidated zone above the water table, was calibrated using the percolation rate of 9.42 cm/yr (3.6 inches/year) as the calibration target. The model was calibrated against the percolation rate by varying the intrinsic permeability and keeping all other AOC-specific geotechnical parameters fixed. The final hydrogeologic parameter values used in this modeling are shown in Table 6-3. The soil porosity was set to the AOC-specific value. The intrinsic permeability, calibrated in SESOIL to the percolation rate (determined from a water balance estimated in HELP), was found to match the AOC-specific measurements from geotechnical samples.

The soil disconnectedness index replaces the moisture retention curves (or characteristic curves) used by other unsaturated zone leaching models. SESOIL's User Guide defines this parameter to be the exponent relating the "wetting" and "drying" time-dependent permeability of soil to its saturated permeability (Hetrick and Scott 1993). This "one variable" approach of using the soil disconnectedness index in SESOIL simplifies the data estimation process and reduces computational time. In addition, this parameter was calibrated for four different soil types ranging from sandy loam to clay (Hetrick et al. 1986), and calibrated values fell within the default range specified in the SESOIL's User Guide.

6.5.2.5 Source Terms

Analytical data from surface and subsurface soil collected at the AOC were used as source terms for SESOIL modeling. All the samples collected at different depth intervals were compiled to provide a detailed loading option for the SESOIL model. The maximum soil concentrations for each CMCOPC, listed in Table 6-1, were used as source term concentrations for the SESOIL model.

6.5.2.6 Application Data

Five different layering schemes were developed for sample locations within the FPA, and five different layering schemes were developed for sample locations within the NPA due to varying thicknesses of the loading and leaching zones that are based on varying soil sample and groundwater depths throughout the AOC. Details of the model layers utilized in this modeling are presented in Appendix E, Table E-11.

Each model in the FPA and NPA consisted of four layers of varying thicknesses. Depending on the analyte being modeled, the top three layers varied between contaminant loading layers and leaching layers. The top layer (Layer 1) for each model consisted of the loading zone and varied in thickness from 0.25–3 ft thick. The thicknesses and purpose of Layers 2 and 3 varied between the models for each chemical. Layer 4 served as a leaching zone in each model. Layer 4 was a thin layer (0.25 ft thick) included just above the water table to read output results at the water table/vadose zone interface (i.e., leachate concentration entering groundwater).

6.5.3 SESOIL Modeling Results

SESOIL modeling was performed for initial CMCOPCs [arsenic, barium, cobalt, zinc, PETN, benzo(b)fluoranthene, and naphthalene from the FPA; arsenic, cobalt, manganese, nickel, PETN, and naphthalene from the NPA] that have the potential to reach the water table within 1,000 years based on the soil screening analysis results (Table 6-1). Table 6-4 presents the predicted peak leachate concentrations beneath the source areas relative to the discrete sample locations corresponding to the time of peak leachate concentrations. The Resident Receptor Adult FWCUGs; RVAAP facility-wide background concentrations; and MCL/RSL values for the CMCOPCs, if available, are also shown in this table for comparison purposes. PETN at the NPA was excluded as a final CMCOPC based on the results of the SESOIL modeling, as the predicted leachate concentration below the source and just above the water table did not exceed its screening criteria. All of the initial CMCOPCs, except PETN at the NPA, modeled in SESOIL were selected as final CMCOPCs based on the results of the SESOIL modeling, as their predicted leachate concentrations below the source and just above the water table exceeded their respective screening criteria. These leachate concentrations do not reflect the groundwater concentrations beneath the source. When the leachate reaches the water table, dilution by groundwater further reduces leachate concentrations. In Appendix E, Figures E-2 through E-13 show the leachate mass flux versus time plots generated by SESOIL that were used as input to the AT123D models.

6.5.4 AT123D Modeling in the Saturated Zone

The fate and transport processes accounted for in the AT123D model include advection, dispersion, adsorption/retardation, and decay. This model can be used as a tool for estimating the dissolved concentration of a chemical in three dimensions in the groundwater resulting from a mass release over a source area (i.e., point, line, area, or volume source). The model can handle instantaneous, as well as continuous, source loadings of CMCOPC concentrations. AT123D is frequently used by the scientific and technical community to perform quick, conservative estimates of groundwater plume

movements in space and time. SESOIL and AT123D are linked in a software package (RISKPRO) so that mass loading to the groundwater predicted by SESOIL can be transferred directly to AT123D.

Therefore, AT123D was chosen to predict the maximum concentration of contaminants in groundwater after mixing with the leachate and the future concentrations for the contaminants in groundwater at the receptor locations.

The hydrogeologic parameter values used in this modeling are shown in Table 6-3. Most of the parameters presented in this table are AOC-specific values, unless otherwise indicated. The chemical-specific parameters used for AT123D are presented in Appendix E, Table E-12. A discussion of model assumptions and limitations is presented in Section 6.5.6.

Figures E-14 through E-31 show the predicted concentration versus time curves based on AT123D modeling at the source as well as at the receptor location. The screening criteria (MCL/RSL and Resident Receptor Adult FWCUG) are also plotted in these figures as a constant line.

6.5.5 AT123D Modeling Results

Results of AT123D modeling for final soil and sediment CMCOPCs are shown in Table 6-5. The results show predicted groundwater concentrations for CMCOPCs beneath the source area and at the selected downgradient receptor location (i.e., Sand Creek). Using the closest available downgradient monitoring wells for comparison with modeled source areas, observed groundwater concentrations from AOC monitoring wells are included in Table 6-5. The observed groundwater concentrations were added for comparison, not for screening criteria. The distances to the downgradient receptors were based on the distance along the groundwater flow direction to the closest surface water body.

For final soil CMCOPCs, the maximum predicted concentrations of all of the CMCOPCs, except barium at the FPA [i.e., arsenic, cobalt, zinc, PETN, benzo(b)fluoranthene, and naphthalene] and nickel at the NPA (i.e., arsenic, cobalt, manganese, and naphthalene), were predicted to exceed the screening criteria in groundwater beneath the source area and were, therefore, modeled to the downgradient receptor locations (i.e., Sand Creek). The identified sediment CMCOPCs [i.e., benz(a)anthracene and naphthalene at the East Ditch aggregate and antimony, benz(a)anthracene, and benzo(b)fluoranthene at the West Ditch aggregate] were also predicted by analytical solutions to exceed screening criteria in groundwater beneath the source and were retained for lateral transport modeling using AT123D.

Lateral transport modeling showed that none of the soil or sediment CMCOPCs were predicted to exceed the screening criteria at the downgradient receptor location (i.e., Sand Creek). Figure 6-5 presents CMCOCs identified based on AT123D modeling.

6.5.6 Limitations/Assumptions

In general, a conservative modeling approach was used, which may overestimate the contaminant concentration in the leachate for migration from observed soil concentrations. Listed below are important assumptions used in this analysis:

- The contaminant fate and transport evaluation included not only chemicals identified as being
 previously used during historical operations, but also included all chemicals identified as soil
 SRCs during the data screening and reduction process.
- Some soil SRCs were identified due to the absence of a background concentration or as having limited or slight exceedances of the established background concentrations.
- Chemical and biological degradation rates for organic CMCOPCs were not considered in the SESOIL and AT123D models.
- Using K_d and R to describe the reaction term of the transport equation assumes that an equilibrium relationship exists between the solid-phase and solution-phase concentrations and that the relationship is linear and reversible.
- Since AOC-specific data are not available, the K_d and K_{oc} values used in this analysis for all CMCOPCs represent literature or calculated values and may not represent conditions at the AOC.
- The K_d for inorganic chemicals used in the modeling evaluation assumed a pH of 6.8 [i.e., the middle value in the USEPA's evaluation presented in the soil screening guidance document (USEPA 1996b)]. The K_d for inorganic chemicals varies with pH (i.e., generally decreasing with decreasing pH, although there are few exceptions); therefore, if AOC-specific pH measurements are greater or less than 6.8, the K_d and calculated screening parameters (such as R) will deviate from those presented here.
- Flow and transport in the vadose zone is one-dimensional (i.e., only in the vertical direction).
- This modeling used the current soil concentrations that were collected approximately 65 years after historical operations were terminated at the AOC. Therefore, it does not account for constituents that have already leached to groundwater.
- Flow and transport are not affected by density variations. Variability in the distribution of soil contamination and overall impacts to predicted groundwater concentrations were not considered in the models. A realistic distribution of soil contamination was not considered. The maximum concentration value was used as the source term concentration for SESOIL model layers; this is a highly conservative assumption that is expected to produce higher leachate concentrations for the CMCOPCs than the average condition. The horizontal distribution of soil contamination was assumed based on concentration levels from nearby sample locations as opposed to taking into account the entire area.
- The water balance represents an overall average rainwater recharge and assumes an even distribution of infiltration in the modeled area. An average water balance assumes some areas will have higher or lower recharge based on the heterogeneity of the soil and varying topography.

The inherent uncertainties associated with using these assumptions must be recognized. K_d values are highly sensitive to changes in the major chemistry of the solution phase. Therefore, it is important

that the values be measured or estimated under conditions that will closely represent those of the contaminant plume. Deviations of actual AOC-specific parameter values from assumed literature values may significantly affect contaminant fate predictions. It is also important to note that the contaminant plume will change over time and will be affected by multiple solutes present at the AOC. The effects of heterogeneity and anisotropy are not addressed in these simulations. The discrepancy between the contaminant concentrations measured in the field and the values predicted by the model could be investigated by performing sensitivity analyses on the model input parameters that have the most influence on the model predictions.

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These parameters are: (1) biodegradation rate constants for organic chemicals; (2) saturated hydraulic conductivity; (3) soil porosity; (4) f_{oc} for organic chemicals; (5) K_d for inorganic chemicals; and (6) longitudinal, transverse, and vertical dispersivity values. Generally, higher biodegradation rates will produce lower concentrations, and lower rates will produce higher concentrations for organic chemicals without impacting the results of the inorganic chemicals. In this analysis, biodegradation rates for organic chemicals have been assumed to be zero, thereby predicting higher concentrations for organic chemicals in groundwater than would be expected to be observed in the future. Higher hydraulic conductivity and dispersivity causes higher advection and dispersion, thereby producing lower peaks near the source area, but increasing the migration distance. The reverse will be true with lower hydraulic conductivity and dispersivity values. Higher f_{oc} values have a similar effect on organic chemicals as higher K_d has on inorganic chemicals: they decrease the mobility of the chemicals as well as produce lower concentrations in groundwater.

6.6 EVALUATION TO IDENTIFY CMCOCS

This evaluation of contaminant fate and transport uses a soil screening analysis to identify SRCs that have potential to leach to groundwater, performs SESOIL modeling to conservatively estimate final CMCOPC leachate concentrations before the SRCs enter the groundwater system beneath the sources with highest level of contamination, and uses AT123D modeling to present a conservative maximum concentration in groundwater of final CMCOPCs beneath the sources and at downgradient receptor locations.

This analysis also includes a sediment screening analysis that was performed for sediment samples at the AOC. Chemical-specific DAFs were calculated using co-located surface water and sediment concentrations for identified sediment SRCs. These DAFs were used in the sediment screening analysis to identify CMCOPCs for predictive modeling and further evaluation.

The limitations and assumptions of the overall process are presented in Section 6.5.6. The text below provides a list of the remaining final CMCOPCs in soil and sediment and a qualitative assessment of the results and considerations of the limitations and assumptions.

6.6.1 Evaluation of Remaining Soil CMCOPCs

6.6.1.1 Former Production Area

Arsenic – The maximum soil concentration for arsenic at the FPA was 30.2 mg/kg at LL11ss-022-0001-SO. Only 8 of 67 soil samples within the FPA had concentrations of arsenic above the subsurface soil background concentration of 19.8 mg/kg. Arsenic was not identified as a soil COC in the HHRA. The modeling estimates that arsenic concentrations in groundwater beneath the source areas in the FPA could potentially exceed its MCL at about 35 years with peak concentrations occurring at approximately 300 years or less. Arsenic is not predicted to migrate to the downgradient receptor location at a detectable concentration within 1,000 years (see Table 6-5). Arsenic concentrations in AOC groundwater samples collected from 2009-2013 were below its MCL in all the wells except LL11mw-007. The observed arsenic concentration in this well exceeded the MCL in all the events during this period with the maximum observed in the most recent event (i.e., 2013) at a concentration of 0.022 mg/L (Table 6-5). However, it should be noted here that the background concentration of arsenic in unconsolidated groundwater also exceeds the MCL and is very close to the observed concentration in LL11mw-007, indicating that the observed concentration of arsenic at LL11mw-007 is background related, and not due to contamination from the site.

Barium – The maximum soil concentration for barium at the FPA (i.e., 127 mg/kg at LL11cs-012-0001-SO) was well below the soil industrial RSL (22,000 mg/kg) and the Resident Receptor Adult FWCUG at a HQ of 0.1 (8,966 mg/kg), and it was not identified as a soil COC in the HHRA. This maximum concentration was only slightly above the subsurface soil background concentration of 124 mg/kg. Barium was not identified as a potential contaminant from historical usage at Load Line 11 and did not have widespread contamination. The modeling estimates that barium concentrations in groundwater beneath the source areas could potentially exceed its RSL at about 100 years or less with peak concentrations occurring at approximately 200 years, and barium is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (see Table 6-5). Barium concentrations in AOC groundwater samples collected from 2009-2013 were below its MCL (Table 6-5).

Cobalt – The maximum soil concentration for cobalt at the FPA (i.e., 33.8 mg/kg at LL11sd-025-0001-SD) was below the soil industrial RSL (35 mg/kg) and the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (803 mg/kg), and it was not identified as a soil COC in the HHRA. This maximum concentration was only slightly above the subsurface soil background concentration of 23.2 mg/kg. Cobalt was not identified as a potential contaminant from historical usage at Load Line 11 and did not have widespread contamination. The modeling estimates that cobalt concentrations in groundwater beneath the source areas could potentially exceed its RSL at about 150 years or less with peak concentrations occurring at approximately 400 years, and cobalt is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (see Table 6-5). Cobalt concentrations in AOC groundwater samples collected from 2010-2013 were below the RSL (Table 6-5).

 Zinc – The maximum soil concentration for zinc (i.e., 465 mg/kg at LL11cs-025-0001-SD) was below the soil residential RSL (2300 mg/kg), and it was not identified as a soil COPC in the HHRA. Zinc was not identified as a potential contaminant from historical usage at Load Line 11 and did not have widespread contamination. The modeling estimates that zinc concentrations in groundwater beneath the source areas could potentially exceed its RSL at about 250 years or less with peak concentrations occurring at approximately 350 years, and zinc is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (see Table 6-5). Zinc concentrations in AOC groundwater samples collected from 2010-2013 were well below the RSL (Table 6-5).

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PETN – The single detection for PETN at the FPA (i.e., 0.049 mg/kg at LL11sb-061-5555-SO, with frequency of detection 1/16) was well below the soil residential RSL (13 mg/kg), and it was not considered a soil COPC in the HHRA. The modeling estimates that PETN concentrations in groundwater beneath the source areas could potentially exceed its RSL at about 5 years or less with peak concentrations occurring at approximately 10 years or less, and PETN is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (Table 6-5). Based on the AOC period of operations, PETN should have already been detected in groundwater exceeding its RSL. However, PETN has not been detected in the AOC groundwater samples collected from 2010-2013 (Table 6-5). Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and PETN would be expected to be below its RSL based on its site-specific attenuation and biodegradation rates.

Benzo(b)fluoranthene – The maximum soil concentration for benzo(b)fluoranthene at the FPA (i.e., 0.56 mg/kg at LL11sb-060-5551-SO) was slightly above the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (0.221 mg/kg); it was not considered a COC in the HHRA. The modeling estimates that benzo(b)fluoranthene concentrations in groundwater beneath the source areas could potentially exceed its RSL at about 700 years or less with peak concentrations occurring in more than 1,000 years. Benzo(b)fluoranthene is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (Table 6-5). Benzo(b)fluoranthene has not been detected in the AOC groundwater samples collected from 2010-2013 mainly because benzo(b)fluoranthene is expected to degrade before reaching the water table. Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and benzo(b)fluoranthene would be expected to be below its RSL based on its site-specific attenuation and biodegradation rates.

Naphthalene – The maximum soil concentration for naphthalene at the FPA (i.e., 0.017 mg/kg at LL11sb-061-5555-SO) was below the soil residential RSL (3.8 mg/kg) and the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (368 mg/kg), and it was not considered a COPC in the HHRA. The modeling estimates that naphthalene concentrations in groundwater beneath the source areas could potentially exceed its RSL in about 3 years or less with peak concentrations occurring in about 12 years or less. Naphthalene is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (Table 6-5). Naphthalene was detected in the AOC groundwater samples collected from 2010-2013 at concentrations below the RSL (Table 6-5). Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and

naphthalene would be expected to be below its RSL based on its site-specific attenuation and biodegradation rates.

6.6.1.2 Non-Production Area

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Arsenic – The maximum soil concentration for arsenic at the NPA was 44.1 mg/kg at LL11cs-045-0001-SO. Only 8 of 47 soil samples within the NPA had concentrations of arsenic above the subsurface soil background concentration of 19.8 mg/kg. Arsenic was not identified as a soil COC in the HHRA. The modeling estimates that arsenic concentrations in groundwater beneath the source areas in the NPA could potentially exceed its MCL within about 40 years or less with peak concentrations occurring at approximately 150 years. Arsenic is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (see Table 6-5). Arsenic concentrations in AOC groundwater samples collected from 2009-2013 were below its MCL in all wells except in LL11mw-007. The observed arsenic concentration in this well exceeded the MCL in all events with the maximum observed in 2013 at a concentration of 0.022 mg/L (Table 6-5). However, it should be noted here that background concentration of arsenic in unconsolidated groundwater also exceeds the MCL and is very close to the observed concentration in LL11mw-007, indicating that the observed concentration of arsenic at LL11mw-007 is background related, and not due to contamination from the site.

Cobalt – The maximum soil concentration for cobalt at the NPA (i.e., 24.7 mg/kg at LL11cs-036-0001-SO) was below the soil industrial RSL (35 mg/kg) and the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (803 mg/kg), and it was not identified as a soil COC in the HHRA. This maximum concentration was barely above the subsurface soil background concentration of 23.2 mg/kg. Cobalt was not identified as a potential contaminant from historical usage at Load Line 11 and did not have widespread contamination. The modeling estimates that cobalt concentrations in groundwater beneath the source areas could potentially exceed its RSL in 24 years or less (and the Resident Receptor Adult FWCUG within 100 years) with peak concentrations occurring at approximately 250 years or less, and cobalt is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (see Table 6-5). Cobalt concentrations in AOC groundwater samples collected from 2010-2013 were below the RSL (Table 6-5). Based on the AOC period of operations, this indicates that the model-predicted concentrations are conservative.

Manganese – The maximum soil concentration for manganese at the NPA (i.e., 1,930 mg/kg at LL11cs-036-0001-SO) was below the soil industrial RSL (2,600 mg/kg), and it was not identified as a soil COC in the HHRA. This maximum concentration was below the subsurface soil background concentration of 3,030 mg/kg. Manganese was not identified as a potential contaminant from historical usage at Load Line 11 and did not have widespread contamination. The modeling estimates that manganese concentrations in groundwater beneath the source areas could potentially exceed its RSL in 150 years or less (and the unconsolidated background concentration within 200 years) with peak concentrations occurring at approximately 850 years, and manganese is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (see Table 6-5). Manganese concentrations in AOC groundwater samples collected from 2009-2013 were above its RSL and FWCUG in several wells; however, they were below its background concentration,

indicating that the observed groundwater concentrations of manganese are background related, and not due to contamination from the site.

Naphthalene – The maximum soil concentration for naphthalene at the NPA (i.e., 0.01 mg/kg at LL11sb-063-5563-SO) was below the soil residential RSL (3.8 mg/kg) and the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (368 mg/kg), and it was not considered a COPC in the HHRA. Naphthalene was infrequently detected at the NPA (2/50, less than 5% detections). The modeling estimates that naphthalene concentrations in groundwater beneath the source areas could potentially exceed its RSL in less than 15 years with peak concentrations occurring in about 24 years or less. Naphthalene is not predicted to migrate to the downgradient receptor location at detectable concentrations within 1,000 years (Table 6-5). Naphthalene was detected in the AOC groundwater samples collected from 2010-2013 at concentrations below the RSL (Table 6-5). Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and naphthalene would be expected to be below its RSL based on its site-specific attenuation and biodegradation rates.

6.6.2 Evaluation of Remaining Sediment CMCOPCs

6.6.2.1 East Ditch Aggregate

Benz(a)anthracene – There was only one sediment sample (LL11sd-083) for organic SRCs at the East Ditch, and the detected concentration of benz(a)anthracene (0.11 mg/kg) was below the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (0.221 mg/kg). Benz(a)anthracene was not identified as a sediment COPC in the HHRA for this area and was not detected above the SL in the AOC groundwater samples collected from 2010–2013 (Table 6-5). Benz(a)anthracene modeling results for sediment in the East Ditch indicate concentrations decreasing below the SL after approximately 10 years and continuing to decrease into the future, and it is not predicted to migrate to the downgradient receptor location at detectable concentrations (see Table 6-5). It should be noted here that the reason benz(a)anthracene is predicted to be in groundwater is because the modeling assumed that the sediment sample is directly in contact with the underlying groundwater and that no biodegradation is occurring. Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and benz(a)anthracene would be expected to be below its SL based on its estimated site-specific biodegradation rate and accounting for the vertical leaching distance.

Naphthalene – There was only one sediment sample (LL11sd-083) for organic SRCs at the East Ditch, and although there is no sediment-specific Resident Receptor Adult FWCUG for naphthalene, the sediment concentration of naphthalene (i.e., 0.01 mg/kg) was well below the soil Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (368 mg/kg). Naphthalene was not identified as a sediment COPC in the HHRA for this area and was not detected above the RSL in the AOC groundwater samples collected from 2010–2013 (Table 6-5). Naphthalene modeling results for sediment indicate concentrations decreasing below the RSL after approximately five years and continuing to decrease into the future, and it is not predicted to migrate to the downgradient receptor location at detectable concentrations (see Table 6-5). It should be noted here that the reason naphthalene is predicted to be in groundwater is because the modeling assumed that the sediment sample is directly in contact with the underlying groundwater and that no biodegradation is occurring.

Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and naphthalene would be expected to be below its SL based on its estimated site-specific biodegradation rate and accounting for the vertical leaching distance.

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6.6.2.2 West Ditch Aggregate

Antimony – There were only three sediment samples for antimony at the West Ditch, and the maximum concentration (i.e., 0.9 mg/kg at LL11sd-024) was below the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (13.6 mg/kg). Antimony was not identified as a sediment COPC in the HHRA for this area and was not detected above the SL in the AOC groundwater samples collected from 2010–2013 (Table 6-5). Antimony modeling results for sediment in the West Ditch indicate concentrations decreasing below the SL after approximately one year and continuing to decrease into the future, and it is not predicted to migrate to the downgradient receptor location at detectable concentrations (see Table 6-5). It should be noted here that the reason antimony is predicted to be in groundwater is because the modeling assumed that the sediment sample is directly in contact with the underlying groundwater. Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and antimony would be expected to be below its SL based on accounting for the vertical leaching distance.

Benz(a)anthracene – There were only two sediment samples for benz(a)anthracene at the West Ditch, and the maximum concentration (i.e., 0.077 mg/kg at LL11sd-096) was below the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (0.221 mg/kg). Benz(a)anthracene was not identified as a sediment COPC in the HHRA for this area and was not detected in the AOC groundwater samples collected from 2010–2013 (Table 6-5). Benz(a)anthracene modeling results for sediment in the West Ditch indicate concentrations decreasing below the SL after approximately five years and continuing to decrease into the future, and it is not predicted to migrate to the downgradient receptor location at detectable concentrations (see Table 6-5). It should be noted here that the reason benz(a)anthracene is predicted to be in groundwater is because the modeling assumed that the sediment sample is directly in contact with the underlying groundwater and that no biodegradation is occurring. Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and benz(a)anthracene would be expected to be below its SL based on its estimated site-specific biodegradation rate and accounting for the vertical leaching distance.

Benzo(b)fluoranthene – There were only two sediment samples for benzo(b)fluoranthene at the West Ditch, and the maximum concentration (i.e., 0.14 mg/kg at LL11sd-096) was below the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (0.221 mg/kg). Benzo(b)fluoranthene was not identified as a sediment COPC in the HHRA for this area and was not detected in the AOC groundwater samples collected from 2010–2013 (Table 6-5). Benzo(b)fluoranthene modeling results for sediment in the West Ditch indicate concentrations decreasing below the SL after approximately 50 years and continuing to decrease into the future, and it is not predicted to migrate to the downgradient receptor location at detectable concentrations (see Table 6-5). It should be noted here that the reason benzo(b)fluoranthene is predicted to be in groundwater is because the modeling assumed that the sediment sample is directly in contact with the underlying groundwater and that no biodegradation is occurring. Therefore, this evaluation concludes that the model-predicted

concentrations are conservative, and benzo(b)fluoranthene would be expected to be below its SL based on its estimated site-specific biodegradation rate and accounting for the vertical leaching distance.

Dibenz(a,h)anthracene – There were only two sediment samples for dibenz(a,h)anthracene at the West Ditch, and the maximum concentration (i.e., 0.016 mg/kg at LL11sd-096) was below the Resident Receptor Adult FWCUG at a TR of 1E-06, HQ of 0.1 (0.022 mg/kg). Dibenz(a,h)anthracene was not identified as a sediment COPC in the HHRA for this area and was not detected in the AOC groundwater samples collected from 2010–2013 (Table 6-5). Dibenz(a,h)anthracene modeling results for sediment in the West Ditch indicate concentrations decreasing below the SL in less than four years and continuing to decrease into the future, and it is not predicted to migrate to the downgradient receptor location at detectable concentrations (see Table 6-5). It should be noted here that the reason dibenz(a,h)anthracene is predicted to be in groundwater is because the modeling assumed that the sediment sample is directly in contact with the underlying groundwater and that no biodegradation is occurring. Therefore, this evaluation concludes that the model-predicted concentrations are conservative, and dibenz(a,h)anthracene would be expected to be below its SL based on its estimated site-specific biodegradation rate and accounting for the vertical leaching distance.

This qualitative assessment concludes that the soil and sediment contaminants identified as final CMCOPCs for evaluation, due to predicted groundwater concentrations beneath a source, are not adversely impacting groundwater quality based on current data and are not predicted to have future impacts. Potential additional investigation under the Facility-wide Groundwater AOC may be warranted, but based on the fate and transport evaluation, CMCOCs were not identified for Load Line 11, and no further action is required of soil and sediment to be protective of groundwater.

6.7 SUMMARY AND CONCLUSIONS

Inorganic and organic SRCs exist in surface soil, subsurface soil, and sediment at Load Line 11. These SRCs include chemicals that were identified as potential contaminants from previous site usage and chemicals that were identified from the SRC screening process using available data. All SRCs were further evaluated to determine if residual concentrations in soil and sediment may potentially impact groundwater quality and warrant evaluation in an FS.

All SRCs identified in the surface and subsurface soil and sediment at Load Line 11 were evaluated through the stepwise fate and transport evaluation. Evaluation of modeling results with respect to current AOC groundwater data and model limitations identified the following CMCOPCs:

- Arsenic, barium, cobalt, zinc, PETN, benzo(b)fluoranthene, and naphthalene at the FPA and
 arsenic, cobalt, manganese, and naphthalene at the NPA among the soil CMCOPCs were
 predicted to exceed the screening criteria in groundwater beneath the source area; however,
 none of these CMCOPCs were predicted to be above their respective groundwater criteria at
 the downgradient receptor location.
- Among the sediment CMCOPCs, benz(a)anthracene and naphthalene at the East Ditch aggregate and antimony, benz(a)anthracene, benzo(b)fluoranthene, and dibenz(a,h)anthracene

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at the West Ditch aggregate were predicted to exceed the screening criteria in groundwater beneath the source area; however, none of these CMCOPCs were predicted to be above criteria at the downgradient receptor location.

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A qualitative assessment of the sample results and considerations of the limitations and assumptions of the models were performed to identify if any CMCOCs are present in soil and sediment at Load Line 11 that may impact the groundwater beneath the source or at the downstream receptor location. This qualitative assessment concluded that there were no CMCOCs present in soil and sediment that may impact the groundwater beneath the source or at the downstream receptor location. No further action is required of soil and sediment at Load Line 11 to be protective of groundwater.

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Table 6-1. Initial CMCOPCs Evaluated with SESOIL Modeling

| SRCs | Maximum Concentrations (mg/kg) | Discrete Sample Location | Sample Depth (ft bgs) | Leachate Modeling Required? (Yes/No) | | | | | | | | | |
|---|--------------------------------------|-----------------------------|-----------------------------|---|--|--|--|--|--|--|--|--|--|
| | Former Production Area | | | | | | | | | | | | |
| Inorganic Chemicals | | | | | | | | | | | | | |
| Arsenic 30.2 LL11ss-022-0001-SO 0-1 Yes | | | | | | | | | | | | | |
| Barium | 127 | LL11cs-012-0001-SO | 1–2 | Yes | | | | | | | | | |
| Cobalt | 33.8 | LL11sd-025-0001-SD | 0–1 | Yes | | | | | | | | | |
| Zinc | 465 | LL11cs-025-0001-SD | 0–1 | Yes | | | | | | | | | |
| | Semi-volat | tile organic compounds | | | | | | | | | | | |
| Benzo(b)fluoranthene | 0.56 | LL11sb-060-5551-SO | 0–1 | Yes | | | | | | | | | |
| Naphthalene | 0.017 | LL11sb-061-5555-SO | 0–1 | Yes | | | | | | | | | |
| | | Explosives | | | | | | | | | | | |
| PETN | 0.049 | LL11sb-061-5555-SO | 0–1 | Yes | | | | | | | | | |
| | Non- | Production Area | | | | | | | | | | | |
| | Inor | ganic Chemicals | | | | | | | | | | | |
| Arsenic | 44.1 | LL11cs-045-0001-SO | 7–8 | Yes | | | | | | | | | |
| Cobalt | 24.7 | LL11cs-036-0001-SO | 1–2 | Yes | | | | | | | | | |
| Manganese | 1930 | LL11cs-036-0001-SO | 1–2 | Yes | | | | | | | | | |
| Nickel | 33.5 | LL11cs-036-0001-SO | 1–2 | Yes | | | | | | | | | |
| Semi-volatile organic compounds | | | | | | | | | | | | | |
| Naphthalene | 0.01 | LL11sb-063-5563-SO | 0–1 | Yes | | | | | | | | | |
| | | Explosives | | | | | | | | | | | |
| PETN | 0.036 | LL11sb-067-5581-SO | 0–1 | Yes | | | | | | | | | |

bgs = Below ground surface.

CMCOPC = Contaminant migration contaminant of potential concern.

ft = Feet.

mg/kg = Milligrams per kilogram.

PETN = Pentaerythritol tetranitrate.

SESOIL = Seasonal Soil Compartment Model.

SRC = Site-related contaminant.

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Table 6–2. Sediment Screening Results for Load Line 11

| | | | | Table 6–2. Sedime | nt Screening Kes | suits for | Load Line 11 | | | | | | | |
|------------------------|------------|---|--|--------------------|------------------------|-----------|-----------------------|-----------|--|------------------|---|-------------------|----------------|---------------------|
| Analyte | CAS Number | Background Criteria (mg/kg) ^a | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | K _{oc} (L/kg) | Reference | K _d (L/kg) | Reference | Maximum Groundwater Concentration (mg/L) ^b | DAF ^c | Maximum Groundwater Concentration (mg/L)/DAF | MCL or RSL (mg/L) | MCL or RSL? | CMCOPC? (Yes/No) |
| | | | |] | East Ditch Aggreg | gate | | | | | | | | |
| | | | | I | norganic Compou | ınds | | | | | | | | |
| Aluminum | 7429-90-5 | 1.39E+04 | 2.10E+04 | LL11cs-031-0001-FD | NA | NA | 1.50E+03 | d | 1.40E+01 | 15 | 9.40E-01 | 2.00E+01 | RSL | No |
| Antimony | 7440-36-0 | 0.00E+00 | 4.20E-01 | LL11sd-017-0001-FD | NA | NA | 4.50E+01 | d | 9.33E-03 | 15 | 6.26E-04 | 6.00E-03 | MCL | No |
| Arsenic | 7440-38-2 | 1.95E+01 | 2.11E+01 | LL11sd-017-0001-SD | NA | NA | 2.90E+01 | d | 7.28E-01 | 1171 | 6.21E-04 | 1.00E-02 | MCL | No |
| Beryllium | 7440-41-7 | 3.80E-01 | 1.00E+00 | LL11sd-083-5594-SD | NA | NA | 7.90E+02 | d | 1.27E-03 | 15 | 8.50E-05 | 4.00E-03 | MCL | No |
| Cadmium | 7440-43-9 | 0.00E+00 | 4.70E-01 | LL11cs-031-0001-SD | NA | NA | 7.50E+01 | d | 6.27E-03 | 15 | 4.21E-04 | 5.00E-03 | MCL | No |
| Chromium | 7440-47-3 | 1.81E+01 | 2.71E+01 | LL11cs-031-0001-FD | NA | NA | 1.90E+01 | d | 1.43E+00 | 1255 | 1.14E-03 | 1.00E-01 | MCL | No |
| Cobalt | 7440-48-4 | 9.10E+00 | 1.74E+01 | LL11sd-083-5594-SD | NA | NA | 4.50E+01 | d | 3.87E-01 | 2275 | 1.70E-04 | 6.00E-03 | RSL | No |
| Copper | 7440-50-8 | 2.76E+01 | 3.14E+01 | LL11sd-083-5594-SD | NA | NA | 3.50E+01 | d | 8.97E-01 | 641 | 1.40E-03 | 1.30E+00 | MCL | No |
| Nickel | 7440-02-0 | 1.77E+01 | 3.26E+01 | LL11sd-083-5594-SD | NA | NA | 6.50E+01 | d | 5.02E-01 | 551 | 9.10E-04 | 3.90E-01 | RSL | No |
| Selenium | 7782-49-2 | 1.70E+00 | 3.20E+00 | LL11sd-083-5594-SD | NA | NA | 5.00E+00 | d | 6.40E-01 | 2462 | 2.60E-04 | 5.00E-02 | MCL | No |
| | | | | | olatile Organic C | ompoun | | | | | | | | |
| 2-Methylnaphthalene | 91-57-6 | None | 1.30E-02 | LL11sd-083-5594-SD | 2.78E+03 | d | 3.06E+00 | e | 4.25E-03 | 15 | 2.85E-04 | 3.60E-03 | RSL | No |
| Anthracene | 120-12-7 | None | 2.50E-02 | LL11sd-083-5594-SD | 1.64E+04 | d | 1.80E+01 | e | 1.39E-03 | 15 | 9.32E-05 | 1.80E+00 | RSL | No |
| Benz(a)anthracene | 56-55-3 | None | 1.10E-01 | LL11sd-083-5594-SD | 1.77E+05 | d | 1.95E+02 | e | 5.65E-04 | 15 | 3.79E-05 | 1.20E-05 | RSL | Yes |
| Benzo(a)pyrene | 50-32-8 | None | 9.80E-02 | LL11sd-083-5594-SD | 5.87E+05 | d | 6.46E+02 | e | 1.52E-04 | 15 | 1.02E-05 | 2.00E-04 | MCL | No |
| Benzo(b)fluoranthene | 205-99-2 | None | 1.30E-01 | LL11sd-083-5594-SD | 5.99E+05 | d | 6.59E+02 | e | 1.97E-04 | 15 | 1.32E-05 | 3.40E-05 | RSL | No |
| Benzo(ghi)perylene | 191-24-2 | None | 6.40E-02 | LL11sd-083-5594-SD | 1.07E+07 | d | 1.18E+04 | e | 5.44E-06 | 15 | 3.65E-07 | 1.20E-01 | RSL | No |
| Benzo(k)fluoranthene | 207-08-9 | None | 5.60E-02 | LL11sd-083-5594-SD | 5.87E+05 | d | 6.46E+02 | e | 8.67E-05 | 15 | 5.82E-06 | 3.40E-04 | RSL | No |
| Chrysene | 218-01-9 | None | 9.60E-02 | LL11sd-083-5594-SD | 1.81E+05 | d | 1.99E+02 | e | 4.84E-04 | 15 | 3.25E-05 | 3.40E-03 | RSL | No |
| Dibenz(a,h)anthracene | 53-70-3 | None | 1.70E-02 | LL11sd-083-5594-SD | 1.91E+06 | d | 2.10E+03 | e | 8.08E-06 | 15 | 5.42E-07 | 3.40E-06 | RSL | No |
| Fluoranthene | 206-44-0 | None | 2.30E-01 | LL11sd-083-5594-SD | 5.55E+04 | d | 6.10E+01 | e | 3.77E-03 | 15 | 2.53E-04 | 8.00E-01 | RSL | No |
| Fluorene | 86-73-7 | None | 1.20E-02 | LL11sd-083-5594-SD | 9.16E+03 | d | 1.01E+01 | e | 1.19E-03 | 15 | 7.99E-05 | 2.90E-01 | RSL | No |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | None | 5.60E-02 | LL11sd-083-5594-SD | 1.95E+06 | d | 2.15E+03 | e | 2.61E-05 | 15 | 1.75E-06 | 3.40E-05 | RSL | No |
| Naphthalene | 91-20-3 | None | 1.00E-02 | LL11sd-083-5594-SD | 1.54E+03 | d | 1.69E+00 | e | 5.90E-03 | 15 | 3.96E-04 | 1.70E-04 | RSL | Yes |
| Phenanthrene | 85-01-8 | None | 1.00E-01 | LL11sd-083-5594-SD | 1.82E+04 | d | 2.00E+01 | e | 5.00E-03 | 15 | 3.35E-04 | 1.20E-01 | RSL | No |
| Pyrene | 129-00-0 | None | 1.80E-01 | LL11sd-083-5594-SD | 5.43E+04 | d | 5.98E+01 | e | 3.01E-03 | 15 | 2.02E-04 | 1.20E-01 | RSL | No |
| | | | | V | Vest Ditch Aggre | gate | | | | | | | | |
| | | | | | norganic Compou | | | _ | | | | | | |
| Aluminum | 7429-90-5 | 13900.00 | 21400 | LL11cs-015-0001-SD | NA | NA | 1.50E+03 | d | 1.43E+01 | 2 | 7.13E+00 | 2.00E+01 | RSL | No |
| Antimony | 7440-36-0 | 0.00 | 0.9 | LL11sd-024-0001-SD | NA | NA | 4.50E+01 | d | 2.00E-02 | 2 | 1.00E-02 | 6.00E-03 | MCL | Yes |
| Beryllium | 7440-41-7 | 0.38 | 1 | LL11sd-024-0001-SD | NA | NA | 7.90E+02 | d | 1.27E-03 | 7 | 1.80E-04 | 4.00E-03 | MCL | No |
| Cadmium | 7440-43-9 | 0 | 0.51 | LL11sd-096-5874-SD | NA | NA | 7.50E+01 | d | 6.80E-03 | 2 | 3.40E-03 | 5.00E-03 | MCL | No |
| Chromium | 7440-47-3 | 18.10 | 26.2 | LL11sd-030-0001-SD | NA | NA | 1.90E+01 | d | 1.38E+00 | 195 | 7.07E-03 | 1.00E-01 | MCL | No |
| Cobalt | 7440-48-4 | 9.10 | 14.1 | LL11sd-024-0001-SD | NA | NA | 4.50E+01 | d | 3.13E-01 | 83 | 3.77E-03 | 6.00E-03 | RSL | No |
| Mercury | 7439-97-6 | 0.059 | 0.1 | LL11sd-096-5874-SD | NA | NA | 5.20E+01 | d | 1.92E-03 | 2 | 9.62E-04 | 2.00E-03 | MCL | No |
| Nickel | 7440-02-0 | 17.70 | 27.7 | LL11sd-030-0001-SD | NA | NA | 6.50E+01 | d | 4.26E-01 | 62 | 6.93E-03 | 3.90E-01 | RSL | No |
| Silver | 7440-22-4 | 0 | 0.072 | LL11sd-096-5874-SD | NA | NA | 8.30E+00 | d | 8.67E-03 | 2 | 4.34E-03 | 9.40E-02 | RSL | No |
| Vanadium | 7440-62-2 | 26.10 | 35.6 | LL11sd-030-0001-SD | NA | NA | 1.00E+03 | d | 3.56E-02 | 2 | 1.78E-02 | 8.60E-02 | RSL | No |
| Sulfide | 18496-25-8 | None | 43.4 | LL11sd-024-0001-SD | NA | NA | NA | NA | NA | 2 | NA | NA | NA | No |
| | | T | T | | Explosives | 1 . | | | | | | T 222-11 | | |
| Nitrocellulose | 9004-70-0 | None | 1.4 | LL11sd-084-5595-SD | 1.00E+01 | d | 1.10E-02 | e | 1.27E+02 | 2 | 6.36E+01 | 6.00E+04 | RSL | No |
| | 0 | T 27 | | | olatile Organic C | T . | | | | | 4.6= | | — | |
| Acenaphthene | 83-32-9 | None | 0.013 | LL11sd-096-5874-SD | 5.03E+03 | d | 5.53E+00 | e | 2.35E-03 | 2 | 1.18E-03 | 5.30E-01 | RSL | No |
| Anthracene | 120-12-7 | None | 0.02 | LL11sd-096-5874-SD | 1.64E+04 | d | 1.80E+01 | e | 1.11E-03 | 2 | 5.56E-04 | 1.80E+00 | RSL | No |
| Benz(a)anthracene | 56-55-3 | None | 0.077 | LL11sd-096-5874-SD | 1.77E+05 | d | 1.95E+02 | e | 6.68E-05 | 2 | 3.34E-05 | 1.20E-05 | RSL | Yes |
| Benzenemethanol | 100-51-6 | None | 0.054 | LL11sd-096-5874-SD | 2.15E+01 | d | 2.36E-02 | e | 2.29E+00 | 2 | 1.14E+00 | 2.00E+00 | RSL | No |

Table 6–2. Sediment Screening Results for Load Line 11 (continued)

| Analyte | CAS Number | Background Criteria (mg/kg) ^a | Maximum Sediment Concentration (mg/kg) | Sediment Sample ID | K _{oc} (L/kg) | Reference | K _d (L/kg) | Reference | Maximum Groundwater Concentration (mg/L) ^b | DAF ^c | Maximum Groundwater Concentration (mg/L)/DAF | MCL or RSL (mg/L) | MCL or RSL? | CMCOPC? (Yes/No) |
|------------------------|---|---|---|--------------------|------------------------|-----------|-----------------------|-----------|--|------------------|---|-------------------|----------------|---------------------|
| Benzo(a)pyrene | 50-32-8 | None | 0.09 | LL11sd-096-5874-SD | 5.87E+05 | d | 6.46E+02 | e | 3.87E-05 | 2 | 6.97E-05 | 2.00E-04 | MCL | No |
| Benzo(b)fluoranthene | 205-99-2 | None | 0.14 | LL11sd-096-5874-SD | 5.99E+05 | d | 6.59E+02 | e | 1.67E-04 | 2 | 1.06E-04 | 3.40E-05 | RSL | Yes |
| Benzo(ghi)perylene | 191-24-2 | None | 0.068 | LL11sd-096-5874-SD | 1.07E+07 | d | 1.18E+04 | e | 8.33E-06 | 2 | 2.89E-06 | 1.20E-01 | RSL | No |
| Benzo(k)fluoranthene | 207-08-9 | None | 0.066 | LL11sd-096-5874-SD | 5.87E+05 | d | 6.46E+02 | e | 2.01E-04 | 2 | 5.11E-05 | 3.40E-04 | RSL | No |
| Chrysene | 218-01-9 | None | 0.099 | LL11sd-096-5874-SD | 1.81E+05 | d | 1.99E+02 | e | 3.22E-04 | 2 | 2.49E-04 | 3.40E-03 | RSL | No |
| Dibenz(a,h)anthracene | 53-70-3 | None | 0.016 | LL11sd-096-5874-SD | 1.91E+06 | d | 2.10E+03 | e | 7.61E-06 | 2 | 3.80E-06 | 3.40E-06 | RSL | Yes |
| Fluoranthene | 206-44-0 | None | 0.21 | LL11sd-096-5874-SD | 5.55E+04 | d | 6.10E+01 | e | 9.18E-04 | 2 | 1.72E-03 | 8.00E-01 | RSL | No |
| Indeno(1,2,3-cd)pyrene | 193-39-5 | None | 0.061 | LL11sd-096-5874-SD | 1.95E+06 | d | 2.15E+03 | e | 4.47E-05 | 2 | 1.42E-05 | 3.40E-05 | RSL | No |
| | Semi-volatile Organic Compounds (continued) | | | | | | | | | | | | | |
| Phenanthrene | 85-01-8 | None | 0.098 | LL11sd-096-5874-SD | 1.82E+04 | d | 2.00E+01 | e | 8.49E-04 | 2 | 2.45E-03 | 1.20E-01 | RSL | No |
| Pyrene | 129-00-0 | None | 0.16 | LL11sd-096-5874-SD | 5.43E+04 | d | 5.98E+01 | e | 3.85E-03 | 2 | 1.34E-03 | 1.20E-01 | RSL | No |

^a Background criteria for sediment from final facility-wide background values for Ravenna Army Ammunition Plant (RVAAP), published in the Final Phase II Remedial Investigation Report for Winklepeck Burning Grounds at Ravenna Army Ammunition Plant, Ravenna, Ohio (USACE 2001).

CMCOPC = Contaminant migration contaminant of potential concern

DAF = Dilution attenuation factor

 K_d = Distribution coefficient.

 K_{oc} = Organic carbon distribution coefficient.

L/kg = Liters per kilogram

MCL = Maximum contaminant level.

mg/kg = Milligrams per kilogram.

mg/L = Milligrams per liter.

NA = Not applicable.

RSL = Regional screening level.

Bold = Final CMCOPC to be modeled with Analytical Transient 1-, 2-, 3-Dimensional modeling

^b Maximum groundwater concentration = maximum sediment concentration divided by the distribution coefficient.

^c An aggregate-specific DAF was calculated based on the sediment and co-located surface water concentrations. The lowest calculated DAF (15 for aluminum in the East Ditch aggregate, 2 for aluminum in the West Ditch aggregate) was used for analytes that did not have an aggregate-specific DAF.

^d U.S. Environmental Protection Agency RSL generic tables June 2015; found at: http://www2.epa.gov/risk/risk-based-screening-table-generic-tables.

^eK_d value for organic chemicals calculated by multiplying K_{oc} by fraction organic carbon (f_{oc}) of 0.0011 (average of the Performance Based Acquisition 2008 Remedial Investigation geotechnical samples LL11SB-085-5567-SO and LL11SB-085-5568-SO). CAS = Chemical Abstract Service.

Table 6-3. Unit-Specific Parameters Used in SESOIL and AT123D Modeling

| Parameters | Symbol | Units | Value | Source for Value |
|--|-----------------------|-----------------|---|---|
| | | | | Seasonal Soil Compartment Model (SESOIL) |
| Percolation Rate (Recharge Rate) | q | m/yr | 9.42E-02 | 0.1 * SESOIL Precipitation for Youngstown, Ohio |
| Horizontal Area of Aggregate | A_p | cm ² | 4.73E+08 for the FPA 2.03E+09 for the NPA | Calculated from Load Line 11 Aggregate Location Map |
| Intrinsic Permeability - clayey sand | р | cm ² | 1.05E-10 | Calibrated from SESOIL model |
| Disconnectedness Index | c | unitless | 11 | Calibrated from SESOIL model |
| Freundlich Equation Exponent | n | unitless | 1 | SESOIL default |
| Fraction Organic Carbon | f_{oc} | unitless | 1.10E-03 | |
| Bulk Density | ρ_{b} | kg/L | 1.75 | |
| Moisture Content | W | wt % | 19.7 | Average from the DDA00 DI Gestachnical Complex II 115D 005 5567 SO and II 115D 005 5560 SO |
| Water-filled Soil Porosity | Tw | unitless | 0.344 | Average from the PBA08 RI Geotechnical Samples LL11SB-085-5567-SO and LL11SB-085-5568-SO |
| Air-filled Soil Porosity | Ta | unitless | 0.014 | |
| Porosity - total | n_{T} | unitless | 0.358 | |
| Vadose Zone Thickness | Vz | m | 0.3 to 4.88 for the FPA 0.3 to 2.44 for the NPA | Average based on ground surface elevations and depth to water table from Figure 3-1 and soil boring logs (MKM 2005a and Appendix A) |
| Leaching Zone Thickness | Th | m | 0.0 to 2.13 for the FPA 0.0 to 2.44 for the NPA | Average based on ground surface elevations and depth to water table from Figure 3-1 and soil boring logs (MKM 2005a and Appendix A) |
| | • | | Analy | tical Transient 1-,2-,3-Dimensional Model (AT123D) |
| Aquifer Thickness | h | m | 6 | Facility-wide assumption for the unconsolidated aquifer presented the Load Line 1 investigation (USACE 2003b) |
| Hydraulic Conductivity in Saturated Zone | K_{S} | cm/s | 3.49E-05 | Average of slug test results from MKM (2005a) |
| Hydraulic Gradient | i | unitless | 0.0131 for the FPA 0.0215 for the NPA | Average gradients determined from Figure 3-1 |
| Effective porosity | n _e | unitless | 0.2 | Assumed for sandstone (USEPA 1985) |
| Dispersivity, longitudinal | $\alpha_{ m L}$ | m | 10 | Assumed |
| Dispersivity, transverse | α_{T} | m | 1 | $0.1 \alpha_{\rm L}$ |
| Dispersivity, vertical | $\alpha_{ m V}$ | m | 0.1 | $0.01 \alpha_{\rm L}$ |
| Retardation factor | R_d | unitless | chemical-specific | Presented in Table E-7, Appendix E |

MKM 2005a. Report for the Remedial Investigation at Load Line 11 (AOC 44).

USACE 2003b. Phase II Remedial Investigation Report for the Load Line 1 at the Ravenna Army Ammunition Plant, Ravenna, Ohio. June 2003.

USEPA 1985. Water Quality Assessment: A Screening Procedure for Toxic and Conventional Pollutants in Surface and Ground Water, Revised 1985 Parts 1 and 2, EPA/600/6-85/002. Office of Research and Development, Environmental Research Laboratory, Athens, Georgia. September 1985.

 cm^2 = Square centimeters.

 $cm/s = \hat{C}entimeters per second.$

FPA = Former production area.

kg/L = Kilograms per liter.

m = Meter.

m/yr = Meters per year.

NPA = Non-production area. PBA08 RI = Performance Based Acquisition 2008 Remedial Investigation.

wt % = Weight percent.

Table 6-4. Summary of SESOIL Modeling Results

| Initial CMCOPC | Maximum Soil Concentration (mg/kg) | Discrete Sample Location | Maximum Depth of Contamination (ft bgs) | Depth to Groundwater (ft bgs) | Predicted C _{leachate, max} Beneath Source (mg/L) | Time Required to Reach C _{leachate, max} (years) | MCL/RSL (mg/L) | Resident Receptor Adult FWCUG ^a (mg/L) | Facility-wide Background Unconsolidated Groundwater (mg/L) | Final CMCOPC? ^b (yes/no) |
|----------------------|---------------------------------------|-----------------------------|--|-------------------------------------|---|---|----------------|--|--|---|
| | | | | | Former Production Area | a | | | | |
| | | | | | Inorganic Chemicals | | | | | |
| Arsenic | 30.2 | LL11ss-022 | 3 | 3.25 | 1.25E+00 | 181 | 1.00E-02 | 5.60E-05 | 1.17E-02 | Yes |
| Barium | 127 | LL11cs-012 | 2 | 2.25 | 4.90E+00 | 96 | 2.00E+00 | 7.06E-01 | 8.00E-02 | Yes |
| Cobalt | 33.8 | LL11sd-025 | 1 | 2 | 9.34E-01 | 226 | 6.00E-03 | 7.29E-02 | 0.00E+00 | Yes |
| Zinc | 465 | LL11cs-025 | 1 | 1.25 | 2,25E+01 | 202 | 6.00E+00 | 1.09E+00 | 6.09E-02 | Yes |
| | | | | Sen | ni-volatile Organic Compo | ounds | | | | |
| Benzo(b)fluoranthene | 0.56 | LL11sb-060 | 1 | 1.25 | 8.93E-04 | 1000 | 3.40E-05 | 2.00E-06 | None | Yes |
| Naphthalene | 0.017 | LL11sb-061 | 1 | 1.3 | 1.71E-02 | 7 | 1.70E-04 | None | None | Yes |
| | | | | | Explosives | | | | | |
| PETN | 0.049 | LL11sb-061 | 1 | 2 | 7.01E-02 | 5 | 1.90E-02 | None | None | Yes |
| | | | | | Non-Production Area | | | | | |
| | | | | | Inorganic Chemicals | | | | | |
| Arsenic | 44.1 | LL11cs-045 | 8 | 8.25 | 1.88E+00 | 73 | 1.00E-02 | 5.60E-05 | 1.17E-02 | Yes |
| Cobalt | 24.7 | LL11cs-036 | 2 | 2.25 | 8.70E-01 | 105 | 6.00E-03 | 7.29E-02 | 0.00E+00 | Yes |
| Manganese | 1930 | LL11cs-036 | 2 | 2.25 | 1.63E+01 | 384 | 4.30E-01 | 1.58E-01 | 1.02E+00 | Yes |
| Nickel | 33.5 | LL11cs-036 | 1 | 2 | 4.97E-01 | 325 | 3.90E-01 | 7.29E-02 | 0.00E+00 | Yes |
| | | | | Sen | ni-volatile Organic Compo | ounds | | | | |
| Naphthalene | 0.01 | LL11sb-063 | 1 | 3 | 3.74E-03 | 15 | 1.70E-04 | None | None | Yes |
| | | | | <u> </u> | Explosives | | | | | |
| PETN | 0.036 | LL11sb-067 | 1 | 9 | 9.66E-03 | 22 | 1.90E-02 | None | None | No |

CMCOPC = Contaminant migration chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

ft = Feet.

MCL = Maximum contaminant level.

mg/kg = Milligrams per kilogram. mg/L = Milligrams per liter.

NA = Not available.

PETN = Pentaerythritol tetranitrate. RSL = Regional screening level.

SESOIL = Seasonal soil compartment model. **Bold** = CMCOPCs exceeding MCL/RSL, Resident Receptor Adult FWCUGs, and facility-wide background concentrations.

^a The Resident Receptor Adult FWCUG is based on a target risk of 1E-06 and a hazard quotient of 0.1.
^b The Final CMCOPC was identified comparing predicted maximum leachate concentration to MCL/RSL, Resident Receptor Adult FWCUGs, and facility-wide background concentrations. A constituent is a CMCOPC if its predicted leachate concentration exceeds its MCL/RSL within 1,000 years. bgs = Below ground surface.

Table 6-5. Summary of AT123D Modeling Results

| Final CMCOPC | Maximum Leachate Concentration ^a (mg/L) | Predicted Max Groundwater ^b (C _{MAX}) Beneath Source (mg/L) | Predicted Max Groundwater ^b (C _{MAX}) Downgradient Receptor (mg/L) | Distance to Downgradient Receptor (ft) | Observed Maximum Groundwater Concentrations ^c (mg/L) | MCL/RSL (mg/L) | Resident Adult FWCUG ^d (mg/L) | Facility-wide Background Unconsolidated Groundwater (mg/L) | CMCOC for Further WOE Evaluation? ^e (yes/no) |
|-----------------------|---|--|---|--|---|-------------------|--|--|---|
| That civicor c | Concentration (mg/L) | Defication Source (mg/L) | Receptor (mg/1) | Final CMCOPCs in So | | WICE/RSE (IIIG/E) | | Groundwater (mg/L) | (yes/110) |
| | | | | Former Production Ar | | | | | |
| | | | | Inorganic chemicals | | | | | |
| Arsenic | 1.25E+00 | 7.12E-01 | 0.00E+00 | 605 | 2.20E-02 | 1.00E-02 | 5.60E-05 | 1.17E-02 | Yes |
| Barium | 4.90E+00 | 3.46E+00 | 0.00E+00 | 825 | 8.30E-02 | 2.00E+00 | 7.06E-01 | 8.00E-02 | Yes |
| Cobalt | 9.34E-01 | 7.13E-01 | 0.00E+00 | 705 | 1.60E-03 | 6.00E-03 | 7.29E-02 | 0.00E+00 | Yes |
| Zinc | 2.25E+01 | 1.43E+01 | 0.00E+00 | 1140 | 2.00E-02 | 6.00E+00 | 1.09E+00 | 6.09E-02 | Yes |
| | | | Å | Semi-volatile Organic Comp | ounds | | | | |
| Benzo(b) | | | | | | | | | |
| fluoranthene | 8.93E-04 | 1.59E-04 | 0.00E+00 | 1265 | ND | 3.40E-05 | 2.00E-06 | None | Yes |
| Naphthalene | 1.71E-02 | 7.17E-03 | 0.00E+00 | 950 | 1.40E-04 | 1.70E-04 | None | None | Yes |
| | | | | Explosives | | | | | |
| PETN | 7.01E-02 | 3.14E-02 | 2.67E-08 | 950 | ND | 1.90E-02 | None | None | Yes |
| | | | | Non-Production Area | ı | | | | |
| | | | | Inorganic chemicals | | | | | |
| Arsenic | 1.88E+00 | 9.12E-01 | 0.00E+00 | 558 | 2.20E-02 | 1.00E-02 | 5.60E-05 | 1.17E-02 | Yes |
| Cobalt | 8.70E-01 | 4.90E-01 | 0.00E+00 | 720 | 1.60E-03 | 6.00E-03 | 7.29E-02 | 0.00E+00 | Yes |
| Manganese | 1.63E+01 | 9.24E+00 | 0.00E+00 | 929 | 9.67E-01 | 4.30E-01 | 1.58E-01 | 1.02E+00 | Yes |
| Nickel | 4.97E-01 | 3.04E-01 | 0.00E+00 | 720 | 3.50E-02 | 3.90E-01 | 7.29E-02 | 0.00E+00 | No |
| | | | | Semi-volatile Organic Comp | | | | | |
| Naphthalene | 3.74E-03 | 1.47E-03 | 5.27E-10 | 825 | 1.40E-04 | 1.70E-04 | None | None | Yes |
| | | | | Final CMCOPCs in Sedin | | | | | |
| | | | | East Ditch Aggregate | r | | | | |
| | | | | Semi-volatile Organic Comp | | | | | |
| Benz(a)anthracene | 3.79E-05 | 3.78E-05 | 0.00E+00 | 650 | ND | 1.20E-05 | 4.00E-06 | None | Yes |
| Naphthalene | 3.96E-04 | 3.69E-04 | 3.10E-12 | 650 | 1.40E-04 | 1.70E-04 | None | None | Yes |
| | | | | West Ditch Aggregate | | | | | |
| | | | | Inorganic chemicals | | | | | |
| Antimony | 1.00E-02 | 1.00E-02 | 0.00E+00 | 700 | 1.50E-04 | 6.00E-03 | 1.30E-03 | 0.00E+00 | Yes |
| | | | | Semi-volatile Organic Comp | | | | | |
| Benz(a)anthracene | 3.34E-05 | 3.27E-05 | 0.00E+00 | 700 | ND | 1.20E-05 | 4.00E-06 | None | Yes |
| Benzo(b)fluoranthene | 1.06E-04 | 1.04E-04 | 0.00E+00 | 700 | ND | 3.40E-05 | 2.00E-06 | None | Yes |
| Dibenz(a,h)anthracene | 3.80E-06 artment model (SESOIL) predicted 1 | 3.73E-06 | 0.00E+00 | 700 | ND | 3.40E-06 | 1.50E-07 | None | Yes |

Represents seasonal soil compartment model (SESOIL) predicted maximum leachate concentration just above the water table.

^b The predicted maximum groundwater concentration was estimated using the results from SESOIL and applying AT123D model.

^cObserved groundwater concentrations were reported in the RVAAP Facility-wide Groundwater Monitoring Program 2009 Annual Report (EQM 2010) and sampling events through 2013.

^d The Resident Receptor Adult FWCUG is based on a target risk of 1E-06 and a hazard quotient of 0.1.

eThe CMCOC was identified comparing predicted concentration in groundwater to MCL/RSL, Resident Receptor Adult FWCUGs, and facility-wide background concentrations. A constituent is a CMCOC if its predicted concentration in groundwater exceeds all its screening criteria within 1,000 years.

^f Maximum groundwater concentration = maximum sediment concentration divided by the distribution coefficient.

AT123D = Analytical transient 1-,2-,3-dimensional model.

CMCOC = Contaminant migration chemical of concern.

CMCOPC = Contaminant migration chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

ft = Feet.

MCL = Maximum contaminant level.

mg/L = Milligrams per liter.

ND = Not detected.

PETN = Pentaerythritol tetranitrate.

RSL = Regional screening level.

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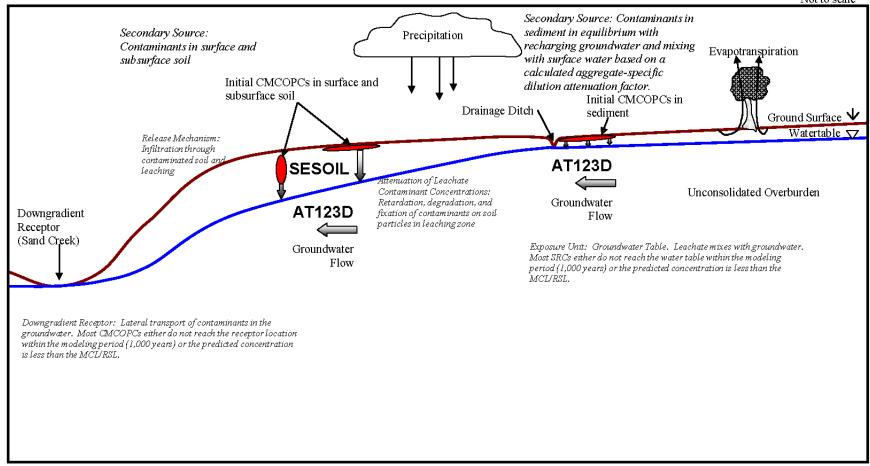


Figure 6-1. Contaminant Migration Conceptual Model

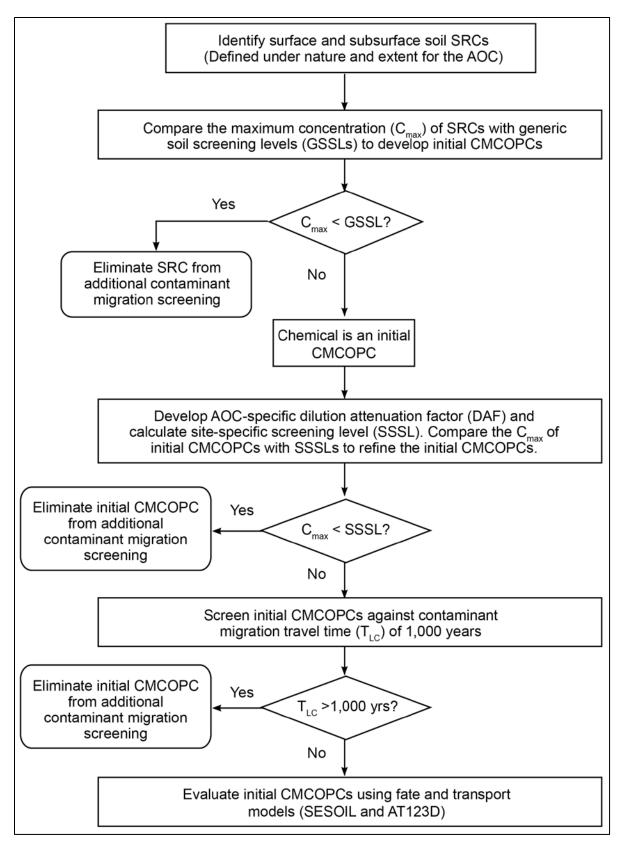


Figure 6–2. Initial CMCOPCs Identified in Soil Screening Analysis for SESOIL Evaluation

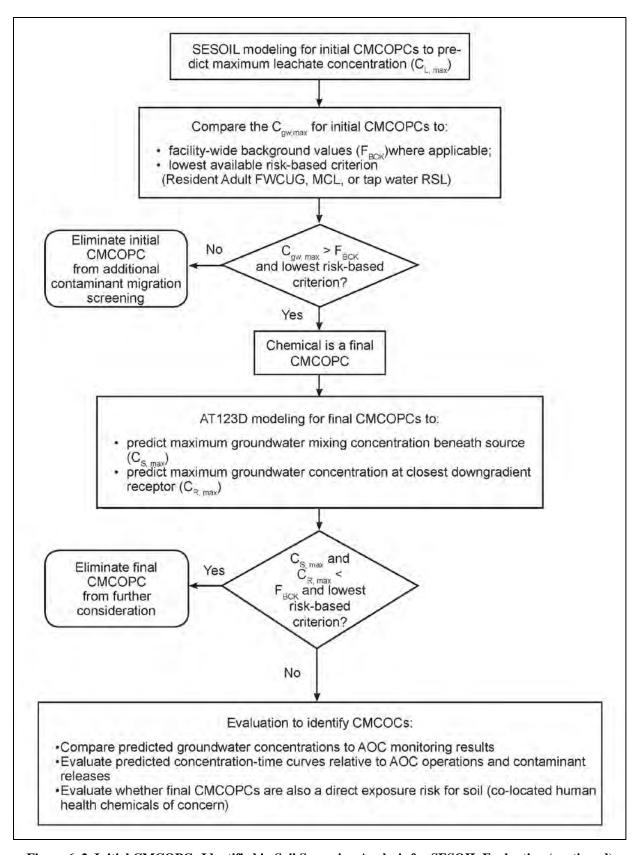


Figure 6–2. Initial CMCOPCs Identified in Soil Screening Analysis for SESOIL Evaluation (continued)

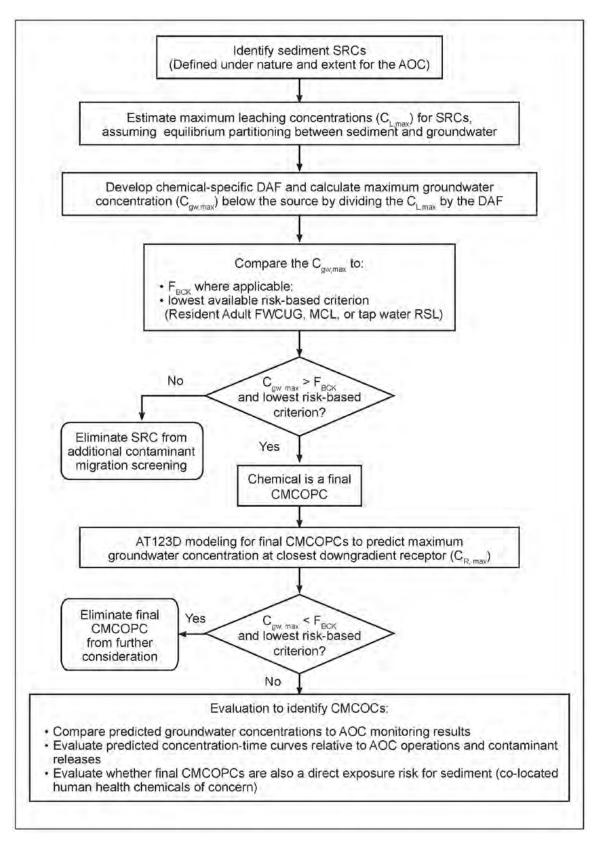


Figure 6-3. AOC Fate and Transport Modeling Approach - Sediment

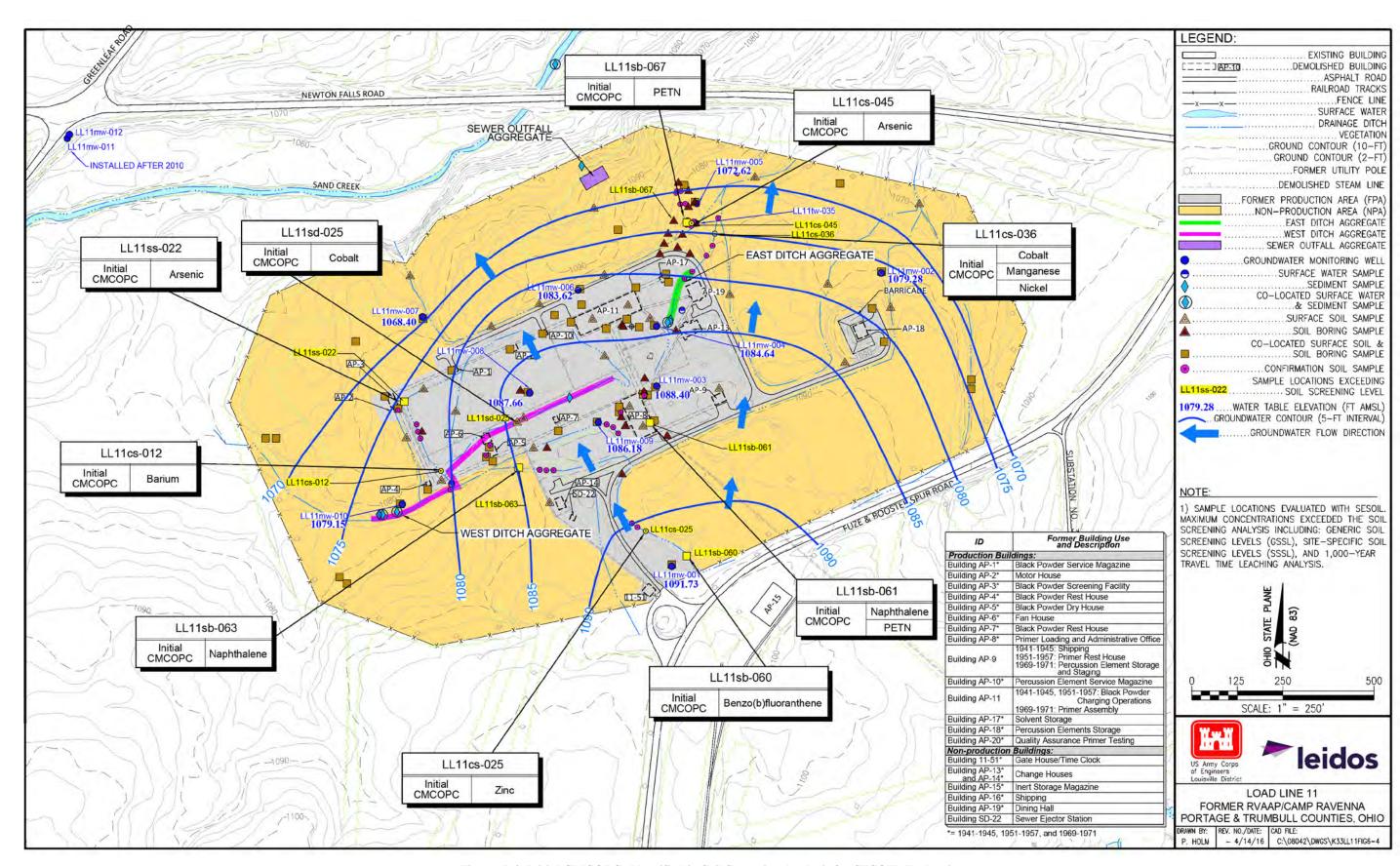


Figure 6-4. Initial CMCOPCs Identified in Soil Screening Analysis for SESOIL Evaluation

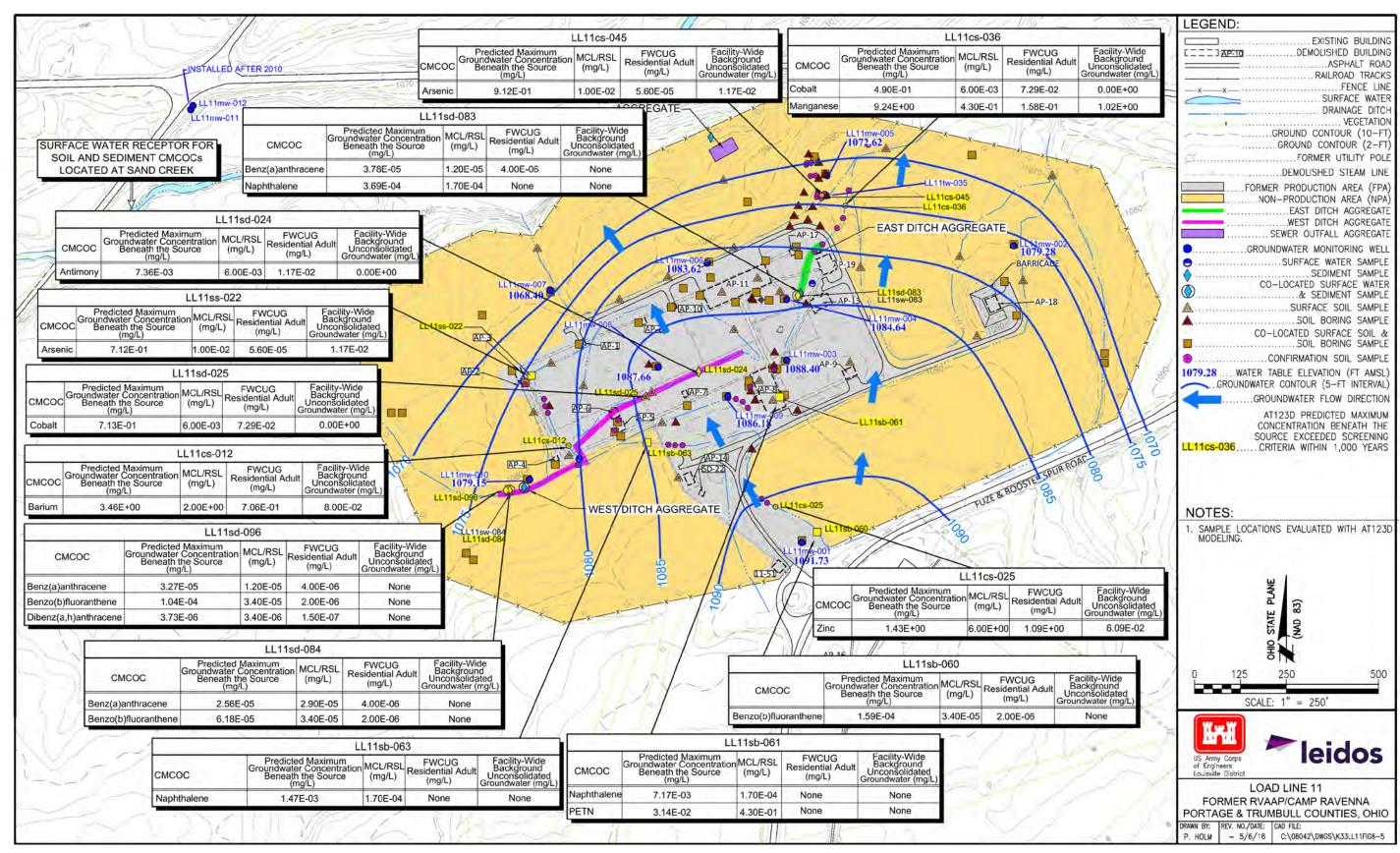


Figure 6-5. CMCOCs Identified for further WOE Evaluation based on AT123D Modeling

7.0 RISK ASSESSMENT

7.1 DATA EVALUATION FOR HUMAN HEALTH AND ECOLOGICAL RISK ASSESSMENTS

The purpose of this data evaluation is to develop a set of chemical data suitable and representative of current conditions for use in the HHRA and ERA. As indicated in Section 4.2, an IRA was performed as an early response to contamination in primary pathways at Load Line 11. This IRA was conducted in unison with the Phase I RI. This IRA included removing sumps, 230 yd³ of contaminated media in drainage ditches, and a hot spot of petroleum-contaminated soil (MKM 2004a). The older sample locations removed during these source removal activities are not included in the data set; however, the numerous confirmation samples collected after source removals are included in these risk assessments.

Load Line 11 data were evaluated to establish data aggregates and to identify a list of SRCs.

7.1.1 Data Aggregates

This section describes the data aggregates for the media for which human and ecological receptors are potentially exposed, followed by a summary of SRCs in Section 7.1.2. Section 4.0 includes a summary of available data. Data collected at Load Line 11 were aggregated by environmental medium, exposure depth (e.g., surface soil), and EU. Samples included in the risk assessment data sets for soil are listed in Tables 7-1 and 7-2. Samples included in the risk assessment data sets for sediment and surface water are listed in Tables 7-3 and 7-4.

7.1.1.1 **Soil Data**

EUs were established at Load Line 11 as part of the data aggregation prior to the risk assessment evaluations conducted for this Phase II RI Report. The EUs take into account how the areas were previously used and the extent of potential contamination within a given area. In establishing EUs at Load Line 11, the area within the perimeter road is assumed to have the maximum potential contamination; therefore, that area was identified as a separate EU (i.e., FPA) from the rest of the AOC (i.e., NPA) as shown in Figure 2-3.

Soil data were aggregated by the following depth intervals:

• Surface soil with an exposure depth of 0–1 ft bgs was evaluated for the Resident Receptor (Adult and Child) and for potential risk to ecological receptors, as this layer is the most active biological zone (USACE 2003a). Table 7-1 presents the risk assessment data set for surface soil (0–1 ft bgs) data. For this risk assessment, discrete surface soil (0–1 ft bgs) samples collected in August 2000 through March 2001 as part of the IRA (MKM 2004a) and Phase I RI (MKM 2005a) and March through April 2010 for the PBA08 RI were used to characterize surface soil. The buildings at Load Line 11 were demolished and removed in 2004 and 2005 after the fieldwork associated with the Phase I RI Report (MKM 2005a) was completed. Soil

- in the vicinity of former production buildings was extensively disturbed during building demolition activities. It is likely that some of the areas sampled in 2000 and 2001 as part of the Phase I RI Report (MKM 2005a) were disturbed during demolition activities.
- Subsurface soil with an exposure depth of 1–13 ft bgs was evaluated for the Resident Receptor (Adult and Child). Discrete soil data from samples collected in 2000, 2001, and 2010 with a starting depth within this interval were used to evaluate subsurface soil. Table 7-2 presents the risk assessment data for subsurface soil (1–13 ft bgs).

7.1.1.2 Surface Water and Sediment Data

No perennial surface water features are present within the AOC or in the immediate vicinity. Intermittent surface water flows in small drainage ditches bordering roads and within the FPA (Figure 3-1). These ditches contain water for short periods of time only during precipitation events or periods of snow melt. On-site surface water and sediment samples were collected at Load Line 11 from the East Ditch that flows north and the West Ditch that flows west/northwest and eventually to Sand Creek. Each of these locations represents separate, isolated drainage and were evaluated as a separate EU.

In addition to these ditches, a sediment sample was collected from sewer outfall location LL11sd-021, and sediment, and surface water samples were collected at off-AOC location LL11sd/sw-082. Risk assessments for these locations were not conducted for this report; rather, these locations were only included as part of the nature and extent evaluation for this Phase II RI. The sewer outfall location (LL11sd-021) will be evaluated as part of the RVAAP-67 Facility-wide Sewers AOC.

Discrete surface water and sediment samples collected in February 2010 per the PBA08 SAP were used to characterize risks from exposure to chemicals in these EUs. Five older surface water and two older sediment samples collected in 2000 (MKM 2005a) were not used in the risk assessment because more recent (2010) data were collected from these sample locations and are more representative of current AOC conditions. Older data not included in the risk assessment were used in characterizing nature and extent of contamination (Section 5.0) to evaluate trends over time.

Samples included in the risk assessment data sets for surface water and sediment are listed in Tables 7-3 (surface water) and 7-4 (sediment). The small size of the surface water and sediment data sets reflects the small size of these EUs.

7.1.2 Identification of SRCs

Section 4.5 presents the statistical methods and screening criteria used to identify SRCs. The purpose of identifying SRCs is to determine the presence or absence of contamination that is above naturally occurring levels.

The 1978 Installation Assessment identified the major contaminants of the former RVAAP to be TNT, composition B (a combination of TNT and RDX), sulfates, nitrates, lead styphnate, and lead azide. Additional site-specific contaminants at Load Line 11 include TNT, RDX, HMX,

nitroglycerine, nitrocellulose, nitroguanidine, and heavy metals (lead, chromium, mercury, and arsenic) from munitions assembly activities; VOCs and petroleum from former Building AP-17 that was utilized for solvent storage and at former Building AP-11 where lacquer sealing materials were used on finished primers; and PCBs from on-site transformers. The evaluation of historical chemical contamination is not limited to these chemicals; rather, this evaluation is expanded to include all eligible chemical data that is available.

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The SRC screen was not limited to only contaminants that may have been a product of previous site use. Rather, the SRC screen followed the three steps outlined in the FWCUG Report, as summarized below, using all chemical data available:

- Background screening: MDCs of naturally occurring inorganic chemicals were compared to
 the facility-wide background concentrations for RVAAP, which are summarized in the
 FWCUG Report. Inorganic constituents detected above facility-wide background
 concentrations or having no background concentrations were retained as SRCs. All detected
 organic chemicals were retained as SRCs.
- Screening of essential human nutrients: Chemicals considered essential nutrients (e.g., calcium, chloride, iodine, iron, magnesium, potassium, phosphorous, and sodium) are an integral part of the human food supply and are often added to foods as supplements. USEPA recommends these chemicals not be evaluated provided they are: (1) present at low concentrations (i.e., only slightly elevated above naturally occurring levels) and (2) toxic at only very high doses (i.e., much higher than those that could be associated with contact at the AOC) (USEPA 1989). Essential nutrients detected near or below their RDA/RDI-based SLs were eliminated as SRCs.
- Frequency of detection screening: In accordance with the FWCUG Report and as revised in the Final (Revised) United States Army Corps of Engineers RVAAP Position Paper for the Application and Use of Facility-Wide Human Health Cleanup Goals (USACE 2012b) (hereafter referred to as the Position Paper for Human Health CUGs), analytes detected in less than 5% of the samples are screened out from further consideration with the exception of explosives and propellants).

Details of the SRC screening for each exposure medium are provided in Tables G-1 through G-4 of Appendix G. The SRCs identified for Load Line 11 are summarized in Table 7-5.

7.2 HUMAN HEALTH RISK ASSESSMENT

This HHRA identifies COCs that may pose potential health risks to humans resulting from exposure to contamination at Load Line 11. This HHRA was conducted as part of the PBA08 RI and is based on the methods from the following guidance documents:

- FWHHRAM (USACE 2005b),
- FWCUG Report (USACE 2010a),
- Position Paper for Human Health CUGs (USACE 2012b), and
 - Technical Memorandum (ARNG 2014).

- 1 To accomplish the goal of streamlined risk-based decision making, the FWCUG Report was
- 2 developed to support risk assessments of the remaining AOCs within the former RVAAP. The
- 3 FWCUG Report contains calculated FWCUGs and guidance for applying the FWCUGs to accelerate
- 4 the risk assessment process. This approach takes advantage of the many risk assessment inputs and
- 5 decisions that have previously been accepted by stakeholders through the application of the CERCLA
- 6 process at the former RVAAP.

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- Most of the agreed upon risk assessment methodology has been documented in the FWHHRAM (USACE 2005b) and follows standard USEPA-approved risk assessment guidance. This includes the
- 10 process to identify RVAAP COPCs (presented in Figure 4-6), a TR of 1E-06 and HQ of 0.1 to
- identify COPCs, and a TR of 1E-05 and HQ of 1 to identify COCs.

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- 13 Other approaches, such as calculating the sum-of-ratios (SOR), were developed in the FWCUG
- Report (USACE 2010a) and Position Paper for Human Health CUGs (USACE 2012b). The Technical
- 15 Memorandum (ARNG 2014) amends the risk assessment process to establish future Land Uses and
- applicable receptors to be evaluated in an RI.

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The approach to risk-based decision making is as follows:

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- 1. **Develop FWCUGs** Use the risk assessment process presented in the FWHHRAM to develop FWCUGs for all COPCs identified from the facility-wide data set at RVAAP. This process was completed in the FWCUG Report.
- 2. *RI Characterization Sampling* Perform sampling and analysis to characterize an AOC and establish baseline chemical concentrations. A summary and the results of the RI characterization sampling for Load Line 11 are presented in Section 4.0.
- 3. *Mapping and Data Analysis to Identify SRCs and COPCs* Follow the requirements specified in the FWHHRAM and the Position Paper for Human Health CUGs (USACE 2012b), perform data analysis and mapping to identify SRCs and COPCs, establish EUs, and calculate exposure point concentrations (EPCs) for each COPC. The results of the mapping and data analysis for Load Line 11 to identify SRCs are presented in Section 4.0 and 5.0 and are summarized in Section 7.1.
- 4. *Identification of COCs* Compare EPCs to FWCUGs to determine COCs.
- 5. Address Identified COCs Develop FS, PP, and ROD to address any COCs requiring remedy.

- 36 Identifying COPCs and COCs follows the four steps for a streamlined risk assessment established in
- 37 the FWCUG Report: identify media of concern, identify COPCs, present AOC Land Use and
- 38 appropriate receptors, and compare to appropriate FWCUGs to identify COCs. These steps are
- 39 discussed in the following subsections.

7.2.1 Identify Media of Concern

Media of concern at Load Line 11 are surface and subsurface soil, sediment, and surface water, as defined in Section 7.1.1. Groundwater is present at this AOC and will be evaluated (including risk assessment) in a separate document, as explained in Section 1.2.

7.2.2 Identify COPCs

Section 4.5 presents the statistical methods and screening criteria used to identify SRCs. COPCs are a subset of the SRCs in each exposure medium present at concentrations that indicate the potential for impacts to human receptors. The COPC screen follows the approach specified in the FWCUG Report and is summarized in this section.

To identify COPCs, the MDC of all SRCs was screened against the most stringent chemical-specific FWCUG of all RVAAP receptors at a target cancer risk level of 1E-06 and non-carcinogenic target HQ of 0.1 for the Resident Receptor (Adult and Child) and National Guard Trainee. If no FWCUGs existed for an SRC, the USEPA residential RSL (from RSL table dated June 2015) was used for this screen. No reference dose (RfD) or cancer potency factors are available for acenaphthylene, benzo(ghi)perylene, and phenanthrene; therefore, the RSL for pyrene was used for these PAHs (NDEP 2006).

Hexavalent chromium was detected in two of three surface soil samples collected at Load Line 11 for chromium speciation. Since hexavalent chromium was detected, as part of the conservative screening approach for identifying COPCs, the FWCUG for hexavalent chromium (the more toxic of the two chromium species evaluated) was used at this stage.

Details of the COPC screening for each exposure medium are provided in Appendix G, Tables G-1 through G-4. The COPCs identified for the media of concern at Load Line 11 are presented in Table 7-6 and are summarized below.

7.2.2.1 COPCs in Surface Soil

Former Production Area. Of the 48 chemicals detected in surface soil (0-1 ft bgs) samples at the FPA, 39 (17 inorganic chemicals, 16 SVOCs, 1 VOC, 1 PCB, and 4 explosives) were identified as SRCs. Risk-based screening identified seven inorganic chemicals (aluminum, arsenic, chromium, cobalt, cyanide, manganese, and sulfide), one PCB (PCB-1254), and five SVOCs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] as COPCs in surface soil (0-1 ft bgs).

Non-Production Area. Of the 45 chemicals detected in surface soil (0-1 ft bgs) samples at the NPA, 35 (16 inorganic chemicals, 3 explosives, 1 PCB, and 15 SVOCs) were identified as SRCs. Risk-based screening identified six inorganic chemicals (arsenic, chromium, cobalt, cyanide, manganese, and sulfide) and four SVOCs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and dibenz(a,h)anthracene] as COPCs in surface soil (0-1 ft bgs).

7.2.2.2 COPCs in Subsurface Soil

Former Production Area. Of the 36 chemicals detected in subsurface soil (1-13 ft bgs) samples at the FPA, 14 (12 inorganic chemicals, 1 explosive, and 1 PCB) were identified as SRCs. Risk-based screening identified four inorganic chemicals (arsenic, cobalt, sulfate, and sulfide), and one PCB (PCB-1254) as COPCs in subsurface soil.

Non-Production Area. Of the 41 chemicals detected in subsurface soil (1-13 ft bgs) samples at the NPA, 25 (16 inorganic chemicals, 2 miscellaneous organic chemicals (TPH-DRO and TPH-GRO), 1 explosive, and 6 SVOCs) were identified as SRCs. Risk-based screening identified eight inorganic chemicals (aluminum, arsenic, barium, cobalt, cyanide, sulfate, sulfide, and thallium), two miscellaneous organic chemicals (TPH-GRO and TPH-DRO), and one PAH [benzo(a)pyrene] as COPCs in subsurface soil at the NPA.

7.2.2.3 COPCs in Sediment

East Ditch. Of the 36 chemicals detected in sediment samples collected at the East Ditch, 25 (10 inorganic chemicals and 15 SVOCs) were identified as SRCs. Risk-based screening identified five COPCs in sediment: four inorganic chemicals (aluminum, arsenic, chromium, and cobalt) and one SVOC [benzo(a)pyrene].

West Ditch. Of the 32 chemicals detected in sediment samples collected at the West Ditch, 18 (8 inorganic chemicals and 10 SVOCs) were identified as SRCs. Risk-based screening identified four inorganic chemicals (aluminum, chromium, cobalt, and sulfide) as COPCs in sediment.

7.2.2.4 COPCs in Surface Water

East Ditch. Of the 18 chemicals detected in surface water samples collected at the East Ditch, 8 (6 inorganic chemicals and 2 pesticides) were identified as SRCs. Risk-based screening identified one inorganic chemical (cobalt) as a COPC in East Ditch surface water.

West Ditch. Of the 30 chemicals detected in surface water samples collected at the West Ditch, 23 (12 inorganic chemicals, 1 pesticide, and 10 SVOCs) were identified as SRCs. Risk-based screening identified two inorganic chemicals (arsenic and cobalt) and four SVOCs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] as COPCs in West Ditch surface water.

7.2.3 Land Use and Representative Receptors

Camp Ravenna is a controlled-access facility. Load Line 11 is located in the south-central portion of the facility and is not currently used for training. Three Land Uses for the RVAAP restoration program are specified in the Technical Memorandum (ARNG 2014) for consideration in the RI along with the following Representative Receptors:

- 1. Unrestricted (Residential) Land Use Resident Receptor (Adult and Child).
- 2. Military Training Land Use National Guard Trainee.
- 3. Commercial/Industrial Land Use Industrial Receptor (USEPA's Composite Worker).

Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, then the AOC is also considered to have met the requirements of the other Land Uses (i.e., Commercial/Industrial and Military Training), and those other Land Uses do not require evaluation.

As stated in Paragraph 6.d of the Technical Memorandum, if an AOC fails to meet the Unrestricted (Residential) Land Use, then an FS will be completed that evaluates cleanup options for all three Land Uses [i.e., Unrestricted (Residential) Land Use, Military Training Land Use, and Commercial/Industrial Land Use]. Remedial alternatives for meeting each Land Use are to be evaluated per the current guidelines for selecting a remedy for the AOC. The preferred remedy is one that would meet Unrestricted (Residential) Land Use. RI/FS Reports in progress at the time of the Technical Memorandum's approval on February 11, 2014 will not be revised to include an evaluation of the Commercial/Industrial Land Use as an alternative if it achieves no further action for Unrestricted (Residential) Land Use.

7.2.4 Compare to Appropriate FWCUGs

Previous sections have outlined the process for identifying SRCs and COPCs. Comparing COPC exposure concentrations to FWCUGs and determining COCs follows guidance presented in the Position Paper for Human Health CUGs (USACE 2012b) and Technical Memorandum (ARNG 2014).

The COC determination process is as follows:

- Report all carcinogenic- and non-carcinogenic based FWCUGs corresponding to a TR of 1E-05 and target HQ of 1 using the most stringent of the Resident Receptor (Adult and Child) FWCUGs to evaluate Unrestricted (Residential) Land Use for each COPC. If no FWCUG is available for a COPC, the Residential RSL, adjusted to represent a TR of 1E-05 or target HQ of 1, is used.

- Report critical effect and target organ for each non-carcinogenic based FWCUG.
- Compare the selected FWCUG to the EPC, including an SOR.

- For non-carcinogens, compare the EPC to the target HQ FWCUG. Sum the ratios of EPC/FWCUG for COPCs that affect similar target organs or do not have an identified target organ.
- o For carcinogens, compare the EPC to the TR FWCUG. Sum the ratios of EPC/FWCUG for all carcinogens.
- Identify the COPC as a COC if:

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- o The EPC exceeds the most stringent of the Resident Receptor (Adult and Child) FWCUGs for either the 1E-05 target cancer risk or the 1 target HQ; or
- o The SOR for all carcinogens or non-carcinogens that may affect the same organ is greater than one. Chemicals contributing at least 10% to the SOR are also considered COCs. In accordance with the Position Paper for Human Health CUGs (USACE 2012b), chemicals contributing greater than 5% but less than 10% to the SOR must be further evaluated before being eliminated as COCs.

The process for calculating FWCUGs rearranges cancer risk or non-cancer hazard equations in order to obtain a concentration that will produce a specific risk or hazard level (USEPA 1991, USACE 2010a). For example, the FWCUG for arsenic at the cancer risk level of 1E-05 for the Resident Receptor Adult is the concentration of arsenic that produces a risk of 1E-05 when using the exposure parameters specific to the Resident Receptor Adult.

For carcinogens, risk is expressed as the probability that an individual will develop cancer over a lifetime as a result of exposure to the carcinogen. Cancer risk from exposure to contamination is expressed as the increased chance of cancer above the normal background rate. In the United States, the background chance of contracting cancer is a little more than 3 in 10 for women and a little less than 5 in 10 for men, or 3E-01 to 5E-01 (American Cancer Society 2015). The calculated incremental lifetime cancer risks (ILCRs) are compared to the range specified in the NCP of 10⁻⁶ to 10⁻⁴, or 1 in a million to 1 in 10,000 exposed persons developing cancer (USEPA 1990). Cancer risks below 10⁻⁶ are considered acceptable; cancer risks above 10⁻⁴ are considered unacceptable. The range between 10⁻⁶ and 10⁻⁴ is of concern, and any decisions to address risks further in this range, either through additional study or engineered control measures, should account for the uncertainty in the risk estimates. The Ohio EPA Division of Environmental Response and Revitalization (DERR) program has adopted a human health cumulative ILCR goal within this range of 1E-05 to be used as the level of acceptable excess cancer risk and for developing remediation goals for the site. The DERR notes that the defined risk goal should be applied as a goal, recognizing the need to retain flexibility during the evaluation and selection of remedial alternatives.

In addition to developing cancer from exposure to chemicals, an individual may experience other adverse effects. The term "adverse effects" is used here to describe a wide variety of systemic effects ranging from minor irritations, such as eye irritation and headaches, to more substantial effects, such as kidney or liver disease and neurological damage. The risk associated with toxic (i.e., non-carcinogenic) chemicals is evaluated by comparing an estimated exposure (i.e., intake or dose) from AOC media to an acceptable exposure expressed as an RfD. The RfD is the threshold level below which no adverse effects are expected to occur in a population, including sensitive subpopulations. The ratio of intake over the RfD is the HQ (USEPA 1989).

The SOR is used to account for the potential additive effects from exposure to multiple chemicals that 1 2 can cause the same effect (e.g., cancer) or affect the same target organ. Cancer risk is assumed to be 3 additive for all carcinogens. Non-cancer risk is assumed to be additive for chemicals with similar sites 4 of toxicological action (i.e., target organ such as liver or critical effect such as adversely affecting the 5 ability to reproduce). This approach compares the EPC of each COPC to the FWCUG to determine a ratio. The sum of these individual ratios is then compared to one. The SOR method is based on the 6 principle that a ratio greater than one represents unacceptable cumulative exposure (i.e., above 7 8 FWCUGs if adjusted for exposure to multiple COPCs), and a ratio less than or equal to one represents 9 acceptable cumulative exposure (i.e., below FWCUGs if adjusted for exposure to multiple COPCs). 10 The FWCUGs for some chemical/receptor combinations are less than the background concentration. 11 In these instances the chemical concentrations are compared to background concentrations to identify 12 COCs. Since the background concentration is not risk-based, these chemicals are not included in the 13 SOR calculations. COCs identified by comparing EPCs to FWCUGs are further evaluated in an 14 uncertainty analysis to identify COCs requiring evaluation in the FS.

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The selection of FWCUGs, calculation of EPCs for comparison to the FWCUGs, and the resulting risk-based COCs are detailed in the following sections.

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7.2.4.1 Selection of Appropriate FWCUGs

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As specified in the Technical Memorandum (ARNG 2014), EPCs for each AOC should initially be evaluated using the most stringent of the Resident Receptor (Adult and Child) FWCUGs to determine if no further action is necessary at an AOC to attain Unrestricted (Residential) Land Use. If this assessment indicates COCs exist that prevent Unrestricted (Residential) Land Use, an FS must be completed to evaluate cleanup options for all three Land Uses [i.e., Unrestricted (Residential) Land Use, Military Training Land Use, and Commercial/Industrial Land Use].

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Unrestricted (Residential) Land Use is evaluated using FWCUGs for the Resident Receptor (Adult and Child). These FWCUGs are provided in Tables 7-7 and 7-8 and are the lower of the Resident Receptor (Adult and Child) values for each COPC and endpoint (non-cancer and cancer). The critical effect or target organ associated with the toxicity values used to calculate the FWCUGs are also provided in these tables.

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The SLs provided in Tables 7-7 and 7-8 for the Resident Receptor (Adult and Child) are the FWCUGs corresponding to a TR of 1E-05 and target HQ of 1. If no FWCUG is available for a COPC, the residential RSLs, adjusted to represent a TR of 1E-05 or target HQ of 1, are used for the Resident Receptor.

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Chromium Speciation

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41 FWCUGs are available for hexavalent chromium and trivalent chromium. Existing data at other 42 AOCs, such as the Building 1200 and Anchor Test Area AOCs (USACE 2012c, USACE 2012d), indicate chromium exists predominantly in the trivalent state rather than the more toxic hexavalent

state. The implementation of the chromium speciation process per the PBA08 SAP is discussed below.

- Hexavalent chromium and total chromium sample collection and results To determine whether the FWCUGs for trivalent or hexavalent chromium are most applicable to Load Line 11 and to support risk management decisions, three discrete surface soil samples were collected and analyzed for hexavalent chromium and total chromium. Three samples were collected in April 2010 per the PBA08 SAP. Samples were collected from the three sampling areas having the highest, mid-point, and lowest total historical chromium results. No hexavalent chromium was detected in one of the three samples. Hexavalent chromium was detected in two speciation samples at concentrations of 0.44 and 0.71 mg/kg. Total chromium and hexavalent chromium results for these three samples are summarized in Table 7-9.
- Percent hexavalent chromium in the chromium speciation samples As documented in the PBA08 SAP, "Chromium speciation evaluates the concentration ratio of hexavalent chromium to total chromium. This ratio will be calculated by collecting and analyzing three samples per AOC for both hexavalent chromium and total chromium." The three chromium speciation samples at Load Line 11 contained no detectable hexavalent chromium in one sample and 4.4% and 4.6% in the remaining two samples, with the highest percent hexavalent chromium associated with the sample having the second highest concentration of total chromium (LL11ss-075). The total chromium concentration (15.4 mg/kg) in sample LL11ss-075 is below the facility-wide background concentrations of 17.4 mg/kg in surface soil and 27.3 mg/kg in subsurface soil. The FWCUG for hexavalent chromium is based on a cancer unit risk factor (URF) calculated using a chromium mixture containing 14% hexavalent chromium and 86% trivalent chromium. The Load Line 11 sample results are below the 14% hexavalent chromium used as the basis for the cancer URF, which was used to calculate the hexavalent chromium FWCUGs.
- Speciation samples to the residential RSL for hexavalent chromium The FWCUG for hexavalent chromium is more appropriately applied to total chromium because it was calculated from a cancer URF based on a chromium mixture containing 1/7 (14%) hexavalent chromium (USEPA 2010a). The residential RSL (3 mg/kg based on a TR of 1E-05) for hexavalent chromium is specific to hexavalent chromium (i.e., it has been adjusted for the chromium mixture used in the toxicity study). The detected concentrations of hexavalent chromium in the chromium speciation samples (0.44 and 0.71 mg/kg) are less than 3 mg/kg, indicating hexavalent chromium is not present above the residential RSL for hexavalent chromium.
- Comparison of the concentration of total chromium to the FWCUG for trivalent chromium After implementing the chromium speciation process specified in the PBA08 SAP, hexavalent chromium was determined to be present at a very low concentration (i.e., below the residential RSL for hexavalent chromium), and the percent hexavalent chromium is less than 14% in of the chromium speciation samples. Therefore, hexavalent chromium is not of concern at Load Line 11, and the reported concentrations of total chromium were compared to the FWCUGs for trivalent chromium for identifying COCs at this AOC.

7.2.4.2 Exposure Point Concentrations for Comparison to FWCUGs

Soil EPCs

EPCs were calculated for each depth interval and EU using analytical results from the discrete samples presented in Tables 7-1 and 7-2. Per the FWHHRAM, the EPC is either the 95% upper confidence limit (UCL) of the mean or the MDC, whichever value is lowest. If the 95% UCL could not be determined, the EPC is the MDC.

Sediment and Surface Water EPCs

Two discrete sediment samples and one discrete surface water sample collected from the East Ditch and three discrete sediment samples and one discrete surface water sample collected from the West Ditch were used to characterize risks from exposure to sediment and surface water. Because of the small number of samples, each ditch was evaluated individually, and the EPC was equal to the detected concentration in each sample.

Identification of COCs for Unrestricted (Residential) Land Use

Load Line 11 COCs for Unrestricted (Residential) Land Use, as represented by the Resident Receptor (Adult and Child), are presented below.

COCs for Surface Soil (0-1 ft bgs)

COC screening for surface soil (0-1 ft bgs) for the Resident Receptor (Adult and Child) is detailed in Appendix G, Tables G-5 through G-7. Sulfide and five PAHs were identified as COCs for the Resident Receptor (Adult and Child), as explained below:

29 <u>COPCs with EPCs lower than the Resident Receptor (Adult and Child) FWCUG:</u> All aluminum, 30 chromium, cobalt, cyanide, manganese, benz(a)anthracene, benzo(b)fluoranthene, 31 dibenz(a,h)anthracene, indeno(1,2,3-cd)pyrene, and PCB-1254 EPCs are lower than the Resident 32 Receptor (Adult and Child) FWCUG.

COPCs with no FWCUG: Sulfide was detected in surface soil at the FPA and NPA. No FWCUGs or RSLs are available for this anion. The potential impact of the lack of screening values for this anion is discussed in the uncertainty assessment (Section 7.2.5).

38 <u>COPCs with EPCs exceeding the Resident Receptor (Adult and Child) FWCUG</u>: EPCs of arsenic at the FPA and NPA and benzo(a)pyrene at the NPA exceed the FWCUG.

• The EPCs for arsenic in surface soil (0–1 ft bgs) at the FPA (14.5 mg/kg) and NPA (14.1 mg/kg) exceed the FWCUG of 4.25 mg/kg but are below the surface soil facility-wide background concentration of 15.4 mg/kg. Because the FWCUG is less than the background

- concentration, the background concentration is used as the cleanup goal for this inorganic chemical. Thus, arsenic is representative of background and is not a COC in surface soil.
- The EPC for benzo(a)pyrene in surface soil at the NPA (0.28 mg/kg) is approximately 1.3 times its FWCUG of 0.221 mg/kg. The EPC for this data aggregate is the MDC (at sample LL11sb-067); benzo(a)pyrene concentrations in the remaining NPA samples ranged from a non-detectable concentration to 0.11 mg/kg. Sample LL11sb-067 was collected in 2010 from the northeastern portion of Load Line 11.

SOR Analysis: Four additional PAHs were identified as COCs based on the SOR analysis summarized below:

- Five COPCs (aluminum, arsenic, chromium, cobalt, cyanide, manganese, and PCB-1254) identified in surface soil have FWCUGs for non-cancer endpoints. The EPCs for aluminum, arsenic, chromium, cobalt, and manganese were below the facility-wide background concentrations; therefore, these inorganic chemicals were not included in the SOR. The total SORs (Table G-6), regardless of endpoint, were less than or equal to one; therefore, no additional COCs were identified using this analysis.
- Eight COPCs [arsenic, cobalt, benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, indeno(1,2,3-cd)pyrene, and PCB-1254] identified in surface soil have FWCUGs for the cancer endpoint (chromium was evaluated for non-carcinogenic effects as trivalent chromium as previously discussed). An SOR was calculated for each EU (Table G-7). The EPCs for arsenic and cobalt were below the facility-wide background concentrations; therefore, arsenic and cobalt were not included in the SORs. The SORs for the FPA and NPA are both two, due primarily to benzo(a)pyrene. The EPC for benzo(a)pyrene in the FPA (0.219 mg/kg) was slightly less than the FWCUG (0.221 mg/kg). The benzo(a)pyrene EPC for the NPA (0.28 mg/kg) was slightly above the FWCUG. COPCs contributing at least 5% to these SORs are benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and dibenz(a,h)anthracene (Table G-7) in the FPA and NPA and indeno(1,2,3-cd)pyrene in the FPA.

COCs for Subsurface Soil (1-13 ft bgs)

COC screening for subsurface soil (1-13 ft bgs) for the Resident Receptor (Adult and Child) is detailed in Appendix G, Tables G-8 through G-10. No COCs were identified for the Resident Receptor (Adult and Child), as explained below:

COPCs with EPCs lower than the Resident Receptor (Adult and Child) FWCUG: All aluminum, barium, cobalt, cyanide, thallium, PCB-1254, and benzo(a)pyrene EPCs are lower than the Resident Receptor (Adult and Child) FWCUG.

COPCs with no FWCUG: Sulfate, sulfide, TPH-DRO, and TPH-GRO were detected in subsurface soil at the FPA and/or NPA. No FWCUGs or RSLs are available for these constituents. The potential impact of the lack of screening values is discussed in the uncertainty assessment (Section 7.2.5).

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COPCs with EPCs exceeding the Resident Receptor (Adult and Child) FWCUG: EPCs of arsenic at the FPA and NPA exceed the FWCUG. The EPCs for arsenic in subsurface soil (1-13 ft bgs) at the FPA (14.3 mg/kg) and NPA (16.3 mg/kg) exceed the FWCUG of 4.25 mg/kg but are below the surface soil facility-wide background concentration of 19.8 mg/kg. Because the FWCUG is less than the background concentration, the background concentration was used as the cleanup goal for this inorganic chemical. Thus, arsenic is representative of background and is not a COC in subsurface soil.

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SOR Analysis: No additional COCs were identified based on the SOR analysis summarized below:

- Three COPCs (arsenic, cobalt, and PCB-1254) identified in subsurface soil at the FPA have FWCUGs for non-cancer endpoints. The EPCs for arsenic and cobalt are less than the facility-wide background concentrations for subsurface soil; therefore, no SOR was calculated for the FPA. Six COPCs [aluminum, arsenic, barium, cobalt, cyanide, and thallium] identified in the NPA subsurface soil have FWCUGs for non-cancer endpoints. The EPCs for aluminum, arsenic, cobalt, and thallium are less than the facility-wide background concentrations for subsurface soil. These inorganic chemicals were not included in the SOR. The total SOR (Table G-9), regardless of endpoint, was less than one; therefore, no additional COCs were identified using this analysis.
- Four COPCs [arsenic, cobalt, PCB-1254, and benzo(a)pyrene] identified in subsurface soil have FWCUGs for the cancer endpoint. The EPCs for arsenic and cobalt are less than the facility-wide background concentrations for subsurface soil; therefore, these metals were not included in the SOR. SORs were calculated for the remaining COPCs (Table G-10). The

COCs for Sediment

calculated SORs are less than one.

COC screening for sediment for the Resident (Adult and Child) Receptor is detailed in Appendix G, Tables G-11 though G-13. No COCs were identified for the Resident Receptor (Adult and Child), as explained below:

- The detected concentration of arsenic (19.7 mg/kg) in one of two sediment samples from the East Ditch exceeded the FWCUG for the Resident Receptor (Adult and Child) of 4.25 mg/kg. The concentration in this sample is just above the sediment background screening concentration of 19.5 mg/kg. All of the other sediment samples from the East Ditch and West Ditch were below background concentrations. Because the reported concentration in one sediment sample is essentially the same as the background concentration and all other samples are less than background concentrations, arsenic was not identified as a COC in sediment.
- Sulfide was detected in sediment at the West Ditch; however, no FWCUGs or RSLs are available for this anion. The potential impact of the lack of screening value is discussed in the uncertainty assessment (Section 7.2.5).
 - The detected concentrations of all other COPCs were less than their respective FWCUGs.

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- Four COPCs (aluminum, arsenic, chromium, and cobalt) identified in sediment have FWCUGs for non-cancer endpoints. Detected concentrations of aluminum, chromium, and cobalt less than the facility-wide background concentrations for sediment were not included in the SOR for that sample location. Total SORs (Table G-12), regardless of endpoint, were less than or equal to one; therefore, no additional COCs were identified using this analysis.
- Three COPCs [arsenic, cobalt, and benzo(a)pyrene] identified in sediment have FWCUGs for the cancer endpoint. SORs (Table G-13) were less than or equal to one; therefore, no additional COCs were identified using this analysis.

COCs for Surface Water

- COC screening for sediment for the Resident (Adult and Child) Receptor is detailed in Appendix G, Tables G-14 though G-16. No COPCs and, therefore, no COCs were identified in the surface water collected from the East Ditch. Four PAHs were identified as COCs in the West Ditch for the Resident Receptor (Adult and Child), as explained below:
 - The detected concentrations of benz(a)anthracene (0.00035 mg/L), benzo(b)fluoranthene (0.00041 mg/L), and indeno(1,2,3-cd)pyrene (0.00021) in surface water collected from the West Ditch range from approximately two to four times their FWCUG of 0.0001 mg/L. The detected concentration of benzo(a)pyrene (0.00033 mg/L) is approximately 41 times the FWCUG of 0.000008 mg/L. As a result, these PAHs were identified as COCs in surface water of the West Ditch.
 - The detected concentrations of all other COPCs were less than their respective FWCUGs in the surface water collected from the West Ditch.
- No additional COCs were identified based on the SOR analysis as summarized below:
 - Two COPCs (arsenic and cobalt) identified in West Ditch surface water have FWCUGs for non-cancer endpoints. The SOR (Table G-15), regardless of endpoint, was less than one; therefore, no additional COCs were identified using this analysis.
 - Five COPCs [arsenic, benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] identified in West Ditch surface water have FWCUGs for the cancer endpoint. The SOR (Table G-16) was 51, due primarily to benzo(a)pyrene. COPCs contributing at least 5% to the SOR are benz(a)anthracene, benzo(a)pyrene, and benzo(b)fluoranthene (Table G-16). These COPCs were detected in the West Ditch surface water at concentrations above their FWCUGs; therefore, no additional COCs were identified using this analysis.

7.2.5 Uncertainty Assessment

The sources of uncertainty, as well as the potential bias they impart to the risk assessment (i.e., whether conservatism is increased or decreased) and approaches for minimizing their impact on the conclusions of the RI, are briefly discussed below.

7.2.5.1 Uncertainty in Estimating Potential Exposure

Sources of uncertainty in estimating potential human exposure include sampling and analysis limitations, comparison to background concentrations to identify SRCs, and estimation of EPCs.

Sampling Limitations. Uncertainties arise from limits on the media sampled, the total number and specific locations that can be sampled, and the parameters chosen for analysis to characterize the AOC. A total of 82 surface (0-1 ft bgs) soil samples were available for the HHRA. Samples were collected from areas biased toward areas anticipated to have the highest level of potential contamination (i.e., around former buildings, ditches, and a former petroleum contamination removal area) to delineate potential sources as well as from areas where excavation was performed to remove contamination during the 2001 IRA. The results of surface soil sampling were used to efficiently guide selection of locations for subsurface soil sampling with a bias toward the areas of highest potential contamination.

Analytical Limitations. Uncertainty is associated with the chemical concentrations detected and reported by the analytical laboratory. The quality of the analytical data used in the risk assessment was maximized and uncertainty was minimized by implementing QA/QC procedures that specify how samples are selected and handled; however, sampling errors, laboratory analysis errors, and data analysis errors can occur. Beyond the potential for errors, there is normal variability in analytical results.

Some current analytical methods are limited in their ability to achieve detection limits at or below risk-based SLs. Under these circumstances, it is uncertain whether the true concentration is above or below the SLs, which are protective of human health. When analytes have a mixture of detected and non-detected concentrations, EPC calculations may be affected by these detection limits. Risks may be overestimated as a result of some sample concentrations being reported as non-detected at the method detection limit (MDL), when the actual concentration may be much smaller than the MDL. Risks may also be underestimated if some analytes that were not detected in any sample were removed from the COPC list. If these analyte concentrations are below the MDL but are above the SL, the risk from these analytes would not be included in the risk assessment results.

Identification of SRCs. Part of determining SRCs is to identify chemicals detected above the established RVAAP facility-wide background concentrations. This screen does not account for the potential sources of chemicals, and background values are only available for inorganic chemicals.

Uncertainty associated with screening against background concentrations results from statistical limitations and natural variation in background concentrations. Because of these variations, inorganic

chemical concentrations below the background concentration are likely representative of background conditions. Inorganic chemical concentrations above the background concentration may be above background conditions or may reflect natural variation. This is especially true for measured concentrations close to the background concentration.

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At Load Line 11, 16 inorganic chemicals (aluminum, arsenic, barium, beryllium, calcium, chromium, cobalt, copper, iron, lead, magnesium, manganese, mercury, nickel, selenium, and vanadium) had MDCs in surface and/or subsurface soil that were above but close to (i.e., less than 2 times) the background concentration. The consequences of carrying most of these inorganic chemicals forward as SRCs, even if they are actually representative of background concentrations, is negligible because they are not toxic at near background concentrations. By contrast, naturally occurring (background) arsenic and manganese in soil exceeds risk-based cleanup goals. Therefore, the consequence of identifying arsenic or manganese as an SRC if it is, in fact, representative of background can significantly impact the conclusions of the risk assessment. The EPCs of these two metals are less than background concentrations in soil.

Other studies indicate arsenic may be naturally occurring in Ohio soils at greater than 20 mg/kg. For example, an environmental study of three locations in Cuyahoga County performed for Ohio EPA (Weston 2012) showed arsenic ranged from 4.6–25.2 mg/kg (22.9 mg/kg excluding statistical outliers) in surface soil (0–2 ft bgs) and 5.3–34.8 mg/kg (22.6 mg/kg excluding statistical outliers) in subsurface soil (2–4 ft bgs). Also, Vosnakis and Perry (2009) published the results of arsenic concentration studies that included 313 samples of Ohio soil. Naturally occurring arsenic in these samples ranged from 1.6–71.3 mg/kg with 95th percentiles of 21.7 mg/kg in surface soil, 25.5 mg/kg in subsurface soil, and upper tolerance limits of 22.8 mg/kg for surface soil and 29.6 mg/kg for subsurface soil. In other studies, native soil concentrations of arsenic in Ohio have been reported as ranging from 0.5–56 mg/kg (Ohio EPA 1996), and the U.S. Geological Survey's Certificate of Analysis of the Devonian Ohio Shale estimates arsenic concentrations of 68.5 mg/kg are naturally present in bedrock shale (USGS 2004).

The MDC of arsenic in sediment in the East Ditch was 19.7 mg/kg. The RVAAP background concentration for arsenic in sediment is 19.5 mg/kg. All of the other sediment samples from the East Ditch and West Ditch were below background concentrations. Because the reported concentration in one sediment sample is essentially the same as the background concentration, and all other samples are less than background concentrations, arsenic in sediment appears to be present at Load Line 11 at naturally occurring concentrations.

Organic chemicals were not screened against background concentrations even though some organic compounds are present in the environment as a result of natural or human activities not related to the CERCLA releases at the AOC. For example, PAHs are present in the environment as a result of burning fossil fuels and as a component of road dust, vehicle exhaust, tire wear particles, pavement, and slag used as railroad ballast and fill. Samples collected near roadways or parking areas may represent normal "urban" sources of PAHs. These issues represent significant sources of uncertainty at sites where low levels of PAHs are found over large areas of the AOC. At Load Line 11, PAHs were detected across the entire AOC, as one or more PAHs were detected in 22 of 28 surface soil

samples analyzed for SVOCs. PAH concentrations were less than the Resident Receptor (Adult and

- 2 Child) FWCUGs in all but four sample locations (LL11cs-020, LL11sb-060, LL11sb-061, and
- 3 LL11sb-067) where concentrations were up to two times the FWCUG of 0.221 mg/kg for

4 benzo(a)pyrene.

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- 6 Although no background concentrations for PAHs were established for RVAAP, the Phase II
- 7 Remedial Investigation Report for Winklepeck Burning Grounds (USACE 2001b), which established
- 8 the background concentrations for inorganics, included a characterization of naturally occurring
- 9 background metal concentrations in surface and subsurface soil at Camp Ravenna using samples from
- outside the process areas. Surface soil samples were collected at 15 locations on the eastern half of
- 11 Camp Ravenna. These background locations were chosen using aerial photographs and site visits with
- 12 the concurrence of Ohio EPA and USACE to reflect areas not impacted by RVAAP activities and
- establish background values that are unaffected by any human activity. The background locations
- were situated upgradient and generally upwind of known or suspected contaminant sources.

- 16 Background sampling was conducted in April and May 1998. All background samples were analyzed
- 17 for TAL metals, cyanide, and SVOCs. Two of the background samples were also analyzed for VOCs
- and pesticides/PCBs. The background soil sampling effort established concentrations for naturally
- 19 occurring metals in soil at RVAAP.
- 20 In establishing the background concentrations for naturally occurring metals, data were screened to
- 21 identify outliers in the inorganic chemical results. Ohio EPA guidance (Comment Resolution
- Meeting, December 2, 1998) called for using upper and lower cutoff limits based on quartiles to
- 23 identify outliers. The upper cutoff limit is the third quartile (75th percentile) plus one and a half times
- 24 the interquartile range. All results that exceeded the upper cutoff limit were examined to determine if
- 25 the results should be used in establishing the background concentrations for naturally occurring
- 26 metals. Outliers were removed so that background values would most nearly represent natural
- 27 conditions and exclude human disturbance whether from RVAAP or pre-RVAAP activities.
- 28 Statistical outliers were identified in 5 surface soil and 15 subsurface soil samples. All analytical
- 29 results for four of these samples [BKGss-011(b)-0794-SO, BKGss-012(b)-0795-SO, BKGss-015(b)-
- 30 0798-SO, and BKGss-005(b)-0788-SO] were removed from the surface soil background data set.
- The primary reason for eliminating these four samples from the surface soil background data set was
- 32 that PAHs were elevated and these sampling locations were near pre-existing homes or farms and
- 33 could have been influenced by activities associated with those structures (e.g., burning wood and
- fossil fuels, vehicle exhaust, or building materials such as slag used as fill or tar paper and shingles).
- 35 The other 11 outlier samples were not excluded from the background calculations primarily because
- 36 no SVOCs were detected in those samples and thus the outliers did not appear to be associated with
- 37 human activities.
- 38 Concentrations of PAHs in surface soil at Camp Ravenna associated with pre-RVAAP anthropogenic
- 39 sources calculated from the 15 RVAAP background surface soil samples are shown in Table 7-10.
- 40 Since the purpose here is to identify PAH levels associated with anthropogenic activities unrelated to
- 41 CERCLA releases from operations at RVAAP, it is appropriate to include all 15 background samples

in these calculations. The following criteria were used per the method used in establishing the background concentrations for naturally occurring metals (USACE 2001b):

• For analytes with a frequency of detection greater than 50%, a distribution (determined using the Shapiro-Wilk test) that is neither normal nor log-normal, and a sample size of 59 or less, the maximum result represents the nonparametric 95% upper tolerance limit and was identified as the background concentration for naturally occurring metals. These conditions apply to four of the PAHs detected in background samples [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and chrysene].

 For analytes with a frequency of detection between 0 and 50% with a sample size of 15, the maximum result represents the 99th percentile value and was identified as the background concentration for naturally occurring metals. These conditions apply to the remainder of the PAHs detected in background samples.

These results demonstrate the large variability in environmental concentrations of PAHs. For example, benzo(a)pyrene was detected in 8 of 15 background surface soil samples at concentrations ranging from 0.058–3.7 mg/kg.

Other studies of environmental concentrations of PAHs in Ohio soils show similar variability. For example, in the environmental study of three locations in Cuyahoga County performed for Ohio EPA (Weston 2012), PAHs were detected in only 1 of 36 surface soil samples with a reported concentration of benzo(a)pyrene of 1.33 mg/kg. Aerial photographs indicate this sample was collected near an old road or trail, but no other sources of PAHs are apparent.

In addition to these RVAAP and Cuyahoga County studies, numerous other environmental studies have been conducted that examine environmental levels of PAHs in rural and urban surface soil (e.g., ATSDR 1995, Bradley et al. 1994, IEPA 2005, MADEP 2002, and Teaf et al. 2008). Reported minimum, maximum, and 95th percentile concentrations of benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and dibenz(a,h)anthracene from numerous studies are shown in Table 7-11. These studies further demonstrate the high variability in environmental levels of PAHs within a single study area and among multiple studies.

The lack of established RVAAP background concentrations for identifying SRCs for PAHs is a source of uncertainty. Evaluating potential RVAAP process-related sources and other common anthropogenic (non-CERCLA) sources using available PAH environmental data minimizes the impact of this uncertainty on the conclusions of the RI (see Section 7.2.6).

Exposure Point Concentrations. Generally, the 95% UCL on the arithmetic mean was adopted as the EPC for discrete sample results and was considered to represent a conservative estimate of the average concentration. This imparts a small but intentional conservative bias to the risk assessment, provided the sampling captured the most highly contaminated areas. Representative EPCs for the EU were calculated from discrete data based on the assumption that the samples collected from the EU were truly random samples. This assumption is not true for Load Line 11 where sample locations

were biased to identify areas of highest contaminant concentrations. Therefore, EPCs generated from these data are likely to represent an upper bound of potential exposure concentrations.

In addition to calculating EPCs for each EU, individual discrete sample results above FWCUGs were evaluated to identify whether potential hot spots are present as a result of specific source areas.

Arsenic - The EPCs of arsenic range from 14.1–16.3 mg/kg and are less than RVAAP-specific background concentrations but reported concentrations exceed RVAAP-specific background at a few sample locations. The MDCs of arsenic in surface soil at Load Line 11 are 30.2 mg/kg at the FPA and 40.1 mg/kg at the NPA. The MDCs of arsenic in subsurface soil are 25.8 mg/kg at the FPA and 44.1 mg/kg at the NPA. The RVAAP-specific background value for arsenic in subsurface soil is 19.8 mg/kg. The subsurface background value is applicable for comparison to surface and subsurface results at Load Line 11 due to the extensive soil disturbance and mixing during building demolition and IRA activities at the Load Line. As previously described, naturally occurring arsenic concentrations in Ohio soils are often measured at greater than 20 mg/kg and the U.S. Geological Survey's Certificate of Analysis of the Devonian Ohio Shale estimates arsenic concentrations of 68.5 mg/kg are naturally present in bedrock shale (USGS 2004).

Individual arsenic results greater than 19.8 mg/kg are described below.

Surface Soil at the FPA – 56 surface soil samples (including field duplicates) were collected at the FPA. Reported concentrations in 49 of these samples range from 2.7–18.7 mg/kg. Reported concentrations in seven samples ranged from 20.2–30.2 mg/kg. Field duplicate results are available for four of these seven samples. Arsenic concentrations in these field duplicates ranged from 13.2–18.7 mg/kg (i.e., less than 19.8 mg/kg). These seven samples are scattered across the FPA and arsenic concentrations in nearby samples are less than background.

• Subsurface Soil at the FPA - 72 subsurface soil samples (including field duplicates) were collected at the FPA. Reported concentrations in 66 of these samples range from non-detect to 19.5 mg/kg. Reported concentrations in six samples collected in 2000 and 2001 ranged from 20–25.8 mg/kg. Four of these six samples (LL11ss-002, LL11sb-043, LL11ss-003, and LL11sb-009) are scattered across the FPA and arsenic concentrations in nearby samples are less than background. Two of these samples (LL11sb-042 and LL11sb-022) are near each other; however, additional samples collected near and adjacent to these two samples, including confirmation samples collected following sump removal performed during the IRA (MKM 2004a), have arsenic concentrations less than 19.8 mg/kg.

• Surface Soil at the NPA - 32 surface soil samples (including field duplicates) were collected at the NPA. Reported concentrations in 31 of these samples range from non-detect to 18.8 mg/kg. The reported concentration in one sample (LL11ss-070) was 40.4 mg/kg. This sample was collected from a drainage ditch at the road leading to Building AP-18.

• Subsurface Soil at the NPA - 51 subsurface soil samples (including field duplicates) were collected at the NPA. Reported concentrations in 43 of these samples range from non-detect to 19.7 mg/kg. Reported concentrations in eight samples ranged from 20.1–44.1 mg/kg. Three of these eight samples (LL11sb-065, LL11ss-028, LL11ss-031) with arsenic concentrations

ranging from 20.1–24.7 mg/kg are scattered across the NPA and arsenic concentrations in nearby samples are less than background. An additional three samples (LL11cs-050, LL11sb-039, LLsb-040) with arsenic concentrations of 22.3–33.6 mg/kg were collected near each other near the site of the hotspot removal of petroleum-contaminated soil north of Building AP-17 conducted during the 2001 IRA. Additional samples collected near and adjacent to these three samples have arsenic concentrations less than 19.8 mg/kg. Two samples (LL11cs-045 and its field duplicate) with arsenic concentrations of 44.1 and 36.4 mg/kg are also surrounded by several samples with arsenic concentrations less than 19.8 mg/kg.

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While several discrete soil samples had arsenic concentrations greater than 19.8 mg/kg, these results do not represent hot spots because:

- With the exception of three samples collected from the area near the site of the hotspot removal of petroleum-contaminated soil north of Building AP-17, these samples are scattered widely across the 48-acre AOC (i.e., they are not clustered) and are surrounded by samples with lower (less than 19.8 mg/kg) arsenic concentrations.
- Three samples from the area near site of the hotspot removal are clustered together; however, there are several other samples in this same area with lower concentrations.
- All samples having arsenic concentrations greater than the RVAAP-specific background concentration of 19.8 mg/kg were evaluated; however, several studies indicate arsenic may be naturally occurring in Ohio soils at greater than 20 mg/kg. For example, an environmental study of three locations in Cuyahoga County performed for Ohio EPA (Weston 2012) showed arsenic concentrations as high as 34.8 mg/kg and the U.S. Geological Survey's Certificate of Analysis of the Devonian Ohio Shale estimates arsenic concentrations of 68.5 mg/kg are naturally present in bedrock shale (USGS 2004).

Based on this information, arsenic appears to be present at Load Line 11 at naturally occurring concentrations.

The facility-wide background concentrations for arsenic at RVAAP are 15.4 mg/kg for surface soil and 19.8 mg/kg for subsurface soil; however, earthmoving (e.g., building removal, grading, and other building/demolition activities) has blurred the distinction between surface and subsurface soil. Vosnakis and Perry (2009) recently published the results of arsenic background studies that included 313 samples of Ohio soil. Naturally occurring arsenic in these samples ranged from 1.6–71.3 mg/kg with 95th percentiles of 21.7 mg/kg in surface soil 25.5 mg/kg in subsurface soil, and upper tolerance limits of 22.8 and 29.6 mg/kg for surface and subsurface soil, respectively. In other studies, native soil concentrations of arsenic in Ohio have been reported as ranging from 0.5-56 mg/kg (Ohio EPA 1996) and the U.S. Geological Survey's Certificate of Analysis of the Devonian Ohio Shale estimates arsenic concentrations of 68.5 mg/kg are naturally present in bedrock shales (USGS 2004). Based on this evaluation, arsenic does not represent a hot spot and was not identified as a COC.

 Benzo(a)pyrene – The EPC of benzo(a)pyrene at the FPA (0.219 mg/kg) is less than the FWCUG of 0.221 mg/kg but the concentrations in three samples exceed the FWCUG. The MDC of benzo(a)pyrene in surface soil at Load Line 11 is 0.45 mg/kg at LL11sb-060 collected in 2010 from

the ditch line immediately east of the entry road to Load Line 11, upstream of the FPA. Few samples were collected in this area where no activities took place. No benzo(a)pyrene was detected in the only other sample in this area (LL11sb-001). A benzo(a)pyrene concentration of 0.39 mg/kg was reported in sample LL11cs-020 collected in 2001 from the drainage ditch north of Building AP-14 as a confirmation sample following the IRA for removal of metals, SVOCs, and pesticides. Benzo(a)pyrene concentrations in the other four confirmation samples collected from this IRA area (LL11cs-016, -017, -018, and -019) range from non-detect to 0.2 mg/kg. A benzo(a)pyrene concentration of 0.31 mg/kg was reported in sample LL11sb-061 collected in 2010 the area of the former Building AP-8. Building AP-8 was used for primer loading and as an administrative building. No other samples collected from the footprint of this building were analyzed for SVOCs because they were not previously used at this building. Benzo(a)pyrene was not identified as a COPC in subsurface soil at the FPA.

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The EPC of benzo(a)pyrene at the NPA (0.28 mg/kg) is the MDC. The next highest detected concentration was 0.11 mg/kg. The MDC was reported sample LLsb-067 collected North of Building AP-17 in the area of the IRA to remove a hot spot of petroleum-contaminated soil. Benzo(a)pyrene was detected in one of the six confirmation samples collected following this removal action at a concentration of 0.11 mg/kg. Thus, the previously identified petroleum hot spot in this area has been successfully remediated.

7.2.5.2 Uncertainty in Use of FWCUGs

Sources of uncertainty in the FWCUGs used to identify COCs include selecting appropriate receptors and the exposure parameters, exposure models, and toxicity values used in calculating FWCUGs.

Selection of Representative Receptors. Load Line 11 is not currently used for training. While residential Land Use is unlikely, an evaluation using Resident Receptor (Adult and Child) FWCUGs is included to provide an Unrestricted (Residential) Land Use evaluation as required by CERCLA and the Army. As stated in Paragraph 6.d of the Technical Memorandum, if an AOC fails to meet the Unrestricted (Residential) Land Use, then all three Land Uses [i.e., Unrestricted (Residential) Land Use, Military Training Land Use, and Commercial/Industrial Land Use] will be evaluated.

Exposure Parameters and Exposure Models. For each primary exposure pathway included in the FWCUGs, assumptions are made concerning the exposure parameters (e.g., amount of contaminated media a receptor can be exposed to and intake rates for different routes of exposure) and the routes of exposure. Most exposure parameters have been selected so that errors occur on the side of human health protection. When several of these upper-bound values are combined in estimating exposure for any one pathway, the resulting risks can be in excess of the 99th percentile and, therefore, outside of the range that may be reasonably expected. Therefore, the consistent selection of upper-bound parameters generally leads to overestimation of the potential risks.

Toxicity Values. The toxicity of chemicals is under constant study and values change from time to time. The toxicity values used in calculating the FWCUGs were the most recent values available at

the time (September 2008). These values are designed to be conservative and provide an upper-bound estimate of risk.

The toxicity and mobility of many inorganic chemicals in the environment is dependent on the chemical species present. Two important examples are arsenic and chromium. The toxicity values used in developing the FWCUGs are for inorganic arsenic and do not distinguish between arsenite and arsenate.

Chromium is generally present in the environment as either the trivalent (Cr+3) or hexavalent (Cr+6) species, with the trivalent form generally being more stable and, therefore, more common. FWCUGs are available for hexavalent chromium and trivalent chromium. Trivalent chromium has not been shown to be carcinogenic. It is an essential micronutrient but can also be toxic at high doses (i.e., above the RfD used to calculate the FWCUG). The FWCUGs for trivalent chromium are based on non-cancerous effects. Hexavalent chromium is much more toxic than trivalent chromium. It is classified as a "known human carcinogen" and may also cause non-cancerous effects. The cancer URF for hexavalent chromium published in USEPA's Integrated Risk Information System (IRIS) is based on epidemiological data on lung cancer in workers associated with chromate production. Workers in the chromate industry are exposed to trivalent and hexavalent compounds of chromium. The cancer mortality in the study used to establish the URF was assumed to be due to hexavalent chromium. It was further assumed that hexavalent chromium constituted no less than 1/7 of the total chromium in air that the workers were exposed to. As noted in IRIS, the assumption that the ratio of hexavalent to trivalent chromium was 1:6 in this study may lead to a sevenfold underestimation of risk when using this URF to evaluate exposure to hexavalent chromium alone.

To avoid the underestimation of risk, the selection of the FWCUG for chromium includes a step that compares the maximum concentration of hexavalent chromium detected in chromium speciation samples to the residential RSL for hexavalent chromium. The detected concentrations of hexavalent chromium in the chromium speciation samples are less than the residential RSL for hexavalent chromium, and support using the trivalent chromium FWCUGs for evaluating total chromium results. Using speciation samples to identify the appropriate FWCUG minimizes the associated uncertainty.

No FWCUGs or RSLs were available for TPH-DRO, TPH-GRO, sulfate, and sulfide in soil and/or sediment because no toxicity values are available for calculating FWCUGs. These COPCs were evaluated qualitatively based on comparison to SLs available for other chemicals. A qualitative evaluation of these chemicals indicates the lack of screening values does not represent a significant source of uncertainty as described below.

• After removing the 30 by 30 by 8 ft area of the petroleum-contaminated soil hot spot, confirmation samples were collected. Of the six confirmation samples collected from that excavation footprint, only two samples had petroleum hydrocarbon fractions (TPH-DRO and TPH-GRO) detected at low concentrations in deep surface soil. TPH-DRO and TPH-GRO were identified as COPCs at the NPA because no screening values (FWCUG or RSL) are available. TPH is a mixture of compounds. Many of the compounds included in the reported TPH fractions are also measured individually (e.g., VOCs such as benzene and SVOCs such

as PAHs). All individual VOCs and SVOCs associated with TPH were either not detected or were detected below FWCUGs. Some states have developed screening values for TPH. Ohio EPA has published toxicity values for TPH using guidance from other states, including Massachusetts and Louisiana (Ohio EPA 2004b). Louisiana Department of Environmental Quality has published risk-based screening values for TPH-DRO (65 mg/kg) and TPH-GRO (65 mg/kg) in soil (LDEQ 2000). The MDCs for TPH-DRO (34 mg/kg) and TPH-GRO (0.054 mg/kg) at Load Line 11 are below the Louisiana risk-based screening values for these fractions. Therefore, TPH-DRO and TPH-GRO are not COCs for soil.

• Sulfide was identified as a COPC in the FPA and NPA surface and subsurface soil and West Ditch sediment as a COPC and sulfate was identified as a COPC in the FPA and NPA subsurface soil because no screening values (FWCUG or RSL) are available. Sulfate and sulfide have low toxicity to humans, and appropriate toxicity data specific to these anions are not available for calculating FWCUGs. Soil RSLs are available for these anions in combination with other inorganic molecules. Specifically, RSLs are available for hydrogen sulfide (2,800,000 mg/kg) and selenium sulfide (390 mg/kg). The sulfate and sulfide EPCs for surface (0-1 ft bgs) and subsurface soil (1-13 ft bgs) range from a non-detectable concentration to 30.6 mg/kg for sulfate and 27.8-65.3 mg/kg for sulfide. The EPC for sulfate in sediment was 43.4 mg/kg. These EPCs are well below the RSLs available for sulfide combined with other inorganic molecules; therefore, sulfide is not a COC in soil or sediment. Although toxicity data do not exist for sulfate, it is present in naturally occurring minerals contained in soils. As sulfate is naturally occurring and there is no source (e.g., history of agriculture, mine drainage, urban runoff, or industrial processes generating sulfate) of sulfate contamination at Load Line 11, sulfate was eliminated from consideration as a COC.

FWCUGs Below Background Concentrations. One purpose of the HHRA process is to identify COCs and cleanup goals for evaluating remedial alternatives for remediating residual contamination that has resulted from process operations at the AOC. The FWCUGs are risk-based values. In some cases, natural or anthropogenic background concentrations, unrelated to process operations, exceed the risk-based FWCUGs. For naturally occurring inorganic chemicals, this problem is addressed by using the background concentration as the cleanup goal. This introduces uncertainty in the chosen cleanup goal because there is uncertainty in assigning a specific value to background, which can be highly variable.

No background concentrations are available for organic chemicals, although PAHs are often present in the environment from natural and anthropogenic sources and regulatory standards are often much lower than baseline levels of PAHs in urban and rural surface soil, especially near areas of vehicle traffic (e.g., roads and parking areas). Given their frequent presence in environmental media, and especially in areas influenced by vehicle exhaust and tire particles, it is important to compare risk-based cleanup levels with typical environmental concentrations before utilizing unrealistically low cleanup targets. Numerous studies have been conducted that examine ambient levels of PAHs in rural and urban surface soil (e.g., ATSDR 1995, Bradley et al. 1994, MADEP 2002, and Teaf et al. 2008). These studies indicate that given the multitude of non-point mobile sources for PAHs, it is not uncommon for ambient concentrations to exceed health-based regulatory recommendations. Some states have begun to consider ambient anthropogenic levels by establishing minimum SLs based on

environmental studies. For example, the New York State Department of Environmental Conservation has established a minimum soil cleanup objective of 1 mg/kg for benz(a)anthracene, benzo(a)pyrene, and benzo(b)fluoranthene and 0.1 for dibenz(a,h)anthracene based on the 95th percentile concentrations of these PAHs in rural areas near roads (NYSDEC 2006).

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7.2.5.3 Uncertainty in the Identification of COCs

All of the sources of uncertainty described in the previous sections potentially impact the identification of COCs. The exposure and toxicity values used to calculate FWCUGs as well as the approach for identifying SRCs, COPCs, and ultimately COCs based on the FWCUGs were designed to ensure the over rather than underestimation of potential risk. The uncertainty assessment attempts to put the identified COCs in perspective to facilitate informed risk management decisions for the AOC.

The SOR is used to account for the potential additive effects from exposure to multiple chemicals that can cause the same critical effect or affect the same target organ. Cancer risk is assumed to be additive for all carcinogens. Non-cancer risk is assumed to be additive for chemicals with similar sites of toxicological action. In the event that any combination of COPCs results in synergistic effects, risk might be underestimated. Conversely, the assumption of additivity would overestimate risk if a combination of COPCs acted antagonistically. It is unclear whether the potential for chemical interaction has been inadvertently understated or overstated. It seems unlikely that the potential for chemical interaction contributes significant uncertainty to the conclusions of the risk assessment.

7.2.6 Identification of COCs for Potential Remediation

As indicated in Section 4.2, an IRA was performed in 2001 to remove contamination in primary migration pathways at Load Line 11. The IRA was performed in parallel with the Phase I RI activities described in Section 4.3. This IRA included removing sump water from production buildings, grouting selected sanitary sewer manholes, excavating contaminated drainage ditches, and removing a hot spot petroleum-contaminated soil.

Using data and information that reflects the current status of Load Line 11, COCs were identified in Section 7.2.4 as any COPC having an EPC greater than an applicable FWCUG or contributing more than 5–10% to an SOR greater than one. For inorganic chemicals with FWCUGs or RSLs below background concentrations, the background concentration was used as the point of comparison. The TR for the FWCUGs used to identify COCs is 1E-05 per the Ohio EPA DERR program, which has adopted a human health cumulative ILCR goal of 1E-05 to be used as the level of acceptable excess cancer risk and for developing site remediation goals.

The results of the COC screening (Section 7.2.4) are combined with the results of the uncertainty assessment (Section 7.2.5) to identify COCs to be carried forward for potential remediation.

No COCs were identified in subsurface soil or sediment based on comparison to FWCUGs and the uncertainty assessment (including evaluation of individual data points).

COCs for Surface Soil (0–1 ft bgs) for Potential Remediation

Five PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] were identified as COCs for the Resident Receptor (Adult and Child) in surface soil. The EPC for benzo(a)pyrene in surface soil at the NPA (0.28 mg/kg) is approximately 1.3 times its FWCUG of 0.221 mg/kg. The EPC for this data aggregate is the MDC (at sample LL11sb-067); benzo(a)pyrene concentrations in the remaining NPA samples ranged from a non-detectable concentration to 0.11 mg/kg. Sample LL11sb-067 was collected in 2010 from the northeastern portion of Load Line 11. The EPC for benzo(a)pyrene in the FPA (0.219 mg/kg) was slightly less than the FWCUG (0.221 mg/kg). Other PAHs were identified as COCs because they contribute to SORs of two at the FPA and NPA. The EPCs of benzo(a)pyrene (0.28 and 0.221 mg/kg) are slightly above and below the FWCUG of 0.221 and the SORs are two, indicating cumulative cancer risks to a hypothetical future resident on the order of 1E-05. Reported benzo(a)pyrene concentrations above the FWCUG are scattered (i.e., no hot spots were identified); therefore, no COCs are identified for potential remediation.

COCs for Surface Water for Potential Remediation

Four PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] were identified as COCs for the Resident Receptor (Adult and Child) in surface water collected from the West Ditch. The detected concentrations of benz(a)anthracene (0.00035 mg/L), benzo(b)fluoranthene (0.00041 mg/L), and indeno(1,2,3-cd)pyrene (0.00021) in surface water collected from the West Ditch range from two to four times their FWCUG of 0.0001 mg/L, and the detected concentration of benzo(a)pyrene (0.00033 mg/L) was 41 times the FWCUG of 0.000008 mg/L. All reported concentrations included a "J" qualifier, indicating they were estimated results below the detection limit of the analytical method. PAHs are present in the environment from natural and anthropogenic sources. Since surface runoff is the potential source of contamination to surface water in the West Ditch, the fact that PAH concentrations are not significantly elevated in soil or sediment indicates there is no identifiable source beyond normal levels of these chemicals to surface water due to runoff from roads and other traffic areas. In addition, incidental exposures of the Resident Receptor (Adult and Child) to surface water in the West Ditch that only intermittently holds water are much less than the ingestion of drinking water (i.e., 2 L/day for an adult) and dermal contact while swimming and wading exposures incorporated into developing the FWCUGs. Thus, these low, estimated concentrations of PAHs were not identified as COCs for remediation in surface water.

7.2.7 Summary of HHRA

This HHRA documents COCs that may pose potential health risks to human receptors resulting from exposure to contamination at Load Line 11. This HHRA was conducted as part of the RI and was based on the streamlined approach described in the FWCUG Report (USACE 2010a), Position Paper for Human Health FWCUGs (USACE 2012b), and Technical Memorandum (ARNG 2014). The components of the risk assessment (receptors, exposure media, EPCs, and results) are summarized below.

Page 7-25

Receptors. Camp Ravenna is a controlled access facility. Load Line 11 is located in the south-central portion of the facility and is not currently used for training. An evaluation using Resident Receptor (Adult and Child) FWCUGs is used to provide an Unrestricted (Residential) Land Use evaluation as required by CERCLA and the Army. Unrestricted (Residential) Land Use is considered protective for all categories of Land Use on the former RVAAP/Camp Ravenna.

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Exposure Media. Media of concern at Load Line 11 are surface soil, subsurface soil, surface water, and sediment. Soil data associated with Load Line 11 were aggregated into surface and subsurface soil at the FPA and NPA.

Estimation of EPCs. The EPCs for soil were calculated from the results of all of the discrete samples collected from a given depth interval at each EU. The EPC was either the 95% UCL of the mean or the MDC, whichever value is lowest. The limited number of sediment and surface water samples were evaluated individually (i.e., the number of samples was too small to calculate 95% UCLs).

Results of Human Health Risk Assessment. No COCs for evaluation in an FS were identified for soil, sediment, or surface water at Load Line 11.

7.3 ECOLOGICAL RISK ASSESSMENT

7.3.1 Introduction

The ERA presented in this RI Report follows a unified approach of methods integrating Army, Ohio EPA, and USEPA guidance. This ERA approach is consistent with the general approach by these agencies and primarily follows the Level I Scoping ERA, Level II Screening ERA, and Level III Baseline ERA outlined in the *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA 2008), with specific application of components from the FWERWP (USACE 2003a), *Risk Assessment Handbook Volume II: Environmental Evaluation* (USACE 2010b), and *Ecological Risk Assessment Guidance for Superfund: Process for Designing and Conducting Ecological Risk Assessments* (USEPA 1997). The process implemented in this RI Report combines these guidance documents to meet requirements of the Ohio EPA and Army, while following previously accepted methods established for RVAAP. This unified approach resulted from coordination between USACE and Ohio EPA during the summer of 2011.

7.3.1.1 Scope and Objective

- Load Line 11 contains habitat that supports ecological receptors. The habitat has known chemical contamination (MKM 2005a). Habitat types and an assessment of the ecological resources found at
- 39 Load Line 11 are presented in subsequent subsections.

Additionally, the conclusions of a historical ERA [SERA performed as part of the Phase I RI (MKM 2005a)] state the following:

"No additional action is recommended based on ecological risk because, based on the Step 3a refinement step, it was determined that site-related risks were not great enough to warrant proceeding further into the ecological risk assessment process."

The historical ERA and the PBA08 RI results were combined for this RI Report to evaluate the conclusion of the historical ERA to determine whether a qualitative ERA (Level I) is sufficient, based on the quality of the habitat and the presence of contamination; or whether a more rigorous ERA (Level II or III) should be conducted.

7.3.2 Level I: Scoping Level Ecological Risk Assessment

The ERA method for Level I follows guidance documents listed in Section 7.3.1. Level I is intended to evaluate if the AOC had past releases or the potential for current contamination and if there are important ecological resources on or near the AOC.

The following two questions should be answered when the Level I ERA is complete:

1. Are current or past releases suspected at the AOC? Current or past releases are determined by evidence that chemical contaminants or COPECs are present.

2. Are important ecological resources present at or in the locality of the AOC? Important ecological resources are defined in the *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA 2008) and the *Technical Document for Ecological Risk Assessment: Process for Developing Management Goals* (BTAG 2005).

If an AOC has contaminants but lacks important ecological resources, this indicates the ERA process can stop at Level I. Contamination and important ecological resources must both be present to proceed to a Level II Screening Level ERA.

7.3.2.1 AOC Description and Land Use

Load Line 11 is approximately 48 acres. The buildings within the Load Line 11 fence line have been removed. The habitat is mostly field and shrubland with some forest and is large enough to completely support cover and food for small birds and mammals that typically require approximately 1 acre of habitat (USEPA 1993). The habitat area at Load Line 11 represents 0.22% of the 21,683 acres at Camp Ravenna.

Future use at Load Line 11 is anticipated to be within the Military Training or Commercial/Industrial
Land Use scenarios.

7.3.2.2 Evidence of Historical Chemical Contamination

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The 1978 Installation Assessment identified the major contaminants of the former RVAAP as TNT, composition B (a combination of TNT and RDX), sulfates, nitrates, lead styphnate, and lead azide. Additional site-specific contaminants at Load Line 11 include black powder, TNT, RDX, HMX, nitroglycerine, nitrocellulose, nitroguanidine, heavy metals (lead, chromium, mercury, and arsenic) from munitions assembly activities, VOCs from storing solvents (Building AP-17) and lacquer sealing materials used on finished primers (Building AP-11), and PCBs from on-site transformers. The evaluation of historical chemical contamination is not limited to these chemicals; rather, this evaluation is expanded to include all eligible chemical data that is available.

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As indicated in Section 4.2, in 2001 an IRA was performed as an early response to contamination in primary pathways at Load Line 11. This IRA was conducted in unison with the Phase I RI. This IRA included removing sumps, 230 yd³ of contaminated media in drainage ditches, and a hot spot of petroleum-contaminated soil (MKM 2004a). Confirmation samples from these source removal activities are included in this risk assessment, as applicable.

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The goal of the historical ERA (MKM 2005a) was to identify COPECs in soil, sediment, and surface water for Load Line 11 that existed after the IRA was conducted. Bioaccumulative compounds were identified and HQs based on food chain models were calculated in the Phase I RI (MKM 2005a); this will be discussed if this ERA continues to Level III. The Phase I RI followed instructions presented in the Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2003) and included the first two of six steps listed in Figure III of the FWERWP (USACE 2003a). These two steps identified the evaluation procedures, which were used to determine AOC-related COPECs. First, the MDC of each chemical was compared to its respective facility-wide background concentration. Chemicals were not considered COPECs if the MDC was below the background concentration. For all chemicals detected above background concentrations, the MDC was compared to an ESV. The hierarchy of screening values was based on the guidance included in the FWERWP and Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2003). In addition to the ESV comparison, it was determined if the chemical was a persistent, bioaccumulative, and toxic (PBT) compound. For sediment, chemicals were also compared with the Ohio Sediment Reference Value (SRV). Chemicals were retained as COPECs if they exceeded background concentrations and the ESV, if the chemical exceeded background concentrations and had no toxicity information, or if the chemical was considered a PBT compound. Groundwater was not included in the historical ERA.

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The historical ERA tables for soil, sediment, and surface water are provided in Appendix H, Tables H-1, H-2, and H-3 and include the following applicable components to the Level I assessment:

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- Frequency of detection,
- Minimum concentration,
- 41 MDC,
 - Mean concentration,
 - Average of positive detects,
 - Sample of maximum detect,

- Background concentration,
- Comparison of MDC to background (SRC determination),
- SRVs (sediment only),
 - Screening levels (ESVs),
 - HQ (MDC/ESV ratio),
 - PBT compound identification (soil and surface water only),
 - COPEC determination, and
 - COPEC rationale.

Historical COPECs for Soil. The historical ERA conducted as part of the Phase I RI reported 35 chemicals in surface soil (0-1 ft) at Load Line 11 (MKM 2005a). Of the 35 chemicals detected, 5 (aluminum, calcium, magnesium, potassium, and sodium) were essential nutrients and were excluded from the COPEC screen (Appendix H, Table H-1). A total of 16 inorganic chemicals, 2 anions, and 10 organic chemicals were determined to be SRCs because they exceeded background concentrations or did not have an associated background concentration for comparison. Nine inorganic chemicals (arsenic, chromium, cyanide, lead, manganese, mercury, nickel, vanadium, and zinc), one anion (sulfide), and one organic chemical (PCB-1254) exceeded their ESVs and were identified as COPECs (Table 7-12). Three chemicals (iron, nitrate, and nitrocellulose) were identified as COPECs due to a lack of ESVs. Two COPECs that exceeded their ESVs (mercury and PCB-1254) were also PBT compounds. Appendix H, Table H-1 presents the Phase I RI ecological screening for surface soil at Load Line 11.

Historical COPECs for Sediment. The historical ERA conducted as part of the Phase I RI reported 38 chemicals in sediment at Load Line 11 (MKM 2005a). Of the 38 chemicals detected, 4 chemicals (calcium, magnesium, potassium, and sodium) were essential nutrients and were excluded from the COPEC screen (Appendix H, Table H-2). A total of 6 inorganic chemicals, 3 anions, and 13 organic chemicals were determined to be SRCs because they exceeded background concentrations and the SRV or they did not have an associated background concentration for comparison. Of the 22 SRCs, 3 inorganic chemicals (copper, lead, and nickel) and 9 organic chemicals [benz(a)anthracene, benzo(a)pyrene, benzo(ghi)perylene, chrysene, fluoranthene, indeno(1,2,3-cd)pyrene, phenanthrene, pyrene, and total PAHs] exceeded their ESVs and were identified as COPECs (Table 7-12). In addition, two inorganic chemicals (beryllium and manganese), three anions (nitrate, sulfate, and sulfide), and one organic chemical (nitrocellulose) were selected as COPECs because they did not have an ESV for comparison. Two detected chemicals were also PBT compounds (mercury and PCB-1254). Mercury was eliminated as a COPEC because it did not exceed its background concentration. PCB-1254 did not exceed its ESV but was retained as a COPEC because it is a PBT compound.

Historical COPECs for Surface Water. The historical ERA conducted as part of the Phase I RI reported 12 chemicals in surface water at Load Line 11 (MKM 2005a). Of the 12 chemicals detected, 4 (calcium, magnesium, potassium, and sodium) were essential nutrients and were excluded from the COPEC screen (Appendix H, Table H-3). Two inorganic chemicals, one anion, and two organic chemicals were determined to be SRCs because they exceeded background concentrations or did not have an associated background concentration for comparison. Of the five SRCs, one inorganic

chemical (manganese) and one anion (sulfate) were identified as COPECs (Table 7-12) because they did not have an ESV for comparison. No PBT compounds were detected in surface water.

Summary of Historical ERA. An IRA was performed for sumps, contaminated drainage ditches, and a hot spot of petroleum-contaminated soil at Load Line 11 as an early response to contamination in primary pathways. Using information and data from the IRA and Phase I RI, a historical ERA was performed to determine COPECs at Load Line 11 in surface soil, sediment, and surface water. Table 7-12 summarizes the COPECs by media. Based on the identified COPECs in the historical Level I ERA, ecological risk was predicted (MKM 2005a).

7.3.2.3 Ecological Significance

Sources of data and information about the ecological resources at Load Line 11 include the *Integrated Natural Resources Management Plan* (INRMP; OHARNG 2014), *Facility-Wide Biological and Water Quality Study* (USACE 2005a), previous characterization work (e.g., the Phase I RI), IRA (MKM 2004a), and visits to Load Line 11 conducted for the PBA08 RI.

One of the two key questions to answer in the Level I Scoping ERA is whether there are ecologically important and especially ecologically significant resources at Load Line 11. Ecological importance is defined as a place or resource that exhibits unique, special, or other attributes that makes it of great value. Ecological significance is defined as an important resource found at an AOC or in its vicinity that is subject to contaminant exposure. The underlying basis for this distinction can be found in *Ecological Significance and Selection of Candidate Assessment Endpoints* (USEPA 1996a), and is stated as follows:

"A critical element in the ERA process requires distinguishing important environmental responses to chemical releases from those that are inconsequential to the ecosystem in which the site resides: in other words, determining the ecological significance of past, current, or projected site-related effects."

Important places and resources identified by the Army and Ohio EPA (Appendix H, Table H-4) include wetlands, terrestrial areas used for breeding by large or dense aggregations of animals, habitat known to be used by threatened or endangered species, state land designated for wildlife or game management, locally important ecological places, and state parks. The Army and Ohio EPA recognize 17 important places and resources. The Army recognizes an additional 16 important places (BTAG 2005), and the Ohio EPA recognizes another 6 important places (Ohio EPA 2008). In total, there are 39 important places. Presence or absence of an ecologically important place can be determined by comparing environmental facts and characteristics of Load Line 11 with each of the important places and resources listed in Appendix H, Table H-4.

Presence of an important ecological resource or place and proximity to contamination at an AOC makes a resource ecologically significant. Thus, any important places and resources listed in Appendix H, Table H-4 are elevated to ecologically significant when present on the AOC and there is exposure to contaminants. For all 39 important places and resources, it is relatively clear that the

ecological place or resource is either present or absent on the AOC; therefore, the decision process is objective. If no important or significant resource is present at an AOC, the evaluation will not proceed to Level II regardless of the presence of contamination. Instead, the Level I Scoping ERA would acknowledge that there are important ecological places, but that those resources are not ecologically significant, and no further evaluation is required.

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Management Goals for the AOC. Regardless of whether the evaluation is concluded at Level I or continues to Level II, there is another level of environmental protection for Load Line 11 through the natural resource management goals expressed in the INRMP (OHARNG 2014). OHARNG manages the ecological and natural resources at Camp Ravenna to maintain or enhance the current integrity of the natural resources and ecosystems at the facility. Natural resource management activities in place at Camp Ravenna may also be applicable to any degradation noted from contamination.

Some natural resource management goals of OHARNG (listed in Appendix H, Table H-5) benefit Load Line 11. For example, Goal 1 states natural resources need to be managed in a compatible way with the military mission, and Goal 5 requires the Army to sustain usable training lands and native natural resources by implementing a natural resource management plan which incorporates invasive species management and utilizes native species mixes for revegetation after ground disturbance activities. These management goals help detect degradation (whether from training activities or historical contamination). While the applicability of the remaining 10 management goals to Load Line 11 varies, all of the management goals are intended to monitor, maintain, or enhance RVAAP natural resources and its ecosystem. While these goals are for managing all types of resources at and near Load Line 11, they do not affect the decisions concerning the presence or absence of important or significant ecological places or resources at Load Line 11.

Important Places and Resources. Ecological importance means a place or resource that exhibits a unique, special, or other attribute that makes it of great value. Examples of important places and resources include wetlands, terrestrial areas used for breeding by large or dense aggregations of animals, and habitat of rare species. An important resource becomes significant when found on an AOC and there is contaminant exposure. The wetlands are important/significant ecological resources at Load Line 11 (Appendix H, Table H-4).

Terrestrial Resources. Load Line 11 is dominated by terrestrial resources, as described below.

Habitat Descriptions and Species. The INRMP and AOC visits by SAIC scientists indicated Load Line 11 consists of five vegetation types (Figure 7-1). The habitat area is dominated by dry, mid-successional, cold-deciduous shrubland (Figure 7-1 and Photograph 7-1). Small areas of three forest types are represented along the western, northern, and eastern boundaries. The western boundary runs through a small amount of mixed, cold-deciduous, successional forest and American beech (Fagus grandifolia), oak (Quercus spp.), and maple (Acer spp.) forest alliance. The northern boundary runs through a small strip of green ash (Fraxinus pennsylvanica), American elm (Ulmus americana), and hackberry (Celtis occidentalis) temporarily-flooded forest alliance. The eastern boundary runs through a small amount of mixed, cold-deciduous, successional forest. The southern boundary runs through a small amount of a fourth forest type – American beech (Fagus grandifolia), sugar maple

(*Acer saccharum*), and American tulip tree (*Liriodendron tulipifera*) forest alliance. This characterization was originally established by a vegetation study using aerial photography and field verification (USACE 1999) and was later used in the INRMP (OHARNG 2008).

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SAIC scientists conducted a field survey at Load Line 11 on May 19, 2010 and determined there have been changes in vegetation at the AOC, including (1) the recent development (post 1999) of a dry, early-successional, herbaceous field habitat; (2) a decrease in dry, mid-successional, cold-deciduous shrubland habitat; and (3) small increases in the four forest habitats. The herbaceous habitat has developed since 1999 and is largely located within the loop road around the former load line. The demolition activities associated with removal of the buildings and other infrastructure at Load Line 11 also removed shrubland, which has been replaced by grassland. Dominant plants include an assortment of grasses, forbs, and seedlings of trees and shrubs. Common species include several species of goldenrod (*Solidago* spp.), clasping-leaf dogbane (*Apocynum cannabinum*), yarrow (*Achillea millefolium*), gray dogwood (*Cornus racemosa*), blackberry (*Rubus allegheniensis*), autumn olive (*Elaeagnus umbellata*), and multiflora rose (*Rosa multiflora*).

The decrease in shrubland is attributable to the removal actions and plant succession. In general, the lost shrubland habitat has been replaced by modest increases in the edge forest and central herbaceous field habitat types. The remaining shrubland areas occur along the boundary and the perimeter road that encircles the former load line. Shrubs and small trees remain the dominant habitat type at the AOC. Common species include various willows (*Salix* spp.), gray dogwood (*Cornus racemosa*), autumn olive (*Elaeagnus umbellata*), blackberry (*Rubus allegheniensis*), hawthorn (*Crataegus* spp.), and multiflora rose (*Rosa multiflora*).

The increase in forested area is attributed primarily to plant succession, as saplings in the shrubland habitat grow into more mature trees and other tree species colonize the habitat. Forested habitat has expanded across the knoll in the north-central portion of the AOC, almost to the perimeter road. There is also a small area of forested habitat near the locations of former Buildings AP-4, AP-7, and AP-8 (Figure 7-1). The forests include small open areas and understory that result in multi-story vegetation. The vegetation provides multiple layers for the various foraging height preferences of birds, mammals, insects, and other organisms.

The habitats at Load Line 11 were assessed to be healthy and functioning based on May 2008 and 2010 observations by SAIC biologists (Photographs 7-1 through 7-3). Functional habitat was determined by noting the absence of large bare spots and dead vegetation or other obvious visual signs of an unhealthy ecosystem. Additional habitat photographs are provided in Appendix H.

Threatened and Endangered and Other Rare Species. The northern long-eared bat (Myotis septentrionalis; federally threatened) exists at Camp Ravenna. There are no other federally listed species and no critical habitat on Camp Ravenna. Load Line 11 has not been previously surveyed for rare, threatened, or endangered species; however, there have been no documented sightings of rare, threatened, or endangered species at the AOC (OHARNG 2014).

Other Terrestrial Resources. While there are no other known important terrestrial places and resources (Appendix H, Table H-4), there are other resources at or near Load Line 11 (e.g., vegetation, animals) that interact in their ecosystems and support nutrient cycling and energy flow. For example, wildlife such as wild turkey (Meleagris gallopavo) and white-tailed deer (Odocoileus virginianus) could use the area. The INRMP provides information about species and habitat surveys at Camp Ravenna (e.g., timber and ecological succession) (OHARNG 2014). There are no other reported surveys of habitats and wildlife at Load Line 11 beyond those summarized in the INRMP (OHARNG 2014).





Photograph 7-1. Herbaceous Growth, Shrubland, and Forest in the Habitat Area (May 20, 2008)



Photograph 7-2. Portion of the East Ditch Network (May 19, 2010)



Photograph 7-3. Portion of the West Ditch Network Facing Northeast Across the Central Part of the AOC (May 19, 2010)

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Aquatic Resources. Load Line 11 has few aquatic resources. Although there are no ponds or streams at Load Line 11, there are two shallow drainage ditches and two small wetlands.

Habitat Descriptions and Species. The water features at Load Line 11 include the East and West Ditches (Figures 2-1 and 7-1). The East Ditch drains the northeastern portion of the AOC. The ditch is well-vegetated with shrubs and herbaceous plants, and there is no evidence of erosion. Despite several days of moderately heavy rain at the East and West Ditches, there was only a small amount of water flowing in the East Ditch during a May 19, 2010 site visit. This small amount of water flow was due to a small drainage area and a blocked culvert (or other impediment). Water from the East Ditch eventually discharges onto the Sand Creek floodplain near Newton Falls Road. The West Ditch drains the central portion of the AOC and flows southwest to an unnamed tributary to Sand Creek. The ditch is well-vegetated with a mixture of shrub and herbaceous species, and there is no evidence of erosion. Water in the West Ditch was flowing at the time of the site visit. The West Ditch is fed by a network of small ditches; most of the flow appeared to be coming from a small forested area between former Buildings AP-4, AP-7, and AP-8 rather than the main part of the AOC. Within the central portion of the AOC, water was present in the entire network of smaller ditches that feed into the West Ditch, but very little flow was evident.

Intermittent surface water flows in drainage ditches at Load Line 11. During several inspections and sampling visits over a three year period, SAIC scientists found no water in the West Ditch or East Ditch throughout most of the year. However, due to a blocked culvert or other impediment in the ditch, a very small wetland has developed in part of the East Ditch. Precipitation data from Camp Ravenna are provided in Section 3.5. The storm frequency is 35 days per year, and precipitation occurs 154 days per year; field observations have determined this is not sufficient to create and maintain aquatic habitat in the unblocked portion of the drainage ditches.

The May 19, 2010 site visit also included an inspection of an old sewer outfall in the north-central part of the AOC boundary. The outfall consisted of an 8-inch ceramic tile pipe partially blocked by leaves and other organic debris. There was no evidence that any sediment had been moving recently, because the opening of the pipe was effectively covered with leaves and other debris. Any discharge from the sewer outfall would eventually flow onto the Sand Creek floodplain. The outfall assessment was completed and provided in the *Remedial Investigation/Feasibility Study Report for RVAAP-67 Facility-Wide Sewers* (USACE 2012a).

Wetlands. Important wetland resources exist at the AOC. Wetlands are important habitats with water-saturated soil or sediment and plant life that can survive saturation. Wetlands are home to many different species and are also chemical sinks that can serve as detoxifiers and natural water purifiers. It is expected the wetlands at Load Line 11 perform these and other related functions.

A planning level survey [i.e., based on desktop surveys conducted for the OHARNG of wetlands data and resources [e.g., National Wetlands Inventory maps, aerials] identified one small wetland (Wetland 1) that straddles the western boundary of the AOC (Figure 7-1) (OHARNG 2014). There is one other wetland (Wetland 2) not previously identified on any wetlands surveys; it is associated with the East Ditch EU. Wetland 2 is not present on any of the wetland maps provided by the OHARNG

and was once occupied by a building. No jurisdictional wetlands determination has been conducted at this AOC. For any wetland at the AOC potentially affected by remedial activities, a jurisdictional determination by USACE would be required to determine the regulatory status.

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An SAIC Professional Wetland Scientist used the Ohio Rapid Assessment Method (ORAM) (Ohio EPA 2001) in December 2011 to assess the condition of the two wetlands within the habitat area for Load Line 11 and to determine the potential ecological importance to those wetlands (Appendix H, Figure H-1). Using ORAM, wetlands are classified into three categories:

- Category 1 wetlands are described as "limited quality waters." They are considered to be a resource that has been degraded, has limited potential for restoration, or is of such low functionality that lower standards for avoidance, minimization, and mitigation can be applied. Scores range from 1–29.
- Category 2 includes wetlands of moderate quality and wetlands that are degraded but exhibit reasonable potential for restoration. Scores range from 30–59.
- Category 3 includes wetlands of very high quality and wetlands of concern regionally and/or statewide, such as wetlands that provide habitat for species listed as threatened or endangered. Scores range from 60–100.

The field sheets detailing the ORAM at Load Line 11 are presented in Appendix H, Figures H-1 and H-2. Figure 7-1 shows the wetland locations. Based on the ORAM, Wetland 1 is classified as Category 1 (with a final score of 28), indicating low wetland quality, with degradation of wetland functions and conditions (Appendix H, Figure H-1). Wetland 2 is classified as Category 1 (with a final score of 21), indicating low wetland quality, with degradation of wetland functions (Appendix H, Figure H-2).

Wetland 1 is a small emergent/scrub-shrub planning level survey wetland that straddles the western boundary of Load Line 11. Wetland 1 is 0.24 acres, with 0.13 acres inside the AOC. The wetland is located on the western boundary of the AOC between the extension of the West Ditch and a small wet weather tributary to Sand Creek. Dominant vegetation includes silky dogwood (*Cornus amomum*), arrowwood (*Viburnum dentatum*), red maple (*Acer rubrum*), green ash (*Fraxinus pennsylvanica*), blackberry (*Rubus* spp.), cutgrass (*Leersia* spp.), and cattails (*Typha* spp.).

Wetland 2 is a very small emergent/scrub-shrub wetland (0.02 acres) that has formed in the center of the East Ditch EU. The wetland has likely formed as a result of blocked drainage in the ditch. Dominant vegetation includes silky dogwood, arrowwood, red maple, green ash, softstem bulrush (*Schoenoplectus tabernaemontani*), reed canary grass (*Phalaris arundinacea*), an unidentified bulrush (*Scirpus* spp.), cattails, cutgrass, and primrose willow (*Ludwigia* spp.). Wetland 2 has not been delineated.

Another wetland, located just outside the western and northern boundaries of the AOC is a moderately large, forested wetland, associated with the floodplains of Sand Creek and an unnamed tributary to Sand Creek. The wetland receives runoff from Load Line 11 via the East and West Ditches, some small conveyances/drainageways, and seeps along the northern boundary of the AOC.

Dominant vegetation includes black willow (*Salix nigra*), red maple, cottonwood (*Populus deltoides*), silky dogwood, golden ragwort (*Senecio aureus*), and poison ivy (*Toxicodendron radicans*).

Threatened and Endangered and Other Rare Species. The northern long-eared bat (Myotis septentrionalis; federally threatened) exists at Camp Ravenna. There are no other federally listed species and no critical habitat on Camp Ravenna. Load Line 11 has not been previously surveyed for rare, threatened, or endangered species; however, there have been no documented sightings of rare, threatened, or endangered species at the AOC (OHARNG 2014).

Other Aquatic Resources. There are no other known aquatic resources (Appendix H, Table H-4) at or near Load Line 11 (e.g., vegetation, animals). There are no other reported surveys of habitats and wildlife at Load Line 11 beyond those summarized in the INRMP (OHARNG 2014). There are two nearby biological and water quality stations. The following subsections provide a summary of the biological and water quality stations in the vicinity of Load Line 11.

Biological/Water Quality Sampling Stations. Ohio EPA and USACE investigated several streams at RVAAP in a network of various biological/water quality sampling stations (USACE 2005a). The purpose of this investigation was to document ecological effects of AOCs on stream or pond biota and conditions. Two sampling stations were located in the vicinity of Load Line 11. Station S-1 was located upstream of the AOC, and Station S-2 was located downstream of the AOC. The upstream biological/water quality station (S-1) provides information about potential contamination from upstream AOCs and if upstream AOCs may be contributing to adverse biological, chemical, and physical measurements in the vicinity of Load Line 11. The downstream sampling station (S-2) provides information about potential contamination from Load Line 11 and upstream AOCs. A positive rating (e.g., good, excellent, full attainment, and other positive terms reported in the study) at a downstream sampling station suggests Load Line 11 and other upstream AOCs are not adversely impacting the quality of Sand Creek.

According to the *Facility-Wide Biological and Water Quality Study* (USACE 2005a), each sampling location included a sediment sampling/assessment, surface water sampling/assessment, fish and macroinvertebrate community assessment, and habitat assessment. The sampling reach for stream sampling stations ranged 395-690 ft.

Sediment evaluations were conducted in June 2003 using guidelines established in *Development and Evaluation of Consensus-based Sediment Quality Guidelines for Freshwater Ecosystems* (MacDonald et al. 2000), SRVs for inorganic chemicals (Ohio EPA 2003), and USEPA Region 5 ecological screening levels (ESLs) (USEPA 2003a). Sediment samples were analyzed for SVOCs, pesticides, PCBs, TAL metals, explosives, percent solids, cyanide, ammonia, nitrate, and phosphorus. Surface water grab samples collected in June and September 2003 were evaluated using comparisons to Ohio Water Quality Standards criteria, reference conditions, or cited in the *Facility-Wide Biological and Water Quality Study*. Surface water samples were analyzed for TAL metals, pesticides, PCBs, explosives, SVOCs, and several nutrients.

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Fish and macroinvertebrate sampling and assessments occurred in August and September 2003. Fish were sampled using electrofishing methods. Macroinvertebrate communities were assessed using artificial substrates (quantitative sampling), supplemented with a composite natural substrate sample (qualitative sampling). The fish and macroinvertebrate community assessments followed the methods contained in *Biological Criteria for the Protection of Aquatic Life: Volume III, Standardized Biological Field Sampling and Laboratory Methods for Assessing Fish and Macroinvertebrate Communities* (Ohio EPA 1989).

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The physical habitat assessment was conducted in June 2003 and used the Qualitative Habitat Evaluation Index (QHEI) developed by the Ohio EPA (Rankin 1989, 1995). The types(s) and quality of substrates, amount and quality of in-stream cover, channel morphology, extent and quality of riparian vegetation, pool, run, riffle development and quality, and gradient are some of the habitat characteristics used to determine the QHEI score.

Sampling Station Locations. Station S-1 [River Mile (RM) 7.0] is located at the first and eastern intersection of Sand Creek and Newton-Falls Road, approximately 4,500 ft west of Load Line 11. Station S-1 provides the closest upstream reference point for the AOC. Station S-2 (RM 5.9) is located at the second and western intersection of Sand Creek and Newton Falls Road, approximately 300 ft directly north of Load Line 11 (Figure 7-1). The station is immediately downstream from Load Line 11 and provides the closest downstream sampling point to the AOC.

Summary of Sampling/Assessment Results. Table 7-13 shows the ratings of the attributes for sampling stations S-1 and S-2. Review of the Facility-Wide Biological and Water Quality Study (USACE 2005a) data from the two stations showed many positive attribute ratings (e.g., good, excellent, full attainment) and no sign of aquatic impairment. Each station was rated at Full Use Attainment Status, which indicated that all indices met the Ohio EPA biological criteria.

At S-1, all inorganic chemicals tested in sediment were below Ohio SRVs and threshold effects concentration (TEC) levels. All tested explosives, pesticides, and PCBs were not detected in sediment samples collected from S-1. The few SVOCs were measured at low levels, with all concentrations below TEC or ESL guidelines. Ammonia and total phosphorus levels were measured below screening guidelines. None of the surface water chemical concentrations at S-1 exceeded Ohio Water Quality Standards aquatic life maximum or average water quality criteria (WQC), and none of the chemicals measured exceeded criteria protective of the Warm Water Habitat aquatic life use (USACE 2005a). Overall, the sediment quality and water quality at S-1 was rated "excellent." The fish community at S-1 was rated "good." The Index of Biotic Integrity score was 41; 15 species were reported. The macroinvertebrate community at S-1 was rated "exceptional." Based on the fish and macroinvertebrate community assessment, no biological impairment associated with chemical contaminants was observed at S-1. The physical habitat was also evaluated at S-1, and the QHEI score was 74.0, indicating "good" stream habitat capable of supporting Warm Water Habitat biological communities.

At S-2, all inorganic chemicals tested in sediment were below Ohio SRVs and TEC levels. All tested explosives, pesticides, and PCBs were not detected in sediment samples collected from S-2. The few

SVOCs were measured at low levels, with all concentrations below TEC or ESL guidelines. Ammonia and total phosphorus levels were measured below screening guidelines. None of the surface water chemical concentrations at S-2 exceeded Ohio Water Quality Standards aquatic life 4 maximum or average WQC, and none of the chemicals measured exceeded criteria protective of the Warm Water Habitat aquatic life use (USACE 2005a). Overall, the sediment quality and water quality at S-2 was rated "excellent." The fish community at S-2 was rated "good." The Index of Biotic Integrity score was 40, and 15 species were reported. The macroinvertebrate community at S-2 was rated "exceptional." The physical habitat was also evaluated at S-2, and the QHEI score was 78.5, indicating "excellent" stream habitat capable of supporting Warm Water Habitat biological communities. These favorable sediment/water quality findings at S-2 support the observation that Load Line 11 is not contributing contamination to Sand Creek.

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Ecosystem and Landscape Roles and Relationships. Four spatial areas were evaluated to assess the ecosystem and landscape roles and relationships at Load Line 11: the actual AOC, vicinity of the AOC, entire Camp Ravenna, and ecoregion of northeastern Ohio. Information about the first spatial area (AOC) was provided in the section above on terrestrial and aquatic resources.

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Vicinity of the AOC. Four forest communities border Load Line 11 (Figure 7-1). There are no apparent differences in habitat quality of these forest communities inside or outside of the AOC. The types and qualities of habitat are not unique and can be found at many other areas at Camp Ravenna.

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Figure 7-1 shows there is one wetland along the western border of the AOC and one in the vicinity of the East Ditch. Other wetlands are located outside of Load Line 11, to the north along Sand Creek, southeast of Fuze & Booster Spur Road, and southwest of the AOC boundary. No perennial surface water features exist in AOC boundary. Sand Creek is located at least 125 ft north of the AOC (Figure 7-1).

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The closest recorded rare species [caddisfly (Psilotreta indecisa)] is located approximately 900 ft west of the AOC (Table 7-14). It is a state threatened species. The next closest rare species [pale sedge (Carex pallescens)] is located approximately 1,100 ft west of the AOC. It is a state potentially threatened species.

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No beaver dams are in or near the AOC. There is a 100-year floodplain along Sand Creek that reaches the northern boundary of the AOC. There is a biological and water quality station (stream sample station) within 300 ft of the AOC.

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The Entire Camp Ravenna. Load Line 11 is approximately 48 acres, which represents 0.22% of the total area of Camp Ravenna (21,683 acres). There are approximately 2,310 acres of forest type FL1 [temporarily flooded forest alliance (e.g., green ash and American elm)] at Camp Ravenna (OHARNG 2014); FL1 represents 10.7% of the habitat at Camp Ravenna. There are approximately 1,440 acres of forest type FU1 (American beech/sugar maple/American tulip tree) at Camp Ravenna (OHARNG 2014); FU1 represents 6.6% of the habitat at Camp Ravenna. There are approximately 2,290 acres of forest type FU2 (American beech/oak/maple) (OHARNG 2014); FU2 represents 0.6% of the habitat at Camp Ravenna. There are approximately 1,650 acres of forest type FU5 (white

ash/black cherry/red maple) (OHARNG 2014); FU5 represents 7.6% of the habitat at Camp Ravenna. There are approximately 2,050 acres of vegetation type HU1 (goldenrod/clasping-leaf dogbane) (OHARNG 2014); HU1 represents 9.5% of the habitat at Camp Ravenna. There are approximately 2,900 acres of vegetation type SU1 (gray dogwood/northern arrowwood) (OHARNG 2014); SU1 represents 13.4% of the habitat at Camp Ravenna. There are approximately 1,970 acres of wetlands (jurisdictional and planning level survey) as defined in the INRMP (OHARNG 2014); wetlands represent about 9% of the habitat at Camp Ravenna. These types of resources are abundant and are not unique to Load Line 11 at Camp Ravenna.

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> Ecoregion. In the area surrounding Camp Ravenna, forests occupy a high percentage of the terrain. Ohio's forests cover approximately 8,000,000 acres or 30% of the state (USDA 2009). The Erie/Ontario Drift and Lake Plain ecoregion (USGS 1998) is located in the northeastern part of Ohio, and both contain the communities of temporarily flooded forest alliance (e.g., green ash and American elm), American beech/sugar maple/American tulip tree forest alliance and American beech/oak/maple forest alliance, and mixed cold-deciduous successional forest (e.g., white ash and red maple). The Erie/Ontario Drift and Lake Plain ecoregion exhibits rolling to level terrain formed by lacustrine and low lime drift deposits. Lakes, wetlands, and swampy streams occur where stream networks converge or where the land is flat and clayey (USGS 1998). The U.S. Forest Service has a Forest Inventory Data Online tool that was queried for the forest types in the surrounding counties in or near Camp Ravenna (USFS 2011). In 2009, approximately 93,900 acres of forest type FL1, 171,380 acres of forest type FU1, 621,100 acres of forest type FU2, and 210,340 acres of forest type FU5 were found throughout northwestern Ohio in Cuyahoga, Geauga, Mahoning, Portage, Stark, Summit, and Trumbull counties that surround RVAAP (USFS 2011). The herbaceous field and shrubland was not individually found in this query because it is not classified as a main group of trees in the forest inventory data tool. However, herbaceous field (HU1) and shrubland (SU1) are common across the ecoregion (USDA 2011). Wetlands across the ecoregion make up 207,800 acres (USEPA 1999a). The vegetation and wetland communities at Load Line 11 are also found in the surrounding counties in the ecoregion of northeastern Ohio.

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In summary, the current vegetation types of: (1) temporarily flooded forest alliance (e.g., green ash and American elm); (2) American beech/sugar maple/American tulip tree forest alliance; (3) American beech/oak/maple forest alliance; (4) mixed, cold-deciduous, successional forest; (5) dry, early-successional, herbaceous field; (6) dry, mid-successional, cold-deciduous shrubland; and (7) wetlands are found in the vicinity of Load Line 11. The forest types, herbaceous field, shrubland, and wetlands are abundant at Camp Ravenna and the larger surrounding local ecoregion. There is no known unique resource at Load Line 11 that cannot be found in the immediate vicinity of the AOC, Camp Ravenna, and in a large part of the ecoregion of northeastern Ohio.

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7.3.2.4 Evaluation of Historical Chemical Contamination and Ecological Significance

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Based on the historical ESV screening, 14 soil, 19 sediment, and 2 surface water COPECs were identified at Load Line 11. These COPECs are listed on Table 7-12.

- 1 The Army and Ohio EPA provide a checklist of important ecological places and resources to
- determine if such ecological resources are present in (or nearby) an AOC. The only resource on this
- 3 list present at Load Line 11 is wetlands. Environmental management goals and objectives of
- 4 OHARNG are applicable to Load Line 11, including Goal 1 requiring management of natural
- 5 resources to be compatible with military mission, and Goal 5 requiring the Army to sustain usable
- 6 training grounds and natural resources.

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Load Line 11 is made up of approximately 48 acres of herbaceous field, shrubland, and forest communities. Two wetlands are found at the AOC. The vegetation types and wetlands at Load Line 11 are found nearby at RVAAP and in the ecoregion.

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Because there is contamination at Load Line 11 and important or significant resources (i.e., wetlands) are present, this ERA continues to a Level II Screening Level ERA.

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7.3.3 Level II: Screening Level Ecological Risk Assessment

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18 19 The Level II method follows the guidance documents listed in Section 7.3.1. The Level II method identifies evaluation procedures used for problem formulation and ecological effects evaluation to determine AOC-related COPECs. This work includes defining habitats/environmental setting, suspected contaminants, possible pathways, and mechanisms for ecotoxicity and contaminant transport. Level II also includes establishing screening values.

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In addition, technical and refinement factors can be used to assess outcomes of the above procedures. The factors include using mean exposure concentrations and discussing approved ESVs, wetland quality at the AOC, and other topics that evaluate and refine the COPECs from the Level II Screening ERA. This type of assessment is Step 3A in the ERA process (USEPA 1997).

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7.3.3.1 Generic Ecological Conceptual Exposure Model

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The conceptual site exposure model identifies the interconnections of contaminant sources and transport mechanisms for contaminant migration through the environment to the receptors. The conceptual site exposure model provides an understanding of the relationships of all sources, release and transport pathways, potential exposure media, and receptors and includes:

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• **Source Media.** Based on historical AOC information, operations associated with various former buildings at Load Line 11 are the contaminant source. The operations contributed chemicals to the surrounding soil, sediment, and surface water. Some contaminated source media was removed during the 2001 IRA: sumps, 230 yd³ of contaminated media in drainage ditches, and a hot spot of petroleum-contaminated soil (MKM 2004a).

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• Transport Mechanisms. Material in soil can migrate via erosion and leaching. Migration to sediment and surface water via erosion and leaching is controlled by the amount of precipitation, type of ground cover, and topography of the AOC. Little erosion is expected to occur at the AOC because, although parts of the land are relatively sloped, the slopes have a high percentage of vegetative ground cover that will further help to increase infiltration and

- decrease erosion. This extensive vegetative cover includes herbaceous fields, shrublands, and forests. While much of the precipitation landing on this area is expected to infiltrate the soil, some rainfall will leave the AOC as runoff. For example, there are shallow drainages (East Ditch and West Ditch) that provide a likely flow pathway from the location of the former buildings to the small wetlands or Sand Creek. The ditches also provide a possible pathway to transport leachate from the soil to the wetlands or Sand Creek.
- Exposure Media. These are media where contaminants are available for exposure to ecological receptors. Potential exposure media at Load Line 11 are soil, sediment, surface water, vegetation, and animals.
- Exposure Pathways. A main exposure pathway is ingestion of contaminated food. Other exposure pathways may include ingesting soil, sediment, and water and dermal contact by receptors with soil, sediment, or water.
- **Ecological Receptors.** A variety of ecological receptors, such as terrestrial birds and mammals, are present in the area. Receptors associated with various published toxicological endpoints (e.g., reproduction, physiology) are assumed to represent these various plants and animals.

7.3.3.2 <u>Habitats and Species (Including Generic Receptors)</u>

Habitats and species at Load Line 11 were defined in the Level I ERA (Section 7.3.2). Five types of habitats were described. Habitats, species, and other resources were analyzed, and it was determined that important or significant ecological resources are present at Load Line 11. One small planning level survey wetland and one emergent wetland (e.g., marsh) associated with the East Ditch are located at Load Line 11. Because contamination is present, a Level II analysis is needed. Level II assumes ecological receptors are sensitive to various chemicals based on a variety of toxicological data from field-observed effects and laboratory tests. The ESV is utilized as a toxicity metric representing multiple generic receptors, including plants, microorganisms, and animals.

7.3.3.3 Procedure to Identify COPECs

The SL approach to evaluate sample results from the PBA08 RI followed a similar approach used in the historical ERA. Section 5.3 details chemical concentration data. The PBA08 RI included collecting discrete surface soil (0-1 ft bgs), sediment, and surface water samples. Some locations were different from the historical sample locations, and some were co-located with previous samples (Figures 5-2 through 5-11).

As discussed in Section 7.1.1, the soil data within the geographic area of Load Line 11 were subdivided into two EUs: FPA and NPA. The FPA includes the area inside the gravel perimeter road and all former buildings and operational areas. The NPA includes the area outside the production area to the fence line. The FPA contains all known or potential primary contaminant sources, such as the former operational buildings. The former buildings located within the NPA were limited to administrative and storage functions. The sediment and surface water samples were divided into two EUs: East Ditch and West Ditch. This ERA uses updated SRVs (Appendix H, Table H-6) and ESVs

that follow the revised *Ecological Risk Assessment Guidance* (Ohio EPA 2008), as provided in Appendix H, Tables H-7 through H-9.

The hierarchy of ESVs is based on the information found in the Ohio EPA risk assessment guidance (Ohio EPA 2008) and FWERWP (USACE 2003a). The MDC of each chemical is compared to its respective facility-wide background concentration. Sediment concentrations are also compared to the SRV. Chemicals are not considered site-related if the MDC is below the background concentration. For all chemicals detected above background concentrations, the MDC is compared to the chemical-specific ESV. In addition to the ESV comparison, it was determined if the chemical is a PBT compound. Chemicals are retained as COPECs if they exceed background concentrations (and SRVs for sediment) and the ESV, if the chemical exceeds background concentrations (and SRVs for sediment) and had no toxicity information, or if the chemical is considered a PBT compound. MDC to ESV ratios are used to determine the integrated COPECs that result from the combined current and historical data sets. A ratio greater than one suggests a possible environmental consequence. Any chemicals with ratios greater than one are identified as integrated COPECs.

Based on comment resolution with Ohio EPA in July 2014, the selection of integrated COPECs for surface water was modified to include two screens. Along with comparing the MDC to the Ohio EPA outside mixing zone maximum (OMZM) ESV, when available, Ohio EPA requested the average concentration also be compared to the Ohio EPA outside mixing zone average (OMZA) ESV when available (Appendix H, Table H-9). However, there is only one surface water sample for each surface water EU at Load Line 11, so average concentrations were not available to compare against the OMZA. As a result, the MDC was compared against the OMZA and OMZM at the East and West Ditches.

Maximum Detected Concentrations. The MDCs were compared to the background concentrations and ESVs (Appendix H, Tables H-10 through H-15) for detected chemicals. These comparisons are provided for soil in the FPA (Appendix H, Table H-10), soil in the NPA (Appendix H, Table H-11), sediment in the East Ditch (Appendix H, Table H-12), sediment in the West Ditch (Appendix H, Table H-13), surface water in the East Ditch (Appendix H, Table H-14), and surface water in the West Ditch (Appendix H, Table H-15).

Ecological Screening Values. Although the historical ERA used ESVs from the 2003 version of the *Guidance for Conducting Ecological Risk Assessments* (Ohio EPA 2003), this ERA uses updated ESVs from the 2008 version of the guidance document. The hierarchy for soil is ecological soil screening levels (EcoSSLs), PRGs, and ESLs. The hierarchy for sediment is the Sediment Quality Guidelines, followed by ESLs. The hierarchy for surface water is the Ohio EPA WQC, National Recommended WQC, and ESLs. Appendix H provides for values and sources for ESVs in Tables H-7 through H-9.

7.3.3.4 Integrated COPECs for Surface Soil (0-1 ft bgs)

As discussed in Section 7.1.1, the soil data within the geographic area of Load Line 11 were subdivided into two EUs: FPA and NPA. A summary of the integrated COPECs identified in soil at the FPA and NPA follows.

Integrated COPECs for Surface Soil (0-1 ft) at the Former Production Area. During the PBA08 RI, 48 chemicals were detected in surface soil at the FPA. Five chemicals (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were excluded as SRCs. A total of 16inorganic chemicals, 2 anions, and 22 organic chemicals were determined to be SRCs because they exceeded background concentrations or did not have an associated background concentration for comparison. Of the 40 SRCs, 12 inorganic chemicals (aluminum, arsenic, cadmium, chromium, cobalt, copper, cyanide, lead, manganese, mercury, vanadium, and zinc), 1 anion (sulfide), and 1 organic chemical (PCB-1254) exceeded their ESVs and were identified as integrated COPECs (Table 7-15). In addition, one anion (nitrate) and four organic chemicals (nitrocellulose, nitroguanidine, PETN, and tetryl) were selected as integrated COPECs because they do not have an ESV for comparison. Two of the integrated COPECs (mercury and PCB-1254) were also PBT compounds. Table 7-15 shows the calculated ratio of MDC to ESV for each integrated COPEC. Appendix H, Table H-10 presents the details of the ESV comparisons for surface soil at the FPA.

Most inorganic COPECs reported in the historical ERA (Table 7-12) for soil are also identified in this ERA. The historical ERA (MKM 2005a) identified two inorganic COPECs (iron and nickel) that are not integrated COPECs for the FPA. Nickel at the FPA was below the new, less conservative ESV (Ohio EPA 2008) (Appendix H, Table H-10), and iron was considered an essential nutrient in the new data set. Seven new integrated COPECs (aluminum, cadmium, cobalt, copper, nitroguanidine, PETN, and tetryl) were identified in this ERA. Aluminum was considered an essential nutrient in the historical ERA; however, aluminum is not considered an essential nutrient in the PBA08 ERA and is identified as an integrated COPEC. Three other inorganic COPECs (cadmium, cobalt, and copper) were identified in the FPA due to detections above new, more conservative ESVs (Ohio EPA 2008). Four new integrated COPECs (nitroguanidine, PETN, tetryl, and PCB-1254) were identified due to detections of these organic chemicals in samples collected during the PBA08 RI.

Integrated COPECs for Surface Soil (0-1 ft) at the Non-Production Area. During the PBA08 RI, 45 chemicals were detected in surface soil at the NPA. Five chemicals (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were excluded as SRCs. A total of 14 inorganic chemicals, 2 anions, and 19 organic chemicals were determined to be SRCs because they either exceeded their background concentrations or did not have an associated background concentration for comparison. Of the 35 SRCs, 7 inorganic chemicals (arsenic, cadmium, cobalt, lead, manganese, mercury, and zinc) and 1 anion (sulfide) exceeded the ESVs and were identified as integrated COPECs (Table 7-16). In addition, one anion (nitrate) and three organic chemicals (HMX, nitrocellulose, and PETN) were selected as integrated COPECs because they do not have an ESV for comparison. One PBT compound (PCB-1254) was also identified as a COPEC, even though it did not exceed its ESV. One integrated COPEC (mercury) was also a PBT compound. Table 7-16 shows the

calculated ratio of MDC to ESV for each integrated COPEC. Appendix H, Table H-11 presents the details of the ESV comparisons for surface soil at the NPA.

Some inorganic COPECs reported in the historical ERA (Table 7-12) for soil are also identified in this ERA. The historical ERA (MKM 2005a) identified five COPECs (chromium, cyanide, iron, nickel, and vanadium) that are not integrated COPECs for the NPA. Iron was considered an essential nutrient in the new data set (Appendix H, Table H-11); chromium and nickel were below their new, less conservative ESVs (Ohio EPA 2008); and vanadium was below its background concentration. Cyanide was not detected in NPA soil above its ESV. Four new integrated COPECs (cadmium, cobalt, HMX, and PETN) were identified in this ERA. Cadmium and cobalt were identified in the NPA due to detections above new, more conservative ESVs (Ohio EPA 2008). Two new integrated COPECs (HMX and PETN) were identified due to detections of these organic chemicals in samples collected during the PBA08 RI.

7.3.3.5 <u>Integrated COPECs for Sediment</u>

As discussed in Section 7.1.1, the sediment data collected within the geographic area of Load Line 11 were subdivided into two EUs: East Ditch and West Ditch. A summary of the integrated COPECs identified in sediment at the East Ditch and West Ditch is presented below.

Integrated COPECs for Sediment at the East Ditch. During the PBA08 RI, 36 chemicals were detected in sediment at the East Ditch. Five chemicals (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were excluded as SRCs. A total of 3 inorganic chemicals and 15 organic chemicals were determined to be SRCs because they either exceeded their background concentrations or did not have an associated background concentration for comparison. Of the 18 SRCs, 1 organic chemical [benz(a)anthracene] exceeded its ESV and was identified as an integrated COPEC (Table 7-17). In addition, two inorganic chemicals (beryllium and selenium) were selected as integrated COPECs because they do not have an ESV for comparison. No PBT compounds were identified in the East Ditch sediment. Table 7-17 shows the calculated ratio of MDC to ESV for each integrated COPEC. Appendix H, Table H-12 presents the details of the ESV comparisons for sediment at the East Ditch.

Two COPECs [beryllium and benz(a)anthracene] reported in the historical ERA (Table 7-12) for sediment are also identified in this ERA for the East Ditch. The historical ERA (MKM 2005a) identified 4 inorganic COPECs (copper, lead, manganese, and nickel), 3 anions (nitrate, sulfate, and sulfide), and 10 organic COPECs [benzo(a)pyrene, benzo(ghi)perylene, chrysene, fluoranthene, indeno(1,2,3-cd)pyrene, phenanthrene, pyrene, total PAHs, PCB-1254, and nitrocellulose] that are not integrated COPECs for East Ditch sediment because concentrations of these chemicals in the East Ditch did not exceed their ESVs, SRVs, or background concentrations (Appendix H, Table H-12). One new integrated COPEC (selenium) was identified for East Ditch sediment in the PBA08 RI. The addition of selenium was due to a detection above its background concentration.

Integrated COPECs for Sediment at the West Ditch. During the PBA08 RI, 32 chemicals were detected in sediment at the West Ditch. Five chemicals (calcium, iron, magnesium, potassium, and

sodium) were essential nutrients and were excluded as SRCs. A total of 3 inorganic chemicals, 1 anion, and 10 organic chemicals were determined to be SRCs because they exceeded their background concentrations, did not have an associated background concentration for comparison, or were PBT compounds. Of the 14 SRCs, 2 chemicals (beryllium and sulfide) were selected as integrated COPECs because they do not have an ESV for comparison. Mercury did not exceed its ESV; however, it was selected as an integrated COPEC because it is a PBT compound. Table 7-18 shows the calculated ratio of MDC to ESV for each integrated COPEC. Appendix H, Table H-13 presents the details of the ESV comparisons for sediment at the West Ditch.

Two COPECs (beryllium and sulfide) reported in the historical ERA (Table 7-12) for sediment are also identified in this ERA for the West Ditch. The historical ERA (MKM 2005a) identified 4 inorganic COPECs (copper, lead, manganese, and nickel), 2 anions (nitrate and sulfate), and 11 organic COPECs [benz(a)anthracene, benzo(a)pyrene, benzo(ghi)perylene, chrysene, fluoranthene, indeno(1,2,3-cd)pyrene, phenanthrene, pyrene, total PAHs, PCB-1254, and nitrocellulose] that are not integrated COPECs for West Ditch sediment because concentrations of these chemicals in the West Ditch did not exceed their ESVs, SRVs, or background concentrations (Appendix H, Table H-13). One PBT compound (mercury) was also identified as an integrated COPEC for the West Ditch, even though it did not exceed its ESV.

7.3.3.6 Integrated COPECs for Surface Water

As discussed in Section 7.1.1, the surface water data within the geographic area of Load Line 11 were subdivided into two EUs: East Ditch and West Ditch. A summary of the integrated COPECs identified in surface water at the East Ditch and West Ditch is presented below.

Integrated COPECs for Surface Water at the East Ditch. During the PBA08 RI, 18 chemicals were detected in surface water at the East Ditch. Five chemicals (calcium, iron, magnesium, potassium, and sodium) were essential nutrients and were excluded as SRCs. Six inorganic chemicals and two organic chemicals were determined to be SRCs because they exceeded their background concentrations, did not have an associated background concentration for comparison, or were PBT compounds. Of the eight SRCs, two PBT compounds (beta-BHC and gamma-chlordane) were identified as integrated COPECs (Table 7-19), even though they did not exceed their ESVs. No other SRCs exceeded their ESVs. Table 7-19 shows the calculated ratio of MDC to ESV for each integrated COPEC. Appendix H, Table H-14 presents the details of the MDC to OMZM ESV comparisons and Appendix H, Table H-16 presents the details of the MDC to OMZA ESV comparisons for surface water at the East Ditch. These comparisons used the OMZM and OMZA from Ohio EPA WQC, when available, as the ESV; when the OMZM or OMZA were not available, the hierarchy of preferred sources was followed, as presented in Section 7.3.3.3.3.

The historical ERA (MKM 2005a) identified two COPECs (manganese and sulfate) that are not integrated COPECs for surface water in the East Ditch. Manganese was below its background concentration (Appendix H, Table H-14), and sulfate was not detected in East Ditch surface water.

Integrated COPECs for Surface Water at the West Ditch. During the PBA08 RI, 30 chemicals 1 2 were detected in surface water at the West Ditch. Five chemicals (calcium, iron, magnesium, 3 potassium, and sodium) were essential nutrients and were excluded as SRCs. A total of 12 inorganic 4 chemicals and 11 organic chemicals were determined to be SRCs because they exceeded their 5 background concentrations, did not have an associated background concentration for comparison, or were PBT compounds. Of the 23 SRCs, 1 inorganic chemical (manganese) and 1 organic chemical 6 7 [benzo(a)pyrene] exceeded their ESVs and were identified as integrated COPECs (Table 7-20). In 8 addition, one organic chemical [benzo(k)fluoranthene] was selected as an integrated COPEC because 9 it does not have an ESV for comparison. Beta-BHC did not exceed its ESV; however, it was selected 10 as an integrated COPEC because it is a PBT compound. Table 7-20 shows the calculated ratio of 11 MDC to ESV for each integrated COPEC. Appendix H, Table H-15 presents the details of the MDC 12 to OMZM ESV comparisons and Appendix H, Table H-17 presents the details of the MDC to OMZA 13 ESV comparisons for surface water at the West Ditch. These comparisons used the OMZM and 14 OMZA from Ohio EPA WQC, when available, as the ESV; when the OMZM or OMZA were not 15 available, the hierarchy of preferred sources was followed, as presented in Section 7.3.3.3.

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One of the COPECs (manganese) reported in the historical ERA (Table 7-12) for surface water is also identified in this ERA for the West Ditch. The historical ERA (MKM 2005a) identified one other COPEC (sulfate) that was not detected in West Ditch surface water. Three new integrated COPECs [benzo(a)pyrene, benzo(k)fluoranthene and beta-BHC] were identified in this ERA.

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7.3.3.7 Step 3A: Refinement of Integrated COPECs

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Step 3A refines the list of integrated COPECs to determine if: (1) there are final COPECs requiring further evaluation in Level III or remediation to protect ecological receptors; or (2) integrated COPECs can be eliminated from further consideration. This section evaluates and applies refinement factors to the integrated COPECs for the AOC. This evaluation is an important part of Level II and is adapted from USEPA Step 3A, as outlined in the *Ecological Risk Assessment Guidance for Superfund: Process for Designing and Conducting Ecological Risk Assessments* (USEPA 1997) and *Risk Assessment Handbook Volume II: Environmental Evaluation* (USACE 2010b). The purpose of Step 3A is stated as follows by the Army (BTAG 2005):

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"The results of Step 3A will be used to determine if threats to ecological receptors are negligible and an appropriate risk management decision may be made to end the ERA process, or potential threats are still indicated and a baseline ecological risk assessment should be initiated."

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The evaluation and refinement factors used in Step 3A are as follows:

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- Comparing average (i.e., mean) concentration to ESV.
- Comparing mean concentration to background concentration,
- Comparing background concentration to ESV,
 - Determining frequency of chemical occurrence relative to ESV,
 - Observing magnitude of ESV exceedance (ratio of ESV to chemical concentrations),

- Discussing Ohio EPA approved and preferred ESVs,
 - Evaluating qualitative relationship of exposure area to general home range,
- Categorizing wetland quality inside the AOC,
 - Evaluating geographical relationship of on-site wetlands to AOC exceedance area,
- Obtaining information about on-site migration of chemicals to on-site wetlands, and
 - Evaluating off-site migration of chemicals at biological/water quality stations.

Surface soil at Load Line 11 is divided into two EUs: FPA and NPA. For the FPA, there are 19 integrated COPECs in surface soil. Aluminum, arsenic, cadmium, chromium, cobalt, copper, cyanide, lead, manganese, mercury, vanadium, zinc, sulfide, and PCB-1254 are the 14 integrated COPECs that exceeded their background concentrations and ESVs. Mercury and PCB-1254 are also PBT compounds. Five additional chemicals (nitrate, nitrocellulose, nitroguanidine, PETN, and tetryl) are integrated COPECs because they do not have an ESV. For the NPA, there are 13 integrated COPECs in surface soil. Arsenic, cadmium, cobalt, lead, manganese, mercury, zinc, and sulfide are the eight integrated COPECs that exceeded their background concentrations and ESVs. PCB-1254 did not exceed its ESV but was identified as a COPEC because it is a PBT compound. One other COPEC (mercury) was also identified as a PBT compound. Four additional chemicals (nitrate, HMX, nitrocellulose, and PETN) are integrated COPECs because they do not have an ESV.

Sediment at Load Line 11 is divided into two EUs: East Ditch and West Ditch. In the East Ditch, there are three integrated COPECs in sediment [beryllium, selenium, and benz(a)anthracene]. Benz(a)anthracene exceeded its ESV, and beryllium and selenium were integrated COPECs by default (i.e., the chemical does not have an ESV). In the West Ditch, there are three integrated COPECs in sediment (beryllium, mercury, and sulfide). Beryllium and sulfide were integrated COPECs because they do not have an ESV, and mercury is a PBT compound.

Surface water at Load Line 11 is divided into two EUs: East Ditch and West Ditch. In the East Ditch, there are two integrated COPECs in surface water (beta-BHC and gamma-chlordane). Beta-BHC and gamma-chlordane did not exceed their ESVs but were identified as COPECs because they are PBT compounds. In the West Ditch, there are four integrated COPECs in surface water [manganese, benzo(a)pyrene, benzo(k)fluoranthene, and beta-BHC]. Manganese and benzo(a)pyrene are the only integrated COPECs that exceeded their background concentrations and ESVs [benzo(a)pyrene has no background value available], benzo(k)fluoranthene was an integrated surface water COPEC because it does not have an ESV, and beta-BHC is a PBT compound that did not exceed its ESV.

Chemicals with no ESVs are discussed later in Step 3A and in Section 7.3.3.9. PBT compounds are discussed later in Step 3A. All integrated COPECs that have an MDC to ESV ratio greater than one are evaluated based on a series of evaluation or refinement factors. The first four evaluation and refinement factors are organized to compare key quantitative information and are listed below.

- Comparing average (i.e., mean) concentration to ESV,
- Comparing mean concentration to background concentration,
- Comparing background concentration to ESV, and
 - Determining frequency of chemical occurrence relative to ESV.

Multiple evaluation factors can be used to define whether an integrated COPEC should be retained or eliminated from further consideration. There are two types of comparisons and associated decisions in the first steps of the refinement process (Table 7-21).

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Comparing background concentration to an ESV is also an important consideration in this part of the evaluation. Additionally, frequency of detection is provided. The evaluations are presented by type of decision (Table 7-21) on a COPEC-by-COPEC basis. These evaluations are followed by the applying additional evaluation and refinement factors, when necessary. The COPECs for each soil EU are defined separately.

Former Production Area – Comparison of Mean Concentration to ESV. Seven integrated COPECs (arsenic, cadmium, chromium, cobalt, copper, cyanide, and PCB-1254) in soil are eliminated in this step because the mean concentration is smaller than the ESV (Appendix H, Table H-16). Each eliminated integrated COPEC is discussed below relative to the mean concentration being smaller than the ESV and the related evaluation and refinement factors. Table 7-22 shows the relevant data and various comparisons.

Arsenic. Arsenic was detected in all 47 discrete samples analyzed for inorganic chemicals at the FPA. Only 11 samples had detections above the background concentration, and 8 samples had arsenic concentrations exceeding the ESV (Table 7-22). Although the MDC for arsenic in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV and background concentration. Therefore, arsenic was eliminated from further consideration and will not be a final COPEC.

Cadmium. Cadmium was detected in 26 of 47 discrete soil samples analyzed for inorganic chemicals at the FPA, but only 7 samples had cadmium concentrations exceeding the ESV (Table 7-22). Although the MDC for cadmium in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV. Therefore, cadmium was eliminated from further consideration and will not be a final COPEC.

Chromium. Chromium was detected in all 47 discrete samples analyzed for inorganic chemicals at the FPA. A total of 10 samples exceeded the background concentration, and only 2 samples had chromium concentrations exceeding the ESV (Table 7-22). Although the MDC for chromium in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV and background concentration. Therefore, chromium was eliminated from further consideration and will not be a final COPEC.

Cobalt. Cobalt was detected in all 47 discrete samples analyzed for inorganic chemicals at the FPA. Six samples had detections above the background concentration, but only two of these samples had cobalt concentrations exceeding the ESV (Table 7-22). Although the MDC for cobalt in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV and background concentration. Therefore, cobalt was eliminated from further consideration and will not be a final COPEC.

Copper. Copper was detected in all 47 discrete samples analyzed for inorganic chemicals at the FPA. 1 2

Nineteen samples were above the background concentration, and only three samples had copper

concentrations exceeding the ESV (Table 7-22). Although the MDC for copper in surface soil

4 exceeds the background concentration and ESV, the mean concentration is less than the ESV and

close to the background concentration. Therefore, copper was eliminated from further consideration

and will not be a final COPEC.

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Cyanide. Cyanide was detected in 3 of 34 discrete soil samples analyzed for inorganic chemicals at the FPA, but only 1 sample had a cyanide concentration exceeding the ESV (Table 7-22). Although the MDC for cyanide in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV. Therefore, cyanide was eliminated from further consideration and will not be a final COPEC.

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PCB-1254. PCB-1254 was detected in 4 of 12 discrete samples analyzed for PCBs at the FPA. There is no background concentration for comparison, but only one sample had a concentration exceeding the ESV (Table 7-22). Although the MDC for PCB-1254 in surface soil exceeds the ESV, the mean concentration is less than the ESV. While PCB-1254 is a PBT compound, the ESV accounted for bioaccumulation (DOE 1997); therefore, PCB-1254 was eliminated from further consideration and will not be a final COPEC. Additional information about PBT compounds is provided later in Step 3A.

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Of the 14 integrated FPA surface soil COPECs with MDCs exceeding the ESV, 7 COPECs (arsenic, cadmium, chromium, cobalt, copper, cyanide, and PCB-1254) were eliminated from further consideration. Seven remaining integrated COPECs with MDCs greater than the ESV (aluminum, lead, manganese, mercury, vanadium, zinc, and sulfide) in FPA surface soil have mean concentrations larger than the ESV.

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Non-Production Area - Comparison of Mean Concentration to ESV. Four integrated COPECs (arsenic, cadmium, cobalt, and PCB-1254) in soil are eliminated in this step because the mean concentration is smaller than the ESV (Appendix H, Table H-17). Each eliminated integrated COPEC is discussed relative to the mean concentration being smaller than the ESV and the related evaluation and refinement factors. Table 7-23 shows the relevant data and various comparisons.

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Arsenic. Arsenic was detected in 29 of 30 discrete samples analyzed for inorganic chemicals at the NPA. Only five samples had detections above the background concentration, and only three samples had arsenic concentrations exceeding the ESV (Table 7-23). Although the MDC for arsenic in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV and background concentration. Therefore, arsenic was eliminated from further consideration and will not be a final COPEC.

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Cadmium. Cadmium was detected in 17 of 30 discrete soil samples analyzed for inorganic chemicals at the NPA. All 17 detections were above the background concentrations, but only 4 samples had cadmium concentrations exceeding the ESV (Table 7-23). Although the MDC for cadmium in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV. Therefore, cadmium was eliminated from further consideration and will not be a final COPEC.

Cobalt. Cobalt was detected in all 30 discrete samples analyzed for inorganic chemicals at the NPA. Seven samples had detections above the background concentration, but only one of these samples had a cobalt concentration exceeding the ESV (Table 7-23). Although the MDC for cobalt in surface soil exceeds the background concentration and ESV, the mean concentration is less than the ESV and background concentration. Therefore, cobalt was eliminated from further consideration and will not be a final COPEC.

PCB-1254. PCB-1254 was detected in 1 of 14 discrete samples analyzed for PCBs at the NPA. The detected PCB-1254 concentration did not exceed its ESV in surface soil in the NPA but was retained as a COPEC because it is a PBT compound (Table 7-23). While PCB-1254 is a PBT compound, the ESV accounted for bioaccumulation (DOE 1997); therefore, PCB-1254 was eliminated from further consideration and will not be a final COPEC. Additional information about PBT compounds is provided later in Step 3A.

Of the nine integrated NPA surface soil COPECs with MDCs exceeding the ESV, four COPECs (arsenic, cadmium, cobalt, and PCB-1254) were eliminated from further consideration. Five remaining integrated COPECs with MDCs greater than the ESV (lead, manganese, mercury, zinc, and sulfide) in NPA surface soil have mean concentrations larger than the ESV.

Former Production Area – Comparison of Mean Concentration Above ESV to Background Concentration. Three integrated COPECs (aluminum, manganese, and vanadium) in surface soil at the FPA have mean concentrations larger than the ESV but are eliminated in this step because the mean concentration is smaller than the background concentration. Each eliminated integrated COPEC is discussed relative to the various evaluation and refinement factors.

Aluminum. Aluminum in surface soil at the FPA has a mean concentration less than the background concentration (Table 7-22). The background concentration is more than 350 times greater than the ESV, so the ESV can be considered conservative. Although aluminum was detected in all 47 discrete samples at concentrations above the ESV, only 3 samples had detections exceeding the background concentration. Having only three samples exceed the background concentration suggests the concentration of aluminum in surface soil is not likely a concern. Additionally, aluminum is not a concern because the soil pH is too high to dissociate the chemical. A typical soil pH is 6-7 at one of the nearby load lines at RVAAP (USACE 2004). Regarding aluminum chemistry and ecological risk in soil, the USEPA states, "aluminum is identified as a COPC only for soil with a pH less than 5.5" (USEPA 1999a). Therefore, because the mean concentration is less than the background concentration, the ESV is conservative, and the soil pH at RVAAP is higher than the USEPA dissociation limit, aluminum was eliminated from further consideration and will not be a final COPEC.

Manganese. Manganese in surface soil in the FPA has a mean concentration less than the background concentration, and the background concentration is more than six times greater than the ESV

(Table 7-22). Because the ESV is lower than the background concentration, the ESV can be considered conservative. Manganese was detected in 42 of 47 discrete samples at concentrations above the ESV; however, only 2 samples had detections which (slightly) exceeded the background concentration. Although the MDC for manganese in surface soil exceeds the background concentration and the ESV, the MDC is just slightly greater than background concentration, the mean concentration is below background concentration, and the ESV is conservative. Therefore, manganese was eliminated from further consideration and will not be a final COPEC.

Vanadium. Vanadium in surface soil in the FPA has a mean concentration less than the background concentration, and the background concentration is approximately four times greater than the ESV (Table 7-22). Because the ESV is lower than the background concentration, the ESV can be considered conservative. Vanadium was detected in 46 of 47 discrete samples at concentrations above the ESV; however, only 2 samples had detections which (slightly) exceeded the background concentration. Although the MDC for vanadium in surface soil exceeds the background concentration and the ESV, the MDC is slightly greater than background concentration, the mean concentration is below background concentration, and the ESV is conservative. Therefore, vanadium was eliminated from further consideration and will not be a final COPEC.

Non-Production Area – Comparison of Mean Concentration Above ESV to Background Concentration. Three integrated COPECs (lead, manganese, and mercury) in surface soil at the NPA are eliminated in this step because the mean concentration is larger than the ESV but smaller than the background concentration. Each eliminated integrated COPEC is discussed relative to the various evaluation and refinement factors.

Lead. Lead in surface soil in the NPA has a mean concentration less than the background concentration, and the background concentration is more than twice the ESV (Table 7-23). Because the ESV is lower than the background concentration, the ESV can be considered conservative. Lead was detected in 27 of 29 discrete samples at concentrations above the ESV; however, only 4 samples had detections exceeding the background concentration. Although the MDC for lead in surface soil exceeds the background concentration and the ESV, the mean concentration is below the background concentration, and the ESV is conservative. Therefore, lead was eliminated from further consideration and will not be a final COPEC.

Manganese. Manganese in surface soil in the NPA has a mean concentration less than the background concentration, and the background concentration is more than six times greater than the ESV (Table 7-23). Because the ESV is lower than the background concentration, the ESV can be considered conservative. Manganese was detected in 29 of 30 discrete samples at concentrations above the ESV; however, only 2 samples had detections exceeding the background concentration. Although the MDC for manganese in surface soil exceeds the background concentration and the ESV, the mean concentration is below the background concentration, and the ESV is conservative. Therefore, manganese was eliminated from further consideration and will not be a final COPEC.

Mercury. Mercury in surface soil in the NPA has a mean concentration less than the background concentration, and the background concentration is 70 times greater than the ESV (Table 7-23).

Because the ESV is lower than the background concentration, the ESV can be considered conservative. Mercury was detected in 14 of 30 discrete samples, with all detections above the ESV; 11 of 30 samples had detections exceeding the background concentration. Although the MDC for mercury in surface soil exceeds the background concentration and the ESV, the mean concentration is below the background concentration, and the ESV is conservative. Therefore, mercury was eliminated from further consideration and will not be a final COPEC.

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Former Production Area – Continued Evaluations. The remaining four integrated COPECs (lead, mercury, zinc, and sulfide) at the FPA in surface soil have mean concentrations greater than the ESV and the background concentration (Table 7-24). Each remaining integrated COPEC is discussed below relative to the first four and related evaluation and refinement factors.

Lead. Lead in FPA surface soil has a mean concentration greater than the background concentration and ESV. It was detected above the background concentration in 21 of 47 discrete samples and above the ESV in 44 of 47 samples (Table 7-24). This is because the ESV is lower than the background concentration, which indicates the ESV for lead may be conservative. Although the ESV may be conservative, the mean concentration for lead in surface soil exceeds the background concentration and ESV. Lead requires further evaluation as a COPEC.

Mercury. Mercury in FPA surface soil has a mean concentration greater than the background concentration and ESV. Mercury was detected above the background concentration in 13 of 17 discrete samples and above the ESV in all 17 samples (Table 7-24). This is because the background concentration is 70 times greater than the ESV, suggesting the ESV may be very conservative. Although mercury has a very conservative ESV, it is a PBT compound, and the mean concentration exceeds the background concentration and ESV. Mercury requires further evaluation as a COPEC.

Zinc. Zinc in FPA surface soil has a mean concentration greater than the background concentration and ESV. It was detected above the background concentration in 22 of 47 discrete samples and at concentrations above the ESV in 42 of 47 samples (Table 7-24). This is because the ESV is lower than the background concentration, which indicates the ESV for zinc may be conservative. Because the mean concentration in surface soil exceeds the background concentration and the ESV, zinc requires further evaluation as a COPEC.

Sulfide. Sulfide in FPA surface soil has a mean concentration greater than the ESV. There is no background concentration for comparison. Sulfide was detected above the ESV in all 22 samples (Table 7-24). Because the mean concentration in surface soil exceeds the ESV and there is not a background concentration for comparison, sulfide requires further evaluation as a COPEC.

Non-Production Area – Continued Evaluations. The remaining two integrated COPECs (zinc and sulfide) at the NPA in surface soil have mean concentrations greater than the ESV and the background concentration (Table 7-24). Each remaining integrated COPEC is discussed below and discussed relative to the first four and related evaluation and refinement factors.

Zinc. Zinc in NPA surface soil has a mean concentration greater than the background concentration and ESV. It was detected above the background concentration in 12 of 30 discrete samples and at concentrations above the ESV in 25 of 30 samples (Table 7-24). This is because the ESV is lower than the background concentration, which indicates the ESV for zinc may be conservative. Because the mean concentration in surface soil exceeds the background concentration and the ESV, zinc requires further evaluation as a COPEC.

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Sulfide. Sulfide in NPA surface soil has a mean concentration greater than the ESV. There is no background concentration for comparison. It was detected above the ESV in all 11 samples (Table 7-24). Because the mean concentration in surface soil exceeds the ESV and there is not a background concentration for comparison, sulfide requires further evaluation as a COPEC.

Additional Aspects of Continued Evaluations. The second refinement factor comparing the mean concentration to the background concentration evaluates how much higher the mean soil concentration is than the background concentration. Four COPECs (lead, mercury, zinc, and sulfide at the FPA and zinc and sulfide at the NPA) have mean concentrations higher than their background concentrations. If the degree of difference between the mean concentration and the background concentration is small, the integrated COPEC will not be considered a final COPEC. Table 7-24 shows that while the mean concentration exceeds the background concentration, the exceedance is often relatively small. For example, the mean concentration for lead at the FPA is 29.6 mg/kg, while the background concentration is 26.1 mg/kg. Sulfide does not have an established background concentration at RVAAP; therefore, a comparison of mean concentration to background concentration is not available.

Additional Technical and Refinement Factors. The next three evaluation and refinement factors include:

- Observing magnitude of ESV exceedance (ratio of ESV to chemical concentrations),
- Discussing Ohio EPA approved and preferred ESVs, and
- Evaluating qualitative relationship of exposure area to general home range.

Former Production Area – Magnitude of ESV Exceedance. Although the mean concentration to ESV ratios for lead (2.7), mercury (77.8), zinc (1.7), and sulfide (13,770) indicate a possibility of risk, the ratio for zinc is relatively small (Appendix H, Table H-16). The small ratio for zinc indicates the potential for toxicity is relatively low, and this likely supports eliminating zinc as an integrated COPEC. The Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008) states:

"If only minor exceedances are detected and other evidence can substantiate, a claim may be made that some or all of the site-associated soils have not been impacted and no additional ecological investigation of the soils is warranted."

Non-Production Area – Magnitude of ESV Exceedance. Although the mean concentration to ESV ratios for zinc (1.7) and sulfide (11,620) indicate a possibility of risk, the ratio for zinc is relatively

small (Appendix H, Table H-17). The small ratio for zinc indicates the potential for toxicity is relatively low, and this supports eliminating zinc as an integrated COPEC.

Comparison of Ohio EPA Approved and Preferred ESVs. The Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008) gives specific guidance on selecting media screening values (ESVs) for Level II evaluation. For soil, three possible sources of ESV values are listed in order of preference: (1) USEPA EcoSSLs, (2) Preliminary Remediation Goals for Ecological Endpoints (DOE 1997), and (3) Region 5 Ecological Screening Levels (USEPA 2003a). However, it is important to note the preferred source (EcoSSLs) can have up to four values per chemical – one for each receptor type (plants, soil invertebrates, birds, and mammals). Because Ohio EPA does not provide guidance on which value to select of these four, the most conservative (lowest) value was chosen for this ERA. It is possible that the chosen ESV is too conservative. Alternative ESVs are presented below for the remaining four integrated COPECs at the FPA and two integrated COPECs at the NPA.

The Ohio EPA approved and preferred lead ESV used in this ERA is 11 mg/kg. This ESV is from the USEPA EcoSSLs (Appendix H, Table H-7). The lead ESV used in this ERA is the most conservative ESV of the USEPA EcoSSLs. Other EcoSSLs for lead include 56 mg/kg, 120 mg/kg, and 1,700 mg/kg (USEPA 2005). The ESV of 11 mg/kg is about four times lower than the ESV (40.5 mg/kg) from the next source of ESVs preferred by Ohio EPA (Appendix H, Table H-7) (DOE 1997). The preferred ESV used for lead is also lower than the background concentration of 26.1 mg/kg (Appendix H, Table H-10), while all of the other EcoSSLs and the alternate ESV are above the background concentration. This information indicates the Ohio EPA approved and preferred ESV for lead, and the selection of lead as a COPEC, is very conservative.

The Ohio EPA approved and preferred mercury ESV used in this ERA is 0.00051 mg/kg (Appendix H, Table H-7). The ESV is lower than other ESVs for mercury (0.1 mg/kg), the ESV for methylmercury (0.0016 mg/kg) (USEPA 2003b) (Appendix H, Table H-7), and the background concentration (0.036 mg/kg) (Appendix H, Table H-10). The *Preliminary Remediation Goals for Ecological Endpoints* labels the form of mercury as an "inorganic chemical" and notes the ESV for mercury is "so low that it may often be within background soil concentrations" (DOE 1997). Although mercury can bioaccumulate in food chains as a PBT compound, the Ohio EPA approved and preferred ESV is very low; therefore, the selection of mercury as a COPEC is very conservative.

The Ohio EPA approved and preferred zinc ESV used in this ERA is 46 mg/kg. This value is from the USEPA EcoSSLs (Appendix H, Table H-7). The zinc ESV used in this ERA is the most conservative ESV of the USEPA EcoSSLs. Other EcoSSLs for zinc include 79 mg/kg, 120 mg/kg, and 160 mg/kg (USEPA 2007a). The ESV of 46 mg/kg is about five times greater than the ESV (8.5 mg/kg) from the next source of ESVs preferred by Ohio EPA (Appendix H, Table H-7) (DOE 1997). The preferred ESV used for zinc is also lower than the background concentration of 61.8 mg/kg (Appendix H, Table H-10), while all of the other EcoSSLs are above the background concentration. These factors indicate the Ohio EPA approved and preferred ESV for zinc is somewhat conservative; thus, the selection of zinc as a COPEC is somewhat conservative.

The Ohio EPA approved and preferred sulfide ESV used in this ERA is 0.00358 mg/kg. This value is from the USEPA Region 5 ESLs (USEPA 2003a) (Appendix H, Table H-7). Sulfide does not have any known alternative ESVs.

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The above information about alternative ESVs shows there are less conservative ESVs that could be chosen for the Level II work. Table 7-25 shows the ratio of ESV-to-mean concentration for the preferred ESV and an alternative ESV. This alternative ESV is the ESV with the closest concentration to the preferred ESV that is above the background concentration. For the four remaining integrated COPECs at the FPA, using the alternative ESV would decrease ratios to less than one for lead, mercury, and zinc. Using the alternative ESV for mercury would decrease the ratio to 0.4. For the two remaining integrated COPECs at the NPA, using the alternative ESV would decrease the ratio to less than one for zinc. If the alternative ESVs were used, lead, mercury, and zinc would be eliminated from further consideration and would not be final COPECs. However, sulfide still needs further evaluation.

Qualitative Relationship of Exposure Area to General Home Range. A majority of the highest lead, mercury, zinc, and sulfide concentrations in surface soil (0-1 ft bgs) are scattered near the former buildings and roads at Load Line 11 (Figures 5-2 and 5-4). The highest concentrations are as follows:

- Lead at the FPA: 102 mg/kg at LL11cs-006 and 97.1 mg/kg at LL11ss-016.
- Mercury at the FPA: 0.34 mg/kg at LL11ss-013 and 0.08 mg/kg at LL11cs-025 and LL11ss-003.
- Zinc at the FPA: 465 mg/kg at LL11cs-025 and 219 mg/kg at LL11cs-021.
- Zinc at the NPA: 478 mg/kg at LL11ss-078 and 176 mg/kg at LL11cs-039.
- Sulfide at the FPA: 244 mg/kg at LL11ss-001 and 226 mg/kg at LL11ss-014.
- Sulfide at the NPA: 85.5 mg/kg at LL11ss-030 and 79.7 mg/kg at LL11sd-014.

Wildlife receptors, especially those with small one-acre home ranges (e.g., small birds and mammals such as robins and field mice), could be exposed to the highest detected concentrations of lead, mercury, zinc, or sulfide at the FPA and zinc or sulfide at the NPA. The highest concentrations of integrated COPECs are interspersed with large areas of lower concentrations over many acres. For example, the two highest mercury concentrations are about 600 ft apart, with lower concentrations between the peaks. This suggests the exposure area with highly elevated concentrations would be small compared to the home ranges of small receptors (robins and field mice) and even smaller for large receptors (turkey and deer) that roam over hundreds of acres. Likewise, the amount of area with highest concentrations to which immobile receptors would be exposed is small.

An examination of the number of concentrations detected above the alternative ESVs (discussed in the previous section) indicates receptor exposures would be limited to these elevated concentrations. Specifically, in the FPA, only 1 of the 47 mercury results is above the alternative ESV. Lead (8 of 47 detections) and zinc (11 of 47 detections) at the FPA, and zinc (4 of 30 detections) at the NPA also have relatively limited exceedances of the alternative ESVs compared to the overall number of detections. Thus, COPEC exposure to receptors would be relatively small, suggesting that lead, mercury, and zinc can be eliminated from further consideration and would not be final COPECs.

Sulfide needs to be further evaluated because all detections exceeded the only available ESV of 0.00358 mg/kg.

Other Considerations. Sulfide was only measured to better characterize the soil and not for risk purposes in the Phase I RI (MKM 2005a). However, once the measurements were available, a risk assessment was conducted by MKM Engineers, Inc. (MKM). Sulfide became a historical COPEC because the MDC (244 mg/kg) exceeded the ESV (0.00358 mg/kg). Given the absence of background concentrations for sulfide at RVAAP and in Ohio, it was not possible to determine if AOC concentrations were similar to background concentrations. To provide perspective, sulfide was detected at a maximum of 928 mg/kg and a median of 117 mg/kg in background samples in Massachusetts (BBL 2002). In Kentucky, soil background concentrations for sulphur were used as a proxy for sulfide; background soil concentrations ranged between 260 and 1,600 mg/kg (DOE 1995). Thus, the MDC for sulfide at Load Line 11 is within the range of these background concentrations. Although there is no other known ESV for sulfide, there are other protective values for sulfides in soil. For example, human health values, often based on animal studies, are greater than 1,000,000 mg/kg for hydrogen sulfide, according to USEPA (2010b). This is several orders of magnitude higher than the 0.00358 mg/kg ESV.

Sulphur and sulfides are naturally occurring in soil and sediment, and sulfides can help reduce toxicity of contaminants. Sulfides bind with metals and make them less bioavailable than their oxide, hydroxide, carbonate, and sulfate forms. Metals in soil and sediment would also show reduced toxicity when they are bound to the sulfides. The above information suggests that sulfide can be eliminated from further consideration and would not be a final COPEC.

Evaluation of Integrated COPECs in Sediment and Surface Water. Integrated COPECs in sediment and surface water were subjected to the same evaluation and refinement factors applied to integrated COPECs for surface soil. Some factors, such as those concerning wetlands and off-site migration, apply to the whole AOC and will be discussed later. Evaluating sediment and surface water includes chemical-specific evaluation and refinement factors and the qualitative relationship of exposure area to general home range.

Sediment. The three integrated COPECs in sediment at the East Ditch are beryllium, selenium, and benz(a)anthracene (Table 7-17), and the three integrated COPECs in sediment at the West Ditch are beryllium, mercury, and sulfide (Table 7-18). Beryllium, selenium, and sulfide were identified as integrated COPECs by default (i.e., the chemicals did not have an ESV) and are discussed later in Step 3A and in Section 7.3.3.9. Mercury did not exceed its ESV but was retained as a COPEC because it is a PBT compound. Mercury is discussed later in Step 3A. One COPEC [benz(a)anthracene] had an MDC that exceeded the ESV and is discussed below.

 Benz(a)anthracene. Benz(a)anthracene was detected in the one discrete sediment sample collected in the East Ditch, and the single detection slightly exceeded the ESV (Table 7-17; Appendix H, Table H-12). Benz(a)anthracene does not have a background concentration for comparison; therefore, it was evaluated further. The magnitude of ESV exceedance was very low; the MDC to ESV ratio was 1.02

(Appendix H, Table H-12). The small ratio indicates risk to ecological receptors is small and this supports eliminating benz(a)anthracene as an integrated COPEC for sediment in the East Ditch.

Surface Water. The two integrated COPECs in surface water at the East Ditch are beta-BHC and gamma-chlordane (Table 7-19) and the four integrated COPECs in surface water at the West Ditch are manganese, benzo(a)pyrene, benzo(k)fluoranthene, and beta-BHC (Table 7-20). Benzo(k)fluoranthene was identified as an integrated COPEC because it did not have an ESV and is discussed later in Step 3A and in Section 7.3.3.9. Beta-BHC and gamma-chlordane did not exceed their ESVs but were retained as COPECs because they are PBT compounds. Beta-BHC and gamma-chlordane are discussed later in Step 3A. Two COPECs (manganese and benzo(a)pyrene) had MDCs that exceeded the ESVs and are discussed below.

Manganese. Manganese (0.559 mg/L) was detected above its background concentration (0.39 mg/L) and the ESV (0.12 mg/L) in the single West Ditch surface water sample included in the PBA08 RI. Although the MDC to ESV ratio for manganese (4.7) in surface water at the West Ditch indicates a possibility of risk, manganese had a concentration less than the background concentration for manganese in the West Ditch sediment, and mean concentrations in both soil EUs were below the background concentration. This information suggests limited source material for the elevated detections in the surface water, and this chemical may not become a final COPEC; however, manganese will undergo further evaluation.

Benzo(a)pyrene. Benzo(a)pyrene (0.00033 mg/L) was detected above its OMZA ESV (0.00006 mg/L) in the single West Ditch surface water sample included in the PBA08 RI. Benzo(a)pyrene does not have a background concentration for comparison. Benzo(a)pyrene was below its OMZM ESV and became a COPEC when compared to the OMZA. Although the MDC to ESV ratio for benzo(a)pyrene (5.5) in surface water at the West Ditch indicates a possibility of risk, the MDC for benzo(a)pyrene was below its ESV in the sediment of the West Ditch and in both soil EUs. This information suggests limited source material for the elevated detection in the surface water, and this chemical may not become a final COPEC; however, benzo(a)pyrene will undergo further evaluation.

Summary. In summary, no sediment COPECs required further evaluation. The only integrated COPECs retained for further evaluation in surface water are manganese and benzo(a)pyrene in the West Ditch.

Comparison of Ohio EPA Approved and Preferred ESVs. The Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008) gives guidance on selecting media screening values (ESVs) for Level II evaluation. For surface water, five possible sources of ESV values are listed in order of preference: (1) Ohio EPA Administrative Code OMZAs, (2) Ohio EPA Administrative Code OMZMs, (3) National Ambient WQC, (4) Tier II values (Suter & Tsao 1996), and (5) Region 5 Ecological Screening Levels (USEPA 2003a). It is possible that the chosen ESV is too conservative. Consequently, alternative ESVs are presented below for the remaining two integrated COPECs in surface water at the West Ditch.

The Ohio EPA approved and preferred manganese ESV used in this ERA is 0.12 mg/L. This value is from Tier II (Suter & Tsao 1996) (Appendix H, Table H-9). Manganese does not have any known alternative surface water ESVs.

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The Ohio EPA approved and preferred benzo(a)pyrene ESV used in this ERA is 0.00006 mg/L. This ESV is from the Ohio EPA Administrative Code OMZAs (Appendix H, Table H-9). The benzo(a)pyrene OMZA ESV used in this ERA is about 10 times lower than the OMZM ESV (0.00054 mg/L) and about 4 times higher than the other available surface water ESV for benzo(a)pyrene (0.000014 mg/L, USEPA 2003a).

Qualitative Relationship of Exposure Area to General Home Range. The West Ditch drains the central portion of the AOC and flows southwest through Wetland 1 to an unnamed tributary to Sand Creek. The West Ditch receives surface water from a network of small ditches. Intermittent surface water flows in drainage ditches at Load Line 11. SAIC scientists have observed no water in the West Ditch during repeated visits to the area during sampling and other activities. The intermittent surface water in the West Ditch is not sufficient to create and maintain aquatic habitat. Any ecological receptor that uses Load Line 11 as part of its home range would have only incidental contact with the West Ditch. The relatively low exposure supports removing manganese and benzo(a)pyrene as integrated COPECs in surface water in the West Ditch.

Wetland Quality, Geographical Information, and On-site Migration of Chemicals. The next three evaluation and refinement factors are concerned with risk to wetlands. The three factors are:

- Categorizing wetland quality inside the AOC,
- Evaluating geographical relationship of on-site wetlands to AOC exceedance area, and
- Obtaining information about on-site migration of chemicals to on-site wetlands.

If the wetland quality is low, it is distant from the AOC exceedance area (i.e., high concentration area), or on-site migration is unlikely, it increases the likelihood that the remaining integrated COPECs in soil, sediment, and surface water at the AOC [lead, mercury, zinc, sulfide, manganese, and benzo(a)pyrene] will not be of ecological concern to wetlands and do not need to be evaluated as final COPECs.

There are two small wetlands at Load Line 11. Wetland 1 is a planning level survey wetland that straddles the western boundary of the AOC (Figure 7-1) (OHARNG 2014). It is a Category 1 wetland (with an ORAM score of 28). Wetland 2 was not previously identified on any wetland surveys and has likely formed as a result of blocked drainage in the East Ditch. It is a Category 1 wetland (with an ORAM score of 21). Category 1 indicates low wetland quality, with degradation of wetland functions.

Wetland 1 is small in size, with 0.13 acres inside the AOC (Figure 7-1). Wetland 2 is about 0.02 acres. RVAAP contains about 1,970 acres of wetlands, and the 0.15 acres of wetlands inside the habitat boundary at Load Line 11 represent 0.008% of the total wetland area of RVAAP (OHARNG 2014). The relatively small area and availability of many more wetland acres at RVAAP lowers the importance of the wetlands at Load Line 11.

The main mechanism for migration of on-site COPECs to the wetlands is via the ditches. Chemicals in soil that migrate into the ditches are then transported to the wetlands in ditch sediment and surface water. The West Ditch drains toward Wetland 1 and becomes less channelized approximately 200 ft southeast of Wetland 1. However, it is possible that contaminants in the ditch or in the soil immediately adjacent to Wetland 1 could migrate into and potentially impact the wetland. Similarly, the East Ditch drains directly into Wetland 2, which has likely formed around a blocked or obstructed culvert in the ditch (Photograph 7-2).

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Any COPECs impacting Wetland 1 most likely originate from the West Ditch sediment and surface water. In the West Ditch, the COPECs are beryllium, mercury, and sulfide in sediment and manganese, benzo(a)pyrene, benzo(k)fluoranthene, and beta-BHC in surface water. No direct chemical concentration measurements were taken in the wetland, but adjacent soil samples were taken at LL11sb-066 and LL11ss-028. Manganese, mercury, and sulfide were found in the adjacent soil samples, but only sulfide was at high concentrations (i.e., average and maximum concentrations above the background concentration and ESV). In the West Ditch sediment closest to the wetland, mercury was at a lower concentration than in the soil, and sulfide was not reported. Mercury was also below its background concentration, SRV, and ESV. Manganese and benzo(a)pyrene were not COPECs in either soil or sediment near Wetland 1. This suggests that the remaining integrated COPECs in soil, sediment, and surface water are unlikely to impact Wetland 1. Importantly, there is a biological/water quality station downstream of Wetland 1 where the five assessment attributes exhibit such rankings as excellent, good, and exceptional (Table 7-13).

Any COPECs impacting Wetland 2 most likely originate from the East Ditch sediment and surface water. In the East Ditch, the COPECs are beryllium, selenium, and benz(a)anthracene in sediment and beta-BHC and gamma-chlordane in surface water. No chemical concentration measurements were taken directly in the small 0.02 acre area. However, samples LL11sd/sw-083, LL11sd-017, and LL11sw-012 were taken directly adjacent to the wetland. The most recent sample (LL11sd/sw-083) was included in the East Ditch data set, but the two older samples it replaced (LL11sd-017 and LL11sw-012) had similar or lower concentrations of sediment and surface water COPECs. None of the East Ditch sediment and surface water COPECs were of ecological concern in the FPA soil (i.e., soil COPECs). Further, although some soil COPECs (arsenic, cadmium, cobalt, lead, manganese, and zinc) were detected in the East Ditch, concentrations in sediment were lower, and none were at levels of ecological concern (i.e., above background concentrations, SRVs, and/or ESVs) in the ditch. Thus, no nearby soil, sediment, or surface water COPECs are likely to adversely affect the small wetland. Importantly, vegetation appears healthy at Wetland 2, and the downstream biological/water quality station exhibits rankings of excellent, good, and exceptional for the five assessment attributes (Table 7-13).

Evaluation of Biological and Water Quality Sampling Stations. The last evaluation and refinement factor is:

• Evaluating off-site migration of chemicals at biological/water quality stations.

Various biological measurements of macroinvertebrates and fish, as well as chemical and physical measurements of surface water and sediment, were taken and assessed for evidence of upgradient and downgradient contamination. These studies were published in the *Facility-wide Biological and Water Quality Study* (USACE 2005a). Monitoring stations are positioned in streams and ponds downgradient of several AOCs. One sampling station (S-1) is upstream of Load Line 11, and one station (S-2) is downstream of the AOC. When the assessment attributes are positively rated (e.g., "good," "excellent," "full attainment status"), this is evidence the downstream sampling station has not been impaired by upstream chemical conditions at the AOC.

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The measurements taken at each station are sediment chemistry and conditions, surface water chemistry and conditions, fish community, benthic macroinvertebrate community, and habitat conditions. Also, a statement about attainment status is provided. Table 7-13 shows the results of the attributes for both sampling stations (S-1 and S-2). Review of the *Facility-Wide Biology and Water Quality Study* (USACE 2005a) data from the two sampling stations showed many positive metrics and no sign of aquatic impairment.

Table 7-13 shows the biological, sediment, and water quality attributes at the downstream location (S-2) indicate little to no impairment (i.e., Full Attainment Status). No inorganic or organic chemicals were detected above their screening criteria; therefore, chemicals at Load Line 11 have not adversely affected downstream ecological conditions. This greatly reduces concern for the remaining COPECs (in soil, sediment, and water) at Load Line 11.

Evaluation of PBT Compounds and COPECs Without ESVs. As discussed in Level II, there are two chemicals that are PBT compounds in surface soil (mercury and PCB-1254), one chemical that is a PBT compound in sediment (mercury), and two chemicals that are PBT compounds in surface water (beta-BHC and gamma-chlordane). Six chemicals (nitrate, HMX, nitrocellulose, nitroguanidine, PETN, and tetryl) are integrated COPECs by default (i.e., the chemicals did not have ESVs) in surface soil, three (beryllium, selenium, and sulfide) are COPECs by default in sediment, and one [benzo(k)fluoranthene] is a COPEC by default in surface water at Load Line 11. These chemicals are briefly evaluated below.

PBT Compounds. The Guidance for Conducting Ecological Risk Assessments (Ohio EPA 2008) includes a PBT compound screen in the Level II ERA. This screen is necessary because not all ESVs account for bioaccumulation; instead, they are derived based primarily on toxicity to endpoint receptors exposed by direct contact (e.g., plants, soil-dwelling invertebrates) or ingesting soil or water (e.g., mammals, birds). For AOCs that move to a Level III baseline ERA, PBT compounds are evaluated in the wildlife food chains. Ohio EPA allows PBTs to be screened out in Level II if the "method used to derive the screening value considered exposure to higher trophic level organisms in the development of the screening value" (Ohio EPA 2008).

For the first two sources of soil ESVs preferred by Ohio EPA (i.e., EcoSSLs and PRGs), bioaccumulation in higher trophic levels is considered in development of the ESV. According to EcoSSL guidance, "wildlife receptors may be exposed to contaminants in soil by two main pathways: incidental ingestion of soil while feeding, and ingestion of food items that have become contaminated

due to uptake from soil" (USEPA 2007b). Deriving EcoSSL values includes uptake equations that account for direct ingestion and food chain bioaccumulation (USEPA 2007b). The same is true of PRGs: "the 90th percentile of the soil-to-biota uptake factor was used as a conservative estimate of the chemical concentrations in wildlife food types (earthworms, plants, or small mammals)," and "the model accounts for the ingestion of soil as well as food" (DOE 1997). It is also important to note that both sources often derive values for multiple receptors, and the most conservative (lowest) value is chosen. Thus, for soil ESVs from these two sources, PBT compounds that have ratios less than one can be dismissed as final COPECs.

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The USEPA Region 5 ESLs (USEPA 2003a) are an Ohio EPA-approved source for soil, sediment, and surface water ESVs (see Appendix H, Tables H-7 to H-9 for hierarchies); they are a source of screening values for some PBT compounds not covered by the EcoSSLs or PRGs. The ESLs also account for bioaccumulation in the food chain, as "development of ESLs focused on mammalian or avian species and identified those chemicals that have the potential for significant bioaccumulation or biomagnification" (USEPA 1999b). Thus, development of ESLs included bioaccumulation in higher tropic levels as a standard component in the equations, and PBT compounds with ESVs in soil, sediment, and surface water from this source that have ratios less than one can be dismissed as final COPECs. For this ERA, this includes PCB-1254 in surface soil at the NPA and beta-BHC in surface water in the East Ditch and West Ditch. PCB-1254 in FPA surface soil was previously evaluated in Step 3A, and it will not be a final COPEC because the MDC (0.42 mg/kg) is just slightly greater than the ESV (0.371 mg/kg) and the mean concentration (0.0562 mg/kg) is less than the ESV.

Mercury was detected in both soil EUs above the ESV, which accounts for bioaccumulation (DOE 1997). These exceedances are not surprising considering the background concentration is 70 times greater than the ESV. At the NPA, the mean concentration of mercury (0.034 mg/kg) is below the background concentration (0.036 mg/kg) (Table 7-23); therefore, mercury was eliminated from further consideration and will not be a final COPEC at NPA. While the mean concentration of mercury (0.0397 mg/kg) at the FPA is slightly above the background concentration (0.036 mg/kg) (Table 7-22), the similarity between the concentrations suggests exposures to mercury at Load Line 11 are no different from background concentrations. Thus, mercury at the FPA was eliminated from further consideration and will not be a final COPEC.

In the West Ditch, mercury is the only chemical in sediment identified as an integrated COPEC because it is a PBT compound. While the mean concentration and MDC are below the ESV, the sediment ESV does not account for bioaccumulation. However, the MDC (0.049 mg/kg) is below the background concentration (0.059 mg/kg) (Appendix H, Table H-13). Mercury was found in 1 of 5 samples in the ditches. Thus, the frequency is low. These facts support the view that mercury does not need to be retained as a final COPEC. Therefore, mercury was eliminated from further consideration and will not be a final COPEC in sediment.

In the East Ditch, gamma-chlordane was identified as an integrated COPEC for surface water because it is a PBT compound. The MDC (0.000015 mg/kg) is below the ESV (0.0043 mg/kg) (Appendix H, Table H-14); therefore, gamma-chlordane was eliminated from further consideration and will not be a final COPEC.

PBT compounds at Load Line 11 are mercury and PCB-1254 in surface soil, mercury in sediment, and beta-BHC and gamma-chlordane in surface water. As discussed above, all PBT compounds are dismissed and will not be final COPECs.

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COPECs without ESVs. The Guidance for Conducting Ecological Risk Assessments specifies chemicals without screening benchmark values should be retained as COPECs (Ohio EPA 2008). While Ohio EPA allows the use of additional screening benchmark values, such values need to be approved prior to submitting the report. For Load Line 11, a search for (and subsequent approval of) additional values was not deemed necessary. Rather, to mitigate concern for the uncertainties associated with COPECs that lack ESVs, a limited additional evaluation was conducted for each medium, focusing on frequency of detection, relationship to background concentration, and other chemical-specific refinement factors.

For soil at the FPA, the integrated COPECs without ESVs are nitrate, nitrocellulose, nitroguanidine, PETN, and tetryl. Nitrate was detected in 7 of 34 samples, nitroguanidine was detected in 1 of 4 samples, PETN was detected in 1 of 7 samples, and tetryl was detected in 1 of 41 samples. Thus, exposure to these chemicals would be limited. While nitrocellulose was detected in 4 of 4 samples, it is essentially non-toxic (USEPA 1987), and this chemical is also not expected to be an ecological concern. For soil at the NPA, the integrated COPECs without ESVs are nitrate, HMX, nitrocellulose, and PETN. Nitrate was detected in 6 of 17 samples, HMX was detected in 1 of 29 samples, and PETN was detected in 1 of 12 samples. Thus, exposure to these chemicals would be limited. While nitrocellulose was detected in 3 of 5 samples, it is essentially non-toxic (USEPA 1987), and this chemical is also not expected to be an ecological concern. In addition, nitrate is unlikely to be of ecological concern. As discussed in the Phase I RI (MKM 2005a), nitrate analysis was conducted primarily for gross characterization of AOC soil conditions and not for risk assessment purposes. Although it is possible that nitrate is related to the AOC, nitrates in soil are not typically considered to be toxic and are often added to soil as fertilizer.

For sediment in the East Ditch, beryllium and selenium were the only identified integrated COPECs without ESVs. Beryllium was detected in two of two sediment samples; however, the mean concentration (0.81 mg/kg) is essentially the same as the SRV (0.8 mg/kg). Selenium was detected in two of two sediment samples; however, the mean concentration (2.6 mg/kg) is similar to the SRV (1.7 mg/kg). Beryllium and selenium are not considered of ecological concern in sediment.

For sediment in the West Ditch, beryllium and sulfide were the only identified integrated COPECs without ESVs. Beryllium was detected in three of three sediment samples; however, the mean concentration (0.81 mg/kg) is essentially the same as the SRV (0.8 mg/kg). Sulfide was detected in one of one sediment sample. Sulfide toxicity varies as a function of its binding with metals. Binding with metals lowers toxicity. Given the number of detected metals in the sediment and the principle of binding by sulfides, it is likely that sulfide toxicity is relatively low. As discussed in the Phase I RI (MKM 2005a), sulfide analysis was conducted primarily for gross characterization of AOC conditions and not for risk assessment purposes. Beryllium and sulfide are not considered of ecological concern in sediment.

Surface water in the East Ditch did not have integrated COPECs without ESVs. For surface water in the West Ditch, one PAH [benzo(k)fluoranthene] was the only identified integrated COPEC without an ESV for comparison. Benzo(k)fluoranthene was detected in one surface water sample at a low concentration of 0.00024 mg/L. While benzo(k)fluoranthene does not have an ESV, it is detected below all ESVs for other detected PAHs in surface water [e.g., 0.023 mg/L ESV for benzo(b)fluoranthene]. In addition, benzo(k)fluoranthene was detected below its ESV in all soil and sediment samples at Load Line 11. As a result, benzo(k)fluoranthene is not likely an ecological concern for surface water.

Summary of Findings in Step 3A. Of the 20 integrated COPECs in surface soil at Load Line 11, the six that did not have ESVs – nitrate, HMX, nitrocellulose, nitroguanidine, PETN, and tetryl – were eliminated as COPECs because they had low frequency of detection or little to no toxicity. Additional integrated COPECs were eliminated from further consideration because the mean concentration is smaller than the ESV (arsenic, cadmium, chromium, cobalt, copper, cyanide, and PCB-1254) or the mean concentration is smaller than the background concentration (aluminum, manganese, and vanadium). One PBT compound (PCB-1254) was eliminated because it had a ratio less than one, using an ESV that accounted for bioaccumulation.

The remaining four integrated COPECs in soil (lead, mercury, zinc, and sulfide) have a combination of factors that together eliminated them from further consideration as presented below.

- 1. Most mean concentrations are only slightly higher than background concentrations.
- 2. Many mean concentration-to-ESV ratios of exceedance are near one.
- 3. Ohio EPA guidance allows alternative ESVs that are less conservative than the ESVs used in this ERA. These alternate ESVs (unlike the preferred ESV) are above background concentrations. If these alternate ESVs were used in lieu of current ESVs, ratios for lead, mercury, and zinc would be below one.
- 4. Samples with highly elevated concentrations are scattered widely, and few ecological receptors are exposed to soil with detected concentrations above the alternative ESVs.
- 5. The wetlands are relatively small and of low quality, and on-site migration is possible.
- 6. Off-site downstream sampling station indicates a healthy environment.

No final COPECs were identified for Load Line 11 surface soil.

Of the five integrated COPECs in sediment, beryllium, selenium, and sulfide have no ESVs. They were eliminated as COPECs because the mean concentration is essentially the same as the SRV for beryllium and selenium. Sulfide was collected for gross soil characterization reasons, concentrations were similar to available background concentrations, and it is often bound with metals; these points helped to eliminate sulfide as a COPEC. Benz(a)anthracene was eliminated because the MDC was similar to the ESV. Despite mercury being a PBT compound with an ESV that did not recognize bioaccumulation, mercury was eliminated due to the similarity of mean concentration and background concentration, low mean-to-ESV ratio, and the positive conditions in the downstream station in nearby Sand Creek. No final COPECs were identified for Load Line 11 sediment.

Of the five integrated COPECs in surface water, benzo(a) fluoranthene had no ESV. PBT compounds beta-BHC and gamma-chlordane were eliminated because the MDCs are considerably lower than the ESVs. Manganese and benzo(a) pyrene lack sources in soil and sediment media that contribute to surface water; mean concentrations in soil and maximum concentrations in sediment for manganese are below background, and maximum concentrations for benzo(a) pyrene in both media are below ESVs. No final COPECs were identified for Load Line 11 surface water.

7.3.3.8 Consideration of Human Health Driven Remediation

After the 2001 IRA was conducted [i.e., removing sumps, 230 yd³ of contaminated media in drainage ditches, and a hot spot of petroleum-contaminated soil (MKM 2004a)], the HHRA did not identify any additional COCs for remediation and recommends no further action for surface soil, sediment, and surface water at Load Line 11. Thus, there will be no remediation implemented to further protect human health that would also protect ecological resources at Load Line 11.

7.3.3.9 Uncertainties and Mitigations

Uncertainties or unknowns are present in both exposure and effects data. To mitigate uncertainty in exposure data, the MDCs of all available and appropriate data were used in Level II. In Level II, the MDC and mean COPEC concentrations were used to mitigate uncertainty concerning exposure data for receptors in the AOC. To mitigate uncertainty in effects information, a site visit for habitat condition was conducted, and the latest INRMP of rare species sightings and jurisdictional wetlands was used (OHARNG 2014). In addition, the ORAM was applied to the wetlands. Conservative ESVs, which are typically based on concentrations observed to have no effect on test species in laboratory studies, were used in Level II to mitigate uncertainty concerning effects on receptors in the AOC. There are some COPECs by default (i.e., the chemicals do not have ESVs). These COPECs are assumed to have limited toxicity given the lack of cause-effect laboratory tests and field-observed effects in the scientific literature.

In Level II, to mitigate uncertainty concerning effects on receptors in the AOC, the ESVs for COPECs are compared to background concentrations. Using ESVs that are lower than background concentrations indicates the conservative nature of the evaluation. Conservative ESVs are appropriate for use as screening thresholds in Level I and II (i.e., soil constituents with an MDC below the ESV need no further consideration in Level II).

7.3.3.10 Summary and Recommendations of Screening Level Ecological Risk Assessment

Integrated COPECs were identified in surface soil, sediment, and surface water at Load Line 11. Most of the soil COPECs identified in the historical ERA were also identified during screening of PBA08 RI data. Three historical COPECs were not identified during the PBA08 RI, and a few new COPECs were identified due to additional sampling data. Some comparable historical sediment and surface water data were available and were considered. Those chemicals retained after screening historical and PBA08 RI data were termed integrated COPECs.

Twenty integrated soil COPECs, five integrated sediment, and five integrated surface water COPECs were further evaluated in Step 3A with technical and refinement factors. All integrated soil, sediment, and surface water COPECs were determined to be of no ecological concern. None require remediation or further evaluation. Consequently, the ERA for Load Line 11 can conclude with a Level II ERA that no further action is necessary to be protective of important ecological receptors.

1 2

7.3.4 Conclusions

As indicated in Section 4.2, in 2001 an IRA was performed as an early response to contamination in primary pathways at Load Line 11. This IRA included removing sumps, 230 yd³ of contaminated media in drainage ditches, and a hot spot of petroleum-contaminated soil (MKM 2004a). This ERA concludes that there is chemical contamination currently present in surface soil, sediment, and surface water at Load Line 11. This contamination was identified using historical and PBA08 RI data. Dry, mid-successional, cold-deciduous shrubland (dominant vegetation type); dry, mid-successional, herbaceous field; and four types of forests were observed on the 48 acres of the AOC. Wetlands are important and significant ecological resources near contamination in the AOC. These findings invoked a Level II assessment.

The Level II assessment evaluated soil, sediment, and surface water using historical and PBA08 RI data, identified integrated COPECs, and evaluated the integrated COPECs using technical and refinement factors in Step 3A. The factors in Step 3A showed there is no further evaluation necessary for integrated COPECs, and there is no ecological concern requiring additional remediation. Consequently, the ERA for Load Line 11 can conclude with Level II ERA that no further action is necessary to be protective of important ecological receptors. This supports the conclusion from the Phase I RI, which stated:

"No additional action is recommended based on ecological risk because, based on the Step 3a refinement step, it was determined that site-related risks were not great enough to warrant proceeding further into the ecological risk assessment process."

Table 7–1. Risk Assessment Data Set for Surface Soil (0-1 ft bgs) Discrete Samples

| Location | Sample ID | Date | Depth (ft bgs) |
|------------|--------------------|------------|----------------|
| | Former Product | ion Area | |
| LL11cs-006 | LL11cs-006-0001-SD | 3/21/2001 | 0 - 1 |
| LL11cs-010 | LL11cs-010-0001-SD | 3/21/2001 | 0 - 1 |
| LL11cs-011 | LL11cs-011-0001-SD | 3/23/2001 | 0 - 1 |
| LL11cs-016 | LL11cs-016-0001-SD | 3/23/2001 | 0 - 1 |
| LL11cs-020 | LL11cs-020-0001-SO | 3/23/2001 | 0 - 1 |
| LL11cs-021 | LL11cs-021-0001-SD | 3/21/2001 | 0 - 1 |
| LL11cs-025 | LL11cs-025-0001-SD | 3/21/2001 | 0 - 1 |
| LL11cs-030 | LL11cs-030-0001-SD | 3/21/2001 | 0 - 1 |
| LL11sb-001 | LL11sb-001-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-003 | LL11sb-003-0001-SO | 10/30/2000 | 0 - 1 |
| LL11sb-004 | LL11sb-004-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-008 | LL11sb-008-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-009 | LL11sb-009-0001-SO | 10/30/2000 | 0 - 1 |
| LL11sb-019 | LL11sb-019-0001-SO | 8/21/2000 | 0 - 1 |
| LL11sb-020 | LL11sb-020-0001-SO | 8/21/2000 | 0 - 1 |
| LL11sb-021 | LL11sb-021-0001-SO | 8/21/2000 | 0 - 1 |
| LL11sb-022 | LL11sb-022-0001-SO | 8/22/2000 | 0 - 1 |
| LL11sb-023 | LL11sb-023-0001-SO | 8/22/2000 | 0 - 1 |
| LL11sb-060 | LL11sb-060-5551-SO | 3/18/2010 | 0 - 1 |
| LL11sb-061 | LL11sb-061-5555-SO | 3/17/2010 | 0 - 1 |
| LL11sb-062 | LL11sb-062-5559-SO | 3/18/2010 | 0 - 1 |
| LL11sb-064 | LL11sb-064-5569-SO | 3/18/2010 | 0 - 1 |
| LL11sd-023 | LL11sd-023-0001-SD | 11/15/2000 | 0 - 1 |
| LL11sd-025 | LL11sd-025-0001-SD | 11/16/2000 | 0 - 1 |
| LL11sd-026 | LL11sd-026-0001-SD | 11/16/2000 | 0 - 1 |
| LL11sd-031 | LL11sd-031-0001-SD | 11/16/2000 | 0 - 1 |
| LL11ss-001 | LL11ss-001-0001-SO | 11/7/2000 | 0 - 1 |
| LL11ss-002 | LL11ss-002-0001-SO | 11/6/2000 | 0 - 1 |
| LL11ss-003 | LL11ss-003-0001-SO | 11/7/2000 | 0 - 1 |
| LL11ss-004 | LL11ss-004-0001-SO | 11/7/2000 | 0 - 1 |
| LL11ss-008 | LL11ss-008-0001-SO | 11/7/2000 | 0 - 1 |
| LL11ss-009 | LL11ss-009-0001-SO | 11/8/2000 | 0 - 1 |
| LL11ss-010 | LL11ss-010-0001-SO | 11/13/2000 | 0 - 1 |
| LL11ss-013 | LL11ss-013-0001-SO | 11/8/2000 | 0 - 1 |
| LL11ss-014 | LL11ss-014-0001-SO | 11/6/2000 | 0 - 1 |
| LL11ss-015 | LL11ss-015-0001-SO | 11/8/2000 | 0 - 1 |
| LL11ss-016 | LL11ss-016-0001-SO | 11/13/2000 | 0 - 1 |
| LL11ss-017 | LL11ss-017-0001-SO | 11/8/2000 | 0 - 1 |
| LL11ss-018 | LL11ss-018-0001-SO | 11/9/2000 | 0 - 1 |
| LL11ss-019 | LL11ss-019-0001-SO | 11/10/2000 | 0 - 1 |
| LL11ss-020 | LL11ss-020-0001-SO | 11/10/2000 | 0 - 1 |
| LL11ss-021 | LL11ss-021-0001-SO | 11/9/2000 | 0 - 1 |
| LL11ss-022 | LL11ss-022-0001-SO | 11/9/2000 | 0 - 1 |
| LL11ss-023 | LL11ss-023-0001-SO | 11/9/2000 | 0 - 1 |
| LL11ss-024 | LL11ss-024-0001-SO | 11/9/2000 | 0 - 1 |
| LL11ss-025 | LL11ss-025-0001-SO | 11/9/2000 | 0 - 1 |
| LL11ss-071 | LL11ss-071-5597-SO | 4/12/2010 | 0 - 1 |

Table 7-1. Risk Assessment Data Set for Surface Soil (0-1 ft bgs) Discrete Samples (continued)

| Location | Sample ID | Date | Depth (ft bgs) |
|-------------------------|--------------------|------------|----------------|
| LL11ss-074 | LL11ss-074-5600-SO | 4/12/2010 | 0 - 1 |
| LL11ss-076 | LL11ss-076-5602-SO | 4/12/2010 | 0 - 1 |
| LL11ss-072 ^a | LL11ss-072-5598-SO | 4/12/2010 | 0 - 1 |
| LL11ss-073 ^a | LL11ss-073-5599-SO | 4/12/2010 | 0 - 1 |
| LL11ss-075 ^a | LL11ss-075-5601-SO | 4/12/2010 | 0 - 1 |
| | Non-Productio | n Area | |
| LL11cs-039 | LL11cs-039-0001-SD | 3/20/2001 | 0 - 1 |
| LL11sb-002 | LL11sb-002-0001-SO | 10/30/2000 | 0 - 1 |
| LL11sb-005 | LL11sb-005-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-006 | LL11sb-006-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-007 | LL11sb-007-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-010 | LL11sb-010-0001-SO | 10/31/2000 | 0 - 1 |
| LL11sb-063 | LL11sb-063-5563-SO | 3/22/2010 | 0 - 1 |
| LL11sb-065 | LL11sb-065-5573-SO | 3/22/2010 | 0 - 1 |
| LL11sb-066 | LL11sb-066-5577-SO | 3/22/2010 | 0 - 1 |
| LL11sb-067 | LL11sb-067-5581-SO | 3/18/2010 | 0 - 1 |
| LL11sb-068 | LL11sb-068-5585-SO | 3/18/2010 | 0 - 1 |
| LL11sb-069 | LL11sb-069-5589-SO | 3/22/2010 | 0 - 1 |
| LL11sd-014 | LL11sd-014-0001-SD | 11/14/2000 | 0 - 1 |
| LL11sd-015 | LL11sd-015-0001-SD | 11/14/2000 | 0 - 1 |
| LL11sd-016 | LL11sd-016-0001-SD | 11/14/2000 | 0 - 1 |
| LL11sd-020 | LL11sd-020-0001-SD | 11/15/2000 | 0 - 1 |
| LL11sd-022 | LL11sd-022-0001-SD | 11/15/2000 | 0 - 1 |
| LL11sd-029 | LL11sd-029-0001-SD | 11/16/2000 | 0 - 1 |
| LL11ss-026 | LL11ss-026-0001-SO | 11/10/2000 | 0 - 1 |
| LL11ss-027 | LL11ss-027-0001-SO | 11/10/2000 | 0 - 1 |
| LL11ss-028 | LL11ss-028-0001-SO | 11/10/2000 | 0 - 1 |
| LL11ss-029 | LL11ss-029-0001-SO | 11/14/2000 | 0 - 1 |
| LL11ss-030 | LL11ss-030-0001-SO | 11/14/2000 | 0 - 1 |
| LL11ss-031 | LL11ss-031-0001-SO | 11/10/2000 | 0 - 1 |
| LL11ss-070 | LL11ss-070-5596-SO | 4/12/2010 | 0 - 1 |
| LL11ss-077 | LL11ss-077-5603-SO | 4/12/2010 | 0 - 1 |
| LL11ss-078 | LL11ss-078-5604-SO | 4/12/2010 | 0 - 1 |
| LL11ss-079 | LL11ss-079-5605-SO | 4/12/2010 | 0 - 1 |
| LL11ss-080 | LL11ss-080-5606-SO | 4/12/2010 | 0 - 1 |
| LL11ss-081 | LL11ss-081-5607-SO | 4/12/2010 | 0 - 1 |

^aChromium speciation samples used to evaluate the presence of hexavalent chromium. bgs = Below ground surface. Ft = Feet.

ID = Identification.

Table 7–2. Risk Assessment Data Set for Subsurface Soil (1-13 ft bgs): Discrete Samples

| Location | Sample ID | Date | Depth (ft bgs) |
|------------|--------------------|------------|----------------|
| Location | Former Produ | | zopen (it ngs) |
| LL11cs-007 | LL11cs-007-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-008 | LL11cs-008-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-009 | LL11cs-009-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-012 | LL11cs-012-0001-SO | 3/23/2001 | 1 - 2 |
| LL11cs-017 | LL11cs-017-0001-SO | 3/23/2001 | 1 - 2 |
| LL11cs-018 | LL11cs-018-0001-SO | 3/23/2001 | 1 - 2 |
| LL11cs-019 | LL11cs-019-0001-SO | 3/23/2001 | 1 - 2 |
| LL11cs-022 | LL11cs-022-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-023 | LL11cs-023-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-024 | LL11cs-024-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-026 | LL11cs-026-0001-SD | 3/21/2001 | 1 - 2 |
| LL11cs-027 | LL11cs-027-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-028 | LL11cs-028-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-029 | LL11cs-029-0001-SO | 3/21/2001 | 1 - 2 |
| LL11cs-032 | LL11cs-032-0001-SO | 3/20/2001 | 1 - 2 |
| LL11cs-033 | LL11cs-033-0001-SO | 3/20/2001 | 1 - 2 |
| LL11sb-060 | LL11sb-060-5552-SO | 3/18/2010 | 1 - 4 |
| LL11sb-061 | LL11sb-061-5556-SO | 3/17/2010 | 1 - 4 |
| LL11sb-062 | LL11sb-062-5560-SO | 3/18/2010 | 1 - 4 |
| LL11sb-064 | LL11sb-064-5570-SO | 3/18/2010 | 1 - 4 |
| LL11ss-001 | LL11ss-001-0002-SO | 11/7/2000 | 1 - 3 |
| LL11ss-002 | LL11ss-002-0002-SO | 11/6/2000 | 1 - 3 |
| LL11ss-003 | LL11ss-003-0002-SO | 11/6/2000 | 1 - 3 |
| LL11ss-004 | LL11ss-004-0002-SO | 11/7/2000 | 1 - 3 |
| LL11ss-007 | LL11ss-007-0002-SO | 11/13/2000 | 1 - 3 |
| LL11ss-008 | LL11ss-008-0002-SO | 11/7/2000 | 1 - 3 |
| LL11ss-009 | LL11ss-009-0002-SO | 11/8/2000 | 1 - 3 |
| LL11ss-010 | LL11ss-010-0002-SO | 11/13/2000 | 1 - 3 |
| LL11ss-013 | LL11ss-013-0002-SO | 11/8/2000 | 1 - 3 |
| LL11ss-014 | LL11ss-014-0002-SO | 11/7/2000 | 1 - 3 |
| LL11ss-015 | LL11ss-015-0002-SO | 11/8/2000 | 1 - 3 |
| LL11ss-016 | LL11ss-016-0002-SO | 11/13/2000 | 1 - 3 |
| LL11ss-017 | LL11ss-017-0002-SO | 11/8/2000 | 1 - 3 |
| LL11ss-018 | LL11ss-018-0002-SO | 11/9/2000 | 1 - 3 |
| LL11ss-019 | LL11ss-019-0002-SO | 11/10/2000 | 1 - 3 |
| LL11ss-020 | LL11ss-020-0002-SO | 11/10/2000 | 1 - 3 |
| LL11ss-021 | LL11ss-021-0002-SO | 11/9/2000 | 1 - 3 |
| LL11ss-022 | LL11ss-022-0002-SO | 11/9/2000 | 1 - 3 |
| LL11ss-023 | LL11ss-023-0002-SO | 11/9/2000 | 1 - 3 |
| LL11ss-024 | LL11ss-024-0002-SO | 11/9/2000 | 1 - 3 |
| LL11ss-025 | LL11ss-025-0002-SO | 11/9/2000 | 1 - 3 |
| LL11cs-001 | LL11cs-001-0001-SO | 1/24/2001 | 5 - 6 |
| LL11cs-002 | LL11cs-002-0001-SO | 1/26/2001 | 5 - 6 |
| LL11cs-003 | LL11cs-003-0001-SO | 1/26/2001 | 5 - 6 |
| LL11cs-004 | LL11cs-004-0001-SO | 2/21/2001 | 5 - 6 |
| LL11cs-005 | LL11cs-005-0001-SO | 2/21/2001 | 5 - 6 |
| LL11sb-009 | LL11sb-009-0002-SO | 11/13/2000 | 4 - 6 |
| LL11sb-011 | LL11sb-011-0001-SO | 8/21/2000 | 4 - 6 |
| LL11sb-017 | LL11sb-017-0001-SO | 8/22/2000 | 4 - 6 |
| LL11sb-018 | LL11sb-018-0001-SO | 8/22/2000 | 4 - 6 |
| LL11sb-019 | LL11sb-019-0002-SO | 8/21/2000 | 6 - 8 |

Table 7-2. Risk Assessment Data Set for Subsurface Soil (1-13 ft bgs): Discrete Samples (continued)

| Location | Sample ID | Date | Depth (ft bgs) |
|------------|--------------------------------|------------|----------------|
| LL11sb-020 | LL11sb-020-0002-SO | 8/21/2000 | 6 - 8 |
| LL11sb-021 | LL11sb-021-0002-SO | 8/21/2000 | 6 - 8 |
| LL11sb-022 | LL11sb-022-0002-SO | 8/22/2000 | 6 - 8 |
| LL11sb-023 | LL11sb-023-0002-SO | 8/22/2000 | 6 - 8 |
| LL11sb-041 | LL11sb-041-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-042 | LL11sb-042-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-043 | LL11sb-043-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-060 | LL11sb-060-5553-SO | 3/18/2010 | 4 - 7 |
| LL11sb-061 | LL11sb-061-5557-SO | 3/17/2010 | 4 - 7 |
| LL11sb-062 | LL11sb-062-5561-SO | 3/18/2010 | 4 - 7 |
| LL11sb-064 | LL11sb-064-5571-SO | 3/18/2010 | 4 - 7 |
| LL11sb-004 | LL11sb-004-0002-SO | 11/8/2000 | 10 - 12 |
| LL11sb-008 | LL11sb-008-0002-SO | 11/8/2000 | 12 - 13 |
| LL11sb-012 | LL11sb-012-0001-SO | 8/22/2000 | 8 - 10 |
| LL11sb-013 | LL11sb-013-0001-SO | 8/22/2000 | 8 - 10 |
| LL11sb-013 | LL11sb-013-0001-SO | 8/22/2000 | 11 - 13 |
| LL11sb-014 | LL11sb-014-0001-SO | 8/22/2000 | 10 - 12 |
| LL11sb-015 | LL11sb-015-0001-SO | | 8.5 - 11.5 |
| | | 8/22/2000 | - 10 |
| LL11sb-064 | LL11sb-064-5572-SO Non-Product | 3/18/2010 | 7 - 13 |
| LL11cs-013 | LL11cs-013-0001-SO | 3/23/2001 | 1 - 2 |
| LL11cs-013 | | † | |
| | LL11cs-014-0001-SO | 3/23/2001 | |
| LL11cs-034 | LL11cs-034-0001-SO | 3/20/2001 | |
| LL11cs-035 | LL11cs-035-0001-SO | 3/20/2001 | 1 - 2 |
| LL11cs-036 | LL11cs-036-0001-SO | 3/20/2001 | 1 - 2 |
| LL11cs-037 | LL11cs-037-0001-SO | 3/20/2001 | 1 - 2 |
| LL11cs-038 | LL11cs-038-0001-SO | 3/20/2001 | 1 - 2 |
| LL11cs-040 | LL11cs-040-0001-SO | 3/22/2001 | 3 - 7 |
| LL11cs-041 | LL11cs-041-0001-SO | 3/22/2001 | 3 - 7 |
| LL11cs-042 | LL11cs-042-0001-SO | 3/22/2001 | 3 - 7 |
| LL11cs-043 | LL11cs-043-0001-SO | 3/22/2001 | 3 - 7 |
| LL11cs-046 | LL11cs-046-0001-SO | 3/22/2001 | 3 - 4 |
| LL11cs-047 | LL11cs-047-0001-SO | 3/22/2001 | 3 - 4 |
| LL11cs-048 | LL11cs-048-0001-SO | 3/22/2001 | 3 - 4 |
| LL11cs-049 | LL11cs-049-0001-SO | 3/22/2001 | 3 - 4 |
| LL11cs-050 | LL11cs-050-0001-SO | 3/22/2001 | 3 - 4 |
| LL11sb-006 | LL11sb-006-0002-SO | 11/13/2000 | 2 - 4 |
| LL11sb-063 | LL11sb-063-5564-SO | 3/22/2010 | 1 - 4 |
| LL11sb-065 | LL11sb-065-5574-SO | 3/22/2010 | 1 - 4 |
| LL11sb-066 | LL11sb-066-5578-SO | 3/22/2010 | 1 - 4 |
| LL11sb-067 | LL11sb-067-5582-SO | 3/18/2010 | 1 - 4 |
| LL11sb-068 | LL11sb-068-5586-SO | 3/18/2010 | 1 - 4 |
| LL11sb-069 | LL11sb-069-5590-SO | 3/22/2010 | 1 - 4 |
| LL11ss-026 | LL11ss-026-0002-SO | 11/10/2000 | 1 - 3 |
| LL11ss-027 | LL11ss-027-0002-SO | 11/10/2000 | 1 - 3 |
| LL11ss-028 | LL11ss-028-0002-SO | 11/10/2000 | 1 - 3 |
| LL11ss-029 | LL11ss-029-0002-SO | 11/14/2000 | 1 - 3 |
| LL11ss-030 | LL11ss-030-0002-SO | 11/14/2000 | 1 - 3 |
| LL11ss-031 | LL11ss-031-0002-SO | 11/10/2000 | 1 - 3 |
| LL11sb-002 | LL11sb-002-0002-SO | 11/10/2000 | 6 - 8 |
| LL11sb-005 | LL11sb-005-0002-SO | 11/9/2000 | 6 - 8 |
| LL11sb-025 | LL11sb-025-0001-SO | 11/17/2000 | 5 - 7 |

Table 7-2. Risk Assessment Data Set for Subsurface Soil (1-13 ft bgs): Discrete Samples (continued)

| Location | Sample ID | Date | Depth (ft bgs) |
|------------|--------------------|------------|----------------|
| LL11sb-026 | LL11sb-026-0001-SO | 11/16/2000 | 4 - 6 |
| LL11sb-027 | LL11sb-027-0001-SO | 11/16/2000 | 6 - 8 |
| LL11sb-028 | LL11sb-028-0001-SO | 11/16/2000 | 4 - 6 |
| LL11sb-029 | LL11sb-029-0001-SO | 11/16/2000 | 6 - 8 |
| LL11sb-035 | LL11sb-035-0001-SO | 11/17/2000 | 4 - 6 |
| LL11sb-037 | LL11sb-037-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-038 | LL11sb-038-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-039 | LL11sb-039-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-040 | LL11sb-040-0001-SO | 3/13/2001 | 6 - 8 |
| LL11sb-063 | LL11sb-063-5565-SO | 3/22/2010 | 4 - 5 |
| LL11sb-065 | LL11sb-065-5575-SO | 3/22/2010 | 4 - 7 |
| LL11sb-066 | LL11sb-066-5579-SO | 3/22/2010 | 4 - 7 |
| LL11sb-067 | LL11sb-067-5583-SO | 3/18/2010 | 4 - 7 |
| LL11sb-068 | LL11sb-068-5587-SO | 3/18/2010 | 4 - 7 |
| LL11sb-069 | LL11sb-069-5591-SO | 3/22/2010 | 4 - 5 |
| LL11cs-044 | LL11cs-044-0001-SO | 3/22/2001 | 7 - 8 |
| LL11cs-045 | LL11cs-045-0001-SO | 3/22/2001 | 7 - 8 |
| LL11sb-024 | LL11sb-024-0001-SO | 11/16/2000 | 7 - 9 |
| LL11sb-030 | LL11sb-030-0001-SO | 11/17/2000 | 8 - 10 |
| LL11sb-065 | LL11sb-065-5576-SO | 3/22/2010 | 7 - 13 |

bgs = Below ground surface.

Table 7-3. Risk Assessment Data Set for Surface Water

| Location | Sample ID | Date | | | | |
|------------|--------------------|-----------|--|--|--|--|
| | East Ditch | | | | | |
| LL11sw-083 | LL11sw-083-5609-SW | 2/25/2010 | | | | |
| West Ditch | | | | | | |
| LL11sw-084 | LL11sw-084-5610-SW | 2/25/2010 | | | | |

ID = Identification.

Table 7-4. Risk Assessment Data Set for Sediment

| Location | Sample ID | Date | Depth (ft bgs) | | | | |
|------------|--------------------|------------|----------------|--|--|--|--|
| East Ditch | | | | | | | |
| LL11cs-031 | LL11cs-031-0001-SD | 3/20/2001 | 0 - 1 | | | | |
| LL11sd-083 | LL11sd-083-5594-SD | 2/25/2010 | 0 - 0.5 | | | | |
| | West Ditch | ı | | | | | |
| LL11cs-015 | LL11cs-015-0001-SD | 3/23/2001 | 0 - 1 | | | | |
| LL11sd-024 | LL11sd-024-0001-SD | 11/16/2000 | 0 - 1 | | | | |
| LL11sd-084 | LL11sd-084-5595-SD | 2/25/2010 | 0 - 0.5 | | | | |

ID = Identification.

Ft = Feet.

ID = Identification.

bgs = Below ground surface.

Table 7–5. Summary of SRCs

| Surface Soil Subsurface Soil | | | | | | | |
|------------------------------|------------------|-------------|-----------------|--------|-------|--------|---------|
| (0-1 ft) | t bgs) | (1-13 | 3 ft bgs) | Sedin | nent | Surfac | e Water |
| | | FPA | NPA | ED | WD | ED | WD |
| Inorga | nic Ch | emicals | <u> </u> | | | | |
| X | | | X | X | X | | |
| | | | | X | X | | X |
| X | | X | | | | | X |
| | | | | | | | X |
| _ | | | | | X | | X |
| _ | | | | | | | X |
| _ | | | | | X | | X |
| _ | | | | | | | X |
| _ | | | | X | | | |
| _ | | | | | | | |
| _ | | | | | | | X |
| _ | | | - | | + - | | X |
| _ | | | - | | + - | | |
| _ | | | + | | | | X |
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| _ | | | | | + | | |
| _ | | | | | + - | | |
| | | | | | | | |
| | | | | | T 1 | X | X |
| | | | | | + - | | |
| _ | | | | | + - | | |
| | | | | | | | |
| | | | | X | l | | |
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| | | | | | | | X |
| | | | | | | | X |
| | | | - | | | | X |
| X | X | | | X | | | |
| | | | | Λ | | | ı |
| | | | - | | v | _ | v |
| X | X | | X | X X | X | | X |
| | FPA Inorga X | Norganic Ch | FPA NPA FPA | FPA | FPA | FPA | FPA |

Table 7-5. Summary of SRCs (continued)

| | Surfac | e Soil | Subsu | rface Soil | | | | |
|---------------|----------|--------|-------|------------|-------|-----|--------|---------|
| | (0-1 f | t bgs) | (1-13 | 3 ft bgs) | Sedim | ent | Surfac | e Water |
| SRC | FPA | NPA | FPA | NPA | ED | WD | ED | WD |
| Naphthalene | X | X | | | X | | | |
| Phenanthrene | X | X | | X | X | X | | X |
| Pyrene | X | X | | X | X | X | | X |
| Volu | atile Or | ganic | Compo | unds | | | | |
| Acetone | X | | - | | | - | | - |
| Miscellaneous | | | | | | | | |
| TPH - DRO | | | | X | | | | - |
| TPH - GRO | | | | X | | | | |

bgs = Below ground surface.

BHC = Hexachlorocyclohexane.

DRO = Diesel range organics.

ED = East Ditch.

FPA = Former Production Area.

Ft = Feet.

1

GRO = Gasoline range organics.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocane.

NPA = Non-Production Area.

PCB = Polychlorinated biphenyl.

PETN = Pentaerythritol tetranitrate.

SRC = Site-related contaminant.

TPH = Total petroleum hydrocarbon.

WD = West Ditch.

X = Chemical is an SRC at this depth interval.

-- = Chemical is not an SRC at this depth interval.

Load Line 11

Table 7-6. Summary of COPCs

| | C | - C-9 | | ırface | | | | | |
|------------------------|---------------------|--------|-------|---------|------|------|---------|---------|--|
| | Surfac | | | oil | | | | | |
| 2072 | (0-1 ft | bgs) | | ft bgs) | Sedi | ment | Surface | e Water | |
| COPC | FPA | NPA | FPA | NPA | ED | WD | ED | WD | |
| | Inorganic Chemicals | | | | | | | | |
| Aluminum | X | | | X | X | X | | - | |
| Arsenic | X | X | X | X | X | | | X | |
| Barium | | | | X | | | | - | |
| Chromium | X | X | | | X | X | | | |
| Cobalt | X | X | X | X | X | X | | X | |
| Cyanide | X | X | | X | | | | | |
| Manganese | X | X | | | | | | | |
| Sulfate ^a | | | X | X | | | | | |
| Sulfide ^a | X | X | X | X | | X | | | |
| Thallium | | | | X | | | | | |
| | Pesticid | e/PCB | | | | | | | |
| PCB-1254 | X | | X | | | | | | |
| Semi-volat | tile Org | anic C | отрои | nds | | | | | |
| Benz(a)anthracene | X | X | | | | | | X | |
| Benzo(a)pyrene | X | X | | X | X | | | X | |
| Benzo(b)fluoranthene | X | X | | | | | | X | |
| Dibenz(a,h)anthracene | X | X | | | | | | | |
| Indeno(1,2,3-cd)pyrene | X | | | | | | | X | |
| Miscellaneous | | | | | | | | | |
| TPH - DRO ^a | | | | X | | | | - | |
| TPH - GRO ^a | | | | X | | | | - | |

^aChemical identified as a COPC because no screening value is available.

bgs = Below ground surface.

COPC = Chemical of potential concern.

DRO = Diesel range organics.

ED = East Ditch.

FPA = Former Production Area.

ft = Feet.

GRO = Gasoline Range Organics.

NPA = Non-Production Area.

PCB = Polychlorinated biphenyl.

TPH = Total petroleum hydrocarbon.

WD = West Ditch.

X = Chemical is a COPC in this medium.

^{-- =} Chemical is not a COPC in this medium.

Table 7-7. FWCUGs Corresponding to an HQ of 1 and TR of 1E-05 for COPCs in Soil and/or Sediment

| | | FWCUG | G (mg/kg) |
|------------------------|------------------------------------|--------------------|-----------------------------|
| | Critical Effect or | | ult and Child) ^a |
| COPC | Target Organ | HQ = 1 | TR=1E-05 |
| | Inorganic Chemicals | | |
| Aluminum | Neurotoxicity in offspring | 73,798 | |
| Arsenic | Skin | 20.2 | 4.25 ^b |
| Barium | Nephropathy | 14,129 | |
| Chromium, trivalent | NOAEL | 81,473 | |
| Cobalt | Not specified | 1,313 | 8,030 |
| Cyanide | Decreased caudal epididymis weight | 2.7° | |
| Manganese | CNS | 2,927 ^b | |
| Sulfate | | | |
| Sulfide | | | |
| Thallium | Multiple systems ^d | 6.12 | |
| | Organic Chemicals-Pesticide/PC | В | |
| PCB-1254 | Eyes ^e | 1.2 | 2.03 |
| | Organic Chemicals-Semi-volatile | ? | |
| Benz(a)anthracene | NA | | 2.21 |
| Benzo(a)pyrene | NA | | 0.221 |
| Benzo(b)fluoranthene | NA | | 2.21 |
| Dibenz(a,h)anthracene | NA | | 0.221 |
| Indeno(1,2,3-cd)pyrene | NA | | 2.21 |
| | Miscellaneous | | |
| TPH – DRO | NA | | |
| TPH – GRO | NA | | |

^aResident FWCUGs are the smaller of the Adult or Child values for each COPC and endpoint (non-cancer and cancer).

CNS = Central nervous system.

COPC = Chemical of potential concern.

DRO = Diesel Range Organics.

FWCUG = Facility-wide cleanup goal.

GRO = Gasoline Range Organics.

HQ = Hazard quotient.

mg/kg = Milligrams per kilogram.

NA = Not available.

NOAEL = No observed adverse effect level.

PCB = Polychlorinated biphenyl.

TPH = Total petroleum hydrocarbon

TR = Target risk.

^bFWCUG value is less than the background screening values for arsenic in surface soil (15.4 mg/kg), subsurface soil (19.8 mg/kg), and sediment (19.5 mg/kg) and for manganese in subsurface soil (3,030 mg/kg).

^eNo FWCUG is available for cyanide. Value is the U.S. Environmental Protection Agency regional screening level.

^dVomiting, diarrhea, temporary hair loss, and effects on the nervous system, lungs, heart, liver, and kidneys.

^eOcular exudate, inflamed and prominent meibomian glands.

⁻⁻ FWCUG could not be quantified based on lack of approved toxicity value.

Table 7-8. FWCUGs Corresponding to an HQ of 1 and TR of 1E-05 for COPCs in Surface Water

| | Critical Effect or | CUG (mg/L) | | | | | |
|------------------------|------------------------|-------------|----------|--|--|--|--|
| COPC | Target Organ | HQ=1 | TR=1E-05 | | | | |
| Inorganic Chemicals | | | | | | | |
| Arsenic | Skin | 0.0463 | 0.011 | | | | |
| Cobalt | Not specified | 0.006^{b} | | | | | |
| | Organic chemical-Semi- | volatile | | | | | |
| Benz(a)anthracene | NA | | 0.0001 | | | | |
| Benzo(a)pyrene | NA | | 8.00E-06 | | | | |
| Benzo(b)fluoranthene | NA | | 0.0001 | | | | |
| Indeno(1,2,3-cd)pyrene | NA | | 0.0001 | | | | |

^aResident Receptor FWCUGs are the smaller of the Adult or Child values for each COPC.

COPC = Chemical of potential concern.

FWCUG = Facility-wide cleanup goal.

HQ = Hazard quotient.

mg/L – Milligrams per liter.

NA = Not applicable.

TR = Target risk.

-- = No value available.

Table 7-9. Total and Hexavalent Chromium Soil Sample Results

| Discrete | Total Chromium | Hexavalent Chromium | Percent Hexavalent |
|---|--------------------------|---------------------------|-------------------------|
| Sample Location ^a LL11ss-072 | (mg/kg) 10.1 | (mg/kg) 0.44J | Chromium (%) 4.4 |
| LL11ss-073 | 17.1 | ND | NA |
| LL11ss-075 | 15.4 | 0.71J | 4.6 |

J = Estimated concentration.

mg/kg = Milligrams per kilogram.

NA = Not applicable; hexavalent chromium not detected in sample.

ND = Not detected.

Table 7–10. Environmental Concentrations of PAHs Measured in Background Surface Soil Samples at RVAAP

| | | | Reported Concent | ration (mg/kg) |
|------------------------|-----------|-----|------------------|----------------|
| Analyte | % Detects | | Minimum | Maximum |
| Acenaphthene | 1/15 | 7% | 0.88 | 0.88 |
| Acenaphthylene | 1/15 | 7% | 0.07 | 0.07 |
| Anthracene | 2/15 | 13% | 0.12 | 1 |
| Benz(a)anthracene | 10/15 | 67% | 0.044 | 4.1 |
| Benzo(a)pyrene | 8/15 | 53% | 0.058 | 3.7 |
| Benzo(b)fluoranthene | 10/15 | 67% | 0.062 | 4.8 |
| Benzo(ghi)perylene | 6/15 | 40% | 0.046 | 1.3 |
| Benzo(k)fluoranthene | 6/15 | 40% | 0.053 | 2.6 |
| Chrysene | 10/15 | 67% | 0.057 | 4 |
| Dibenz(a,h)anthracene | 2/15 | 13% | 0.11 | 0.37 |
| Indeno(1,2,3-cd)pyrene | 5/15 | 33% | 0.054 | 1.5 |

mg/kg = Milligrams per kilogram.

PAH = Polycyclic aromatic hydrocarbon.

RVAAP = Ravenna Army Ammunition Plant.

^bNo FWCUG is available for cobalt. Value is the U.S. Environmental Protection Agency tap water regional screening level.

Table 7–11. Baseline Levels of Benz(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Dibenz(a,h)anthracene, and Indeno(1,2,3-cd)pyrene in Soil from Various Data Sets

| St. L. | Number of | Geometric Mean or | N.C | 95th | Mariana |
|---|-----------|----------------------|----------|-------------------------|-----------------|
| Study | Samples | Median | Minimum | Percentile ^a | Maximum |
| CA/TED : th | | enz(a)anthracen | | 10 | 250 |
| CA/T Project ^b | 872 | 0.33 | 0.045 | 19 | 250 |
| LSPA Project ^b | 490 | 0.563 | ND | | 796 |
| Watertown ^b | 17 | 0.411 | 0.021 | 6.04 | 6.05 |
| Worcester ^b | 68 | | ND | 3.8 | 15 |
| New England ^c | 62 | 0.672 | ND | 1.86 | 15 |
| Illinois ^d Urban | | | | 1.8 | |
| Illinois ^d Rural | | | | 0.72 | |
| ATSDR ^e Urban | | | 0.169 | | 59 |
| ATSDR ^e Rural | | | 0.005 | | 0.02 |
| ATSDR ^e Agricultural | | | 0.056 | | 0.11 |
| NYSDEC Rural Near Roads ^f | 28 | | ND | 1.2 | 2.9 |
| NYSDEC Rural Distant Roads ^f | 118 | | ND | 0.16 | 2.6 |
| | 1 | Benzo(a)pyrene | | | |
| CA/T Project ^b | 873 | 0.3 | 0.031 | 17 | 230 |
| LSPA Project ^b | 489 | 0.44 | ND | | 222 |
| Watertown ^b | 17 | 0.95 | 0.6 | 4.77 | 6.08 |
| Worcester ^b | 67 | | ND | 3.3 | 9.7 |
| New England ^c | 62 | 0.686 | ND | 1.82 | 13 |
| Illinois ^d Urban | | | | 2.1 | |
| Illinois ^d Rural | | | | 0.98 | |
| ATSDR ^e Urban | | | 0.165 | | 0.22 |
| ATSDR ^e Rural | | | 0.002 | | 1.3 |
| ATSDR Raricultural | | | 0.0046 | | 0.9 |
| NYSDEC Rural Near Roads ^f | 28 | | ND | 1.1 | 2.4 |
| NYSDEC Rural Distant Roads ^f | 118 | | ND ND | 0.12 | 3.4 |
| N I SDEC Rufai Distant Roads | | ızo(b)fluoranthe | | 0.12 | J. 4 |
| CA/T Project ^b | 873 | 0.68 | 0.045 | 18 | 270 |
| LSPA Project ^b | 486 | | ND | | 250 |
| Watertown ^b | 17 | 1.4 | 0.6 | 6.79 | 7.08 |
| Worcester ^b | | 1.4 | | | 7.08 |
| | | 0.722 | | | 12 |
| New England ^c | 62 | 0.722 | ND | 1.97 | |
| Illinois ^d Urban | | | | | |
| Illinois ^d Rural | | | 1.5 | 0.7 | |
| ATSDR ^e Urban | | | 15 | | 62 |
| ATSDR ^e Rural | | | 0.02 | | 0.03 |
| ATSDR ^e Agricultural | | | 0.058 | | 0.22 |
| NYSDEC Rural Near Roads ^f | 28 | | ND | 1.2 | 3.3 |
| NYSDEC Rural Distant Roads ^f | 118 | | ND | 0.36 | 4.6 |
| 1 | | enz(a,h)anthrace | | | T |
| CA/T Project ^b | 866 | 0.17 | 0.045 | 2.1 | 39 |
| LSPA Project ^b | | | | | |
| Watertown ^b | 17 | 0.195 | 0.155 | 0.604 | 0.64 |
| Worcester ^b | 68 | | ND | | 1.6 |
| New England ^c | 62 | 0.245 | ND | | 2.9 |
| Illinois ^d Urban | | | | 0.42 | |
| Illinois ^d Rural | | | | 0.15 | |
| ATSDR ^e | | | | | |
| NYSDEC Rural Near Roads ^f | 28 | | | | |
| NYSDEC Rural Distant Roads ^f | 118 | | ND | ND | 0.23 |

Table 7–11. Baseline Levels of Benz(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Dibenz(a,h)anthracene, and Indeno(1,2,3-cd)pyrene in Soil from Various Data Sets (continued)

| | Number of | Geometric Mean or | | 95th | |
|---|-----------|----------------------|---------|-------------------------|---------|
| Study | Samples | Median | Minimum | Percentile ^a | Maximum |
| | Inde | eno(1,2,3-cd)pyr | ene | | |
| CA/T Project ^b | 871 | 0.2 | 0.022 | 7 | 100 |
| LSPA Project ^b | 475 | | ND | | 130 |
| Watertown ^b | 17 | 1.752 | 1.2 | 6.2 | 7.2 |
| Worcester ^b | 68 | | ND | 2 | 6 |
| New England ^b | 62 | 0.532 | ND | 1.29 | 6 |
| Illinois ^c Urban | | | | 1.6 | |
| Illinois ^c Rural | | | | 0.51 | |
| ATSDR ^d Urban | | | 8 | | 61 |
| ATSDR ^d Rural | | | 0.01 | | 0.015 |
| ATSDR ^d Agricultural | | | 0.063 | | 0.1 |
| NYSDEC Rural Near Roads ^f | 28 | | ND | 0.62 | 0.66 |
| NYSDEC Rural Distant Roads ^f | 118 | | ND | 0.076 | 1.4 |

^aLognormal 95th percentile value for all studies except: (1) New England value is 95% upper confidence limit and (2) New York State Department of Environmental Conservation (NYSDEC) values are distribution-free 95th percentile.

ND = Not detected.

^bData reported by Massachusetts Department of Environmental Protection (MADEP 2002)

CA/T = Data collected by Mass Highway Department as part of the Central Artery/Tunnel (CA/T) project. LSPA = Preliminary data compiled by the Massachusetts Licensed Site professional Association (LSPA) from background data submitted by its members in 2001. Water Town and Worcester Site-specific background samples.
^cData from 3 New England locations from Bradley et al. 1994.

^dBackground concentrations of polycyclic aromatic hydrocarbons in Illinois metropolitan statistical areas (urban) and nonmetropolitan statistical areas (rural) as reported by Illinois EPA (IEPA 2005)

^eGeneric background data published by the Agency for Toxic Substances and Disease Registry in Polycyclic Aromatic Hydrocarbons August 1995.

^fDistribution-free 95th percentile values for near roads (less than 10 ft from roads and pavement) and not near roads (more than 15 ft from roads and pavement) from New York State Brownfield Cleanup program Development of Soil Cleanup Objectives Technical Support Document (September 2006), Appendix D.

^{-- =} No value reported for this source.

Table 7–12. Summary of Historical COPECs per the Phase I RI

| Soil | Sediment | Surface Water |
|----------------|------------------------|---------------|
| Arsenic | Beryllium | Manganese |
| Chromium | Copper | Sulfate |
| Iron | Lead | |
| Lead | Manganese | |
| Manganese | Nickel | |
| Mercury | Nitrate | |
| Nickel | Sulfate | |
| Vanadium | Sulfide | |
| Zinc | Benz(a)anthracene | |
| Nitrate | Benzo(a)pyrene | |
| Sulfide | Benzo(ghi)perylene | |
| Cyanide | Chrysene | |
| PCB-1254 | Fluoranthene | |
| Nitrocellulose | Indeno(1,2,3-cd)pyrene | |
| | Phenanthrene | |
| | Pyrene | |
| | Total PAHs | |
| | PCB-1254 | |
| | Nitrocellulose | |

Adapted from the Report for the Remedial Investigation at Load Line 11 (AOC 44) (MKM 2005a).

COPEC = Chemical of potential ecological concern.

PAH = Polycyclic aromatic hydrocarbon.

PCB = Polychlorinated biphenyl.

Table 7-13. Comparison of Five Assessment Attributes at Sampling Stations near Load Line 11

| Attributes | S-1 (RM 7.0) (upstream) | S-2 (RM 5.9) (downstream) | Comments |
|--|-------------------------------|---------------------------------|---|
| Sediment quality | Excellent | Excellent | Downstream station rating is equivalent to upstream station, suggesting no negative impacts from the AOC. |
| Water quality | Excellent | Excellent | Downstream station rating is equivalent to upstream station, suggesting no negative impacts from the AOC. |
| Fish community (IBI) ^a | Good | Good | Downstream station rating is equivalent to upstream station, suggesting no negative impacts from the AOC. |
| Macroinvertebrate community (ICI) ^b | Exceptional | Exceptional | Downstream station rating is equivalent to upstream station, suggesting no negative impacts from the AOC. |
| Habitat (QHEI) ^c | Good | Excellent | Downstream station rating is better than upstream station, suggesting no negative impacts from the AOC. |
| Use Attainment Status ^d | Full | Full | Downstream station rating is equivalent to upstream station, suggesting no negative impacts from the AOC. |

^aFish communities range from 0-60, with <18 being "very poor," 18-27 being "poor," 28-35 being "fair," 36-39 being "marginally good," 40-45 being "good," 46-49 being "very good," and 50-60 being "excellent" (Ohio EPA 2009).

bMacroinvertebrate communities range from 0-60, with <2 being "very poor," 2-12 being "poor," 14-32 being "fair," 34-46

AOC = Area of concern.

IBI = Index of Biotic Integrity.

ICI = Invertebrate Community Index.

QHEI = Qualitative Habitat Evaluation Index.

being "good," and 48-60 being "exceptional" (Ohio EPA 1988).

[&]quot;Habitat ranges from 30 to <100, with <30 being "very poor," 30-44 being "poor," 45-59 being "fair," 60-74 being "good," and 75-100 being "excellent" (Ohio EPA 2009).

^dFull-attainment means all of the applicable indices meet the Ohio Environmental Protection Agency biocriteria (USACE 2005a).

Table 7-14. Survey of Proximity to the AOC of Various Ecological Resources

| Natural Resource | Natural Resources Inside Habitat Area | Proximity Within or Near the AOC | Distances to Nearest Resource of the AOC ^a |
|------------------------------|--|-------------------------------------|---|
| Wetlands | Two small Category 1 | Wetlands associated | Other wetlands are in vicinity of AOC |
| (Planning Level Survey | wetlands – one located | with Sand Creek to the | (Figure 7-1) |
| and Jurisdictional) | along the western boundary | north and other small | |
| | of the AOC and one in the | wetlands located to the | |
| | vicinity of the East Ditch | south | |
| Rare species | No known sightings | None | 900 ft west and |
| _ | | | 1,100 ft west |
| | | | (See text for species names) |
| Beaver dams | None | None | 1,300 ft southeast and |
| | | | 2,100 ft south |
| 100-year floodplain | None | Sand Creek Floodplain | 100-year Floodplain to Sand Creek |
| | | to the north | located immediately north |
| Stream sampling ^b | None | Sampling station S-2 is | Sampling station S-1 is approximately |
| | | approximately 300 ft | 4,500 ft west (upgradient) on Sand |
| | | north (downgradient) | Creek |
| | | on Sand Creek | |
| Pond sampling ^b | None | None | Approximately 3,400 ft southwest |
| | | | (Fuze/Booster Ponds) |

^aMeasurements of distance and direction are taken from the nearest boundary of the AOC to the resource being measured.

ft = Feet.

Table 7-15. Summary of Integrated COPECs for Surface Soil at the FPA

| | MDC | ESV | Ratio of MDC | |
|----------------|---------|---------|--------------|--|
| COPEC | (mg/kg) | (mg/kg) | to ESV | Comments |
| Aluminum | 23,800 | 50 | 476 | None |
| Arsenic | 30.2 | 18 | 1.7 | None |
| Cadmium | 1.3 | 0.36 | 3.6 | None |
| Chromium | 28.4 | 26 | 1.1 | None |
| Cobalt | 33.8 | 13 | 2.6 | None |
| Copper | 45.7 | 28 | 1.6 | None |
| Cyanide | 1.6 | 1.33 | 1.2 | None |
| Lead | 102 | 11 | 9.3 | None |
| Manganese | 1,540 | 220 | 7 | None |
| Mercury | 0.34 | 0.00051 | 667 | Second highest ratio at 667x, PBT compound |
| Vanadium | 38.8 | 7.8 | 5 | None |
| Zinc | 465 | 46 | 10.1 | None |
| Nitrate | 3.2 | No ESV | | None |
| Sulfide | 244 | 0.00358 | 68,156 | Highest ratio at 68,156x |
| Nitrocellulose | 1.1 | No ESV | | None |
| Nitroguanidine | 0.077 | No ESV | | None |
| PETN | 0.049 | No ESV | | None |
| Tetryl | 0.02 | No ESV | | None |
| PCB-1254 | 0.42 | 0.371 | 1.13 | PBT compound |

Table excludes nutrients.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg = Milligrams per kilogram.

PBT = Persistent, bioaccumulative, and toxic.

PCB = Polychlorinated biphenyl.

PETN = Pentaerythritol tetranitrate.

x = Multiplier.

-- = Not applicable, no ESV is available for comparison.

^bStream and pond sampling refers to *Facility-Wide Biological and Water Quality Study 2003* (USACE 2005a). AOC = Area of concern.

Table 7–16. Summary of Integrated COPECs for Surface Soil at the NPA

| COPEC | MDC (mg/kg) | ESV (mg/kg) | Ratio of MDC to ESV | Comments |
|----------------|----------------|----------------|------------------------|--|
| Arsenic | 40.4 | 18 | 2.2 | None |
| Cadmium | 0.92 | 0.36 | 2.6 | None |
| Cobalt | 13.9 | 13 | 1.1 | None |
| Lead | 33.9 | 11 | 3.1 | None |
| Manganese | 1,930 | 220 | 8.8 | None |
| Mercury | 0.08 | 0.00051 | 157 | Second highest ratio at 157x, PBT compound |
| Zinc | 478 | 46 | 10.4 | None |
| Nitrate | 3 | No ESV | | None |
| Sulfide | 85.5 | 0.00358 | 23,883 | Highest ratio at 23,883x |
| HMX | 0.013 | No ESV | | None |
| Nitrocellulose | 1.2 | No ESV | | None |
| PETN | 0.036 | No ESV | | None |
| PCB-1254 | 0.035 | 0.371 | 0.09 | PBT Compound |

Table excludes nutrients.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

HMX = Octahydro-1,3,5,7-tetranitro-1,3,5,7-tetrazocine.

MDC = Maximum detected concentration.

mg/kg = Milligrams per kilogram.

PBT = Persistent, bioaccumulative, and toxic.

PCB = Polychlorinated biphenyl.

PETN = Pentaerythritol tetranitrate.

x = Multiplier.

-- = not applicable, no ESV is available for comparison.

Table 7-17. Summary of Integrated COPECs in Sediment at the East Ditch

| | | | Ratio of | |
|-------------------|-------------------|----------------|-------------------|----------|
| COPEC | MDC (mg/kg) | ESV (mg/kg) | Maximum to ESV | Comments |
| | (Ing/Kg) | | to Es v | |
| Beryllium | 1 | No ESV | | None |
| Selenium | 3.2 | No ESV | | None |
| Benz(a)anthracene | 0.11 | 0.108 | 1.02 | None |

Table excludes nutrients.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg = Milligrams per kilogram.

-- = not applicable, no ESV available.

Table 7-18. Summary of Integrated COPECs in Sediment at the West Ditch

| СОРЕС | MDC (mg/kg) | ESV (mg/kg) | Ratio of Maximum to ESV | Comments |
|-----------|----------------|----------------|-------------------------------|--------------|
| Beryllium | 1 | No ESV | | None |
| Mercury | 0.049 | 0.18 | 0.27 | PBT Compound |
| Sulfide | 43.4 | No ESV | | None |

Table excludes nutrients.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/kg = Milligrams per kilogram.

PBT = Persistent, bioaccumulative, and toxic.

-- = Not applicable, no ESV available.

Table 7–19. Summary of Integrated COPECs in Surface Water at the East Ditch

| | | | Ratio of | |
|-----------------|----------|----------|----------|----------------------------|
| | MDC | ESV | Maximum | |
| COPEC | (mg/L) | (mg/L) | to ESV | Comments |
| beta-BHC | 0.000013 | 0.000495 | 0.03 | ESV is USEPA Region 5; PBT |
| | | | | compound |
| gamma-Chlordane | 0.000015 | 0.0043 | 0.003 | ESV is NAWQC; PBT compound |

Table excludes nutrients.

BHC = Hexachlorocyclohexane.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/L = Milligrams per liter.

NAWQC = National Ambient Water Quality Criteria.

PBT = Persistent, bioaccumulative, and toxic.

USEPA = U.S. Environmental Protection Agency.

Table 7-20. Summary of Integrated COPECs in Surface Water at the West Ditch

| СОРЕС | MDC (mg/L) | ESV (mg/L) | Ratio of Maximum to ESV | Comments |
|----------------------|------------|---------------|-------------------------------|---------------------|
| Manganese | 0.559 | 0.12 | 4.66 | ESV is Tier II |
| Benzo(a)pyrene | 0.00033 | 0.00006 | 5.50 | ESV is OMZA |
| Benzo(k)fluoranthene | 0.00024 | No ESV | | None |
| beta-BHC | 0.0000096 | 0.000495 | 0.02 | ESV is USEPA Region |
| | | | | 5; PBT compound |

Table excludes nutrients.

BHC = Hexachlorocyclohexane.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

MDC = Maximum detected concentration.

mg/L = Milligrams per liter.

OMZA = Ohio Environmental Protection Agency outside mixing zone average.

PBT = Persistent, bioaccumulative, and toxic.

USEPA = U.S. Environmental Protection Agency.

-- = Not applicable, no ESV available.

Table 7–21. Application and Decisions of Selected Evaluation Factors to Integrated COPECs for Surface Soil from Level II

| | Condition for Decision | _ |
|--|--|--|
| Action | to Dismiss or Retain COPEC | Outcome |
| Compare mean concentration to ESV | (A) Mean concentration smaller than or equal to the ESV | COPEC is not a candidate to be a final COPEC. Evaluation ceased. |
| Compare mean concentration above ESV to background | (A) Mean concentration smaller than the background concentration | COPEC is not a candidate to be a final COPEC. Evaluation ceased. |
| concentration | (B) Mean concentration larger than background concentration | Continue evaluation of chemical. |

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

Table 7–22. Summary of Data for Step 3A Refinement of Integrated COPECs in Surface Soil at Load Line 11 FPA

| COPEC | Mean Concentration (mg/kg) | ESV (mg/kg) | Mean Concentration > ESV? (Yes/No) | Background Concentration (mg/kg) | Mean Concentration > Background Concentration? (Yes/No) | ESV > Background Concentration? (Yes/No) | Frequency of Detections | Frequency of Detections ^a > ESV | Frequency of Detections ^a > Background Concentration | Further Evaluation in Level II Required? (Yes/No) |
|-----------|---|-------------|------------------------------------|--|---|---|----------------------------|--|---|---|
| | | - | | CO. | PEC with Mean Concentration | < ESV | | | | |
| Arsenic | 13.2 | 18 | No | 15.4 | No | Yes | 47 / 47 | 8 / 47 | 11 / 47 | No |
| Cadmium | 0.215 | 0.36 | No | 0 | Yes | Yes | 26 / 47 | 7 / 26 | 26 / 26 | No |
| Chromium | 14.6 | 26 | No | 17.4 | No | Yes | 47 / 47 | 2 / 47 | 10 / 47 | No |
| Cobalt | 8.54 | 13 | No | 10.4 | No | Yes | 47 / 47 | 2 / 47 | 6 / 47 | No |
| Copper | 18 | 28 | No | 17.7 | Yes | Yes | 47 / 47 | 3 / 47 | 19 / 47 | No |
| Cyanide | 0.301 | 1.33 | No | 0 | Yes | Yes | 3 / 34 | 1/3 | 3 / 3 | No |
| PCB-1254 | 0.0562 | 0.371 | No | 0 | Yes | Yes | 4 / 12 | 1 / 4 | 4 / 4 | No |
| | | | COPEC | C with Mean Concentrat | ion > ESV and Mean Concentr | ation < Background Concentr | ation | | | |
| Aluminum | 11400 | 50 | Yes | 17700 | No | No | 47 / 47 | 47 / 47 | 3 / 47 | No |
| Manganese | 511 | 220 | Yes | 1450 | No | No | 47 / 47 | 42 / 47 | 2 / 47 | No |
| Vanadium | 20.1 | 7.8 | Yes | 31.1 | No | No | 47 / 47 | 46 / 47 | 2 / 47 | No |
| | COPEC with Mean Concentration > ESV and Mean Concentration > Background Concentration | | | | | | | | | |
| Lead | 29.6 | 11 | Yes | 26.1 | Yes | No | 47 / 47 | 44 / 47 | 21 / 47 | Yes |
| Mercury | 0.0397 | 0.00051 | Yes | 0.036 | Yes | No | 17 / 47 | 17 / 17 | 13 / 17 | Yes |
| Zinc | 77.3 | 46 | Yes | 61.8 | Yes | No | 47 / 47 | 42 / 47 | 22 / 47 | Yes |
| Sulfide | 49.3 | 0.00358 | Yes | 0 | Yes | Yes | 22 / 34 | 22 / 22 | 22 / 22 | Yes |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken.

Table 7-23. Summary of Data for Step 3A Refinement of Integrated COPECs in Surface Soil at Load Line 11 NPA

| СОРЕС | Mean Concentration (mg/kg) | ESV (mg/kg) | Mean Concentration > ESV? (Yes/No) | Background Concentration (mg/kg) | Mean Concentration > Background Concentration? (Yes/No) | ESV > Background Concentration? (Yes/No) | Frequency of Detections | Frequency of Detections ^a > ESV | Frequency of Detections ^a > Background Concentration | Further Evaluation in Level II Required? (Yes/No) |
|-----------|-------------------------------|-------------|------------------------------------|-------------------------------------|---|---|-------------------------|--|---|---|
| | | | | COP | EC with Mean Concentration | < ESV | | | | • |
| Arsenic | 12 | 18 | No | 15.4 | No | Yes | 29 / 30 | 3 / 29 | 5 / 29 | No |
| Cadmium | 0.191 | 0.36 | No | 0 | Yes | Yes | 17 / 30 | 4 / 17 | 17 / 17 | No |
| Cobalt | 8.46 | 13 | No | 10.4 | No | Yes | 30 / 30 | 1 / 30 | 7 / 30 | No |
| PCB-1254 | 0.0206 | 0.371 | No | 0 | Yes | Yes | 1 / 14 | 0 / 1 | 1 /1 | No |
| | | | COP | EC with Mean Concentration | on > ESV and Mean Concentr | ation < Background Concentra | ution | | | |
| Lead | 18.1 | 11 | Yes | 26.1 | No | No | 29 / 30 | 27 / 29 | 4 / 29 | No |
| Manganese | 550 | 220 | Yes | 1450 | No | No | 30 / 30 | 29 / 30 | 2/30 | No |
| Mercury | 0.034 | 0.00051 | Yes | 0.036 | No | No | 14 / 30 | 14 / 14 | 11 / 14 | No |
| • | • | • | COP | EC with Mean Concentration | on > ESV and Mean Concentr | ation > Background Concentra | ition | <u>. </u> | | |
| Zinc | 78.2 | 46 | Yes | 61.8 | Yes | No | 30 / 30 | 25 / 30 | 12 / 30 | Yes |
| Sulfide | 41.6 | 0.00358 | Yes | 0 | Yes | Yes | 11 / 17 | 11 / 11 | 11 / 11 | Yes |

^aFrequency of detection refers to the subset of detections relative to ESV or background concentration as opposed to the standard frequency of detections of total samples taken.

COPEC = Chemical of potential ecological concern. ESV = Ecological screening value.

mg/kg = Milligrams per kilogram. PCB = Polychlorinated biphenyl.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg = Milligrams per kilogram.

PCB = Polychlorinated biphenyl.

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Table 7–24. Summary of Mean Concentrations and Background Concentrations of Remaining Integrated Soil COPECs in the Refinement Factors

| СОРЕС | Mean Concentration (mg/kg) | Background Concentration (mg/kg) | Ratio of Mean Concentration to Background Concentration | Qualitative Assessment of Mean to Background Concentration |
|---------|----------------------------------|--|--|---|
| | | Form | ner Production Area | |
| Lead | 29.6 | 26.1 | 1.13 | Concentrations are similar |
| Mercury | 0.0397 | 0.036 | 1.10 | Concentrations are similar |
| Zinc | 77.3 | 61.8 | 1.25 | Concentrations are somewhat similar |
| Sulfide | 49.3 | 0 | NA | Sulfide does not have an established |
| | | | | background concentration at RVAAP |
| | | No | n-Production Area | · |
| Zinc | 78.2 | 61.8 | 1.27 | Concentrations are somewhat similar |
| Sulfide | 41.6 | 0 | NA | Sulfide does not have an established |
| | | | | background concentration at RVAAP |

COPEC = Chemical of potential ecological concern.

mg/kg = Milligrams per kilogram.

NA = Not available, ratio could not be calculated.

RVAAP = Ravenna Army Ammunition Plant.

Table 7-25. Comparison of Mean Concentration to Alternative ESV for Remaining Integrated COPECs

| COPEC | Mean Concentration (mg/kg) | Preferred ESV (mg/kg) | Alternative ESV ^a (mg/kg) | Ratio of Mean Concentration to Preferred ESV | Ratio of Mean Concentration to Alternative ESV | | | | |
|---------------------|----------------------------------|-----------------------------|--|--|---|--|--|--|--|
| | Former Production Area | | | | | | | | |
| Lead | 29.6 | 11 | 40.5 | 2.7 | 0.7 | | | | |
| Mercury | 0.0397 | 0.00051 | 0.1 | 77.8 | 0.4 | | | | |
| Zinc | 77.3 | 46 | 79 | 1.7 | 0.98 | | | | |
| Sulfide | 49.3 | 0.00358 | None | 13,771 | NA | | | | |
| Non-Production Area | | | | | | | | | |
| Zinc | 78.3 | 46 | 79 | 1.7 | 0.99 | | | | |
| Sulfide | 41.6 | 0.00358 | None | 11,620 | NA | | | | |

^aThe Alternative ESV is the ESV with the closest concentration to the preferred ESV that is above the background concentration.

COPEC = Chemical of potential ecological concern.

ESV = Ecological screening value.

mg/kg = Milligrams per kilogram.

NA = Not available or ratio could not be calculated.

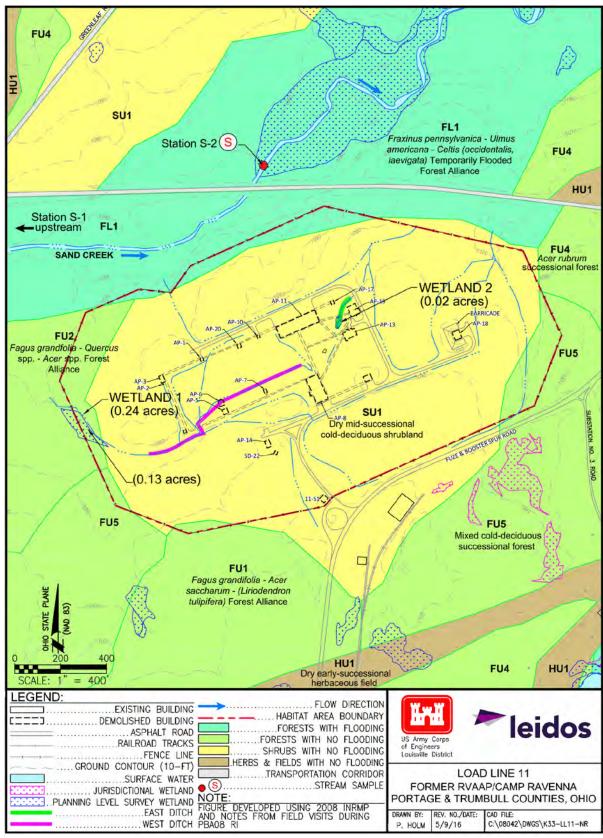


Figure 7-1. Important Ecological Resources Inside and Near Load Line 11

8.0 REMEDIAL INVESTIGATION CONCLUSIONS AND RECOMMENDATIONS

8.1 INTRODUCTION

This Phase II RI Report for Load Line 11 presents a detailed analysis of historical and newly acquired environmental data. The following sections summarize the major findings of the nature and extent of contamination, contaminant fate and transport modeling, HHRA, and ERA. A CSM incorporating all available information is presented to integrate results of prior investigations and the PBA08 RI. The CSM denotes, based on available data where source areas occur, the mechanisms for contaminant migration from source areas to receptor media (e.g., surface water and groundwater), exit pathways from the AOC, and if COCs occur that may require further evaluation in an FS. This section concludes with recommendations for any further characterization under the RI phase of work and, for each of the media evaluated in the RI, whether to proceed to the FS phase of the RI/FS process.

8.2 INTERIM REMOVAL ACTION

In 2001, an IRA was completed at Load Line 11. The IRA was initiated following the Phase I RI activities as an early response action to remove the primary pathways for off-AOC contaminant migration. The IRA included removing sump water from production buildings, grouting selected sanitary sewer manholes, and performing limited excavations from open ditch systems draining the AOC (MKM 2004a).

Sumps located adjacent to Buildings AP-3, AP-5, AP-6, and AP-8 were excavated, removed, and disposed during the 2001 IRA. Approximately 15,000 gal of water were removed from sumps and sewer manholes downgradient of each sump and were filled with bentonite cement to prevent water from infiltrating back into the sumps during excavation and removal operations. Once the water was removed, effluent sewer lines were cut and plugged with mechanical packers and cement grout, which prepared the sumps for removal.

 Six drainage ditch locations were excavated based on Phase I RI analytical data showing exceedances of background concentrations and/or Region 9 PRGs for residential soil. Elevated contaminant concentrations in the six ditch locations included metals in all six ditches, VOCs in two ditches, and SVOCs and/or pesticides in only one ditch. A total of 230 yd³ of contaminated soil was removed during the ditch excavation operations. Confirmation samples were collected to verify that the contaminated soil and sediment were removed.

One 30 by 30 by 8 ft hot spot area, located in an open field north of Building AP-17, was excavated due to petroleum contamination encountered during RI soil boring activities (MKM 2004a). A total of 130 yd³ of petroleum-contaminated soil were excavated. Confirmation samples were used to verify that the petroleum-contaminated soil was removed and the excavations were backfilled with approved off-site soil. During the excavation of the hot spot, it was determined that the area needed to be assessed for UXO. Upon excavating test trenches, cable, scrap metal, bolts, and magnetite containing rock were found. Confirmation samples were then collected from the test trenches. After the

excavation was complete and it was determined no UXO was present, the test trenches were backfilled and leveled to the original ground surface elevation.

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8.3 SUMMARY OF DATA USED IN THE REMEDIAL INVESTIGATION

Quality-assured sample data from the Phase I RI (MKM 2005a), confirmation results from the IRA (MKM 2004a), and 2010 and 2012 PBA08 RI were used to evaluate nature and extent of contamination at Load Line 11. These investigations used discrete sampling methods.

 All available sample data collected at the locations were evaluated to determine suitability for use in various key RI data screens and evaluations (i.e., nature and extent, fate and transport, risk assessment). Evaluation of the data's suitability for use in the PBA08 RI involved two primary considerations: (1) whether the data represented current AOC conditions, and (2) sample collection methods (e.g., discrete vs. ISM).

Samples from the Phase I RI and IRA data sets were evaluated to determine if conditions had changed substantively between earlier characterization efforts and the 2010 and 2012 PBA08 RI, as building demolition activities occurred in 2004-2005 after the 2000 and 2001 Phase I RI and IRA confirmation sampling. The samples collected in 2000-2001 were collected within ditch lines adjacent to former buildings and in areas encompassing, but also extending substantially beyond the footprint of the former buildings. The 2001 confirmation sampling was conducted within the footprints of the removal areas after the contaminated soil was removed. Therefore, both data sets were considered representative of current conditions within and surrounding the footprints of the former buildings and removal areas at Load Line 11.

Data collected in 2010 and 2012 as part of the PBA08 RI focused on delineating the extent of contaminants identified in surface soil (0-1 ft bgs), subsurface soil (1-13 ft bgs), sediment, and surface water. The PBA08 RI sampled locations with the greatest likelihood of contamination (e.g., adjacent to production buildings or within sediment accumulation areas, such as ditches) and analyzed for chemicals identified in historical investigations. Additionally, sediment and surface water samples were collected from drainage ditches exiting the AOC.

8.4 SUMMARY OF NATURE AND EXTENT OF CONTAMINATION

Nature and extent of contamination in surface soil (0-1 ft bgs), subsurface soil (greater than 1 ft bgs), sediment, and surface water subsequent to the 2001 IRA was evaluated in this Phase II RI. Data from the Phase I RI, IRA, and 2010 and 2012 PBA08 RI effectively characterized the nature and extent of the contamination at the AOC. To support the evaluation of nature and extent of contamination, SRC concentrations were compared to SLs corresponding to the lowest FWCUG for the Resident Receptor (Adult and Child) and National Guard Trainee at a target HQ of 0.1 or TR of 1E-06, as presented in the FWCUG Report (USACE 2010a). It can be concluded that the vertical and horizontal extent of contamination is defined, and no further sampling is needed to evaluate Load Line 11.

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8.4.1 Soil

Locations where explosives were identified as potential contaminants from previous use were thoroughly evaluated, including around former process buildings and across the AOC as a whole. The maximum concentrations of the explosives and propellants were all below their respective SLs and were not considered COPCs in the surface and subsurface soil at the FPA and NPA.

Arsenic, chromium, lead, and mercury were identified as potential SRCs and as potentially related to the previous site use. When evaluating these chemicals against their SLs (using the trivalent chromium FWCUG for chromium and the RSL of 400 mg/kg for lead), chromium, lead, and mercury concentrations were all below their SLs; therefore, these chemicals were not considered COPCs for the FPA or NPA in surface and subsurface soil. Arsenic was the only inorganic chemical potentially related to previous AOC operations that is considered a COPC in surface and subsurface soil at the FPA and NPA. Arsenic is considered a COPC in both surface and subsurface soil at the FPA and NPA, with an MDC of 44.1 mg/kg observed from 6–8 ft bgs at sample location LL11cs-045, which evaluated the effectiveness of the hot spot excavation efforts during the IRA.

With the exception of benzo(a)pyrene at four surface soil sample locations, none of the detected concentrations of SVOCs, VOCs, pesticides, or PCBs in surface or subsurface soil were above the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. Benzo(a)pyrene is considered a COPC in the surface soil at the FPA and NPA, with an MDC of 0.45 mg/kg at sample location LL11sb-060, which evaluated the drainage ditch northeast of the Entrance Gate.

8.4.2 Sediment and Surface Water

The East Ditch EU was evaluated with two sediment samples and one surface water sample. No explosives or propellants were detected in the East Ditch sediment and surface water samples. Arsenic was the only inorganic chemical from previous site use that exceeded the SL in sediment and is considered a COPC. No inorganic chemical concentrations detected in the surface water sample exceeded their respective SLs. One PAH [benzo(a)pyrene] concentration detected in sediment exceeded the SL and was identified as a COPC; however, the benzo(a)pyrene concentration was detected at a concentration below the Resident Receptor (Adult and Child) FWCUG at a TR of 1E-05, HQ of 1. SVOCs were not detected in the surface water sample. In addition, no VOCs, pesticides, or PCBs were detected in sediment at the East Ditch. Two pesticides (BHC and gamma-chlordane) were identified as SRCs at low, estimated concentrations. No VOCs or PCBs were detected in the East Ditch surface water.

The West Ditch EU was evaluated with four sediment samples and one surface water sample. No explosives were detected in West Ditch sediment. The propellant nitrocellulose was detected in the sediment sample LL11sd-096 at a concentration below the SL. No explosives or propellants were detected in the surface water sample. No inorganic chemicals related to previous site use exceeded the SLs except for arsenic in the surface water sample. One PAH [benzo(a)pyrene] exceeded the SL in the West Ditch sediment and was identified as a COPC. Four PAHs [benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] exceeded the SLs in the surface

water sample and were identified as COPCs. The PAH concentrations detected in sediment and surface water were all above the Resident Receptor (Adult and Child) FWCUGs at a TR of 1E-05, HQ of 1. No PCBs, pesticides, or VOCs were detected in sediment at the West Ditch. One pesticide (beta-BHC) was identified as an SRC and was detected at a low, estimated concentration below laboratory detection limits. No VOCs or PCBs were detected in the West Ditch surface water.

The Sewer Outfall EU was evaluated with one sediment sample. The only SRCs identified were the inorganic chemicals sulfate and sulfide, neither of which has established background concentrations for comparison.

One sediment sample and one surface water sample were collected at the off-AOC location LL11sd/sw-082 located on Sand Creek. Inorganic chemicals were not detected above background concentrations in sediment; however, six inorganic chemicals were detected above their respective background concentrations for surface water. All the detections were at concentrations below their respective SLs. No explosives, propellants, SVOCs, pesticides, or PCBs were detected in the off-AOC sediment. Two VOCs (carbon disulfide and toluene) were present at low, estimated concentrations below laboratory detection limits.

No explosives, propellants, SVOCs, VOCS, pesticides, or PCBs were detected at the Sand Creek location.

8.5 SUMMARY OF CONTAMINANT FATE AND TRANSPORT

All SRCs identified in the surface soil, subsurface soil, and sediment at Load Line 11 were evaluated through the stepwise contaminant fate and transport evaluation. The evaluation included analyzing leaching and migration from soil and sediment to groundwater and determining whether contamination present in soil and sediment may potentially impact groundwater quality at the site.

Maximum concentrations of SRCs identified in surface and subsurface soil were evaluated using a series of generic screening steps to identify initial CMCOPCs. Initial CMCOPCs for soil were further evaluated using the SESOIL model to predict leaching concentrations and identify final CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria among USEPA MCLs, USEPA tap water RSLs, or RVAAP groundwater FWCUGs for the Resident Receptor Adult. A sediment screening analysis was performed for sediment samples at the AOC. Chemical-specific DAFs were calculated using co-located surface water and sediment concentrations for identified sediment SRCs. These DAFs were used in the sediment screening analysis to identify the final CMCOPCs based on RVAAP facility-wide background concentrations and the lowest risk-based screening criteria. Final CMCOPCs were evaluated using the AT123D model to predict groundwater concentrations beneath source areas and at the nearest downgradient groundwater receptor to the AOC (e.g., stream).

The evaluation of modeling results with respect to current AOC groundwater data and model limitations identified the following CMCOPCs for soil and sediment:

• Arsenic, barium, cobalt, zinc, benzo(b)fluoranthene, naphthalene, and PETN at the FPA and arsenic, cobalt, manganese, and naphthalene at the NPA were the soil CMCOPCs predicted to exceed the screening criteria in groundwater beneath the source area; however, none of these CMCOPCs were predicted to be above their respective groundwater criteria at the downgradient receptor location.

• Benz(a)anthracene and naphthalene at the East Ditch Aggregate and antimony, benz(a)anthracene, benzo(b)fluoranthene, and dibenz(a,h)anthracene at the West Ditch Aggregate were predicted to exceed the screening criteria in groundwater beneath the source area; however, none of these CMCOPCs were predicted to be above criteria at the downgradient receptor location.

A qualitative assessment of the sample results and considerations of the limitations and assumptions of the models were performed to identify if any CMCOCs are present in soil and sediment at Load Line 11 that may impact the groundwater beneath the source or at the downstream receptor location. This qualitative assessment concluded that there were no CMCOCs present in soil and sediment that may impact the groundwater beneath the source or at the downstream receptor location. No further action is required of soil and sediment at Load Line 11 for the protection of groundwater.

8.6 SUMMARY AND CONCLUSIONS OF THE HUMAN HEALTH RISK ASSESSMENT

The HHRA identified COCs and conducted risk management analysis to determine if COCs pose unacceptable risk to the Resident Receptor (Adult and Child). Since the risk management analysis determined there were no unacceptable risks to the Resident Receptor (Adult and Child), it can be concluded that there is no unacceptable risk to the National Guard Trainee and Industrial Receptor.

Media of concern at Load Line 11 are surface soil, subsurface soil, sediment, and surface water. Soil data associated with Load Line 11 were aggregated into surface and subsurface soil at the FPA and NPA. Sediment and surface water were evaluated from ditches within and west of the FPA. The ditches include the East Ditch that flows north and the West Ditch that flows west/northwest and eventually to Sand Creek.

No COCs were identified for the Resident Receptor (Adult and Child) in subsurface soil or sediment. Five PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene] were identified as COCs for the Resident Receptor (Adult and Child) in surface soil. The EPC for benzo(a)pyrene in surface soil at the NPA (0.28 mg/kg) is approximately 1.3 times its FWCUG of 0.221 mg/kg. The EPC for this data aggregate is the MDC (at sample LL11sb-067); benzo(a)pyrene concentrations in the remaining NPA samples ranged from a non-detectable concentration to 0.11 mg/kg. Sample LL11sb-067 was collected in 2010 from the northeastern portion of Load Line 11. No load line operations were conducted in this area. The EPC for benzo(a)pyrene in the FPA (0.219 mg/kg) was slightly less than the FWCUG (0.221 mg/kg). Other PAHs were identified as COCs because they contribute to SORs of two at the FPA and NPA.

The EPCs of benzo(a)pyrene (0.28 mg/kg and 0.221) are slightly above and slightly below the FWCUG of 0.221 and the SORs are two, indicating cumulative cancer risks to a hypothetical future resident on the order of 1E-05. Reported benzo(a)pyrene concentrations above the FWCUG are scattered (i.e., no hot spots were identified); therefore, no COCs were identified for potential additional remediation.

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Four PAHs [benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, and indeno(1,2,3-cd)pyrene] were identified as COCs for the Resident Receptor (Adult and Child) in surface water collected from the West Ditch. The detected concentrations of benz(a)anthracene (0.00035 mg/L), benzo(b)fluoranthene (0.00041 mg/L), and indeno(1,2,3-cd)pyrene (0.00021) in surface water collected from the West Ditch range from 2 to 4 times their FWCUG of 0.0001 mg/L, and the detected concentration of benzo(a)pyrene (0.00033 mg/L) was 41 times the FWCUG of 0.000008 mg/L. All of the reported concentrations included a "J" qualifier, indicating they were estimated results below the detection limit of the analytical method. PAHs are present in the environment from natural and anthropogenic sources. Since surface runoff is the potential source of contamination to surface water in the West Ditch, the fact that PAH concentrations are not significantly elevated in soil or sediment indicates there is no identifiable source beyond normal levels of these chemicals to surface water due to runoff from roads and other traffic areas. In addition, incidental exposures of the Resident Receptor (Adult and Child) to surface water at the West Ditch that only intermittently holds water are much less than the incidental exposure that would occur due to ingesting drinking water (i.e., 2 L/day for an adult) and dermal contact while swimming and wading. These exposures were incorporated into the development of the FWCUGs. Thus, these low, estimated concentrations of PAHs were not identified as COCs for remediation in surface water.

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Based on the risk management analysis, no COCs were identified to be carried forth in an FS for the Resident Receptor (Adult and Child) in any of the media of concern; therefore, no other receptors were evaluated and no further action is recommended from a human health risk perspective.

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8.7 SUMMARY AND CONCLUSIONS OF THE ECOLOGICAL RISK ASSESSMENT

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The Level I ERA presents important ecological resources on or near the AOC and evaluates the potential for current contamination to impact ecological resources. There is chemical contamination present in surface soil, sediment, and surface water at Load Line 11. This contamination was identified using historical and PBA08 RI data. Dry, mid-successional, cold-deciduous shrubland (dominant vegetation type); dry, mid-successional, herbaceous field; and four types of forests were observed on the 48 acres in the AOC. There are important and significant ecological resources in the AOC. Specifically, wetlands and surface water (i.e., two ditches) are present and near contamination. These findings invoked a requirement of a Level II ERA.

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The Level II ERA evaluated integrated COPECs in soil, sediment, and surface water. Integrated COPECs are identified by screening PBA08 RI and historical data sets against ESVs. Twenty integrated COPECs were identified for soil. Five integrated COPECs were identified for surface water. The integrated soil, sediment, and surface water COPECs were further evaluated with technical and refinement factors in Step 3A. The factors

in Step 3A showed there are no integrated COPECs that are of ecological concern requiring remediation or further evaluation. Consequently, the Level II Screening ERA for Load Line 11 concludes with a recommendation that no further action is required to be protective of important ecological resources. This supports the conclusion from the Phase I RI, which stated:

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"No additional action is recommended based on ecological risk because, based on the Step 3a refinement step, it was determined that site-related risks were not great enough to warrant proceeding further into the ecological risk assessment process."

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8.8 UPDATED CONCEPTUAL SITE MODEL

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The updated CSM is presented in this section to incorporate results of this RI. Elements of the CSM include:

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- Primary and secondary contaminant sources and release mechanisms,
- Contaminant migration pathways and discharge or exit points, •
- Potential receptors of risk, and
- Data gaps and uncertainties.

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The following sections describe each of the above elements of the CSM for Load Line 11 and the CSM is presented on Figure 8-1. In addition, figures contained in earlier sections of the report that illustrate AOC features, topography, groundwater and surface water flow directions, and nature and extent of SRCs are cited to assist in visualizing key summary points of the CSM.

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8.8.1 Primary and Secondary Contaminant Sources and Release Mechanisms

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No primary contaminant sources (e.g., operational facilities) are located in the AOC. The IRA conducted in 2001 removed contamination from the primary pathways for off-AOC migration, and all buildings within the Load Line 11 fence line were demolished as of 2009. Remnant contamination in soil and sediment within the AOC is considered a secondary source of contamination.

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Sites where explosives were identified as potential contaminants from previous use were thoroughly evaluated, including around former process buildings and across the AOC as a whole. Explosives were not detected in any of the environmental media sampled (i.e., surface soil, subsurface soil, sediment, and surface water).

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- Surface water drainage generally follows the topography of Load Line 11. The primary drainage routes for surface water are the East Ditch that flows north and the West Ditch that flows westnorthwest, both leading to Sand Creek immediately north of the AOC boundary. Surface water flowing in ditches or other drainage features is the primary migration pathway for contamination to leave the AOC. Three sediment and surface water samples were collected in ditches at Load Line 11.
- 42 Although SRCs were identified in sediment and surface water, none were detected above SLs.

The primary mechanisms for release of chemicals from secondary sources at the AOC are:

- Eroding soil matrices with sorbed chemicals and mobilization in overland surface water storm runoff during heavy rainfall conditions,
- Dissolving soluble chemicals and transport in perennial surface water conveyances and intermittent surface water runoff,
- Re-suspending contaminated sediment during periods of high flow with downstream transport within the surface water system, and
- Leaching contaminants to groundwater.

8.8.2 Contaminant Migration Pathways and Exit Points

8.8.2.1 Surface Water Pathways

Chemical migration from soil sources via surface water occurs primarily by (1) particle-bound chemicals moving through surface water runoff, and (2) dissolved chemicals being transported in surface water. In the case of particle-bound contaminant migration, chemicals will be mobilized during periods of high flow (e.g., rain events). Upon reaching portions of surface water conveyances where flow velocities decrease, chemicals will settle out as sediment accumulation. Sediment-bound chemicals may become re-suspended and migrate during storm events or may partition to dissolved phase in surface water. Sediment and surface water data at Load Line 11 consist of three samples collected from the conveyance ditches. Surface water drainage generally follows the topography of Load Line 11. The primary drainage routes for surface water are the East Ditch that flows north and the West Ditch that flows west-northwest, both leading to Sand Creek immediately north of the AOC boundary. Sand Creek drains to the northeast into the south fork of Eagle Creek. The ditches are predominantly dry but may transport water during periods of heavy rain, although ditches to the south have been known to hold water during periods of time without rain.

In the case of dissolved phase contaminant transport, migration patterns typically reflect a combination of continuous baseflow inputs with superimposed episodic cycles in association with rain events, snow melt, or seasonal precipitation patterns. Such episodic events may temporarily increase dissolved phase contaminant concentrations depending on the source and solubility of SRCs. The events may also dilute and decrease contaminant concentrations if a large influx of comparatively non-contaminated water occurs. As noted in Section 3.0, intermittent surface water from a majority of the AOC flows along various small drainage ditches and exits the AOC along the East Ditch and West Ditch.

8.8.2.2 Groundwater Pathways

The estimated direction of groundwater flow at the AOC is from south to north. This reflects the January 2010 facility-wide potentiometric data presented in the *Facility-Wide Groundwater Monitoring Program Report on the January 2010 Sampling Event* (EQM 2010). Water level elevations at the AOC range from 1,070–1,100 ft amsl, with the highest elevation at LL11mw-001. Potentiometric data indicate the groundwater table occurs within the unconsolidated zone throughout

the AOC. Groundwater discharge to surface water features (e.g., via base flow to streams or springs) does not occur within the AOC boundary. Rather, the closest potential groundwater discharge location is Sand Creek located along the northwest AOC boundary.

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The FWGWMP was initiated in 2005 and involves quarterly sampling of selected wells within the former RVAAP. Between 2009 and 2013, several sampling events under the FWGWMP collected groundwater data at Load Line 11.

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Contaminant leaching pathways from soil and sediment to the water table are through unconsolidated soil representing shallow clay to sand-rich silt tills with interbedded sands scattered throughout. The overall average hydraulic conductivity of the unconsolidated soil is 3.49E-05 cm/s. Conservative transport modeling indicated seven chemicals [arsenic, barium, cobalt, zinc, benz(b)fluoranthene, naphthalene, and PETN] from the FPA and four chemicals (arsenic, cobalt, manganese, and naphthalene) from the NPA may leach from soil, and five chemicals [antimony, benz(a)anthracene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and naphthalene] may leach from sediment and migrate to the groundwater table at concentrations exceeding MCLs/RSLs beneath their respective sources; however, none of these constituents is predicted to migrate laterally and reach the nearest surface water receptor (Sand Creek located along the northwest boundary of Load Line 11) at a concentration exceeding MCL/RSLs. None of these chemicals except arsenic and manganese (that are determined to be background related) were detected in AOC groundwater samples collected from 2009-2013 above their respective groundwater criteria; therefore, this evaluation concludes that the modelpredicted concentrations are conservative. A qualitative assessment of the sample results was performed and the limitations and assumptions of the models were considered to identify if any CMCOCs are present in soil or sediment at Load Line 11 that may potentially impact groundwater at Load Line 11. This qualitative assessment concluded that CMCOCs are not adversely impacting groundwater quality based on current data and are not predicted to have future impacts. No further action is required of soil or sediment to be protective of groundwater.

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8.8.3 Potential Receptors

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In February 2014, the Army and Ohio EPA amended the risk assessment process to address changes in the RVAAP restoration program. The Technical Memorandum identified three Categorical Land Uses and Representative Receptors to be considered during the RI phase of the CERCLA process. These three Land Uses and Representative Receptors are presented below.

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- 1. Unrestricted (Residential) Land Use Resident Receptor (Adult and Child) (formerly called Resident Farmer).
- 2. Military Training Land Use National Guard Trainee.
- 3. Commercial/Industrial Land Use Industrial Receptor (USEPA Composite Worker).

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Unrestricted (Residential) Land Use is considered protective for all three Land Uses at Camp Ravenna. Therefore, if an AOC meets the requirements for Unrestricted (Residential) Land Use, the AOC is also considered to have met the requirements of the other Land Uses (i.e., Commercial/Industrial and Military Training), and the other Land Uses do not require evaluation. The

HHRA did not identify Resident Receptor COCs to be carried forward for potential remediation; therefore, Load Line 11 is considered protective for all potential human health receptors.

Camp Ravenna has a diverse range of vegetation and habitat resources. Habitats present within the facility include large tracts of closed-canopy hardwood forest, scrub/shrub open areas, grasslands, wetlands, open-water ponds and lakes, and semi-improved administration areas. An abundance of wildlife is present on the facility: 35 species of land mammals, 214 species of birds, 41 species of fish, and 34 species of amphibians and reptiles have been identified. The ERA Level I presents important ecological resources on or near the AOC and evaluates the potential for current contamination to impact ecological resources. There is chemical contamination present in soil, sediment, and surface water at Load Line 11, and there are important and significant ecological resources in the AOC. The Level II ERA and the factors in Step 3A showed there are no integrated COPECs that are of ecological concern requiring remediation or further evaluation. Consequently, the Level II Screening ERA concluded that no further action is necessary to be protective of ecological resources.

8.8.4 Uncertainties

Uncertainties are inherent in the CSM depending on the density and availability of data. The CSM for Load Line 11 is overall well defined using existing data, and major data gaps do not remain to be resolved. However, some uncertainties for the CSM for the Load Line 11 include:

Surface water characterization within the drainage ditches on the AOC is subject to some
uncertainty due to the intermittent occurrence of surface water originating within the AOC
during precipitation events. Discharge of surface water from the AOC is generally via ditches
at the AOC.

• Removing primary contaminant sources (e.g., buildings), grading, and continuing vegetation succession within those areas likely have resulted in a lower overall degree of soil erosion and contaminant migration from the former operations area.

• While this RI addresses soil, sediment, and surface water, additional ongoing investigations are being conducted for the Facility-wide Groundwater and Facility-wide Sewers AOCs.

8.9 RECOMMENDATION OF THE REMEDIAL INVESTIGATION

Based on the investigation results, Load Line 11 has been adequately characterized and the recommended path forward is no further action for soil, sediment, and surface water at Load Line 11 to attain Unrestricted (Residential) Land Use. Further investigation is not warranted at this AOC for the following reasons: (1) the IRA conducted in 2001 removed contamination from the primary pathways for off-AOC migration; (2) the current nature and extent of impacted media has been sufficiently characterized; (3) the fate and transport modeling did not identify soil CMCOCs requiring further evaluation or remediation to protect groundwater; (4) there are no CERCLA release-related human health COCs identified in soil, sediment, or surface water requiring further evaluation in an FS or additional remediation; and (5) remedial actions to protect ecological resources are not warranted.

- 1 The next step in the CERCLA process is to prepare a PP to solicit public input with respect to no
- 2 further action for soil, sediment, and surface water. The PP will briefly summarize the history,
- 3 characteristics, risks, and basis for no further action. Comments on the PP received from state and
- 4 federal agencies and the public will be considered in preparing a ROD to document the final remedy.
- 5 The ROD will also include a responsiveness summary addressing comments received on the PP.

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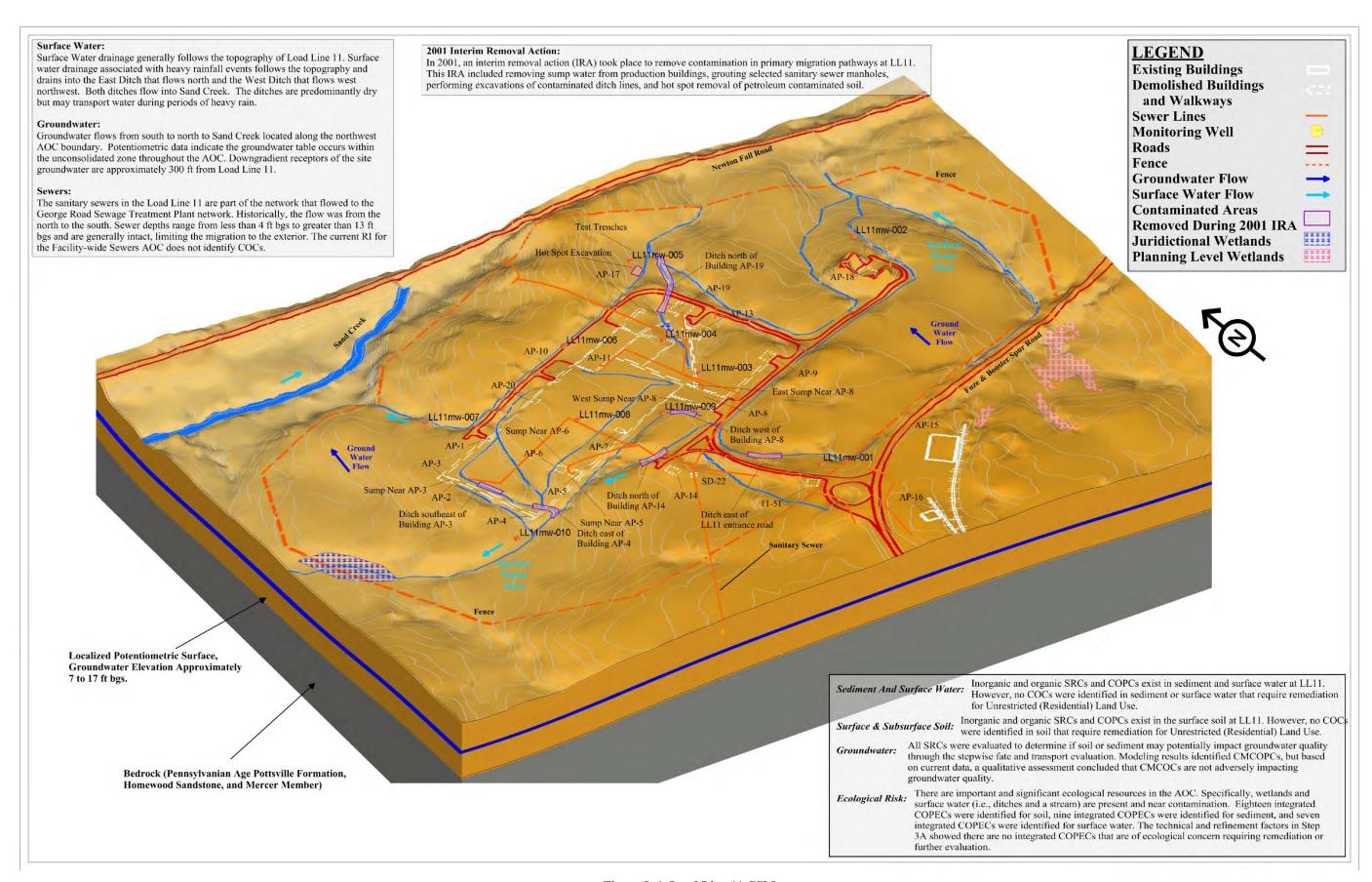


Figure 8-1. Load Line 11 CSM

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9.0 AGENCY COORDINATION AND PUBLIC INVOLVEMENT

The Army is the lead agency responsible for executing the CERCLA process and ultimately completing an approved ROD for soil, sediment, and surface water at Load Line 11. This section reviews actions that have been conducted and presents activities that are planned to ensure the regulatory agencies and members of the public have been provided with appropriate opportunities to stay informed of the progress of Load Line 11 environmental investigation, restoration efforts, and the recommendation of no further action for these media.

9.1 STATE ACCEPTANCE

State acceptance considers comments received from agencies of the state of Ohio on the recommendation for no further action. Ohio EPA is the lead regulatory agency for supporting decisions regarding Load Line 11. This Phase II RI Report has been prepared in consultation with the Ohio EPA.

Ohio EPA has provided input during the ongoing investigation and report development to ensure the recommendation for Load Line 11 meets the needs of the state of Ohio and fulfills the requirements of the DFFO (Ohio EPA 2004a). Ohio EPA provided comments on this Phase II RI Report and will provide comments on the subsequent PP and ROD. The Army will obtain Ohio EPA concurrence prior to the final selection and decision for soil, sediment, and surface water at Load Line 11.

9.2 COMMUNITY ACCEPTANCE

Community acceptance considers comments provided by community members. CERCLA 42 U.S.C. 9617(a) emphasizes early, constant, and responsive community relations. The Army has prepared a *Community Relations Plan for the Ravenna Army Ammunition Plant Restoration Program* (Vista 2015) to facilitate communication between the former RVAAP and the community surrounding Ravenna, Ohio during environmental investigations and potential remedial action. The plan was developed to ensure the public has convenient access to information regarding project progress. The community relations program interacts with the public through news releases, public meetings, public workshops, and Restoration Advisory Board meetings with local officials, interest groups, and the general public.

CERCLA 42 U.S.C. 9617(a) requires an Administrative Record to be established "at or near the facility at issue." Relevant documents regarding the former RVAAP have been made available to the public for review and comment.

| 1 | The Administrative Record for this project is available at the following location: |
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| 2 | |
| 3 | Camp Ravenna |
| 4 | Environmental Office |
| 5 | 1438 State Route 534 SW |
| 6 | Newton Falls, OH 44444 |
| 7 | |
| 8 | Access to Camp Ravenna is restricted but can be obtained by contacting the environmental office at |
| 9 | (614) 336-6136. In addition, an Information Repository of current information and final documents is |
| 10 | available to any interested reader at the following libraries: |
| 11 | |
| 12 | Reed Memorial Library |
| 13 | 167 East Main Street |
| 14 | Ravenna, Ohio 44266 |
| 15 | |
| 16 | Newton Falls Public Library |
| 17 | 204 South Canal Street |
| 18 | Newton Falls, Ohio 44444-1694 |
| 19 | |
| 20 | Additionally, RVAAP has an online resource for restoration news and information. This website is |
| 21 | available at www.rvaap.org. |
| 22 | |
| 23 | Comments will be received from the community upon issuing the RI Report and the PP. As required |
| 24 | by the CERCLA regulatory process and the Community Relations Plan (Vista 2015), the Army will |
| 25 | hold a public meeting and request public comments on the PP for Load Line 11. These comments will |
| 26 | be considered prior to the final selection of a no further action. Responses to these comments will be |
| 27 | addressed in the responsiveness summary of the ROD. |

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